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Article

Patterns of wind dispersal within a forest canopy: An epiphyte perspective

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Mendieta-Leiva et al. (2021), <https://doi.org/10.5281/zenodo.5645775>;

Paton (2019a), <https://doi.org/10.25573/data.10042640.v25>;

Paton (2019b), <https://doi.org/10.25573/data.10042658.v23>;

Paton (2019c), <https://doi.org/10.25573/data.10042688.v23>;

Paton (2019d), <https://doi.org/10.25573/data.10042718.v23>;

22 Paton (2019e), <https://doi.org/10.25573/data.10042694.v23>;
23 Paton (2019f), <https://doi.org/10.25573/data.10042679.v23>;
24 Paton (2020), <https://doi.org/10.25573/data.11799309.v6>;
25 Ramos et al. (2024), <https://doi.org/10.25573/data.24954204.v1>;
26 and Vasquez et al. (2024), <https://doi.org/10.60635/C33W2K>.
27 Additional data used in this study were shared privately by the original data owners and are not
28 publicly available: Janner et al. (in preparation), and data underlying Zotz and Schultz (2008).
29 These privately shared datasets are cited in the manuscript where relevant, and complete
30 references, where available, are provided in the Reference List. Access to these data may be
31 requested from the original data owners, subject to their permission and any applicable
32 restrictions.

33 **Keywords: anemometry; apparent wind; deposition process; dispersal limitation; dispersal**
34 **model; long-distance dispersal (LDD); patchy distribution; terminal velocity; trajectory**
35 **model; tree sway; updrafts; vertical stratification**

Abstract

Plant dispersal theory has largely been developed from a terrestrial perspective, in which interception by vegetation is considered a loss, and success is defined by arrival at suitable ground-level habitat. True epiphytes—non-parasitic plants structurally dependent on a host throughout their lives—represent a striking reversal of this framework: their diaspores must be intercepted by suitable aboveground substrates to disperse successfully. Although dispersal is frequently assumed to underlie the spatially patchy distributions common in epiphyte communities, it remains poorly understood. Here, we quantify airflow along tree vertical profiles in a tropical lowland forest and incorporate these measurements into a mechanistic dispersal model to investigate how airflow, vegetation density, diaspore terminal velocity, and release height jointly shape dispersal within a forest. We examine how these factors influence interception by vegetation—particularly bark, a prerequisite for epiphyte establishment—and promote long-distance dispersal. We further introduce a quantitative metric of dispersal limitation that integrates dispersal success, dispersal distance, and fecundity. Our results reveal aerodynamic stratification within the forest canopy, with weak mean vertical airflow but pronounced turbulence between the upper and lower crowns. This trunk zone may operate as a strong yet permeable filter: although occasional updrafts enable escape above the canopy, most diaspores released in the understorey remain confined within it. Only a fraction is intercepted by bark, typically within 5 m vertically and 15 m horizontally from the release point. This results in strong per-diaspore dispersal limitation and may contribute to spatially patchy epiphyte distributions. However, the degree of dispersal limitation is context-dependent and, when accounting for the high fecundity of many epiphytes, may be less severe than commonly assumed, with post-dispersal demographic filters ultimately shaping distribution patterns. More

59 broadly, our findings highlight a gap in within-forest dispersal theory and underscore the
60 importance of local canopy structure and airflow in shaping diaspore movement in complex,
61 stratified environments. Our model is transferable to other systems—from understorey and
62 invasive plant species to pollen, fungal spores, and ballooning arthropods—and can improve our
63 understanding of connectivity and gene flow within and among forests, informing conservation
64 strategies that enhance dispersal pathways.

Introduction

“Tossing crowns and creaking trunks, and the occasional crash of a falling tree, attest to the strength of the wind aloft, but only an occasional weak gust is felt near the ground.”—

John J. Finnigan (2000)

Plant dispersal theory has long been grounded in a terrestrial perspective, in which **diaspores** (see Table 1 for a glossary of bolded terms) move away from a parent plant to land, ideally, in a suitable ground-level habitat at some distance (Nathan et al., 2002; Nathan, Katul et al., 2011; Tackenberg et al., 2003). Within this framework, atmospheric processes such as turbulence and updrafts are interpreted primarily as modifiers of transport distance, occasionally generating **long-distance dispersal** (LDD) events that support colonisation, gene flow, and range dynamics (Bohrer et al., 2008; Heydel et al., 2014; Mondragón et al., 2015; Soons & Ozinga, 2005). Implicit in this view is a key assumption: vegetation encountered along the dispersal trajectory represents a barrier, limiting dispersal distances or preventing diaspores from reaching their intended destination on the ground (Bohrer et al., 2008; Pouden et al., 2008). Epiphytes fundamentally challenge this assumption: **successful dispersal** does not occur when diaspores avoid vegetation, but when they encounter it.

For epiphytes—non-parasitic plants relying on hosts to complete their life cycle above ground (Zotz, 2016)—soil contact can be detrimental, at least in the wild (e.g., Benzing, 1990; Spicer & Ortega, 2023). Therefore, they face a fundamental challenge during dispersal: their diaspores must not land on the “inhospitable ground” (Benzing, 1990) but instead reach suitable host surfaces where germination and establishment are possible (García-Franco & Rico-Gray, 1988; Zotz, Hietz & Einzmann, 2021). Such **safe sites** (*sensu* Harper et al., 1961) must also occur

within a species' **vertical niche**, across a forest height gradient characterised by steep variation in temperature, moisture, light, and wind regimes. Because dispersal success depends on reaching suitable positions within this vertically structured environment, epiphytes provide an ideal system for examining how dispersal operates within such habitats.

Epiphytes are a defining component of tropical vegetation, contributing up to 50% of local plant diversity (Kelly et al., 2004). Orchids and bromeliads are the most species-rich vascular epiphytic families, but at least 77 other plant families contain epiphytic members, together comprising approximately 31,000 species—roughly 10% of all vascular plant species (Zotz, Weigelt, et al., 2021). When considering only wind-dispersed vascular plants, they make up an even larger share (> 20% globally and ~40–60% in the tropics). Beyond their diversity, epiphytes play key roles in interaction networks, nutrient and water cycling, structural complexity, and habitat provisioning for arboreal fauna (e.g., Ladino et al., 2019; Spicer & Woods, 2022; Thomsen et al., 2018). Despite this ecological importance, they, and specifically their dispersal, remain comparatively understudied (e.g., Spicer & Woods, 2022).

Current dispersal frameworks do not adequately capture the processes governing dispersal within such vertically structured environments. In particular, existing dispersal models (e.g., Bohrer et al., 2008; Heydel et al., 2014; Tackenberg, 2003; Tackenberg et al., 2003) provide limited insight into epiphyte, let alone within-canopy, dispersal. This is striking given that wind dispersal is among the best-studied dispersal types (Cousens et al., 2008), except possibly in tropical forests (Augspurger, 2024). For instance, many models neglect substrate interception altogether (Heydel et al., 2014; Wright et al., 2008; Lee & Yoon, 2025) or consider it only on horizontal surfaces (e.g., Bohrer et al., 2008), despite the central importance of interception on vertical (bark) surfaces for epiphytes—and for limiting dispersal distances in general. Furthermore, dispersal

models are typically based on idealised wind functions or wind measurements from fixed structures such as meteorological towers (Kruijt et al., 2000; Rudnicki et al., 2001), whereas diaspores are released and transported within a dynamic canopy, where they experience **apparent wind** (as opposed to **actual wind**; see Appendix 1: Figure S1 for a conceptual diagram) shaped by local turbulence and tree or branch sway. Empirical measurements of such conditions are rare (Einzmann et al., 2022), and their implications for dispersal remain largely unexplored (Cousens et al., 2008; Murray, 1986; Niklas, 1992; Salisbury, 1942).

In addition, existing dispersal models often initiate dispersal from a **release height** corresponding to the lower part of a plant bearing the diaspores (Greene & Johnson, 1996; Nathan, Horvitz, et al., 2011). In forests, where dispersal studies have largely focused on trees, such assumptions effectively exclude everything below the high crowns (see Appendix 1: Figure S2 for the forest zones we refer to)—sometimes more than two-thirds of the forest height gradient, thereby neglecting wind-dispersed species occurring far below them (though there are exceptions, see e.g., Nathan et al., 2002; Norros et al., 2014). This might reflect the prevailing view that dispersal success is low in forest understories, where very low wind speeds are assumed to limit dispersal (Finnigan, 2000; Schimper, 1888), leading to the expectation that wind dispersal is rare in understorey plants—epiphytic (Kelly, 1985; Vieira & Izar, 1999) or not (Muller-Landau et al., 2008; Roth, 1986). Yet wind-dispersed species can be abundant throughout the understorey and across the full forest height gradient (van der Pijl, 1982; Mendieta-Leiva et al., 2021; see also Appendix 1: Figure S3).

Most epiphytes are wind-dispersed (Madison, 1977; Zotz, 2016). Their diaspores, as other wind-dispersed diaspores, are adapted to very faint winds, being either tiny and extremely light—like the dust seeds of orchids—or plumed with specialised structures such as coma hairs in

tillandsioid diaspores (Schimper, 1888; Benzing, 1990; Chilpa-Galván et al., 2018). Information on **terminal velocity**—the constant speed at which an object falls in still air—is available for only a tiny number of epiphyte species (Bennett, 1988; Chilpa-Galván et al., 2018; Zotz et al., 2016), but available estimates suggest that these traits allow uplift in winds as low as 0.1–0.5 m s⁻¹.

Updrafts are considered important in wind-dispersed (terrestrial) plants, both for **abscission** and LDD (Greene & Quesada, 2011; Heydel et al., 2014; Horn et al., 2012; Tackenberg, 2003; Tackenberg et al., 2003). For epiphytes, however, the role of updrafts may be more fundamental. Because suitable habitat occurs above ground, diaspores that simply descend under gravity risk leaving the species-specific vertical niche, or epiphytic niche altogether. Updrafts allow epiphyte diaspores to remain within, or re-enter, suitable strata after failed interception. Without them, forests might gradually lose epiphytes from upper strata to the ground. Therefore, successful dispersal, **effective dispersal**, and the evolutionary persistence of epiphyte populations are all expected to depend strongly on updrafts.

Most of the few studies on epiphyte dispersal relied on indirect methods, such as nearest-neighbour analyses or colonisation and gene flow rates (e.g., Einzmann & Zotz, 2017a; Gandawijaja & Arditti, 1983; Taylor & Burns, 2015; Zotz & Vollrath, 2003). Only a handful of studies have empirically assessed dispersal distances in epiphytes (e.g., Ackerman et al., 1996; Bennett, 1988; Cascante-Marín et al., 2008, 2009; Mondragón & Calvo-Irabién, 2006), and, together, they reveal a striking paradox. On the one hand, the low terminal velocities of epiphyte diaspores suggest a high potential for transport by updrafts (Schimper, 1908), and thus for LDD (Arditti & Ghani, 2000; Barrington, 1993; Givnish et al., 2016). Indeed, epiphytic orchid species were among the first plants to colonise the islands of Krakatau following the volcanic eruption

(Partomihardjo, 2003). Their high colonisation rates, together with their evolutionary history—colonising and subsequently radiating in distant continents—further support this view (Givnish et al., 2014, 2016; Kessler, 2002; Mendieta-Leiva et al., 2022).

On the other hand, such LDD events are considered exceptional (Givnish et al., 2016; Trapnell et al., 2013). In fact, epiphytes are frequently described as dispersal-limited, often because of their highly clumped (patchy) distributions (Kelly et al., 2004; Nieder & Zotz, 1998). In epiphytes, these patchy distributions may indicate that diaspores often remain near the mother plant, move primarily along branches or stems within the same tree, colonise at best neighbouring trees, and only rarely reach more distant sites (Schimper, 1908; Torres et al., 2018; Zotz & Vollrath, 2003). Accordingly, previous modelling approaches have assumed low dispersal ability, for example through strongly downward-biased dispersal kernels (Petter et al., 2021). However, patchiness alone is weak evidence for **dispersal limitation** unless underlying environmental filters can be excluded, a problem that reflects the broader inconsistent and often implicit use of dispersal limitation (and similar terminology) in plant ecology (Münzbergová & Herben, 2005). Whether patchy distributions in epiphytes actually arise from dispersal limitation is rarely tested explicitly (but see Cascante-Marín et al., 2008, 2009; Mondragón & Calvo-Irabién, 2006).

In contrast, the pronounced **vertical stratification** often observed in epiphytes—where individuals of a species occupy distinct height ranges (Pittendrigh, 1948; Wagner et al., 2013)—is typically attributed to post-dispersal filtering. This distribution pattern primarily reflects the steep microclimatic gradients along the vertical axis (Krömer et al., 2007; Petter et al., 2021; Wagner et al., 2013; Zotz & Hietz, 2001), but could also result from variation in substrate characteristics, including mycorrhizal associations, availability, and substrate age (Francisco et al., 2021; Dislich & Mantovani, 2016; Ruiz-Cordova et al., 2014). This creates a conceptual

asymmetry: while patchy distributions are commonly interpreted as evidence of dispersal limitation, the potential role of dispersal in shaping vertical stratification is rarely considered. Consequently, whether—and how—dispersal contributes to either of these patterns remains an open and surprisingly underexplored question.

This study addresses this gap by explicitly linking vertical airflow dynamics to dispersal outcomes. We quantify vertical wind dynamics directly where epiphytes experience them—along stems and branches across the forest height gradient—and incorporate these measurements into a mechanistic dispersal model for within-forest canopy dispersal (*EpiDisp 1.0*) to simulate diaspore trajectories across a range of terminal velocities and release heights. Specifically, we ask (i) how vertical airflow is structured across forest strata and environmental conditions, (ii) whether lower terminal velocity and higher release height increase dispersal success and escape from the canopy within this stratified forest, and (iii) whether these same conditions, coupled with dispersal distances and fecundity, decrease the degree of dispersal limitation and thereby contribute to spatially patchy and vertically stratified epiphyte communities. Overall, our model represents a pivotal step in quantifying wind dispersal within forests, with relevance extending beyond epiphytes to other wind-dispersed organisms in vertically structured habitats.

Materials and Methods

Study site

The Smithsonian Tropical Research Institute's (STRI) canopy access crane, established in 1997 (9.28° N, 79.97° W), and the associated field station are located within the San Lorenzo National Park, near the Caribbean coast of Panama (Wright et al., 2003). The crane allows access to all layers within a 1 ha permanent plot of this mature lowland tropical moist forest (Figure 1). The

local climate is characterised by a dry season, from approximately mid-December to April, and a wet season, from approximately May to mid-December (Paton, 2020). Mean monthly rainfall increases from below 100 mm in the dry season to around 350 mm in the wet season; mean relative humidity increases from a low of 87% in February to a high of 96% in November; average daily maxima in air temperature, though relatively stable, increase from ca. 26 °C in the dry season to around 29 °C during the wet season (Paton, 2020). Wind also shows marked seasonal differences. Typically, the predominantly northerly winds of the dry season peak at a monthly average of 3.2 m s⁻¹ and a daily maximum of 8 m s⁻¹ in February; the predominantly north-westerly winds of the wet season drop to a monthly average of 1 m s⁻¹ and a daily maximum of 5 m s⁻¹ in October (Paton, 2020).

The closed canopy of forest is between 25 and 35 meters high (Figure 1, also see Appendix 1: Figure S2 for the forest zones we refer to). A repeated census of epiphyte community dynamics at the site has determined a trend of increasing vascular epiphyte abundance from 2002 to 2012, an abundance of nearly 14,000 individuals within a 0.4 ha subplot, and a richness totalling just over 100 species (belonging to 61 genera and 13 families; Mendieta-Leiva et al., 2022).

Though no study explicitly investigated epiphyte phenology at this particular site, studies from Central Panama suggest that many of the wind-dispersed epiphytic (Croat, 1978) and non-epiphytic species (Augspurger & Franson, 1988) release their mature diaspores towards the end of the dry season and the early wet season, when conditions promoting dispersal and successful establishment coincide (Maurer et al., 2013). For this reason, our field work was conducted at the transition from the dry to the wet season.

222 *Terminology*

223 Dispersal is the movement of a diaspore through space (and time, but most epiphytic diaspores
224 will not disperse through time via dormancy). In this study, we consider it successful when the
225 abscised diaspore settles on a suitable substrate (bark). Whether the dispersal is effective
226 (Cousens et al., 2008), depends on whether the site it settles on is a safe site, i.e., whether
227 conditions are suitable for subsequent germination and establishment. Assessing this was beyond
228 the scope of this study, but was assumed, based on prior studies, to fall within a 5–10 m range
229 around the release height (Wagner et al., 2013; Zotz, 2016; Zotz & Schultz, 2008). This area can
230 also be described as a species' vertical niche, meaning the range of heights within the forest
231 where the local microclimate permits epiphyte germination, establishment, and persistence
232 (Krömer et al., 2007; Wagner et al., 2013; Zotz & Hietz, 2001).

233 The dispersal vector—the external force that mediates diaspore movement (Cousens et al.,
234 2008)—of this study is wind. Wind comes in different forms and can be disassembled into
235 different components like a mathematical vector, i.e., magnitude (speed, in m s^{-1}) and direction
236 (in degrees). The direction of horizontal wind usually refers to its origin: a northerly wind comes
237 from the North. Conversely, in the vertical plane, an updraft will indeed flow upward and a
238 downdraft downward.

239 This study differentiates between horizontal and vertical winds as well as horizontal and vertical
240 gusts. Winds are defined as the mean airflow and the circular mean direction measured over a
241 specified time interval in either a horizontal or a vertical plane. The minimum time interval over
242 which the mean was calculated is 1 minute, the maximum 5 minutes. Gusts are defined as a
243 sudden and brief increase in wind speed and are reported as the maximum speeds measured in

the same time interval over which the mean is calculated. However, for specific analyses, their continuous occurrence in shorter time intervals (3 to 10 seconds) will be treated independently.

Updrafts and downdrafts (simply, upward or downward moving air), **ejections** and **sweeps** (see glossary), as well as vertical turbulence (vertically oriented irregular airflow) are common terms to denote different kinds of upward and downward moving air. We will only use these terms when drawing from literature that uses these terms. For our purposes, we will use the terms **up/down-gusts** for maximum and **up/down-winds** for mean air movement in a specified time period. The term **airflow** will comprise both winds and gusts, and the term **wind dynamics** will refer to greater spatial or temporal scales, including both winds and gusts.

All environmental parameters, including temperature, humidity, and wind typically vary along the gradient from forest floor to upper canopy. In the literature, the term “vertical” is often added to specify gradients that are perpendicular to the ground. The *vertical* wind gradient might then refer to the *horizontal* wind at different heights above the ground. In this study, the term forest gradient will imply a vertical gradient, but the terms *vertical* and *horizontal* will be exclusively reserved for denoting the direction of wind. Thus, the *vertical* wind gradient will indeed be the gradient of *vertical* wind at different heights above the forest floor.

Additionally, we distinguish between **release** and abscission. Release refers to the height at which a diaspore is first exposed to the environment, either through the opening of the seed pod or the start of the simulation. Abscission, in contrast, refers to the point at which the diaspore detaches from the seed pod and dispersal begins.

Data collection

Our first aim was to assess wind dynamics across the forest height gradient. The trees, in which vertical airflow was measured (Figure 1), were chosen primarily based on accessibility, while also ensuring that trees of different heights, architectures, and species were included in the sampling. In 15 out of 16 cases, the trees were accessed with the canopy crane, and once through a rope access system. The first replicate was the crane itself, which was securely accessed by climbing its mast. Hereafter, these seventeen replicates will be referred to as vertical profiles.

Field work began on March 8, 2023, and lasted eight weeks (until May 4). Two-dimensional ultrasonic anemometers (HOBO Ultrasonic Wind Speed and Direction Smart Sensor S-WCG-M003, Onset Computer Corporation, Bourne, MA, USA) were installed at various heights and distances along the stems of trees. Where possible, this encompassed the entire forest height gradient from trunk base to upper crown. Often, however, this was not possible as densely growing vegetation prevented access to some parts of the tree. This was particularly often the case from 10–15 m downwards. Then, anemometers were placed where access was possible.

Per vertical profile, four different heights were sampled (once only three). Per height, we installed two anemometers: one near the stem (the “in” position, 0.35 m distance from stem) and one further out (the “out” position, 1.35 m distance from stem). This aimed to investigate whether there are differences in air movement close to the stem and slightly further out. The anemometers were screwed to an aluminium frame, which was strapped to the stem at different points to keep it in a horizontal position (Appendix 1: Figure S4). Importantly, because our focus was on vertical airflow, the anemometers were installed horizontally (even when the tree or branch was growing at an angle). Typically, these anemometers are installed vertically, but prior

286 trials showed that horizontal installation does not influence the measurements (Westphal & Zotz,
287 unpublished). Consequently, instead of aligning the anemometers to the True North using a
288 Northing tool, as is standard practice, they were aligned to face upward (Appendix 1: Figure S5).

289 One logger (HOBO H21-USB Micro Station Data Logger, Onset Computer Corporation,
290 Bourne, MA, USA) per height and per two anemometers was strapped directly onto the stem
291 (Appendix 1: Figure S4). The first three replicates were set to logging intervals of 4, 10, and 5
292 seconds, and thereafter all replicates logged at three second intervals, the maximum resolution of
293 these anemometers. Within their measurement range (0 to 41.2 m s⁻¹), they measure at a
294 resolution of 0.33 m s⁻¹—which also corresponds to the minimum non-zero speed they can
295 accurately measure (measurement threshold)—and with an accuracy of ±0.8 m s⁻¹ or ±4% of the
296 reading, whichever is greater. The anemometers take a reading once every three seconds and
297 then report wind speed (m s⁻¹) as the mean speed calculated over a specified logging interval;
298 gust speed (m s⁻¹) as the maximum 3-second wind speed in that same interval; and wind
299 direction (in degrees) as the unit-vector averaged direction of that interval.

300 At each attachment point, we measured height (meters above ground) using a 50 m measuring
301 tape, facing direction (the cardinal direction, in degrees, toward which the anemometer set is
302 oriented) using a handheld compass, and stem diameter (cm) using a diameter tape. We recorded
303 the total height of the tree (using the same 50 m measuring tape) and its GPS position (using a
304 Garmin eTrex 30, Garmin Ltd., Olathe, Kansas, USA). The trees were identified to species level.
305 All plant names follow Plants of the World Online (POWO; 2026).

306 As there may be minor measuring differences between individual anemometers, their positions
307 and heights were varied at random. During the process of randomisation and reading out data

from the preceding replicate—which was done at the forest floor or at the research station 500 m away from the crane site, using the free software HOBOWare (Onset Computer Corporation, 2022)—the solar-powered anemometers were placed in the sun to recharge their batteries.

During the 59 days of fieldwork, 124 different height × position combinations were assessed, with differing numbers of replicates among combinations. This amounted to approximately 11,000 hours of wind measurements in total, with most sampling effort concentrated at intermediate heights (Appendix 1: Figure S6). We sampled 17 vertical profiles, the canopy access crane itself and nine different tree species: four times a *Brosimum utile* (Kunth) Oken, three times *Socratea exorrhiza* (Mart) H. Wendl. and *Virola fosteri* D. Santam., and once a *Cordia bicolor* A. DC., *Dussia atropurpurea* N. Zamora, R.T. Penn. & C.H. Stirt., *Guatteria lucens* Standl., *Humiriastrum diguense* (Cuatrec.) Cuatrec., *Lonchocarpus heptaphyllus* (Poir.) DC., and a *Pera arborea* Mutis. Except for *S. exorrhiza* and *C. bicolor*, which only reached the mid-canopy, trees were approximately average forest height (ca. 30 m at San Lorenzo) or emergent, i.e., part of the high-crown zone (Appendix 1: Figure S2). Most trees had numerous (hemi-) epiphytes except *Virola* trees, which had very few or none. Rain was registered on eleven days. For an interactive map of the field site with detailed information on the vertical profiles, including imagery, refer to our shinyApp (Appendix 2).

Lidar and topographic data, as well as aerial imagery of the forest plot recorded in May 2023, were obtained from Vasquez et al. (2024). We also used subsets of the long-term climate dataset for sites across Panama hosted by the Smithsonian Tropical Research Institute (STRI; Paton, 2019a–f, 2020). At San Lorenzo, among other devices, a temperature sensor and a cup anemometer are installed near the top of the crane. They log minimum, mean, and maximum temperature, as well as mean and maximum wind speed, and mean wind direction for every 15-

minute interval. However, due to a failure in the anemometer during the sampling period, no wind data were available for San Lorenzo. The nearest functional meteorological station was the Galeta tower site ca. 15 km distance from the San Lorenzo crane.

We modelled the long-term relationship between Galeta and San Lorenzo (14 years of data; Paton 2019a–d) using simple linear regression to assess whether it is justified to impute the missing data. Because the relationships between wind and gust speeds at Lorenzo and Galeta was moderately strong and significant ($R = 0.5$ for wind and $R = 0.53$ for gusts; $p < 0.01$ for both), we replaced the missing data for San Lorenzo using the fitted data of the model.

To relate any obtained patterns of wind dynamics to the dispersal of epiphyte diaspores, we sourced data on the terminal velocity of epiphyte diaspores from primary literature (Bennett, 1988; Mondragón & Calvo-Irabién, 2006; Zotz et al., 2016). Additionally, we measured the falling velocity of three *Tillandsia* species in a custom-built apparatus at the University of Oldenburg, Germany (see Zotz et al., 2016 for specifications).

Data analysis

Lidar data and aerial imagery

Using CloudCompare (2023), only the relevant geographical section of the lidar data (Vasquez et al., 2024) was extracted and transformed to the LAS file type. Using ArcGIS Pro (Esri Inc., 2024), we overlaid the aerial imagery of the San Lorenzo plot with the GPS data of the 17 vertical profiles for post-referencing. In all cases of emergent trees, we were able to locate the crown of the trees in which wind dynamics were assessed using the imagery. Where applicable, we adjusted the GPS positions to match the middle of the crowns. In the case of non-emergent

trees, which did not appear in the aerial imagery, the measured GPS positions were used. We then traced the outline of the tree crowns to calculate crown area. Subsequently, the point locations of trees and the polygons of the crowns were extracted.

The lidar point cloud, point locations and polygons were imported to R (ver. 4.5.2; R Core Team, 2023) and processed, analysed, and visualised using the libraries *lidR* (Roussel et al., 2020), *tidyverse* (Wickham et al., 2019), *raster* (Hijmans, 2025), *sf* (Pebesma, 2018), *sp* (Bivand et al., 2013), and *terra* (Hijmans, 2025). We generated a 2D image of the canopy from the lidar surface (shown in Figure 1). With the crown polygons, we extracted a subset of lidar points that show only the locations of the 17 vertical profiles. We used the anemometers' x, y, and z coordinates to mark their measuring positions in the full point cloud as well as the subset. These can be viewed interactively in our shinyApp (Appendix 2).

Using the *leafR* library (Almeida et al., 2021), we derived **leaf area density (*LAD*)** profiles across the full forest height gradient, averaged a) across the entire plot (i.e., mean *LAD* per height), b) across all vertical profiles, and c) for each profile individually. However, because this method does not distinguish between leaves and wood (or other structures), these estimates are more appropriately interpreted as **plant area density (*PAD*)** rather than true leaf area density. Further, plant or leaf area densities are often criticised for omitting vertical surface components (Olivas et al., 2013), which are formed mainly by tree trunks and provide important attachment surfaces for epiphytes. We therefore refer to the lidar-derived estimates as *horizontal* plant area density (*PAD_h*), representing the horizontal plant surface area per unit volume.

To account for this limitation, we estimated a second component, namely the vertical plant area density (*PAD_v*), from stem surfaces. Specifically, we used diameter at breast height (dbh;

measured at 1.3 m above ground) and stem diameter data collected at higher positions along the stem from previous censuses (Janner & Fernández-Lucero, work in progress; Ramos et al., 2024) alongside our own stem diameter measurements at sensor heights.

We interpolated missing diameters at 1 m height intervals using simple linear regression. For each 1 m stem segment, we calculated vertical surface areas as $2\pi rh$ (where r = stem radius [m] and height h = 1 m) and divided this by the corresponding stem volume (m^3), $\pi r^2 h$, and the crown surface area (m^2) to obtain PAD_v beneath the crown. To reflect our assumption that epiphyte diaspores are most likely to attach only to the side of the stem facing their release point, we multiplied this surface area by 0.5, thereby ignoring redistribution to the opposite side of the stem by turbulent airflows. This yielded a local estimate of PAD_v for the airspace beneath that crown. For a plot-level estimate, we averaged stem diameters within each 1 m height class across all trees. We then converted these mean diameters to stem surface area, scaled them by the number of trees in the plot, and divided by plot area (1 ha), following our previous implementation.

In both approaches, the resulting PAD_v profiles were matched to the lidar-derived horizontal plant area density profile (PAD_h) at 1 m vertical resolution. Thus, total plant area density was defined as the sum of the (lidar-derived) horizontal and the (stem diameter-derived) vertical components of plant area density ($PAD = PAD_h + PAD_v$). Hereafter, we use PAD to refer to this combined measure.

Wind dynamics

To answer our first question, how vertical airflow is structured across forest strata and environmental conditions, the measurement data were imported into R for all subsequent

analyses and combined with site characteristics (e.g., anemometer height, *PAD*, stem diameter etc.). Because the data were not normally distributed (according to Shapiro-Wilk tests, histograms, and qqplots), showed high skewness, high kurtosis, and non-homogenous variances, and had uneven sample sizes among height categories (Appendix 1: Figure S6), we used non-parametric tests. Specifically, we used Kruskal-Wallis (KW) tests followed by Dunn’s post hoc tests with Bonferroni correction, implemented with the *rstatix* library (Kassambara, 2025). Outliers, i.e., points exceeding 1.5 times the interquartile range, were not removed from the data as gusts—the maximum wind speeds in a specified time interval—represent, by definition, extreme values.

To investigate wind dynamics across vertical profiles we pooled the replicates into 5-meter height categories, starting from 0–5 m and ending at 30–35 m, where the highest anemometers were installed. To justify pooling, we assessed the contribution of height to the total variance within each replicate and across all replicates together. Specifically, we calculated the ratio of intra-height variance to inter-height variance, and because inter-height variance exceeded intra-height variance in four of the seven height categories (Appendix 1: Figure S7), supporting our decision to pool replicates by height. Height category was subsequently used as a factor in the analysis.

Effects of height and time of day on gust frequency, mean hourly speed, and relative speed

To assess wind dynamics across the forest height gradient, we used several indices: absolute (i.e., raw) **up-wind** and **up-gust speeds**, **up/down-gust frequency**, **mean up-/down-gust speeds** calculated separately, and **mean vertical gust speed** as an index of net vertical airflow. These indices were calculated either by hour of the day, for some visualisations, or by diel period, for

most statistical tests. A day began at 06:00 and ended at 18:00; the remaining time was treated as night. This roughly corresponded to the sunrise and sunset times at the study site during the sampling period (± 30 minutes). Differences in means among heights, diel period, and position (in/out) were tested using KW-tests followed by Dunn's post hoc tests with Bonferroni correction.

We calculated the frequency of all up- and down-gusts $\geq 0.33 \text{ m s}^{-1}$ (the measurement threshold) for each height, date, and hour-of-day subset. For this calculation, we excluded the first three vertical profiles because their measurement intervals exceeded 3 seconds and were therefore not directly comparable with the remaining data. Relative up-gust and down-gust proportions were calculated relative to the total number of measured gusts per height category. Mean hourly up-gust and down-gust speeds were calculated for the same subsets, with measurements of 0 m s^{-1} divided evenly between up-gusts and down-gusts because they are essentially directionless but still contribute to the mean. To determine whether net vertical airflow was predominantly upward or downward, we calculated mean vertical gust speed per diel period for each height category. For this signed mean, down-gusts were assigned negative values before averaging.

Effects of combined environmental factors on within-canopy gust speed

To test the effect of the environment on within-canopy mean gust speed, we fitted a Bayesian hierarchical model using the brms library (Bürkner, 2017). We calculated the 5-minute mean vertical gust speed (μ), matching the temporal resolution of the above-canopy predictors, which had been interpolated to 5-min intervals as described previously. To reduce computational load and avoid over-representing certain height categories, the data were balanced by down-sampling each height group to the smallest group ($\sim 2,200$ data rows). Residual spread (σ) was modelled

simultaneously to allow variability around the mean to also depend on environmental conditions. Early model testing showed that some relationships were not well described by straight lines, so we allowed moderate non-linearity in the effects of all predictors by using smooth terms. Predictors included above-canopy wind speed, temperature, and vegetation structure, which was represented by two variables: PAD_{local} —defined as PAD at anemometer height—and cumulative PAD above anemometers (PAD_{above}), calculated as the sum of PAD per vertical metre above. We also modelled the linear interaction between above-canopy wind speed and both PAD variables. Sensor height was strongly negatively associated with PAD_{above} (Pearson’s R : -0.9) and therefore excluded. Vertical profile ID was included as a random intercept to account for repeated measurements. All predictors were z-standardised before modelling.

Mean vertical gust speed was modelled with a Student- t distribution to accommodate occasional extreme values:

$$y_i \sim \text{Student } t(\nu, \mu_i, \sigma_i),$$

(Equation 1)

where y_i is the observed mean gust speed for observation i , μ_i is the predicted mean gust speed for observation i , σ_i is the predicted residual scale (i.e. variability) for observation i , and ν is the degrees-of-freedom parameter controlling tail heaviness, allowing the model to accommodate extreme values better than a Gaussian distribution. The predicted mean μ_i was modelled as:

$$\begin{aligned} \mu_i = & \beta_0 + s(\text{above canopy wind}_i, k = 5) + s(PAD_{above_i}, k = 5) + s(PAD_{local_i}, k = 5) + \\ & s(\text{mean temperature}_i, k = 5) + \beta_1(\text{above canopy wind}_i \times PAD_{above_i}) + \beta_2 \\ & (\text{above canopy wind}_i \times PAD_{local_i}) + b_{\text{vertical profile ID}[i]}, \end{aligned}$$

(Equation 2)

where β_0 is the intercept, $s(\cdot)$ denotes a smooth term, and k is the basis dimension controlling the maximum flexibility of the smooth. The coefficients β_1 and β_2 describe interaction effects between above-canopy wind and PAD_{above} , and between above-canopy wind and PAD_{local} , respectively; $b_{\text{vertical profile ID}[i]}$ is a random intercept for vertical profile identity.

The residual scale parameter was modelled as:

$$\log(\sigma_i) = \alpha_0 + s(\text{above canopy wind}_i, k = 5) + s(PAD_{above_i}, k = 5) + s(PAD_{local_i}, k = 5) + s(\text{mean temperature}_i, k = 5),$$

(Equation 3)

where α_0 is the intercept.

Weakly informative regularising priors were used throughout. Because predictors were z-standardised, population-level coefficients were assigned zero-centred normal(mean = 0, SD = 0.15) priors to favour modest effect sizes while remaining broad enough to accommodate stronger effects if supported by the data. The mean intercept was assigned a normal(mean = 0, SD = 0.2) prior, centred on zero because mean vertical gust speed was generally close to zero. Random-effect and smooth-term standard deviations were assigned exponential(10) and exponential(2) priors, respectively, to limit implausibly large between-group variation and excessive complexity in smooth terms while preserving flexibility. In the σ model, the intercept was assigned a normal(mean = log[0.03], SD = 0.5) prior, corresponding to a low but uncertain residual standard deviation under average predictor conditions. The Student- t degrees-of-

freedom parameter was assigned a gamma($\alpha = 2$, $\beta = 0.1$) prior, where α controls the shape of the distribution and β its rate, allowing for heavy-tailed residual variation.

Stable fitting and convergence were achieved after increasing the target acceptance rate to 0.995 and the maximum tree depth to 15. Convergence diagnostics, including trace plots, posterior predictive checks, and Bayesian R^2 , indicated that this sampling effort was sufficient. We report beta coefficients (β) for estimated linear effects and interactions, standard deviations (sds) for the magnitude of non-linear smooth variation, the standard deviation (sd) for random-effect variation, the residual standard deviation for residual spread (σ), the degrees-of-freedom parameter (ν) governing the heaviness of the Student- t residual tails, and 95% credible intervals (CrI).

Additional R libraries used in our analyses are: data.table (Barrett et al., 2025) for working with large data; circular (Agostinelli & Lund, 2025) for calculating circular means; png (Urbanek, 2025) to add background images to the plots; patchwork (Pedersen, 2025) to join plots; PMCMRplus (Pohlert, 2024) and biostat (Gegzna, 2020) to create clds to add to the plots; ggtext (Wilke & Wiernik, 2022) for generating plot annotations and figure titles; ggbreak (Xu et al., 2021) to create axis breaks; viridis (Garnier et al., 2024) for colourblind-friendly colour scales; ggpmisc (Aphalo, 2025) and modelR (Wickham, 2023) for fitting generalised linear models; scales (Wickham et al., 2025) for (pseudo-)log transformation of axes and legends; moments (Komsta & Novomestky, 2022) to test for skewness and kurtosis, and RColorBrewer (Neuwierth, 2022) for the colour grading in Figure 1. Most of the generated figures were further manipulated using the free software Inkscape (Inkscape's contributors, 2023).

Diaspore dispersal simulation

To estimate **primary dispersal** success within the forest height gradient, we developed a trajectory model (*EpiDisp 1.0*) that is based on the winds and gusts measured at the different heights. As in other trajectory models, our model calculates the cumulative vertical and horizontal dispersal distance of a diaspore based on its trajectory during shorter time intervals.

Input parameters

The model incorporates three independent variables: release height (assumed to be the heights at which the anemometers were installed), terminal velocity (V_{term}), and plant area density (PAD). Reported terminal velocities for epiphytes range from 0.06 m s^{-1} in (not truly epiphytic) ferns to 0.09 m s^{-1} in orchids and to 0.5 m s^{-1} in bromeliads (Bennett, 1988; Chilpa-Galván et al., 2018; Zotz et al., 2016; Hoensbroech, pers. data, unpublished); *EpiDisp 1.0* considers a terminal velocity range of 0.01 m s^{-1} to 1 m s^{-1} . Like other trajectory models, this model neglects the short (relative to its subsequent flight period) initial acceleration period of the diaspore (usually $< 0.5 \text{ s}$) and assumes that V_{term} remains constant during the diaspore's flight (Nathan et al., 2002; Nathan, Katul et al., 2011; Tackenberg, 2003).

Plant area density (PAD) affects a diaspores flight: as a rule of thumb, the more obstacles in the form of, e.g., leaves, branches, or stems are present (the higher the PAD), the more likely it is that a diaspore is intercepted by (deposited on) them early during its trajectory. Because bark generally forms the outermost layer of the woody parts of trees, the wood component was used as a proxy for potential bark interception. As the available lidar data did not allow us to distinguish leaves from wood, we estimated this partition from values reported in the literature: in tropical forests, woody surfaces, i.e., bark, contribute approximately 20% to horizontal plant

area density (PAD_h ; Kalacska et al., 2005; Olivas et al., 2013). We therefore assumed that 20% of PAD_h represented bark, while vertical plant area density (PAD_v), being derived from stem surfaces, was assumed to consist entirely of bark. Based on these assumptions, we calculated **bark area density (BAD)** as: $BAD = 0.2 \times PAD_h + PAD_v$, and leaf area density (LAD) as the difference between PAD and BAD ($LAD = PAD - BAD$).

Simulation experiments

EpiDisp 1.0 calculates primary dispersal distance as the sum of the wind/gust speeds (the effective wind) minus the diaspore's falling velocity for every measured time increment (Equations 4a and 4b). Per increment, the maximum absolute gust speed contributes one third, the 1-minute wind average two thirds to the effective wind (the sum of the two components). For the vertical distance, the terminal velocity is subtracted from wind/gust speeds, so that

$$D_w = \sum \begin{cases} \frac{w' + 2\bar{w}}{3} - V_{term}, & \text{if } \frac{w' + 2\bar{w}}{3} > V_{term} \\ \frac{w' + 2\bar{w}}{3}, & \text{if } \frac{w' + 2\bar{w}}{3} < V_{term} \end{cases}$$

$$\text{and } D_u = \sum \left(\frac{1}{3} u' + \frac{2}{3} \bar{u} \right) t$$

(Equations 41a and 4b)

where D_w and D_u are the sum of the, respectively, vertical or horizontal dispersal distances, w' and u' the vertical and horizontal gust speeds, $\bar{w}$ and $\bar{u}$ the 1-minute vertical and horizontal wind averages, t is the time increment and V_{term} the (negative) terminal velocity of a diaspore. Dispersal distances are influenced by terminal velocity only when it exceeds the speed of the downward airflow. When the air moves downward faster than the diaspore's terminal velocity,

the airflow carries the diaspore entirely, effectively cancelling out the influence of terminal velocity on its trajectory (Horn et al., 2012).

Simulations begin at a randomly selected time interval (T) at release height H within a vertical profile. Each diaspore is assigned a terminal velocity (V_{term}) from a predefined set of 12 possible values (0.01, 0.05, 0.10, 0.15, 0.20, 0.25, 0.30, 0.35, 0.40, 0.45, 0.50, and 1 m s⁻¹). At each time step, the model evaluates whether the current wind speed or gust exceeds V_{term} . If yes, the diaspore is abscised and enters flight. If no, the diaspore remains attached and advances to the next time interval (T') without changing height. Once abscised, the diaspore moves horizontally and vertically according to wind vectors and its V_{term} . On its way towards each new height (H'), the diaspore potentially encounters obstacles represented by plant area density (PAD) and its components (BAD and LAD).

$$I_{pw} = 1 - \exp(-PAD |D_w|) \text{ or } I_{pu} = 1 - \exp(-PAD |D_u|)$$

(Equations 5a and 5b)

The probability that a diaspore is intercepted by a substrate during its horizontal (P_u) or vertical (P_v) trajectory is modelled as a function of the absolute horizontal or vertical distance it travels ($|D_u|$ or $|D_v|$) and the plant area density (PAD) at its release height, following Equations 5a and 5b based on Bohrer et al. (2008). To determine whether interception occurs, the model generates a random number (R) between 0 and 1. The diaspore is intercepted by a substrate when $R < P_u$ or $R < P_v$. Dispersal is successful when that substrate is bark ($R < BAD$) and unsuccessful when it is a leaf ($R > BAD$) or when it is dispersed to the forest floor ($H' = 0$). When it is dispersed above the canopy ($H' > \text{crown height}$), possibly initiating LDD, the success remains undetermined. In the latter three cases, the diaspore is removed from the simulation. If none of these conditions are

met, the trajectory continues until termination criteria are reached (see Figure 2, as well as Appendix 3 for a detailed flowchart). The point of interception is calculated as D_u or D_v , i.e., the distance it would have travelled without interception, multiplied by the random number R .

We ran the simulation per subset, defined by vertical profile ID, anemometer position (in/out), and diel period. To capture finer temporal variation, we initially divided the day into four diel periods (early day: 06:00–11:59; late day: 12:00–17:59; early night: 18:00–23:59; late night: 00:00–05:59). For each subset, trajectories were simulated across all V_{term} values using the same randomly selected 1,000 time steps, corresponding to approximately 30 min of wind data, or fewer if the subset was shorter.

Output analysis

To answer our second question, how airflow, forest structure, release height, and diaspore traits jointly determine dispersal success and escape from the canopy, we first quantified the relative frequency of successful and canopy dispersal events. We calculated the relative proportion of the different dispersal outcomes, i.e., the ratio of diaspores per outcome to the total diaspores released per release height, V_{term} , and diel period. Because diel period contributed little explanatory power ($< 0.01\%$), we simplified it to a binary day/night factor for analysis, combining early and late day into day and early and late night into night.

We then assessed whether the probability of successful and canopy dispersal increased with decreasing V_{term} and increasing release height using binomial regression. However, we observed that the binomial regression performed poorly for canopy dispersal, and instead used general additive models (GAMs), which showed a relatively lower Akaike Information Criterion (AIC) indicating better model fit. Both the binomial regression for successful and GAM for canopy

dispersal considered V_{term} , release height, and diel period as predictor variables. Diaspores that dispersed successfully or above the canopy received a value of 1, the remaining outcomes a value of 0. Diaspores still “in dispersal” at the end of the simulation were excluded from both regressions as the outcome remained uncertain. Vertical profiles with a crown height ≤ 15 m were excluded from the canopy dispersal regression as the respective tree crowns were too low to model dispersal above the forest canopy. We ran the model on all V_{term} values to visually identify those that exhibited the most pronounced differences in probabilities. Only those (V_{term} : 0.01, 0.1, 0.3, 0.5, and 1 m s⁻¹) were considered for the manuscript.

We interpreted statistical significance with caution, as inferential outcomes can be strongly influenced by the number of simulated replicates rather than by the biological relevance of an effect. Following Murray and Conner (2009), we therefore emphasized variable importance rather than significance testing by first screening predictors using zero-order associations with the binary response and then quantifying their independent contributions using hierarchical partitioning. For the binomial response models, hierarchical partitioning was implemented with the R library hier.part (Mac Nally & Walsh, 2004), using a binomial error structure and log-likelihood as the goodness-of-fit metric. We report the independent contribution of each predictor and its percentage of the total independent explanatory contribution to describe the dominant tendencies generated by the simulations, rather than treating small p -values as evidence of biological importance.

For bark interception probability, hierarchical partitioning indicated that V_{term} was the dominant predictor of simulated outcomes, followed by diel period and release height. In the full model, V_{term} , diel period, and release height accounted for 82%, 12%, and 6% of the total independent

contribution, respectively. Because removal of release height after zero-order screening produced nearly the same ranking of the remaining predictors, we report the full model in the Results.

For LDD probability, hierarchical partitioning of the full model ranked V_{term} first, release height second, and diel period third, each accounting for 63%, 37%, and $< 0.1\%$ of the total independent contribution, respectively. After removal of diel period following zero-order screening, the reduced model retained the same ranking of the remaining predictors. Again, because removing diel period after zero-order screening produced nearly the same ranking, we present the full model in the Results.

To assess our third question, to what extent these processes generate dispersal limitation and contribute to the spatial and vertical structuring of epiphyte communities, we used linear regression models to quantify the influence of V_{term} , release height, and diel period on vertical and horizontal dispersal distances of successfully dispersed diaspores. We compared the simpler models (release height, V_{term} or diel period alone) with the more complex models (their combination) using AIC to evaluate whether additional predictors improved model fit. In all cases, the more complex model had a lower AIC, indicating they captured the signal better. We repeated the linear regression models on only parts of the forest height gradient to determine whether the degree of influence of predictor variables varied between different forest zones, but the effect was not significant.

We fitted dispersal kernel models to the observed vertical and horizontal dispersal distances of diaspores intercepted by bark using the R `dispfitter` library (Proença-Ferreira et al., 2025). We excluded distances of 0 m and converted negative distances to their absolute values. Dispersal directions (downward, upward, and horizontal) as well as groups ($V_{term} \times$ release height) were

modelled separately. For visualisation, we selected the best-supported model (lowest AIC) and used this model for plotting. We then summarised kernels across groups by calculating the mean kernel estimate and its standard error (SE). We estimated the distances at which 50%, 75%, and 95% of diaspores were predicted to be bark-intercepted by deriving these quantiles from the cumulative bark-interception curve. We additionally used 2d relative densities (i.e., densities per $V_{term} \times$ release height group) to visualise their distribution in the forest post-bark interception.

Finally, we developed a continuous metric of dispersal limitation by combining the proportion of diaspores that successfully dispersed with the distances they moved within the 1 ha forest plot. Vertical and horizontal dispersal distances were expressed relative to their respective maximum possible distances (i.e., from release height to ground or canopy, and across the plot extent, ~140 m, roughly reflecting the maximum dispersed distance of ca. 146 m), describing how broadly diaspores spread away from the parent plant. Both vertical and horizontal displacement were scaled linearly between zero and one.

The relative importance of vertical positioning and horizontal spread was controlled using weighting parameters (α and β , respectively), which determine how strongly each component contributes to the realised dispersal contribution of individual diaspores (ranging from 0 = no influence to 1 = proportional influence, with values >1 increasing emphasis). For example, values around 0.5 indicate moderate importance, whereas values ≥ 0.8 reflect a dominant role of that component.

For each release height $\times$ terminal velocity $\times$ diel period combination, the realised dispersal contribution of an individual diaspore i was defined as:

$$C_i = S_i \cdot (V_i^\alpha) \cdot (H_i^\beta),$$

653 (Equation 6)

654 where C_i is the contribution of diaspore i , S_i is a binary indicator of dispersal success (1 =
655 successful interception, 0 = unsuccessful), V_i is the scaled vertical displacement, and H_i is the
656 scaled horizontal displacement. Both V_i and H_i range from 0 to 1 relative to their maximum
657 achievable distances. The parameters α and β control the relative weighting of vertical and
658 horizontal components.

659 Rather than averaging contributions arithmetically, we aggregated dispersal contributions using a
660 power mean of order q :

662
$$\bar{C}_q = \left(\frac{1}{n} \sum_{i=1}^n C_i^q \right)^{1/q},$$

661 (Equation 7)

663 where $\bar{C}_q$ is the mean realised dispersal contribution and q controls the weighting of contributions
664 across diaspores. When $q = 1$, this reduces to the arithmetic mean, such that all diaspores
665 contribute equally. For $q > 1$, higher-contribution diaspores exert disproportionately greater
666 influence on the aggregate metric, thereby reducing the dominance of abundant low-distance
667 dispersal events and increasing the importance of rare greater dispersal distances.

668 To incorporate fecundity, we scaled the aggregated contribution by a fecundity term, such that
669 dispersal limitation was defined as:

670
$$\text{Dispersal limitation} = (1 - \bar{C}_q)^{\log(1+F)},$$

671 (Equation 8)

where F is fecundity. The logarithmic scaling of fecundity reflects diminishing returns of increasing diaspore supply, such that additional diaspores increase the likelihood of successful dispersal but with progressively smaller marginal gains. This prevents unrealistically rapid collapse of dispersal limitation at high fecundity.

Values of dispersal limitation close to zero indicate weak limitation, whereas values close to one indicate strong limitation. Overall, this framework assumes that dispersal success depends jointly on arrival probability and spatial redistribution, while allowing rare greater dispersal distances to disproportionately influence outcomes through the power-mean formulation.

Results

Effects of height and diel period on gust speed, frequency, and net airflow

Highest raw gust speeds occurred within the trunk zone (~20 m) and lowest gust speeds below the low crown zone (5–10 m; Appendix 4: Figure S1A) The interaction of height and position (Appendix 4: Figure S1A,B) on raw gust speeds was significant ($p < 0.001$, KW- $H_{df=13,n=9,864,421} = 632,588$; Dunn's $p < 0.001$ for all except one combination (at 30–35 m, the difference between the in and out position was $p < 0.005$). Height also significantly increased 1-minute maximum up-gust speeds ($p < 0.001$, KW- $H_{df=3,n=36,946} = 11,895$; Dunn's $p < 0.001$ for all combinations; Appendix 4: Figure S2). Across heights, the majority of up-gusts remained below 1 m s^{-1} , though we also recorded up-gust speeds $> 5 \text{ m s}^{-1}$ at all heights except in the 5–10 m category, where up-gusts did not exceed 2 m s^{-1} , and up-gust speeds $> 10 \text{ m s}^{-1}$ at 0–5 m and 15–20 m (Appendix 4: Figure S2).

692 The frequency of up- and down-gusts increased with height (Figure 3A, Appendix 4: Figure
 693 S3A), peaked at around noon, and plummeted to nearly 0 occurrences at night. Consistent with
 694 this, height ($p < 0.001$, KW- $H_{df=5,n=558,442} = 140,206$), diel period ($p < 0.001$, KW- $H_{df=1,n=558,442} =$
 695 4931), and their interaction ($p < 0.001$, KW- $H_{df=1,n=558,442} = 144,770$) significantly increased
 696 hourly up-gust frequency as well as down-gust frequency (height: $p < 0.001$, KW- $H_{df=5,n=605,061} =$
 697 70,003, diel period: $p < 0.001$, KW- $H_{df=1,n=605,061} = 14,895$, height $\times$ diel period: $p < 0.001$, KW-
 698 $H_{df=11,n=605,061} = 98,232$). For instance, near the forest floor, the frequency of up-gusts peaked at
 699 50 gusts per hour, while at 25–30 m it peaked at more than 200 gusts per hour (Figure 3A).
 700 Down-gusts were relatively more common than up-gusts from the forest floor up to ca. 15 m
 701 (Appendix 4: Figure S3A), where up-gusts became more dominant, increasing from 0.1 near the
 702 forest floor to 0.3–0.4 in the high crowns (25–30 m; difference between heights: $p < 0.001$, KW-
 703 $H_{df=13,n=4728} = 1542$). Yet, overall, horizontal gusts prevailed at all heights (Appendix 4: Figure
 704 S3A).

705 Mean hourly gust speeds showed similar diurnal variation as gust frequency (Figure 3B,
 706 Appendix 4: Figure S3B), with peaks at noon and lows throughout the night at all heights. Again,
 707 height ($p < 0.001$, KW- $H_{df=6,n=2,031,308} = 784,187$), diel period ($p < 0.001$, KW- $H_{df=1,n=2,031,308} =$
 708 75,220), and their interaction ($p < 0.001$, KW- $H_{df=13,n=2,031,308} = 896,367$) significantly increased
 709 the hourly mean of up-gust speeds and of down-gust speeds (height: $p < 0.001$, KW-
 710 $H_{df=6,n=1,841,283} = 523,294$, diel period: $p < 0.001$, KW- $H_{df=1,n=1,841,283} = 8,146$, height $\times$ diel period
 711 $p < 0.001$, KW- $H_{df=13,n=1,841,283} = 567,638$). Both peaked at 25–30 and 30–35 m, though down-
 712 gusts were relatively stronger than up-gusts, and horizontal gusts were stronger yet (Appendix 4:
 713 Figure S3B).

Mean (signed) vertical gust speed, i.e., net airflow, mostly fell within a range of $\pm 0.25 \text{ m s}^{-1}$, though variation was substantial, particularly above 15 m (Figure 3C). Height ($p < 0.001$, KW- $H_{df=6, n=3,872,591} = 414,620$), diel period ($p < 0.001$, KW- $H_{1, 3,872,591} = 22,594$), and their interaction ($p < 0.001$, KW- $H_{df=13, n=3,872,591} = 456,575$) significantly increased net airflow. At night, net airflow was downward throughout the height gradient. During the day, net airflow was downward only below 15 m, increased relatively sharply toward 20 m, and became upward above that. This transition zone coincided with a pronounced decrease in plant area density (PAD) and its components (leaf area density [LAD] and bark area density [BAD]); Figure 3D), which was even more prominent when considering only the seventeen vertical profiles (Appendix 4: Figure S4).

Effects of combined environmental factors on within-canopy gust speed

The Bayesian model explained only a small proportion of the variation in mean signed gust speed (Bayesian $R^2 = 0.02$, 95% CrI = 0.02 to 0.03), indicating weak overall structure in the fitted mean. Smooth-term variation in the mean (μ) model was relatively small for above-canopy wind (sds = 0.03, 95% CrI < 0.01 to 0.11), PAD_{above} (sds = 0.16, 95% CrI = 0.06 to 0.45), PAD_{local} (sds = 0.2, 95% CrI = 0.08 to 0.55), and temperature (sds = 0.12, 95% CrI = 0.04 to 0.35). Predicted mean gust speed nevertheless varied with above-canopy wind speed, and there was evidence that this relationship depended on canopy structure (Figure 4A,B), with a weak positive interaction with PAD_{local} ($\beta < 0.01$, 95% CrI = 0 to < 0.01; Figure 4A) and PAD_{above} ($\beta = 0.01$, 95% CrI = 0 to < 0.01; Figure 4B). The fitted mean decreased towards negative values with higher above-canopy wind, especially when PAD_{local} was intermediate (Figure 4A). When plotted against above-canopy wind speed, μ was more positive at low PAD_{above} but declined more strongly as above-canopy wind speed increased (Figure 4B). At higher PAD_{above} , μ

remained closer to zero across above-canopy wind speeds. Additional interactions can be explored in our shinyApp (Appendix 2).

In contrast, the σ sub-model showed substantially stronger structure, with larger smooth-term variation for above-canopy wind (sds = 1.71, 95% CrI = 0.97 to 2.96), PAD_{above} (sds = 0.79, 95% CrI = 0.35 to 1.68), PAD_{local} (sds = 1.06, 95% CrI = 0.48 to 2.01), and especially temperature (sds = 2.04, 95% CrI = 1.22 to 3.38). Residual spread increased strongly with above-canopy wind speed, and it decreased strongly with increasing PAD_{above} and PAD_{local} across the observed range (Figure 4A,B). Temperature was also strongly associated with σ , indicating overall that predictor variables were more strongly associated with variability around the mean than with the fitted mean itself. Variation among vertical profiles was modest (sd = 0.02, 95% CrI = 0.02 to 0.04), and the low Student- t degrees-of-freedom parameter indicated heavy-tailed residuals (ν = 1.95, 95% CrI = 1.86 to 2.04).

Diaspore dispersal

Based on the wind dynamics at San Lorenzo during the sampling period, *EpiDisp 1.0* predicts that, overall, approximately 74% of diaspores are effectively lost during primary dispersal: 16% of diaspores remain unabsorbed, 59% are intercepted by a leaf, 7% are still in dispersal at the end of the simulation, 2% fall to the ground. Less than 0.1% are dispersed above the canopy, and only 16% are intercepted by bark and, as per our definition, successfully dispersed (Appendix 4: Figure S5). These proportions, however, vary strongly with V_{term} , release height, and diel period (Appendix 4: Figure S5, Table S1), and differences become most evident at the extremes of the predictors' ranges. For example, more than 7% of diaspores with a V_{term} of 0.01 m s⁻¹ released (and absorbed; refer to Table 1 for the difference between released and absorbed) at 30–35 m

during the day dispersed above the canopy—but none below 5 m did. In fact, nearly all diaspores (> 99%) with a V_{term} of 1 m s^{-1} released below 10 m and at night remain unabsorbed. Between 10% and 30% of diaspores released below 5 m end up on the ground, whereas this proportion decreases to < 3% when released above 10 m.

Bark interception probability and long-distance dispersal probability were both affected by V_{term} , release height, and diel period. Higher V_{term} strongly reduced both bark interception probability (Figure 5A, Table 2) and above-canopy emergence probability (Figure 5B, Table 2). Release height and diel period showed contrasting effects: higher release heights and nighttime conditions decreased bark interception probability (Figure 5A, Table 2) but increased long-distance dispersal probability (Figure 5B, Table 2). Interaction terms indicated that these effects depended on one another, particularly for the combinations of release height and V_{term} , V_{term} and nighttime, and the three-way interaction among release height, V_{term} , and nighttime (Table 2).

Vertical and horizontal dispersal distances of bark-intercepted diaspores were also shaped by V_{term} , release height (Figure 6A,B, Appendix 4: Figures S6–S8), and diel period (Table 2).

According to GAMs, higher V_{term} reduced vertical dispersal distance but increased horizontal dispersal distance, whereas nighttime conditions reduced both vertical and horizontal distances (Table 2).

Within-canopy spread after successful dispersal varied with both V_{term} and release height: lower V_{term} and higher release heights were associated with broader distributions across horizontal distance and interception height, whereas higher V_{term} produced more compact distributions closer to the point of release and showed greater spread towards lower heights (Appendix 4: Figure S6). Kernel densities of bark-intercepted diaspores were overall concentrated near the

release point and declined steeply with distance in both the vertical and horizontal dimensions across nearly all release height $\times V_{term}$ groups (Appendix 4: Figure S7A,B). Diaspores released from 30–35 m tended to show a less distinct peak and more often reached greater dispersal distances. Across all groups, kernel density peaked strongly near 1 m from the release point and declined rapidly in all directions (Figure 6A,B). Downward, 75% of intercepted diaspores occurred within < 5 m of the release point, whereas upward, 95% occurred within $< \sim 6$ m of release point (Figure 6A). Horizontally, 50% of intercepted diaspores occurred within 5 m, 75% within 10 m, and 95% within 25 m of the release point (Figure 6B). It seemed that release height added less variation to vertical (Appendix 4: Figure S7A) than to horizontal spread (Appendix 4: Figure S7B), whereas terminal velocity strongly shaped both (Appendix 4: Figure S8A,B).

Dispersal limitation remained generally high across all release height $\times V_{term} \times$ diel period combinations (Figure 7, Appendix 4: Figure S9), typically ranging between $\sim 80\%$ and 100% when horizontal spread was moderately weighted ($\beta > 0.5$). The degree of dispersal limitation declined when neither vertical nor horizontal spread were weighted ($\alpha = 0, \beta = 0$) and with increasing power-mean exponent (q), reflecting the growing influence of rare longer distance dispersal events.

Increasing release height generally reduced dispersal limitation, with a slight peak in the trunk zone. This decline was most pronounced under daytime conditions and for low terminal velocities, particularly when vertical positioning was not considered ($\alpha = 0$). When vertical positioning was included ($\alpha > 0$), higher terminal velocities exhibited comparatively lower dispersal limitation, reflecting their greater downward dispersal.

Discussion

Understanding how dispersal operates within forest canopies is essential for explaining the spatial and vertical structure of epiphyte communities. Our results show that dispersal contributes to the patchy and vertically stratified distribution patterns commonly observed in epiphyte communities but is unlikely to explain these patterns alone. Instead, dispersal in forests emerges as a context-dependent, three-dimensional process in which success, horizontal and vertical dispersal distances, and fecundity jointly determine the degree to which diaspores reach suitable substrates. At the per-diaspore level, epiphytes appear strongly constrained: the probability of being intercepted by bark is low, and most successful diaspores remain close to their parent plant. However, this constraint varies with terminal velocity, release height, and canopy context, and may be substantially reduced at the individual or community level when high fecundity is considered. Thus, dispersal limitation should not be interpreted as a fixed property of epiphytes or other organisms, but as a scale- and context-dependent outcome of dispersal success, dispersal distances, and diaspore supply.

This perspective is important because wind dispersal in forests is often conceptualised from a terrestrial point of view, in which vegetation interception represents loss from the dispersal process and success is defined by arrival at ground-level habitat. Yet many wind-dispersed organisms and particles—canopy soil (and soil organisms), pollen, fungal spores, ballooning spiders—move through forests without necessarily requiring ground deposition. Epiphytes provide a particularly clear reversal of this framework: their diaspores must be intercepted by suitable aboveground substrates, such as bark, to disperse successfully. To capture this process, we developed *EpiDisp 1.0*, a mechanistic within-canopy dispersal model using epiphytes as a model system. In the following, we discuss how canopy context and diaspore traits shape

dispersal outcomes in epiphytes, while highlighting their broader relevance for wind dispersal in vertically structured environments.

Patterns of dispersal in a stratified forest canopy

Dispersal success in forests depends on the target surface of the dispersal unit. For epiphytes, dispersal success is defined by interception on suitable aboveground substrates, here modelled as bark. A central assumption in epiphyte ecology is that dispersal success is extremely low. Seed-trap studies, for example, report retention values of ~0.5% for epiphyte diaspores (Cascante-Marín et al., 2009; Mondragón & Calvo-Irabién, 2006; Paggi et al., 2010). Our results suggest that such methods may underestimate primary dispersal success in epiphytes. Using a mechanistic trajectory model, we estimate that ~16% of diaspores are initially intercepted by bark and thus disperse successfully during primary dispersal. This indicates that the initial dispersal bottleneck may be weaker than previously assumed, although most diaspores still fail to reach suitable substrates. The discrepancy likely reflects methodological differences. Seed-trap studies rely on the chance recovery of tiny diaspores at their final resting positions and may miss a large fraction of dispersal events. In contrast, *EpiDisp 1.0* tracks each diaspore to its first point of interception and therefore quantifies primary dispersal rather than the full dispersal pathway. Incorporating diaspore retention (Tay et al., work in progress) and secondary dispersal could refine these estimates by accounting for how many diaspores are later removed from—or added to—interception sites. Nevertheless, mechanistic estimates of primary interception are likely to remain higher than values inferred from seed-trap studies because they are not constrained by the incomplete recovery of dispersed diaspores. Even this higher estimate, however, indicates that dispersal represents a demographic bottleneck, consistent with earlier suggestions for epiphytes (e.g., Ackerman et al., 1996) and plants more broadly (van der Pijl, 1982).

For other wind-dispersed organisms, interception by vegetation may also constitute successful dispersal. This is evident, for instance, in ballooning arthropods such as spiders, which even show vertical stratification comparable to that in epiphytes (Costa et al., 2025). Unlike epiphyte diaspores, however, spiders may be able to move to a more suitable microhabitat after interception. In foliar pathogens, successful dispersal may occur when wind-borne fungal spores are intercepted directly by susceptible leaf tissue. In ash dieback, for example, wind-borne spores of *Hymenoscyphus fraxineus* enter ash (*Fraxinus* L. spp.) leaves before spreading into and infecting petioles and shoots (e.g., Mansfield et al., 2018). Our results suggest that, when leaves are the target substrate, interception rates can be very high (Appendix 4: Figure S5), which could help explain rapid local transmission in ash dieback. Thus, whether vegetation functions as a dispersal barrier or as a successful destination depends on the organism and, critically, on the surface required for establishment.

Dispersal success, however, captures only whether diaspores reach their destination, not how far they disperse. To understand dispersal in a three-dimensional environment, horizontal and vertical dispersal distances must also be considered. In wind-dispersed plants, dispersal kernels are typically characterised by a strong concentration near the source and a thin tail of rare long-distance events (e.g., Nathan & Muller-Landau, 2000; Qin et al., 2025). Our model reproduces this general pattern for both horizontal and vertical dispersal distances (Figure 6). As a consequence, most successful diaspores move only short distances. In a forest context, these short dispersal distances would often keep diaspores within the same or neighbouring trees and, vertically, within the same or adjacent strata, consistent with earlier observations on epiphyte distributions (Schimper, 1908; Torres et al., 2018; Zotz & Vollrath, 2003). Dispersal can

therefore contribute to the patchy and vertically stratified distribution patterns commonly observed in epiphyte communities.

In addition to bark interception and within-canopy dispersal distances, *EpiDisp 1.0* also quantified above-canopy dispersal, a separate outcome that may initiate long-distance dispersal (LDD). In our simulations, the probability of above-canopy dispersal remained low overall, especially for diaspores released from lower forest strata. Fewer than 0.1% of diaspores released below 10 m dispersed above the canopy, regardless of V_{term} , reinforcing the assumption that diaspores originating from lower forest strata rarely undergo LDD (Bohrer et al., 2008). In contrast, up to 8% of diaspores released between 30 and 35 m dispersed above the canopy (Figure 5B, Appendix 4: Table S1), likely because they experienced a shorter vertical distance to the canopy boundary and more favourable wind conditions in the high crowns, where upward airflow was more common at San Lorenzo (Figure 3C). Thus, although above-canopy dispersal was generally rare, its probability depended on the interaction between release height and V_{term} (discussed below).

Dispersal above the canopy, however, does not guarantee success. On the contrary, the farther a diaspore travels from its parent plant, the lower its chances for survival may become (Winkler et al., 2009), and even surviving diaspores might still fail to land in a suitable site. Consequently, our above-canopy dispersal estimates are only a coarse proxy for LDD, and effective LDD is likely to be lower due to post-dispersal demographic filters. In this sense, our results fully agree with the view that LDD events are rare in epiphytes (Givnish et al., 2016; Trapnell et al., 2013) and understorey plants in general (Bohrer et al., 2008).

Canopy context and diaspore traits shape dispersal

Terminal velocity emerged as the main driver of dispersal outcomes, but its effects depended on canopy context and on which dispersal outcome was considered. For bark interception, terminal velocity was more important than release height. Against our expectation, bark interception probability was broadly similar across release heights, while terminal velocity emerged as the stronger predictor of primary dispersal success (Table 2). This contrasts with studies identifying release height as a dominant driver of wind dispersal in forests (Heydel et al., 2014; Soons et al., 2004), although few studies have modelled release from the understorey, and none considered bark as the final destination. In *EpiDisp 1.0*, especially diaspores with $V_{term} < 0.5 \text{ m s}^{-1}$ were only weakly affected by release height. This may partly reflect our decision to keep vertical bark surface (PAD_v) constant in *EpiDisp 1.0*, which constrained interception probabilities similarly across heights. By contrast, diaspores with $V_{term} > 0.5 \text{ m s}^{-1}$ showed a near-linear increase in dispersal success with increasing release height, possibly because their predominantly downward trajectories (Appendix 4: Figure S6) exposed them to more cumulative bark surface before reaching the ground. However, this pattern may have limited relevance for epiphyte dispersal, because terminal velocities reported so far for wind-dispersed epiphyte diaspores do not appear to exceed 0.4 m s^{-1} —although they may approach higher values when wet or dispersed in clusters (Zotz et al., 2016). Thus, for most wind-dispersed epiphytes, primary dispersal success appears to be governed less by where diaspores are released than by how slowly they fall.

Although terminal velocity remained the main determinant of dispersal outcomes, release height became more important for dispersal distances and above-canopy dispersal than it was for bark interception. However, its effect appeared to be driven less by height per se than by the highly localised airflow associated with different forest zones. Diaspores released from the lower strata

were most dispersal-limited (Figure 7), showing lower dispersal success (Figure 5A), shorter dispersal distances (Figure 6A,B), and rarely dispersing to other strata, neighbouring trees, or above the canopy (Appendix 4: Figures S6–S8). Similar constraints on understorey dispersal have been observed in other forest systems (Heydel et al., 2014; Svahnström et al., 2025), likely reflecting dense vegetation and weak, possibly downward, airflow (Figure 3C). At San Lorenzo, conditions beneath the low crowns (<15 m)—predominantly downward flow, high *PAD*, and a weaker gust regime—would be expected to restrict the upward transport of ascending diaspores while retaining those descending from above, effectively creating a partial one-way filter. Yet this filter was not impermeable: although most diaspores released in the lower forest strata were retained locally, intermittent up-gusts appear to lift some diaspores above the low crowns (Appendix 4: Figure S6), or transport them from the low crowns beyond the canopy (Figure 5B).

In contrast, release from higher strata, and particularly the trunk zone, increased dispersal distances (Appendix 4: Figure S6), consistent with previous studies reporting greater dispersal distances from higher release points (Heydel et al., 2014; Lee & Yoon, 2025; Soons et al., 2004). These longer dispersal distances reduced dispersal limitation (Figure 7), likely due to stronger horizontal airflow in these layers (Appendix 4: Figure S3A,B). Yet stronger or more turbulent airflow did not promote all forms of dispersal.

The trunk zone, despite lower *PAD* (Figure 3D) and stronger, more turbulent airflow (Figure 3A–C), showed neither higher bark interception (Figure 5A) nor greater above-canopy dispersal, which instead declined (Figure 5B). This suggests that open, turbulent parts of the canopy may promote horizontal transport without necessarily facilitating vertical transport out of the forest.

By contrast, densely vegetated crown zones, although characterised by higher *PAD* and reduced airflow, do not necessarily impede dispersal. Instead, tree crowns may facilitate complex vertical

air movements, intermittently enabling uplift events while dampening horizontal winds (Pounden et al., 2008; Qin et al., 2022, 2025). This may explain the relatively high above-canopy dispersal probability from the low and high crown zones (Figure 5B), particularly the small peak in the low crown zone. The reduced prevalence and intensity of down-gusts at night (Figure 3A–C) may likewise help explain the greater nocturnal above-canopy dispersal observed for the diaspores with lowest V_{term} (Figure 5B, Table 2). Thus, intermediate gusts, rather than strong turbulence, may be particularly important for above-canopy dispersal in epiphytes, as also suggested for small diaspores more generally (Norros et al., 2014; Qin et al., 2025).

The effects of terminal velocity were comparatively consistent across these outcomes. Diaspores with low terminal velocities were transported farther horizontally and occasionally upward, whereas increasing terminal velocity progressively shifted dispersal downward, with terminal velocities $>0.5 \text{ m s}^{-1}$ producing the greatest downward dispersal distances (Appendix 4: Figures S6, S8A). Taken together, these results show that wind dispersal in forests is strongly structured by terminal velocity, but that canopy context determines how this trait translates into bark interception, horizontal and vertical dispersal distances, and above-canopy dispersal.

How dispersal-limited are epiphytes really?

In its strict sense, dispersal limitation quantifies the degree to which diaspores fail to arrive at all available suitable sites within their theoretical niche. Yet, in practice, it is often treated as a qualitative label, quantified only indirectly, or inferred from spatial patterns without testing whether environmental heterogeneity could also explain those patterns (Münzbergová & Herben, 2005). Existing quantitative approaches typically focus on expected seed shadows, recruitment limitation, or the proportion of sites reached relative to null expectations, and they explicitly recognise that arrival depends on both dispersal probabilities and diaspore supply (Clark et al.,

1999; Nathan & Muller-Landau, 2000; Terborgh et al., 2011). Building on these approaches, the metric developed here aims to quantify dispersal limitation from mechanistic evidence by integrating the probability of successful dispersal with horizontal and vertical dispersal distances and fecundity. This makes it possible to compare how limitation changes across diaspore traits, canopy positions, and biological scales, while retaining the distinction between low per-diaspore arrival probabilities and potentially high absolute numbers of arriving diaspores.

Because our dispersal limitation metric integrates dispersal success, dispersal distances, and fecundity, the context-dependent dispersal outcomes described above affected the degree of dispersal limitation (Figure 7). Overall, they translated into strong limitation at the per-diaspore level (fecundity = 1). Low bark interception probability and short horizontal and vertical dispersal distances together mean that most diaspores fail to colonise sites far from their parent plant.

However, this does not necessarily imply strong dispersal limitation at the level of individuals or communities. Epiphytes can produce vast numbers of diaspores, from <300 diaspores per individual in some *Tillandsia* spp. to several million diaspores per individual in many orchids (e.g., Arditti & Ghani, 2000; Croat, 1978; Toledo-Aceves, 2012; Zotz, 1997, 1998). The 0.4 ha subplot at San Lorenzo contains nearly 14,000 vascular epiphyte individuals (Mendieta-Leiva et al., 2022; Zotz & Schultz, 2008). Tentatively assuming these are sexually reproductive and produce an average of 100,000 diaspores each, this would imply 1.4 billion diaspores produced every year (see also Gómez Noguez & Flores Galván, 2019). Even if only 1% dispersed beyond 10 m vertically or horizontally, this would still yield 14 million diaspores escaping the immediate vicinity of their parent plant, which is still an intentionally conservative estimate (Figure 6A,B, Appendix 4: Table S1, Figures S7, S8). If only 0.1% dispersed above the canopy,

983 this would still translate to 1.4 million diaspores that can potentially colonise distant forests.

984 Such numbers exceed even reports of relatively high colonisation rates at San Lorenzo
985 (Mendieta-Leiva et al., 2022) and elsewhere (Einzmann & Zotz, 2017a,b), suggesting that strong
986 per-diaspore dispersal limitation can coexist with relatively low community-level dispersal
987 limitation.

988 High fecundity therefore changes the interpretation of dispersal limitation. It does not remove the
989 per-diaspore bottleneck, but it means that low probabilities of successful dispersal, movement
990 between strata, or dispersal to distant trees can still translate into large absolute numbers of
991 arriving diaspores. If many diaspores move beyond their parent plant, neighbouring trees, or
992 vertical zones, then dispersal alone should produce less distinct patches and weaker vertical
993 stratification than are commonly observed. Therefore, dispersal cannot be solely responsible for
994 epiphyte distribution patterns, and later demographic filters must strongly determine which
995 arrivals are ultimately retained, germinate, and establish (Trapnell et al., 2013; Wagner et al.,
996 2013; Winkler et al., 2009). Vertical zonation, as often demonstrated, is likely shaped more
997 strongly by epiphyte physiology, microclimatic gradients, host-tree characteristics, bark
998 properties, and other post-dispersal processes (Petter et al., 2016, 2021; Pittendrigh, 1948;
999 Wagner et al., 2013). Nevertheless, because release height and terminal velocity determine where
1000 diaspores arrive in the first place, dispersal should not be excluded from explanations of vertical
1001 niche preferences.

1002 Altogether, our study calls for a more nuanced and explicitly quantitative view of dispersal
1003 limitation. By integrating mechanistically derived dispersal success, horizontal and vertical
1004 dispersal distances, and fecundity into a single framework, our metric provides a way to quantify
1005 dispersal limitation beyond qualitative labels or indirect inference from spatial patterns. This

1006 approach complements existing concepts based on seed shadows or site occupancy (Clark et al.,
1007 1999; Nathan & Muller-Landau, 2000; Terborgh et al., 2011) by explicitly linking dispersal
1008 processes to spatial dispersal outcomes. Applied to epiphytes, it shows that strong limitation at
1009 the per-diaspore level can coexist with much weaker limitation at the community level once
1010 fecundity is taken into account.

1011 Importantly, this metric is inherently flexible, but its interpretation depends on the spatial scale
1012 and habitat definition considered. In its current form, it captures the dispersal component of
1013 limitation within the modelled plot, including whether diaspores reach a target surface and how
1014 far they disperse horizontally and vertically. However, it does not yet quantify the full spatial
1015 availability or suitability of recruitment sites. For example, it does not distinguish between
1016 occupied and available bark area, nor does it incorporate variation in bark suitability,
1017 microclimate, host-tree identity, or other factors that determine whether a site forms part of the
1018 theoretical niche of the species or dispersal unit of interest. Moreover, because our estimates are
1019 limited to the plot scale, expanding the spatial extent to larger forest areas, landscapes, or
1020 metacommunities would likely change the estimated degree of dispersal limitation. Processes
1021 that appear strongly limiting within a single plot may be less limiting when evaluated across
1022 larger populations with high diaspore production, while rare above-canopy dispersal events may
1023 become more important at broader spatial scales. These processes could be incorporated in future
1024 extensions, thereby moving closer to a strict definition of dispersal limitation as the failure to
1025 reach all suitable sites within the theoretical niche. Developing such an integrated metric
1026 represents a substantial step in itself, but it offers the potential to disentangle the relative
1027 contributions of dispersal, diaspore supply, habitat availability, spatial scale, and post-dispersal
1028 filtering to observed patterns. More broadly, by making these components explicit and

comparable, this framework provides a basis for more consistent, testable, and transferable interpretations of dispersal limitation across plant ecology (Münzbergová & Herben, 2005).

Wind dynamics across forest height and environmental conditions

Gust speed, frequency, and net airflow increased with height, indicating clear vertical structuring within the canopy. This is broadly consistent with previous studies showing that horizontal and vertical air movements generally increase from the forest floor towards the canopy (Dias-Júnior et al., 2017; Finnigan, 2000; Froelich et al., 2005; Heydel et al., 2014; Kruijt et al., 2000; Santana et al., 2018). On average, this increase has been described as exponential within and logarithmic above the canopy (Horn et al., 2001; Rodrigues et al., 2021), although profiles may vary within individual trees (Zhu et al., 2004). Our data diverged from such an average vertical profile and instead peaked in the trunk zone at ~20 m. This suggests that although height matters, its effect is modified, and in some layers even overridden, by local forest structure.

Wind dynamics at San Lorenzo were influenced by both atmospheric conditions, such as above-canopy wind and temperature, and forest structure (Figure 4A,B), although these effects were difficult to disentangle and relationships were generally weak. This is not unexpected given the complexity of the system (Froelich et al., 2005) and the many interacting controls involved, including radiation, temperature or humidity gradients, atmospheric stability (i.e., the resistance of the atmosphere against vertical motion), and topography (Bohrer et al., 2008; Brunet, 2020; Froelich et al., 2005; Tackenberg, 2003). With the diurnal cycle acting as an overarching control on several of these drivers, diel period emerged as a strong predictor of gust activity. Gusts became stronger and more frequent during the day and peaked around noon, consistent with diurnal patterns reported from other forests (Froelich et al., 2005; Maurer et al., 2013).

Mean vertical airflow within the forest was not predominantly upward during the day, as might be expected (Froelich et al., 2005; Maurer et al., 2013). Instead, at heights up to 15 m, mean vertical airflow remained downward both day and night. This discrepancy may partly reflect differences in measurement context, because earlier studies focused on airflow above the canopy and therefore did not capture within-canopy dynamics directly. Another possible explanation is that, at the very low wind speeds common in the lower half of the San Lorenzo forest, directional measurements become unreliable because of meandering (Anfossi et al., 2005). Even so, the daytime intensification of both up- and down-gusts may be important for driving diaspore movement through the forest interior.

Forest structure, expressed here through plant area density (*PAD*), appeared to be one of the key modifiers of wind dynamics in our study (Figures 3C,D, 4A,B). In general, turbulence increased where *PAD* decreased, with the clearest effect in the trunk zone (compare Figure 3C,D). Similar local peaks in airflow have been attributed to channelling or blow-through effects, whereby air is constricted and accelerated as it passes through gaps within or between crowns (e.g., Shaw, 1977; Zeng & Takahashi, 2000; Zhu et al., 2004). This same zone was also dominated by down-gusts (Figure 3B), consistent with studies describing strong sweeps or negative momentum fluxes in comparable forest layers (Kruijt et al., 2000; Rodrigues et al., 2021). At the same time, vertical airflow is often considered a “mean-zero process” (Maurer et al., 2013), in which horizontal turbulence generates both down- and up-gusts that largely cancel each other out over time. Indeed, mean net vertical airflow within the forest remained below $|0.25| \text{ m s}^{-1}$ overall (Figure 3C), as also described within and immediately above the canopy in other forests (Froelich et al., 2005; Heydel et al., 2014). Taken together, these patterns suggest that forest structure, especially gap dynamics, can redirect and intensify local airflow, creating aerodynamic

pathways that may enhance or inhibit dispersal within the canopy. This provides a mechanistic explanation for the contrasting dispersal outcomes among forest zones described above, including increased dispersal distances in the trunk zone but higher above-canopy dispersal from crown zones.

Limitations and perspectives

Several limitations of this study should be acknowledged, primarily with respect to wind-data quality. The temporal resolution of wind measurements, with one reading every 3 to 10 seconds (0.3 to 0.1 Hz), was relatively low. Heydel et al. (2014) recommend a minimum of 10 Hz to capture the effects of continuous turbulence on diaspore trajectories, and other studies have used even higher frequencies (Abdoli et al., 2023; Santana et al., 2018). Moreover, the anemometers used in this study had a detection threshold of 0.33 m s^{-1} , so weaker gusts went unrecorded. This threshold is substantial in the context of epiphyte dispersal, because many epiphyte diaspores have terminal velocities below this value and could potentially travel considerable distances in airflows weaker than those captured by our anemometers. However, because *EpiDisp 1.0* used mean wind speeds averaged over one-minute intervals, some effects of weak but persistent airflow may still have been captured in the model input. Further, horizontal winds were not filtered from the data, which may have obscured signals from vertical wind components (Abdoli et al., 2023). Especially strong horizontal winds passing through the anemometers could have induced turbulence within the sensor plates, falsely registering vertical motion (van der Molen et al., 2004). Additionally, ultrasonic anemometers are known to suffer from (co)sine response errors due to wind angle of attack (van der Molen et al., 2004). These errors should be assessed and corrected for the specific model used. Future studies could improve vertical wind

1096 measurements by shielding anemometers from horizontal winds using shrouds (Abdoli et al.,
1097 2023), although excluding horizontal winds may also obscure important dispersal dynamics.

1098 A particular challenge was the occurrence of extreme up-gusts of up to 32 m s^{-1} . These readings
1099 were an order of magnitude higher than the interpolated above-canopy gusts and may therefore
1100 foster doubt in the accuracy of the measurements. We failed to find comparable data, because
1101 vertical winds are commonly reported as time averages, normalised values, covariance,
1102 skewness, kurtosis, or turbulence intensity rather than as absolute maxima (Dias-Júnior et al.,
1103 2017; Finnigan, 2000; Kruijt et al., 2000; Santana et al., 2018). To the best of our knowledge, the
1104 highest absolute vertical winds reported from a forest under normal conditions, that is, without
1105 wildfires, were 7 m s^{-1} (Dias-Júnior et al., 2017). Some of our extreme gusts may therefore
1106 reflect measurement artefacts, for example if canopy elements such as leaves or twigs entered the
1107 measuring plate.

1108 However, we also tentatively suggest that some extreme values may represent apparent rather
1109 than actual winds. Apparent wind is the wind experienced by an object in motion, whereas actual
1110 wind is relative to the motionless surface of the earth (Appendix 1: Figure S1). Trees and
1111 branches are flexible and can sway considerably (Einzmann et al., 2022; Gardiner, 1995; Mayer,
1112 1987). When a flexible branch bends under external load, elastic potential energy accumulates
1113 and is released as kinetic energy when the load lifts, causing the branch or stem to spring back
1114 and forth until it reaches its resting position. Because the anemometers moved with the stems or
1115 branches to which they were attached, and sometimes independently of them (MH, pers.
1116 observation), they may have recorded apparent winds generated by this motion. When
1117 anemometer movement opposed the actual wind direction, apparent wind speeds could exceed
1118 actual wind speeds (Appendix 1: Figure S1; Video S1 and Video S1 Metadata).

1119 This distinction is biologically relevant because epiphytes growing on moving substrates will
1120 experience some component of the same apparent wind. The magnitude of this movement varies
1121 with tree and attachment height, architecture, foliage (or lack thereof), epiphyte load,
1122 neighbouring trees, and, naturally, wind intensity (Einzmann et al., 2022; Jackson et al., 2021;
1123 Mayer, 1987; Rudnicki et al., 2001). Epiphytes near the stem base should experience weaker
1124 apparent winds than those growing in crowns, where sway is greater (Einzmann et al., 2022).
1125 Reported sway amplitudes range from centimetres to metres (Gardiner, 1995; Jackson et al.,
1126 2021), and branch velocities can reach approximately 6 m s^{-1} (Milne, 1991), although absolute
1127 speeds have rarely been quantified. Such apparent turbulence may trigger diaspore release even
1128 under relatively weak winds, perhaps more effectively than strong but laminar airflow (Niklas,
1129 1992). Dynamic diaspore release under wind loading is a recurring theme in dispersal biology
1130 (Cousens et al., 2008; Murray, 1986; Salisbury, 1942) that, despite its importance for dispersal,
1131 remains poorly investigated. Thus, while apparent winds and possible contact artefacts may not
1132 fully explain all recorded extreme gusts, they highlight the need to consider substrate motion and
1133 apparent wind in dispersal models, because true wind alone may not adequately represent the
1134 aerodynamic conditions governing dispersal within forest canopies.

1135 *EpiDisp 1.0* incorporates apparent winds as epiphytes might experience them, but like other
1136 trajectory models, it relies on several simplifying assumptions. First, the model assumes that the
1137 abscission threshold equals the terminal velocity of diaspores, although gravity alone might
1138 sometimes suffice for abscission. The threshold may also decrease as diaspores mature or with
1139 changing environmental conditions, such as updrafts or a drop in relative humidity, whereas wet
1140 diaspores may not abscise at all (e.g., Cousens et al., 2008; Greene & Quesada, 2011).
1141 Incorporating a dynamic abscission threshold in future versions of *EpiDisp 1.0* or similar models

1142 could improve model accuracy, but experimental work is needed to quantify and parameterize
1143 such thresholds.

1144 Similarly, V_{term} is treated as constant, despite evidence that it can change under different
1145 environmental conditions. Diaspores may change shape when exposed to high humidity, thereby
1146 altering drag, lift forces, and ultimately V_{term} (e.g., Lee & Yoon, 2025; Seale et al., 2022). The
1147 model also assumes single-seed dispersal, although diaspores may disperse in clusters, as seen in
1148 orchids (Zotz et al., 2016) and bromeliads (MH, pers. observation). Such clustering may increase
1149 terminal velocity and, ultimately, dispersal success, although the range of terminal velocities
1150 modelled by *EpiDisp 1.0* may have covered at least some of these events.

1151 Interception probability was also simplified. It was based on mean plant area density, which
1152 cannot fully capture the complex vertical stratification and heterogeneity of forests (Lee & Yoon,
1153 2025). Although this structure was not represented explicitly in the interception term, it was
1154 included indirectly through wind speeds, which reflect local forest structure. We also ignored
1155 particle trapping: slow-falling diaspores, such as fern spores or orchid dust seeds, may remain
1156 trapped within an eddy for longer periods (Heydel et al., 2014; Horn et al., 2001; Nathan, Katul
1157 et al., 2011), potentially reducing interception likelihood. Finally, the model only accounts for
1158 primary dispersal, even though diaspores may later be uplifted by wind (and other agents) once
1159 more. In some cases, this would resemble release from a corresponding height already simulated
1160 by *EpiDisp 1.0*. Future work should therefore assess movement after first surface interception
1161 and may require dedicated secondary dispersal models that incorporate additional vectors such as
1162 water or animals.

1163 Addressing these limitations would enhance the predictive power of future dispersal models.

1164 Priorities include higher-resolution wind measurements, improved quantification of apparent

winds and substrate motion, more detailed forest-structure data, dynamic abscission thresholds, environmental effects on V_{term} , multi-seed dispersal, eddy dynamics, and secondary dispersal. An especially informative empirical research direction would be to test whether epiphyte morphology or substrate movement enhances diaspore abscission through locally generated turbulence.

Given the spatial and temporal complexity of forests, *EpiDisp 1.0* represents an important step towards a realistic approximation of within-canopy (epiphyte) dispersal. Unlike models relying on derived rather than measured wind data, it incorporates empirical within-canopy airflow, an abscission threshold, interception on both horizontal and vertical surfaces, and explicit effects of release height and terminal velocity. This is particularly relevant given the scarcity of mechanistic within-canopy dispersal models. Future work should validate model predictions through empirical approaches, such as trap-based dispersal studies (Mondragón & Calvo-Irabién, 2006) or wind-tunnel experiments (Bennett, 1988), and test whether the framework can be extended to other wind-dispersed canopy guilds, such as trees, lianas, understory plants, **as well as** pollen or fungal spores.

Conclusion

Epiphytes occupy the entire height gradient of many tropical forests, showing they have overcome the first major challenge of their life cycle: their diaspores landing on a suitable aboveground substrate. Yet this crucial aspect of their life history has remained poorly understood. Using a simulative trajectory model based on the wind conditions that epiphytes experience inside a tropical forest, this study provides evidence that dispersal limitation is high at a per-diaspore level: the probability of settling on suitable substrate is relatively low, and successful dispersal occurs mostly within a few meters of release. However, high fecundity

strongly reduces the degree of dispersal limitation at the community level. Our study therefore provides a mechanistic explanation for how dispersal contributes to patchy or vertically stratified epiphyte distributions, while also showing that they cannot be explained by dispersal alone. Overall, our study is a major advance in understanding epiphyte dispersal and within-canopy dispersal more broadly. By integrating tree-scale wind data into a mechanistic trajectory model, it provides a framework for studying dispersal in complex, vertically structured environments. Beyond epiphytes, similar approaches could be applied to other wind-dispersed organisms, including pollen, fungal spores (whether beneficial like mycorrhizae or harmful like those responsible for forest diseases such as ash dieback), understorey herbs, lianas, wind-dispersed trees, invasive plants, and ballooning arthropods. Such models could help inform conservation strategies aimed at improving habitat connectivity through corridors, isolated trees, or stepping-stone habitats that facilitate pollen-mediated gene flow and diaspore dispersal. Our study therefore opens new opportunities to understand how organisms overcome spatial barriers and how dispersal processes shape forest communities.

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1220 **Author contributions**

1221 Marie Hoensbroech and Gerhard Zotz conceived and designed the study. Marie Hoensbroech
1222 collected the wind data, conducted the analyses, and wrote the dispersal model with input from
1223 Juliano Sarmiento Cabral. Marie Hoensbroech wrote the first draft, and all authors subsequently
1224 revised the manuscript.

1225 **Conflict of Interest Statement**

1226 The authors declare no conflicts of interest.

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1228 **References**

- 1229 Abdoli, M., K. Lapo, J. Schneider, J. Olesch, and C. K. Thomas. 2023. “Toward Quantifying
1230 Turbulent Vertical Airflow and Sensible Heat Flux in Tall Forest Canopies Using Fiber-Optic
1231 Distributed Temperature Sensing.” *Atmospheric Measurement Techniques* 16: 809–24.
1232 <https://doi.org/10.5194/amt-16-809-2023>.
- 1233 Ackerman, J. D., A. Sabat, and J. K. Zimmerman. 1996. “Seedling Establishment in an Epiphytic
1234 Orchid: An Experimental Study of Seed Limitation.” *Oecologia* 106: 192–8.
- 1235 Agostinelli, C., and U. Lund. 2025. “R Package circular: Circular Statistics.” R Package Version
1236 0.5-2. <https://CRAN.R-project.org/package=circular>.
- 1237 Almeida, D., S. C. Stark, C. A. Silva, C. Hamamura, and R. Valbuena. 2021. “leafR: Calculates the
1238 Leaf Area Index (LAD) and Other Related Functions.” R Package Version 0.3.5.
1239 <https://CRAN.R-project.org/package=leafR>.
- 1240 Anfossi, D., D. Oettl, G. Degrazia, and A. Goulart. 2005. “An Analysis of Sonic Anemometer
1241 Observations in Low Wind Speed Conditions.” *Boundary-Layer Meteorology* 114: 179–203.
1242 <https://doi.org/10.1007/S10546-004-1984-4>.
- 1243 Aphalo, P. J. 2025. “ggpmisc: Miscellaneous Extensions to ‘ggplot2’.” R Package Version 0.6.2.
1244 <https://CRAN.R-project.org/package=ggpmisc>.
- 1245 Arditti, J., and A. K. A. Ghani. 2000. “Tansley Review No. 110. Numerical and Physical Properties
1246 of Orchid Seeds and Their Biological Implications.” *New Phytologist* 145(3): 367–421.
1247 <https://doi.org/10.1046/j.1469-8137.2000.00587.x>.

- 1248 Augspurger, C. 2024. *Seed Dispersal by Wind*. Washington DC: Smithsonian Institution Scholarly
 1249 Press. <https://doi.org/10.5479/si.26880697>.
- 1250 Augspurger, C. K., and S. E. Franson. 1988. "Input of Wind-Dispersed Seeds into Light-Gaps and
 1251 Forest Sites in a Neotropical Forest." *Journal of Tropical Ecology* 4(3): 239–52.
 1252 <https://doi.org/10.1017/S0266467400002807>.
- 1253 Barrett, T., M. Dowle, A. Srinivasan, J. Gorecki, M. Chirico, T. Hocking, B. Schwendinger, and I.
 1254 Krylov. 2025. "data.Table: Extension of 'data.frame'." R Package Version 1.17.8.
 1255 <https://CRAN.R-project.org/package=data.table>.
- 1256 Barrington, D. S. 1993. "Ecological and Historical Factors in Fern Biogeography." *Journal of*
 1257 *Biogeography*. 20(3): 275–79. <https://doi.org/10.2307/2845635>.
- 1258 Bennett, B. C. 1988. "A Comparison of Life History Traits in Selected Epiphytic and Saxicolous
 1259 Species of *Tillandsia* (Bromeliaceae) from Florida and Peru." PhD diss., University of North
 1260 Carolina.
- 1261 Benzing, D. H. 1990. *Vascular Epiphytes: General Biology and Related Biota*. Cambridge:
 1262 Cambridge University Press.
- 1263 Bivand, R., E. Pebesma, and V. Gomez-Rubio. 2013. *Applied Spatial Data Analysis with R*, 2nd ed.
 1264 New York: Springer.
- 1265 Bohrer, G., G. G. Katul, R. Nathan, R. L. Walko, and R. Avissar. 2008. "Effects of Canopy
 1266 Heterogeneity, Seed Abscission and Inertia on Wind-Driven Dispersal Kernels of Tree Seeds."
 1267 *Journal of Ecology* 96(4): 569–80. <https://doi.org/10.1111/j.1365-2745.2008.01368.x>.

- 1268 Brunet, Y. 2020. “Turbulent Flow in Plant Canopies: Historical Perspective and Overview.”
- 1269 *Boundary-Layer Meteorology* 177: 315–64. <https://doi.org/10.1007/S10546-020-00560-7>.
- 1270 Bürkner, P.-C. 2017. “Brms: An R Package for Bayesian Multilevel Models Using Stan.” *Journal*
- 1271 *of Statistical Software* 80:1–28. <https://doi.org/10.18637/jss.v080.i01>.
- 1272 Cascante-Marín, A., J. H. D. Wolf, J. G. B. Oostermeijer, and J. C. M. Den Nijs. 2008.
- 1273 “Establishment of Epiphytic Bromeliads in Successional Tropical Premontane Forests in Costa
- 1274 Rica.” *Biotropica* 40(4): 441–48. <https://doi.org/10.1111/j.1744-7429.2008.00403.x>.
- 1275 Cascante-Marín, A., N. von Meijenfeldt, H. M. H. de Leeuw, J. H. D. Wolf, J. G. B. Oostermeijer,
- 1276 and J. C. M. den Nijs. 2009. “Dispersal Limitation in Epiphytic Bromeliad Communities in a
- 1277 Costa Rican Fragmented Montane Landscape.” *Journal of Tropical Ecology* 25(1): 63–73.
- 1278 <https://doi.org/10.1017/S0266467408005622>.
- 1279 Chilpa-Galván, N., J. Márquez-Guzmán, G. Zotz, I. Echevarría-Machado, J. L. Andrade, C.
- 1280 Espadas-Manrique, and C. Reyes-García. 2018. “Seed Traits Favouring Dispersal and
- 1281 Establishment of Six Epiphytic *Tillandsia* (Bromeliaceae) Species.” *Seed Science Research*
- 1282 28(4): 349–59. <https://doi.org/10.1017/S0960258518000247>.
- 1283 Clark, J. S., M. Silman, R. Kern, E. Macklin, and J. HilleRisLambers. 1999. “Seed Dispersal near
- 1284 and far: Patterns across Temperate and Tropical Forests.” *Ecology* 80(5): 1475–94.
- 1285 [https://doi.org/10.1890/0012-9658\(1999\)080\[1475:sdnafp\]2.0.co;2](https://doi.org/10.1890/0012-9658(1999)080[1475:sdnafp]2.0.co;2).
- 1286 CloudCompare. 2023. *CloudCompare Version 2.13*. <http://www.cloudcompare.org/>.

- 1287 Cousens, R., C. Dytham, and R. Law. 2008. *Dispersal in Plants: A Population Perspective*. Oxford:
1288 Oxford University Press.
- 1289 Costa, R., N. Macías-Hernández, F. Rigal, P. A. V. Borges, and P. Cardoso. 2025. “Do Island
1290 Spiders Descend From Trees? – A Tale of Island Colonisation and Niche Expansion.” *Journal*
1291 *of Biogeography* 52: e70093. <https://doi.org/10.1111/jbi.70093>.
- 1292 Croat, T. B. 1978. *Flora of Barro Colorado Island*. Sanford: Stanford University Press.
- 1293 Dias-Júnior, C. Q., L. D. A. Sá, E. P. Marques Filho, R. A. Santana, M. Mauder, and A. O. Manzi.
1294 2017. “Turbulence Regimes in the Stable Boundary Layer above and within the Amazon
1295 Forest.” *Agricultural and Forestry Meteorology* 233: 122–32.
1296 <https://doi.org/10.1016/j.agrformet.2016.11.001>.
- 1297 Dislich, R., and W. Mantovani. 2016. “Vascular Epiphyte Assemblages in a Brazilian Atlantic
1298 Forest Fragment: Investigating the Effect of Host Tree Features.” *Plant Ecology* 217: 1–12.
1299 <https://doi.org/10.1007/S11258-015-0553-X>.
- 1300 Einzmann, H. J. R., and G. Zotz. 2017a. “Dispersal and Establishment of Vascular Epiphytes in
1301 Human-Modified Landscapes.” *AoB Plants* 9(6): plx052.
1302 <https://doi.org/10.1093/aobpla/plx052>.
- 1303 Einzmann, H. J. R., and G. Zotz. 2017b. “No Signs of Saturation: Long-Term Dynamics of
1304 Vascular Epiphyte Communities in a Human-Modified Landscape.” *Biodiversity and*
1305 *Conservation* 26: 1393–410. <https://doi.org/10.1007/s10531-017-1306-z>.

- 1306 Einzmann, H. J. R., G. Zotz, and J. Y. L. Tay. 2022. “What Happens to Epiphytic Bromeliads in a
1307 Windy Spot?” *Journal of Tropical Ecology* 38: 158–63.
1308 <https://doi.org/10.1017/S0266467422000037>.
- 1309 Esri Inc. 2024. *ArcGIS Pro Version 3.4.3*. <https://www.arcgis.com>.
- 1310 Finnigan, J. 2000. “Turbulence in Plant Canopies.” *Annual Review of Fluid Mechanics* 32: 519–71.
1311 <https://doi.org/10.1146/annurev.fluid.32.1.519>.
- 1312 Francisco, T. M., D. R. Couto, M. L. Garbin, F. Misaki, and C. R. Ruiz-Miranda. 2021. “Role of
1313 Spatial and Environmental Factors in Structuring Vascular Epiphyte Communities in Two
1314 Neotropical Ecosystems.” *Perspectives in Plant Ecology, Evolution and Systematics* 51:
1315 125621. <https://doi.org/10.1016/J.PPEES.2021.125621>.
- 1316 Froelich, N. J., H. P. Schmid, C. S. B. Grimmond, H.-B. Su, and A. J. Oliphant. 2005. “Flow
1317 Divergence and Density Flows above and below a Deciduous Forest: Part I. Non-Zero Mean
1318 Vertical Wind above Canopy.” *Agricultural and Forestry Meteorology* 133(1–4): 140–52.
1319 <https://doi.org/10.1016/j.agrformet.2005.09.005>.
- 1320 Gandawijaja, D., and J. Arditti. 1983. “The Orchids of Krakatau: Evidence for a Mode of
1321 Transport.” *Annals of Botany* 52(2): 127–30.
1322 <https://doi.org/10.1093/oxfordjournals.aob.a086558>.
- 1323 García-Franco, J. G., and V. Rico-Gray. 1988. “Experiments on Seed Dispersal and Deposition
1324 Patterns of Epiphytes—The Case of *Tillandsia deppeana* Steudel (Bromeliaceae).” *Phytologia*
1325 65: 73–8.

- 1326 Gardiner, B. A. 1995. "The Interactions of Wind and Tree Movement in Forest Canopies." In *Wind*
 1327 *and Trees*, edited by M. P. Coutts and J. Grace, 41–59. Cambridge: Cambridge University
 1328 Press.
- 1329 Garnier, S., N. Ross, R. Rudis, A. P. Camargo, M. Sciaini, and C. Scherer. 2024. "viridis(Lite)–
 1330 Colorblind-Friendly Color Maps for R." R Package Version 0.6.5.
 1331 <https://github.com/sjmgarnier/viridis>.
- 1332 Gegzna, V. 2020. "biostat: Routines for Basic (Bio)Statistics." R Package Version 0.0.20.
 1333 <https://github.com/gegznav/biostat/>.
- 1334 Givnish, T. J., M. H. J. Barfuss, B. Van Ee, R. Riina, K. Schulte, R. Horres, P. A. Gonsiska, et al.
 1335 2014. "Adaptive Radiation, Correlated and Contingent Evolution, and Net Species
 1336 Diversification in Bromeliaceae." *Molecular Phylogenetics and Evolution* 71: 55–78.
 1337 <https://doi.org/10.1016/j.ympev.2013.10.010>.
- 1338 Givnish, T. J., D. Spalink, M. Ames, S. P. Lyon, S. J. Hunter, A. Zuluaga, A. Doucette, et al. 2016.
 1339 "Orchid Historical Biogeography, Diversification, Antarctica and the Paradox of Orchid
 1340 Dispersal." *Journal of Biogeography* 43(10): 1905–16. <https://doi.org/10.1111/jbi.12854>.
- 1341 Gómez Noguez, F., and C. Flores Galván. 2019. "Ferns in the Air." *Bulletin of the American Fern*
 1342 *Society* 46: 1–6.
- 1343 Greene, D. F., and E. A. Johnson. 1996. "Wind Dispersal of Seeds from a Forest into a Clearing."
 1344 *Ecology* 77(2): 595–609. <https://doi.org/10.2307/2265633>.

- 1345 Greene, D. F., and M. Quesada. 2011. “The Differential Effect of Updrafts, Downdrafts and
 1346 Horizontal Winds on the Seed Abscission of *Tragopogon dubius*.” *Functional Ecology* 25(3):
 1347 468–72. <https://doi.org/10.1111/j.1365-2435.2010.01788.x>.
- 1348 Harper, J. L., J. N. Clatworthy, I. H. McNaughton, and G. R. Sagar. 1961. “The Evolution and
 1349 Ecology of Closely Related Species Living in the Same Area.” *Evolution* 15(2): 209–27.
 1350 <https://doi.org/10.1111/j.1558-5646.1961.tb03144.x>.
- 1351 Heydel, F., S. Cunze, M. Bernhardt-Römermann, and O. Tackenberg. 2014. “Long-Distance Seed
 1352 Dispersal by Wind: Disentangling the Effects of Species Traits, Vegetation Types, Vertical
 1353 Turbulence and Wind Speed.” *Ecological Research* 29(4): 641–51.
 1354 <https://doi.org/10.1007/s11284-014-1142-5>.
- 1355 Hijmans, R. J. 2025. “terra: Spatial Data Analysis.” R Package Version 1.8-80. [https://CRAN.R-](https://CRAN.R-project.org/package=terra)
 1356 [project.org/package=terra](https://CRAN.R-project.org/package=terra).
- 1357 Hijmans, R. J. 2025. “raster: Geographic Data Analysis and Modeling.” R Package Version 3.6–32.
 1358 <https://doi.org/10.32614/CRAN.package.raster>.
- 1359 Horn, H. S., R. A. N. Nathan, and S. R. Kaplan. 2001. “Long-Distance Dispersal of Tree Seeds by
 1360 Wind.” *Ecological Research* 16(5): 877–85. [https://doi.org/10.1046/J.1440-](https://doi.org/10.1046/J.1440-1703.2001.00456.X)
 1361 [1703.2001.00456.X](https://doi.org/10.1046/J.1440-1703.2001.00456.X).
- 1362 Horn, S., A. Raabe, H. Will, and O. Tackenberg. 2012. “TurbSeed—A Model for Wind Dispersal of
 1363 Seeds in Turbulent Currents Based on Publicly Available Climate Data.” *Ecological*
 1364 *Modelling* 237–238: 1–10. <https://doi.org/10.1016/J.ECOLMODEL.2012.04.009>.

- 1365 Inkscape's Contributors. 2023. *The Inkscape Project Version 1.3.2*. <https://inkscape.org>.
- 1366 Jackson, T. D., S. Sethi, E. Dellwik, N. Angelou, A. Bunce, T. van Emmerik, M. Duperat, et al.
 1367 2021. "The Motion of Trees in the Wind: A Data Synthesis." *Biogeosciences* 18(13): 4059–72.
 1368 <https://doi.org/10.5194/BG-18-4059-2021>.
- 1369 Kalacska, M., J. C. Calvo-Alvarado, and G. A. Sanchez-Azofeifa. 2005. "Calibration and
 1370 Assessment of Seasonal Changes in Leaf Area Index of a Tropical Dry Forest in Different
 1371 Stages of Succession." *Tree Physiology* 25(6): 733–44.
 1372 <https://doi.org/10.1093/treephys/25.6.733>.
- 1373 Kassambara, A. 2025. "rstatix: Pipe-Friendly Framework for Basic Statistical Tests." R Package
 1374 Version 0.7.3. <https://cran.r-project.org/package=rstatix>.
- 1375 Kelly, D. L. 1985. "Epiphytes and Climbers of a Jamaican Rain Forest: Vertical Distribution, Life
 1376 Forms and Life Histories." *Journal of Biogeography* 12(3): 223–41.
 1377 <https://doi.org/10.2307/2844997>.
- 1378 Kelly, D. L., G. O'Donovan, J. Feehan, S. Murphy, S. O. Drangeid, and L. Marciano-Berti. 2004.
 1379 "The Epiphyte Communities of a Montane Rain Forest in the Andes of Venezuela: Patterns in
 1380 the Distribution of the Flora." *Journal of Tropical Ecology* 20(6): 643–66.
 1381 <https://doi.org/10.1017/S0266467404001671>.
- 1382 Kessler, M. 2002. "Environmental Patterns and Ecological Correlates of Range Size among
 1383 Bromeliad Communities of Andean Forests in Bolivia." *The Botanical Review* 68: 100–27.
 1384 [https://doi.org/10.1663/0006-8101\(2002\)068\[0100:EPAECO\]2.0.CO;2](https://doi.org/10.1663/0006-8101(2002)068[0100:EPAECO]2.0.CO;2).

- 1385 Komsta, L., and F. Novomestky. 2022. “moments: Moments, Cumulants, Skewness, Kurtosis and
1386 Related Tests.” R Package Version 0.14.1. <https://CRAN.R-project.org/package=moments>.
- 1387 Krömer, T., M. Kessler, and S. R. Gradstein. 2007. “Vertical Stratification of Vascular Epiphytes in
1388 Submontane and Montane Forest of the Bolivian Andes: The Importance of the Understory.”
1389 *Plant Ecology* 189: 261–78. <https://doi.org/10.1007/s11258-006-9182-8>.
- 1390 Kruijt, B., Y. Malhi, J. Lloyd, A. D. Norbre, A. C. Miranda, M. G. P. Pereira, A. Culf, and J. Grace.
1391 2000. “Turbulence Statistics above and within Two Amazon Rain Forest Canopies.”
1392 *Boundary-Layer Meteorology* 94: 297–331. <https://doi.org/10.1023/A:1002401829007>.
- 1393 Ladino, G., F. Ospina-Bautista, J. Estévez Varón, L. Jerabkova, and P. Kratina. 2019. “Ecosystem
1394 Services Provided by Bromeliad Plants: A Systematic Review.” *Ecology and Evolution* 9(12):
1395 7360–72. <https://doi.org/10.1002/ECE3.5296>.
- 1396 Lee, S., and T. K. Yoon. 2025. “Why Are Seed Dispersal Models Rarely Used? Limitations of
1397 Scalability and Improvement Measures.” *Forests* 16(5): 851.
1398 <https://doi.org/10.3390/f16050851>.
- 1399 Madison, M. 1977. “Vascular Epiphytes: Their Systematic Occurrence and Salient Features.”
1400 *Selbyana* 2: 1–13.
- 1401 Mac Nally, R., and C. J. Walsh. 2004. “Hierarchical Partitioning Public-Domain Software.”
1402 *Biodiversity and Conservation* 13: 659–60.
1403 <https://doi.org/10.1023/B:BIOC.00000009515.11717.0b>.

- 1404 Mansfield, J. W., N. Galambos, and R. Saville. 2018. “The Use of Ascospores of the Dieback
1405 Fungus *Hymenoscyphus fraxineus* for Infection Assays Reveals a Significant Period of
1406 Biotrophic Interaction in Penetrated Ash Cells.” *Plant Pathology* 67: 1354–61.
1407 <https://doi.org/10.1111/ppa.12844>.
- 1408 Maurer, K. D., G. Bohrer, D. Medvigy, and S. J. Wright. 2013. “The Timing of Abscission Affects
1409 Dispersal Distance in a Wind-Dispersed Tropical Tree.” *Functional Ecology* 27(1): 208–18.
1410 <https://doi.org/10.1111/1365-2435.12028>.
- 1411 Mayer, H. 1987. “Wind-Induced Tree Sways.” *Trees* 1: 195–206.
1412 <https://doi.org/10.1007/BF01816816>.
- 1413 Mendieta-Leiva, G., H. L. Buckley, and G. Zotz. 2022. “Directional Changes over Time in the
1414 Species Composition of Tropical Vascular Epiphyte Assemblages.” *Journal of Ecology*
1415 110(3): 553–68. <https://doi.org/10.1111/1365-2745.13817>.
- 1416 Mendieta-Leiva, G., K. Wagner, and G. Zotz. 2021. “Vascular Epiphyte Census Data at the San
1417 Lorenzo Crane Plot [dataset].” *Zenodo*. <https://doi.org/10.5281/zenodo.5645775>.
- 1418 Microsoft. 2024. *Microsoft 365 tools*. <https://microsoft.com/microsoft-365>.
- 1419 Milne, R. 1991. “Dynamics of Swaying of *Picea sitchensis*.” *Tree Physiology* 9(3): 383–99.
1420 <https://doi.org/10.1093/TREEPHYS/9.3.383>.
- 1421 Mondragón, D., and L. M. Calvo-Irabién. 2006. “Seed Dispersal and Germination of the Epiphyte
1422 *Tillandsia brachycaulos* (Bromeliaceae) in a Tropical Dry Forest, Mexico.” *The Southwestern*
1423 *Naturalist*. 51(4): 462–70. <https://www.jstor.org/stable/20424755>.

- 1424 Mondragón, D., T. Valverde, and M. Hernández-Apolinar. 2015. "Population Ecology of Epiphytic
1425 Angiosperms: A Review." *Tropical Ecology* 56(1): 1–39.
1426 <https://doi.org/10.13140/2.1.4043.5849>.
- 1427 Muller-Landau, H. C., S. J. Wright, O. Calderón, R. Condit, and S. P. Hubbell. 2008. "Interspecific
1428 Variation in Primary Seed Dispersal in a Tropical Forest." *Journal of Ecology* 96(4): 653–67.
1429 <https://doi.org/10.1111/J.1365-2745.2008.01399.X>.
- 1430 Münzbergová, Z., and T. Herben. 2005. "Seed, Dispersal, Microsite, Habitat and Recruitment
1431 Limitation: Identification of Terms and Concepts in Studies of Limitations." *Oecologia* 145:
1432 1–8. <https://doi.org/10.1007/s00442-005-0052-1>.
- 1433 Murray, D. R. 1986. *Seed Dispersal*. Sydney: Academic Press.
- 1434 Murray, K., and M. M. Conner. 2009. "Methods to Quantify Variable Importance: Implications for
1435 the Analysis of Noisy Ecological Data." *Ecology* 90(2): 348–55. [https://doi.org/10.1890/07-](https://doi.org/10.1890/07-1929.1)
1436 1929.1.
- 1437 Nathan, R., H. S. Horn, J. Chave, and S. A. Levin. 2002. "Mechanistic Models for Tree Seed
1438 Dispersal by Wind in Dense Forests and Open Landscapes." In *Seed Dispersal and Frugivory:
1439 Ecology, Evolution and Conservation*, edited by D. J. Levey, W. R. Silva, and M. Galetti, 69–
1440 82. Wallingford: CAB International.
- 1441 Nathan, R., N. Horvitz, Y. He, A. Kuperinen, F. M. Schurr, and G. G. Katul. 2011. "Spread of
1442 North American Wind-Dispersed Trees in Future Environments." *Ecology Letters* 14(3): 211–
1443 19. <https://doi.org/10.1111/J.1461-0248.2010.01573.X>.

- 1444 Nathan, R., G. G. Katul, G. Bohrer, A. Kuparinen, M. B. Soons, S. E. Thomson, A. Trakhtenbrot,
 1445 and H. S. Horn. 2011. “Mechanistic Models of Seed Dispersal by Wind.” *Theoretical Ecology*
 1446 4: 113–32. <https://doi.org/10.1007/s12080-011-0115-3>.
- 1447 Nathan, R., and H. C. Muller-Landau. 2000. “Spatial Patterns of Seed Dispersal, their Determinants
 1448 and Consequences for Recruitment.” *Trends in Ecology & Evolution* 15(7): 278–85.
 1449 [https://doi.org/10.1016/S0169-5347\(00\)01874-7](https://doi.org/10.1016/S0169-5347(00)01874-7).
- 1450 Neuwierth, E. 2022. “RColorBrewer: ColorBrewer Palettes.” R Package Version 1.1–3.
 1451 <https://CRAN.R-project.org/package=RColorBrewer>.
- 1452 Nieder, J., and G. Zotz. 1998. “Methods of Analyzing the Structure and Dynamics of Vascular
 1453 Epiphyte Communities.” *Ecotropica* 4: 33–9.
- 1454 Niklas, K. J. 1992. *Plant Biomechanics: An Engineering Approach to Plant Form and Function*.
 1455 Chicago: University of Chicago Press.
- 1456 Norros, V., Ü. Rannik, T. Hussein, T. Petäjä, T. Vesala, and O. Ovaskainen. 2014. “Do Small
 1457 Spores Disperse Further than Large Spores?” *Ecology* 95(6): 1612–21.
 1458 <https://doi.org/10.1890/13-0877.1>.
- 1459 Olivas, P. C., S. F. Oberbauer, D. B. Clark, D. A. Clark, M. G. Ryan, J. J. O’Brien, and H.
 1460 Ordoñez. 2013. “Comparison of Direct and Indirect Methods for Assessing Leaf Area Index
 1461 Across a Tropical Rain Forest Landscape.” *Agricultural and Forest Meteorology* 177: 110–6.
 1462 <https://doi.org/10.1016/j.agrformet.2013.04.010>.

- 1463 Onset Computer Corporation. 2022. *HOBOWare Version 3.7.26*.
- 1464 <https://www.onsetcomp.com/products/software/hoboware>.
- 1465 Paggi, G. M., J. A. T. Sampaio, M. Bruxel, C. Martini Zanella, M. Göetze, M. Valli Büttow, C.
- 1466 Palma-Silva, and F. Bered. 2010. "Seed Dispersal and Population Structure in *Vriesea*
- 1467 *gigantea*, a Bromeliad from the Brazilian Atlantic Rainforest." *Botanical Journal of the*
- 1468 *Linnean Society* 164: 317–25. <https://doi.org/10.1111/j.1095-8339.2010.01088.x>.
- 1469 Partomihardjo, T. 2003. "Colonisation of Orchids on the Krakatau Islands." *Telopea* 10(1): 299–
- 1470 310. <https://doi.org/10.7751/TELOPEA20035620>.
- 1471 Paton, S. 2019a. "Galeta–Wind Direction [dataset]." *Smithsonian Tropical Research Institute*.
- 1472 <https://doi.org/10.25573/data.10042640.v25>.
- 1473 Paton, S. 2019b. "Galeta–Wind Speed [dataset]." *Smithsonian Tropical Research Institute*.
- 1474 <https://doi.org/10.25573/data.10042658.v23>.
- 1475 Paton, S. 2019c. "San Lorenzo Crane–Wind Speed [dataset]." *Smithsonian Tropical Research*
- 1476 *Institute*. <https://doi.org/10.25573/data.10042688.v23>.
- 1477 Paton, S. 2019d. "San Lorenzo Crane–Wind direction [dataset]." *Smithsonian Tropical Research*
- 1478 *Institute*. <https://doi.org/10.25573/data.10042718.v23>.
- 1479 Paton, S. 2019e. "San Lorenzo Crane–Air Temperature [dataset]." *Smithsonian Tropical Research*
- 1480 *Institute*. <https://doi.org/10.25573/data.10042694.v23>.
- 1481 Paton, S. 2019f. "San Lorenzo Crane–Precipitation [dataset]." *Smithsonian Tropical Research*
- 1482 *Institute*. <https://doi.org/10.25573/data.10042679.v23>.

- 1483 Paton, S. 2020. “Yearly Reports–San Lorenzo Crane [dataset].” *Smithsonian Tropical Research*
 1484 *Institute*. <https://doi.org/10.25573/data.11799309.v6>.
- 1485 Pebesma, E. 2018. “Simple Features for R: Standardized Support for Spatial Vector Data.” *The R*
 1486 *Journal* 10: 439–46. <https://doi.org/10.32614/RJ-2018-009>.
- 1487 Pedersen, T. L. 2025. “patchwork: The Composer of Plots.” R Package Version 1.3.2.
 1488 <https://CRAN.R-project.org/package=patchwork>.
- 1489 Petter, G., K. Wagner, W. Wanek, E. J. Sánchez Delgado, G. Zotz, J. S. Cabral, and H. Kreft. 2016.
 1490 “Functional Leaf Traits of Vascular Epiphytes: Vertical Trends within the Forest, Intra-and
 1491 Interspecific Trait Variability, and Taxonomic Signals.” *Functional Ecology* 30(2): 188–98.
 1492 <https://doi.org/10.1111/1365-2435.12490>.
- 1493 Petter, G., G. Zotz, H. Kreft, and J. S. Cabral. 2021. “Agent-Based Modeling of the Effects of
 1494 Forest Dynamics, Selective Logging, and Fragment Size on Epiphyte Communities.” *Ecology*
 1495 *and Evolution* 11(6): 2937–51. <https://doi.org/10.1002/ece3.7255>.
- 1496 Pittendrigh, C. S. 1948. “The Bromeliad-*Anopheles*-Malaria Complex in Trinidad. I–The
 1497 Bromeliad Flora.” *Evolution* 2(1): 58–89. <https://doi.org/10.1111/j.1558-5646.1948.tb02732.x>.
- 1498 Pohlert, T. 2024. “PMCMRplus: Calculate Pairwise Multiple Comparisons of Mean Rank Sums.”
 1499 R Package Version 1.9.12. <https://CRAN.R-project.org/package=PMCMRplus>.
- 1500 Pouden, E., D. F. Greene, M. Quesada, and J. M. Contreras Sánchez. 2008. “The effect of
 1501 collisions with vegetation elements on the dispersal of winged and plumed seeds.” *Journal of*
 1502 *Ecology* 96: 591–8. <https://doi.org/10.1111/j.1365-2745.2008.01380.x>.

1503 Proença-Ferreira, A., L. Borda-de-Água, M. Porto, F. Moreira, A. Mira, and R. Pita. 2025. “dispfit:
 1504 Fit Distributions to Species Dispersal Data.” R Package Version 0.1.0.
 1505 <https://github.com/apferreira/dispfit>.

1506 Qin, X., W. Liang, M. Liu, Z. Xin, Z. Wang, Q. Zhou, L. Tian, L. Zong, J. Zhu, and Z. Liu. 2025.
 1507 “Effects of Plant Crown Characteristics on Dispersal Kernels of Wind-Dispersed Seeds under
 1508 Low Wind Speeds.” *Journal of Plant Ecology* 18(1): rtæ106.
 1509 <https://doi.org/10.1093/jpe/rtæ106>.

1510 Qin, X., W. Liang, Z. Liu, C. C. Baskin, J. M. Baskin, Z. Xin, Z. Wang, and Q. Zhou. 2022. “Plant
 1511 Canopy May Promote Seed Dispersal by Wind.” *Scientific Reports* 12: 63.
 1512 <https://doi.org/10.1038/s41598-021-03402-9>.

1513 R Core Team. 2023. *R: A Language and Environment for Statistical Computing*. Vienna: R
 1514 Foundation for Statistical Computing. <https://www.R-project.org/>.

1515 Ramos, P., P. Villareal, and H. Muller-Landau. 2024. “Tree Height Allometry Data for 5109 Trees
 1516 in Twenty-Four Small ForestGEO Plots in Panama, 2011–2020 [dataset].” *Smithsonian*
 1517 *Tropical Research Institute*. <https://doi.org/10.25573/data.24954204.v1>.

1518 Rodrigues, A., R. A. Sardinha, and G. Pita. 2021. “Exchange of Energy and Mass over Forest
 1519 Canopies.” In *Fundamental Principles of Environmental Physics*, 105–32. Cham: Springer.
 1520 https://doi.org/10.1007/978-3-030-69025-0_4.

1521 Roth, W. E. 1986. *Stratification of a Tropical Forest as Seen in Dispersal Types*. Dordrecht:
 1522 Springer Netherlands.

- 1523 Roussel, J.-R., D. Auty, N. C. Coops, P. Tompalski, T. R. H. Goodbody, A. Sánchez Meador, J.-F.
 1524 Bourdon, F. de Boissieu, and A. Achim. 2020. "LidR: An R Package for Analysis of Airborne
 1525 Laser Scanning (ALS) data." *Remote Sensing of Environment* 251: 112061.
 1526 <https://doi.org/10.1016/j.rse.2020.112061>.
- 1527 Rudnicki, M., U. Silins, V. J. Liefers, and G. Josi. 2001. "Measure of Simultaneous Tree Sways
 1528 and Estimation of Crown Interactions among a Group of Trees." *Trees* 15: 83–90.
 1529 <https://doi.org/10.1007/s004680000080>.
- 1530 Ruiz-Cordova, J. P., V. H. Toledo-Hernández, and A. Flores-Palacios. 2014. "The Effect of
 1531 Substrate Abundance in the Vertical Stratification of Bromeliad Epiphytes in a Tropical Dry
 1532 Forest (Mexico)." *Flora* 209: 375–84. <https://doi.org/10.1016/J.FLORA.2014.06.003>.
- 1533 Salisbury, E. 1942. *The Reproductive Capacity of Plants*. London: Bells and Sons.
- 1534 Santana, R. A., C. Q. Dias-Júnior, J. Tóta, J. D. Fuentes, R. S. Vale, E. G. Alves, R. M. N. dos
 1535 Santos, and A. O. Manzi. 2018. "Air Turbulence Characteristics at Multiple Sites in and above
 1536 the Amazon Rainforest Canopy." *Agricultural and Forestry Meteorology* 260–261: 41–54.
 1537 <https://doi.org/10.1016/j.agrformet.2018.05.027>.
- 1538 Schimper, A. 1908. *Pflanzen-Geographie auf physiologischer Grundlage*, 2nd ed. Jena: Gustav
 1539 Fischer.
- 1540 Schimper, A. 1888. *Die epiphytische Vegetation Amerikas*. Jena: Gustav Fischer.

1541 Seale, M., O. Zhdanov, M. B. Soons, C. Cummins, E. Kroll, M. R. Blatt, H. Zare-Behtash, et al.
 1542 2022. “Environmental Morphing Enables Informed Dispersal of the Dandelion Diaspore.”
 1543 *eLife* 11: e81962. <https://doi.org/10.7554/eLife.81962>.
 1544 Shaw, R. H. 1977. “Secondary Wind Speed Maxima inside Plant Canopies.” *Journal of Applied*
 1545 *Meteorology* 16: 514–521. [https://doi.org/10.1175/1520-](https://doi.org/10.1175/1520-0450(1977)016%3C0514:SWSMIP%3E2.0.CO;2)
 1546 [0450\(1977\)016%3C0514:SWSMIP%3E2.0.CO;2](https://doi.org/10.1175/1520-0450(1977)016%3C0514:SWSMIP%3E2.0.CO;2).
 1547 Soons, M. B., G. W. Heil, R. Nathan, and G. G. Katul. 2004. “Determinants of Long-Distance
 1548 Dispersal by Wind in Grasslands.” *Ecology* 85(11): 3056–68. <https://doi.org/10.1890/03-0522>.
 1549 Soons, M. B., and W. A. Ozinga. 2005. “How Important is Long-Distance Seed Dispersal for the
 1550 Regional Survival of Plant Species?” *Diversity and Distributions* 11(2): 165–72.
 1551 <https://doi.org/10.1111/J.1366-9516.2005.00148.X>.
 1552 Spicer, M. E., and J. Ortega. 2023. “Source Height and Contact with Terrestrial Soil Drive
 1553 Transplanted Epiphyte Performance.” *Journal of Ecology* 111(11): 2388–400.
 1554 <https://doi.org/10.1111/1365-2745.14187>.
 1555 Spicer, M. E., and C. L. Woods. 2022. “A Case for Studying Biotic Interactions in Epiphyte
 1556 Ecology and Evolution.” *Perspectives in Plant Ecology, Evolution and Systematics* 54:
 1557 125658. <https://doi.org/10.1016/J.PPEES.2021.125658>.
 1558 Svahnström, V. J., E. N. Lughadha, F. Forest, and T. C. C. Leão. 2025. “Geographic Range Size
 1559 and Rarity of Epiphytic Flowering Plants.” *Nature Plants* 11: 1380–89.
 1560 <https://doi.org/10.1038/s41477-025-02022-9>.

- 1561 Tackenberg, O. 2003. "Modeling Long-Distance Dispersal of Plant Diaspores by Wind."
 1562 *Ecological Monographs* 73(2): 173–89. [https://doi.org/10.1890/0012-](https://doi.org/10.1890/0012-9615(2003)073[0173:MLDOPD]2.0.CO;2)
 1563 9615(2003)073[0173:MLDOPD]2.0.CO;2.
- 1564 Tackenberg, O., P. Poschlod, and S. Kahmen. 2003. "Dandelion Seed Dispersal: The Horizontal
 1565 Wind Speed Does Not Matter for Long-Distance Dispersal - It Is Updraft!" *Plant Biology* 5(5):
 1566 451–4. <https://doi.org/10.1055/S-2003-44789>.
- 1567 Taylor, A., and K. Burns. 2015. "Plant Composition Patterns inside an Endemic Birds' Nest Fern
 1568 (*Asplenium goudeyi*) on Lord Howe Island: Effects of Fern Size, Fern Isolation and Plant
 1569 Dispersal Abilities." *Journal of Tropical Ecology* 31(5): 413–21.
 1570 <https://doi.org/10.1017/S0266467415000334>.
- 1571 Terborgh, J., P. Alvarez-Loayza, K. Dexter, F. Cornejo, and C. Carrasco. 2011. "Decomposing
 1572 Dispersal Limitation: Limits on Fecundity or Seed Distribution?" *Journal of Ecology* 99: 935–
 1573 44. <https://dx.doi.org/10.1111/j.1365-2745.2011.01836.x>.
- 1574 Thomsen, M. S., A. H. Altieri, C. Angelini, M. J. Bishop, P. E. Gribben, G. Lear, Q. He, et al.
 1575 2018. "Secondary foundation species enhance biodiversity". *Nature Ecology and Evolution* 2:
 1576 634–639. <https://doi.org/10.1038/s41559-018-0487-5>.
- 1577 Toledo-Aceves, T., J. G. García-Franco, S. Landero Lozada, M. L. León Mateos, and K.
 1578 MacMillan. 2012. "Germination and Seedling Survivorship of Three *Tillandsia* Species in the
 1579 Cloud-Forest Canopy." *Journal of Tropical Ecology* 28(4): 423–26.
 1580 <https://doi.org/10.1017/S0266467412000363>.

1581 Torres, E., M. L. Riofrío, and J. M. Iriondo. 2018. “Complex Fine-Scale Spatial Genetic Structure
 1582 in *Epidendrum rhopalostele*: An Epiphytic Orchid”. *Heredity* 122: 458–67.
 1583 <https://doi.org/10.1038/s41437-018-0139-1>.

1584 Trapnell, D. W., J. L. Hamrick, C. D. Ishibashi, and T. R. Kartzinel. 2013. “Genetic Inference of
 1585 Epiphytic Orchid Colonization; It May Only Take One.” *Molecular Ecology* 22(14): 3680–92.
 1586 <https://doi.org/10.1111/mec.12338>.

1587 Urbanek, S. 2025. “png: Read and Write PNG Images.” R Package Version 0.1-9. [https://CRAN.R-](https://CRAN.R-project.org/package=png)
 1588 [project.org/package=png](https://CRAN.R-project.org/package=png).

1589 van der Molen, M. K., J. H. C. Gash, and J. A. Elbers. 2004. “Sonic Anemometer (co)sine
 1590 Response and Flux Measurement: II. The Effect of Introducing an Angle of Attack Dependent
 1591 Calibration.” *Agricultural and Forest Meteorology* 122(1–2): 95–109.
 1592 <https://doi.org/10.1016/J.AGRFORMET.2003.09.003>.

1593 van der Pijl, L. 1982. *Principles of Dispersal in Higher Plants*, 3rd ed. Berlin: Springer.

1594 Vasquez V., M. Garcia, M. Hernandez, ForestGEO Smithsonian, and H. Muller-Landau. 2024.
 1595 “Smithsonian ForestGEO San Lorenzo and Panama Small Plots Aerial Photogrammetry
 1596 Orthomosaics, Digital Surface Models, Point Clouds and Raw Images for 2015–2024.
 1597 [dataset]” *Smithsonian Research Data Repository*. <https://doi.org/10.60635/C33W2K>.

1598 Vieira, E. M., and P. Izar. 1999. “Interactions Between Aroids and Arboreal Mammals in the
 1599 Brazilian Atlantic Rainforest.” *Plant Ecology* 145: 75–82.
 1600 <https://doi.org/10.1023/A:1009859810148>.

1601 Wagner, K., W. Bogusch, and G. Zotz. 2013. "The Role of the Regeneration Niche for the Vertical
1602 Stratification of Vascular Epiphytes." *Journal of Tropical Ecology* 29(4): 277–90.
1603 <https://doi.org/10.1017/S0266467413000291>.

1604 Wickham, H. 2023. "modelr: Modelling Functions that Work with the Pipe." R Package Version
1605 0.1.11. <https://doi.org/10.32614/CRAN.package.modelr>.

1606 Wickham, H., M. Averick, J. Bryan, W. Chang, L. D'Agostino McGowan, R. François, G.
1607 Grolemund, et al. 2019. "Welcome to the Tidyverse." *Journal of Open Source Software* 4(43):
1608 1686. <https://doi.org/10.21105/joss.01686>.

1609 Wickham, H., T. Pedersen, and D. Seidel. 2025. "scales: Scale Functions for Visualisation." R
1610 Package Version 1.4.0. <https://CRAN.R-project.org/package=scales>.

1611 Wilke, C. O., and B. M. Wiernik. 2022. "ggtext: Improved Text Rendering Support for 'ggplot2'."
1612 R Package Version 0.1.2. <https://CRAN.R-project.org/package=ggtext>.

1613 Winkler, M., K. Hülber, and P. Hietz. 2009. "Population Dynamics of Epiphytic Orchids in a
1614 Metapopulation Context." *Annals of Botany* 104(5): 995–1004.
1615 <https://doi.org/10.1093/AOB/MCP188>.

1616 POWO. 2026. Plants of the World Online. Facilitated by the Royal Botanic Gardens, Kew.
1617 Published on the Internet; <https://powo.science.kew.org/>. (accessed 29 June 2026)

1618 Retrieved 01 July 2026."Wright, S. J., V. Horlyck. Y. Basset, H. Barrios, A. Bethancourt, S. A.
1619 Bohlman, G. S. Gilbert, et al. 2003. "Tropical Canopy Biology Program, Republic of
1620 Panama." In *Studying Forest Canopies from above: The International Canopy Crane Network*

1621 edited by Y. Basset, V. Horlyck, and S. J. Wright, 137–55. New York: United Nations
 1622 Environmental Programme, Smithsonian Tropical Research Institute.

1623 Wright, S. J., A. Trakhtenbrot, G. Bohrer, M. Detto, G. G. Katul, N. Horvitz, H. C. Muller-Landau,
 1624 F. A. Jones, and R. Nathan. 2008. “Understanding Strategies for Seed Dispersal by Wind
 1625 under Contrasting Atmospheric Conditions.” *Proceedings of the National Academy of*
 1626 *Sciences USA* 105(49): 19084–9. <https://doi.org/10.1073/pnas.0802697105>.

1627 Xu, S., M. Chen, T. Feng, L. Zhan, L. Zhou, and G. Yu. 2021. “Use Ggbreak to Effectively Utilize
 1628 Plotting Space to Deal with Large Datasets and Outliers.” *Frontiers in Genetics* 12: 774846.
 1629 <https://doi.org/10.3389/fgene.2021.774846>.

1630 Zeng, P., and H. Takahashi. 2000. “A First-Order Closure Model for the Wind Flow within and
 1631 above Vegetation Canopies.” *Agricultural and Forest Meteorology* 103(3): 301-13.
 1632 [https://doi.org/10.1016/S0168-1923\(00\)00133-7](https://doi.org/10.1016/S0168-1923(00)00133-7).

1633 Zhu, J. J., X. F. Li, G. Yutaka, and M. Takeshi. 2004. “Wind Profiles in and over Trees.” *Journal of*
 1634 *Forestry Research* 15: 305-12. <https://doi.org/10.1007/BF02844959>.

1635 Zotz, G. 1997. “Substrate Use of Three Epiphytic Bromeliads.” *Ecography* 20(3): 264–70.
 1636 <https://doi.org/10.1111/J.1600-0587.1997.TB00370.X>.

1637 Zotz, G. 1998. “Demography of the Epiphytic Orchid, *Dimerandra emarginata*.” *Journal of*
 1638 *Tropical Ecology* 14(6): 725–41. <https://doi.org/10.1017/S0266467498000534>.

1639 Zotz, G. 2016. *Plants on Plants—The Biology of Vascular Epiphytes*. Cham: Springer.

- 1640 Zotz, G., and P. Hietz. 2001. "The Physiological Ecology of Vascular Epiphytes: Current
1641 Knowledge, Open Questions." *Journal of Experimental Botany* 52(364): 2067–78.
1642 <https://doi.org/10.1093/JEXBOT/52.364.2067>.
- 1643 Zotz, G., P. Hietz, and H. J. R. Einzmann. 2021. "Functional Ecology of Vascular Epiphytes."
1644 *Annual Plant Reviews Online* 4(4): 869–906. <https://doi.org/10.1002/9781119312994.apr0777>.
- 1645 Zotz, G., and S. Schultz. 2008. "The Vascular Epiphytes of a Lowland Forest in Panama—Species
1646 Composition and Spatial Structure." *Plant Ecology* 195: 131–41.
1647 <https://doi.org/10.1007/s11258-007-9310-0>.
- 1648 Zotz, G., and B. Vollrath. 2003. "The Epiphyte Vegetation of the Palm *Socratea exorrhiza* -
1649 Correlations with Tree Size, Tree Age and Bryophyte Cover." *Journal of Tropical Ecology*
1650 19(1): 81–90. <https://doi.org/10.1017/S0266467403003092>.
- 1651 Zotz, G., T. Weichgrebe, H. Happatz, and H. J. R. Einzmann. 2016. "Measuring the Terminal
1652 Velocity of Tiny Diaspores." *Seed Science Research* 26(3): 222–30.
1653 <https://doi.org/10.1017/S0960258516000155>.
- 1654 Zotz, G., P. Weigelt, M. Kessler, H. Kreft, and A. Taylor. 2021. "EpiList 1.0: A Global Checklist of
1655 Vascular Epiphytes." *Ecology* 102(6): e03326. <https://doi.org/10.1002/ecy.3326>.

1656 **Boxes**

1657 (none)

1659 **Table 1.** Glossary of key terms used in this paper (in alphabetical order).

Term	Definition
Abscission	The point at which a diaspore detaches from the seed pod and dispersal begins.
Actual wind	The wind relative to the motionless surface of the earth.
Apparent wind	The wind an object in motion experiences.
Bark area density (<i>BAD</i>)	The bark area (unit^2) per forest volume (unit^3).
Diaspore	The plant reproductive unit that is dispersed, consisting of a seed or spore and any additional tissues or structures that facilitate dispersal.
Dispersal limitation	A continuous measure of how strongly dispersal constraints restrict a species from occupying its full theoretical niche, e.g., by limiting the distance or likelihood of reaching suitable sites.
Effective dispersal	Successful dispersal of a diaspore combined with subsequent success in germination and establishment.
Ejection	Gradual upward air motion.
Leaf area density (<i>LAD</i>)	The leaf area (unit^2) per forest volume (unit^3).
Long-distance dispersal (<i>LDD</i>)	The movement of a diaspore outside of the spatial limits of a stand and/or the genetic neighbourhood area of individuals.
Mean up/down-gust speed	The time average of either up-gusts or down-gusts occurring over a specified time interval.
Mean vertical gust speed	The time average of up- and down-gusts over a specified time interval, where up-gusts contribute positively and down-gusts negatively.
Plant area density (<i>PAD</i>)	Total plant area density constituting both leaf and wood components.
Primary dispersal	The first dispersal phase, in which a diaspore moves away from the parent plant immediately after release until its first deposition, as opposed to secondary dispersal, which refers to any subsequent movement after that initial deposition.
Release height	The height at which a diaspore is first exposed to the environment, either through fruit opening or at the start of the simulation.
Safe site	A site at which a dispersed diaspore settles and where conditions are suitable for it to germinate (also see vertical niche.)
Successful dispersal	The settling of an abscised epiphyte diaspore on suitable substrate (here: bark) of a suitable host after primary dispersal,

Term	Definition
	recognising that this state is not necessarily final and may be modified through secondary dispersal.
Sweep	Sudden downward air motion.
Terminal velocity (V_{term})	The constant speed at which a diaspore falls in still air after initial acceleration.
Up/down-gusts	The measured maxima of upward or downward flowing air during a specified time interval.
Up/down-gust frequency	The number of up-gusts or down-gusts during a specified time interval.
Up/down-winds	The time averages of upward or downward flowing air calculated over a 1-minute or 5-minute interval.
Vertical niche	The portion of the forest's vertical profile in which abiotic conditions are suitable for a given epiphyte species to establish and persist; differences among species' vertical niches can produce vertical stratification.
Vertical stratification	Non-random distribution of epiphyte species along host-tree or canopy height gradients, typically attributed to microclimatic variation.

1660 Note: terms appear in when first mentioned in the text.

Table 2. Parameter estimates for predictors of bark interception, canopy dispersal, and dispersal distance in epiphyte diaspores.

Model	Term	Estimate	Std. error	Statistic	p-value
Bark interception probability (GAM)	Intercept (Release height = 0, $V_{term} = 0$, Day)	-1.2	< 0.1	-120.5	< 0.01
	Release height	< 0.1	< 0.1	-4.9	< 0.01
	V_{term}	-1.5	< 0.1	-67.4	< 0.01
	Night	-0.2	< 0.1	-12.6	< 0.01
	Release height $\times V_{term}$	< 0.1	< 0.1	33.1	< 0.01
	Release height $\times$ Night	< 0.1	< 0.1	10.5	< 0.01
	$V_{term} \times$ Night	-0.9	< 0.1	-22.8	< 0.01
	Release height $\times V_{term} \times$ Night	< 0.1	< 0.1	2.1	0.1
Vertical distance at bark interception (GLM)	Intercept (Release height = 0, $V_{term} = 0$, Day)	0.5	< 0.1	19.4	< 0.01
	Release height	-0.1	< 0.1	-21.3	< 0.01
	V_{term}	-2.3	< 0.1	-33.2	< 0.01
	Night	-0.4	< 0.1	-9.0	< 0.01
	Release height $\times V_{term}$	-0.1	< 0.01	-28.0	< 0.01
	Release height $\times$ Night	< 0.1	< 0.01	4.3	< 0.01
	$V_{term} \times$ Night	0.5	0.1	3.8	< 0.01
	Release height $\times V_{term} \times$ Night	< 0.1	< 0.1	-14.0	< 0.01
Horizontal distance at bark interception (GLM)	Intercept (Release height = 0, $V_{term} = 0$, Day)	3.6	< 0.1	49.6	< 0.01
	Release height	< 0.1	< 0.1	24.0	< 0.01
	V_{term}	0.9	0.2	4.7	< 0.01
	Night	< 0.1	0.1	0.1	0.9
	Release height $\times V_{term}$	< 0.1	< 0.1	1.7	0.1
	Release height $\times$ Night	-0.1	< 0.1	-5.6	< 0.01
	$V_{term} \times$ Night	0.8	0.3	2.3	0.1
	Release height $\times V_{term} \times$ Night	< 0.1	< 0.1	-0.1	0.9
Above-canopy emergence probability (GAM)	Intercept (Release height = 0, $V_{term} = 0$, Day)	-7.7	< 0.1	-97.0	< 0.01
	Release height	0.2	< 0.01	51.1	< 0.01
	V_{term}	-5.1	0.4	-12.1	< 0.01
	Night	1.6	0.1	15.1	< 0.01
	Release height $\times V_{term}$	< 0.1	< 0.1	2.1	0.1
	Release height $\times$ Night	-0.1	< 0.1	-10.6	< 0.01
	$V_{term} \times$ Night	-7.3	1.0	-7.7	< 0.01
	Release height $\times V_{term} \times$ Night	< 0.1	< 0.1	2.1	0.1

Note: Generalized additive models (GAMs) tested the effects of terminal velocity (V_{term}), release height, diel period, and their interaction on the probability of bark interception (relative to

1665 landing on the ground, leaf interception, or non-abscission) and canopy dispersal. Generalized
1666 linear models (GLMs) tested the effects of the same predictors on vertical and horizontal
1667 dispersal distances. Shown are estimated regression coefficients, standard errors, test statistics (z
1668 for GAMs, t for GLMs), and p -values.

Figure captions

Figure 1. Map of the study site in the San Lorenzo National Park and its location within Panama (inset). Colours represent the height of the canopy, generated using lidar data; dots mark the locations of the seventeen vertical profiles. White lines and text show altitude contours.

Figure 2. Flow chart of the mechanistic dispersal (trajectory) model *EpiDisp 1.0*. Abbreviations used in the figure: Terminal velocity = V_{term} , time at release = T , time increment after release = T' , height at release = H , height after abscission = H' , plant area density = PAD , bark area density = BAD .

Figure 3. Day–night variation in wind dynamics **(A–B)** and across the height gradient **(C)** alongside different vegetation indices **(D)**. **(A)** Mean hourly frequency of up- and down-gusts $\geq 0.33 \text{ m s}^{-1}$ (the lowest wind speed measurable by the anemometers) per 5 m height category. Because the first three vertical profiles were sampled at longer time intervals, up-gust frequency was computed only from the fourth vertical profile onward, resulting in no 30–35 m category in this panel. **(B)** Mean hourly speed of up- and down-gusts with measurements of 0 m s^{-1} included. **(C)** Mean hourly vertical wind speed, coloured by datetime, with all measurements of 0 m s^{-1} included, with down-gusts contributing negative and up-gusts positive values. The smoothed lines go through the means per 5 m height category. **(D)** Plant area density (PAD) and its components bark area density (BAD) and leaf area density (LAD) along the forest height gradient, averaged across the San Lorenzo 1 ha permanent plot. PAD was calculated from lidar data provided by Vasquez et al. (2024); the derivation BAD and LAD leaned heavily on previous censuses at San Lorenzo (Janner & Fernández-Lucero, work in progress; Ramos et al., 2024). The silhouette in the background illustrates the forest structure and was adapted from a graphic

generated with Microsoft Copilot (Microsoft, 2024). Points and whiskers show means and standard errors (SE) in panels A-C. In most cases, SE is much smaller than symbol size.

Figure 4. Model-predicted mean gust value (μ , solid lines) and residual variation (σ , dashed lines) as functions of above-canopy wind speed and plant area density. **(A)** Predictions across PAD_{local} ($0.1\text{--}0.4\text{ m}^2\text{ m}^{-3}$), with colour indicating above-canopy wind speed. **(B)** Predictions across above-canopy wind speed, with colour indicating PAD_{above} ($1\text{--}4\text{ m}^2\text{ m}^{-3}$). All other covariates were held at their median reference values. Shaded ribbons indicate uncertainty around predictions. PAD indices were calculated from lidar data from the San Lorenzo plot provided by Vasquez et al. (2024), and above-canopy wind speed was derived from Paton (2019a–d).

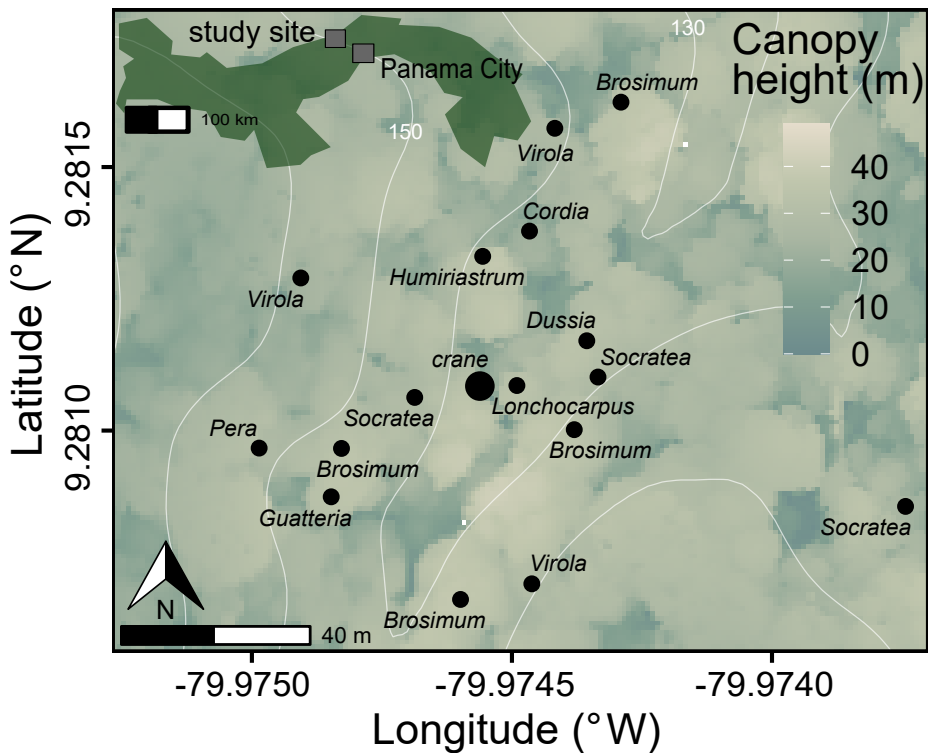
Figure 5. Effects of release height, terminal velocity, and diel period on dispersal outcomes. **(A)** Predicted bark interception probability as a function of release height, shown for each terminal velocity and separated by day and night. **(B)** Predicted above-canopy emergence probability, used here as a proxy for long-distance dispersal probability, as a function of release height, shown for each terminal velocity and separated by day and night. Curves are fitted GAM smooths using cubic regression splines ($k = 5$), and grey ribbons indicate 95% confidence intervals. Note the different y-axis scales.

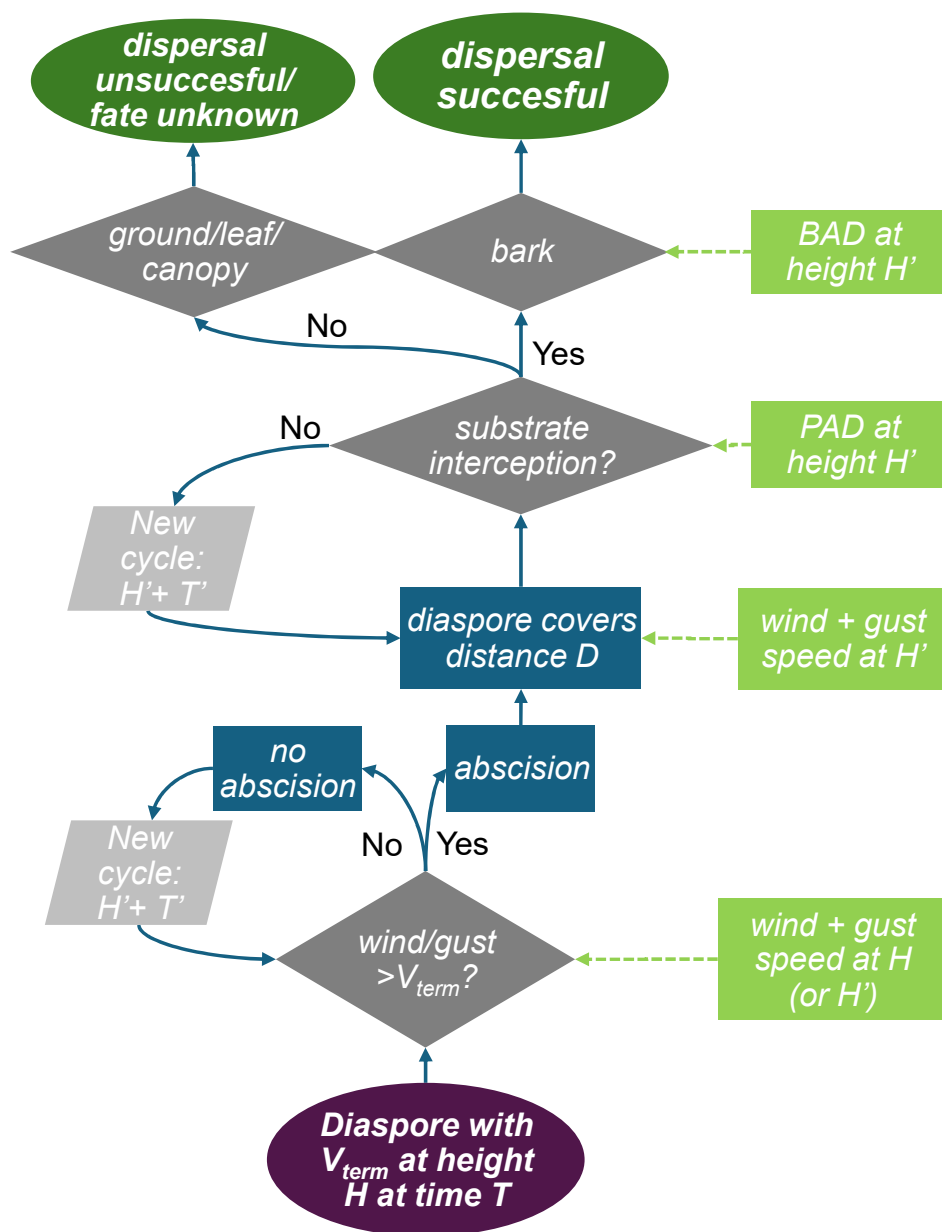
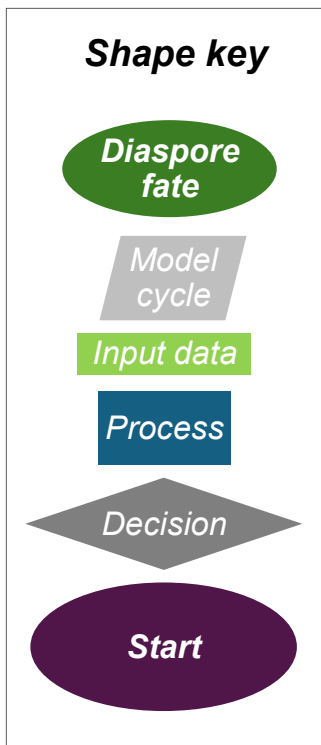
Figure 6. Distance density distribution of bark-intercepted diaspores. **(A)** Vertical and **(B)** horizontal dispersal kernels (black line) with standard errors (grey shading) calculated from the dispersal kernels in Supplement 2: Figure S7. Dashed vertical lines mark the distances at which the cumulative dispersed population reaches 50%, 75%, and 95%, respectively. Note the pseudo-log transformed x-axis and the different axis scales.

1713 **Figure 7.** Dispersal limitation as a function of release height and terminal velocity under varying
1714 fecundity levels. Points represent release height $\times V_{term}$ group simulation outcomes during
1715 daytime and are sized by bark interception probability (%), lines show smoothed trends across
1716 release height, and colours indicate terminal velocity (m s^{-1}). Dispersal limitation was calculated
1717 assuming strong weighting of both vertical and horizontal displacement ($\alpha = 1, \beta = 1$), such that
1718 both vertical and horizontal spread contribute proportionally to dispersal success. Contributions
1719 were aggregated using a power mean of order $q = 2$, which increases the influence of longer
1720 distance dispersal events relative to the arithmetic mean while still retaining contributions from
1721 the broader distribution. Note the truncated y-axis.

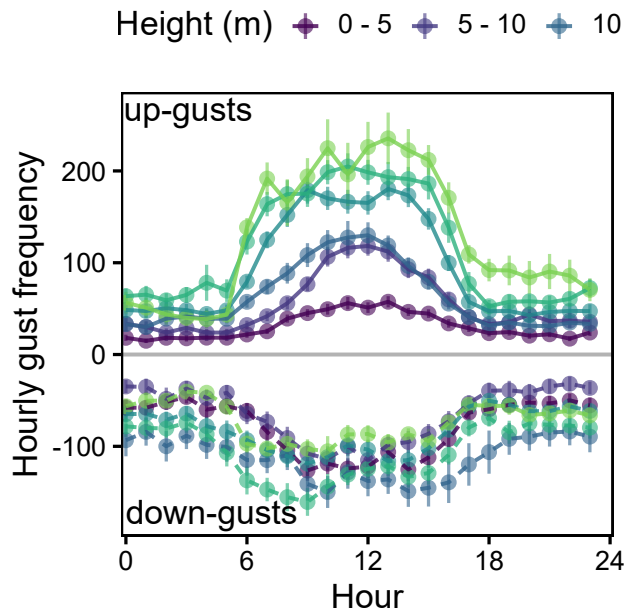
1722 **Figures**

1723 Uploaded as separate files.

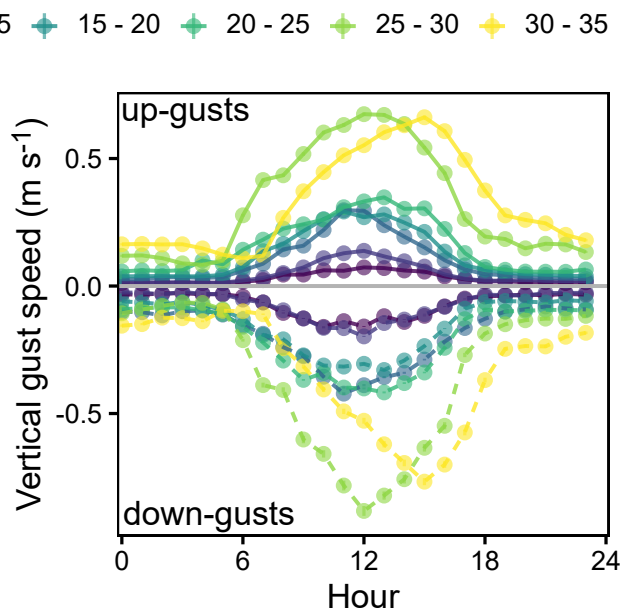




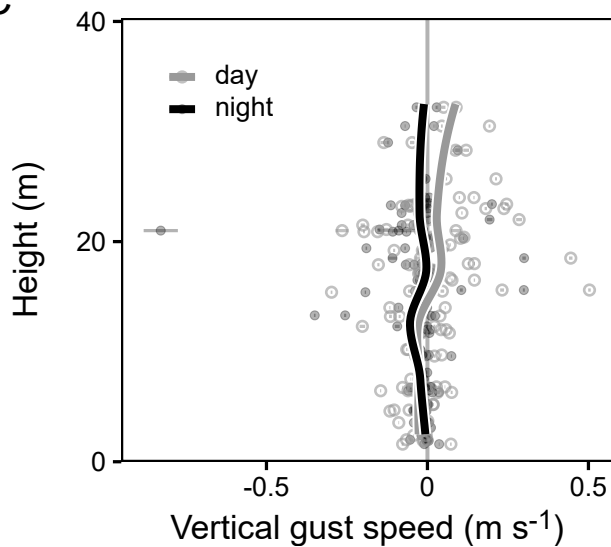
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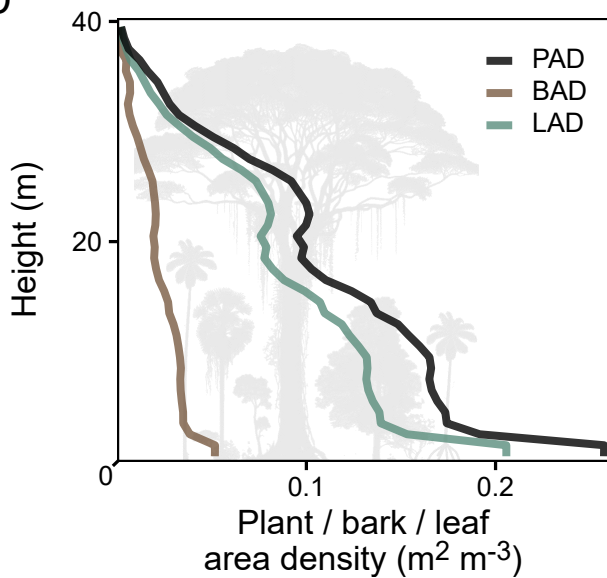
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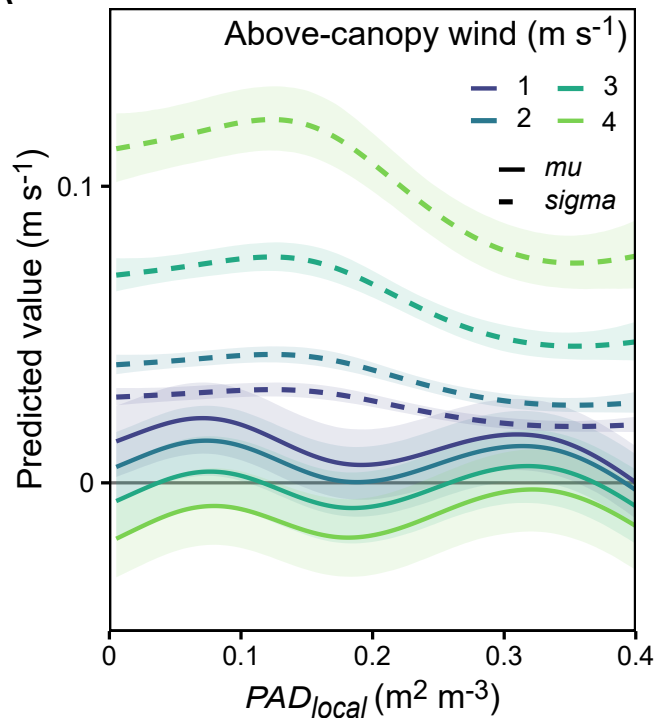
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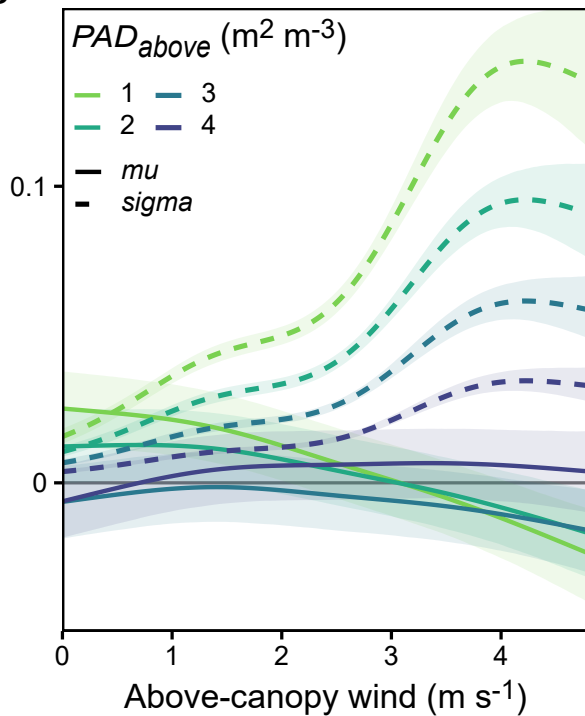
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A

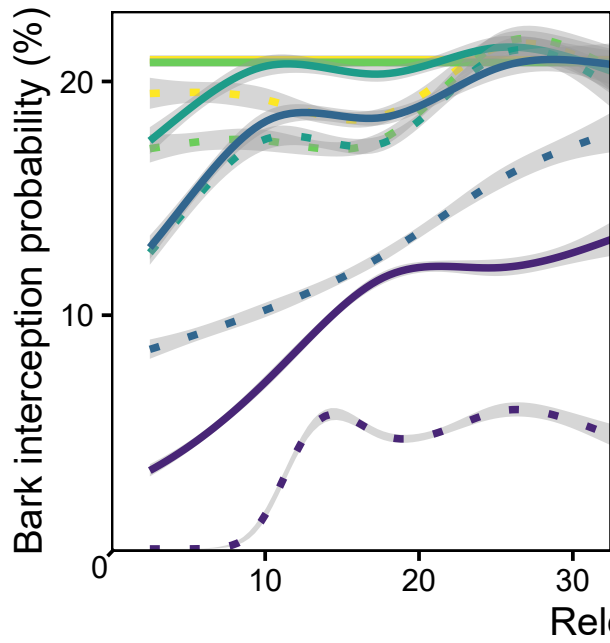


B

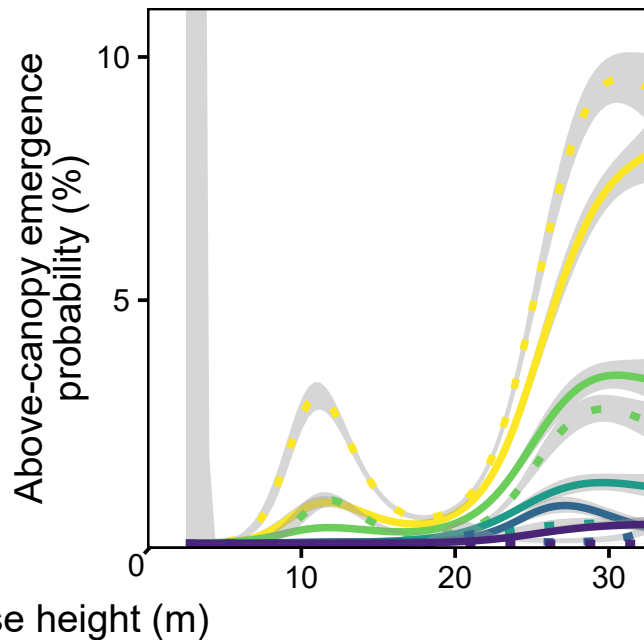


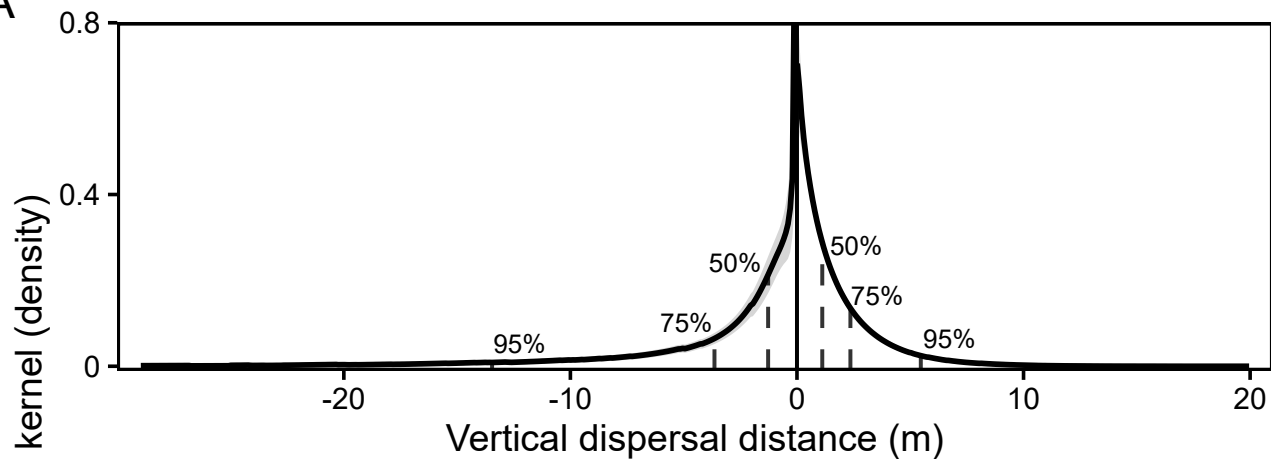
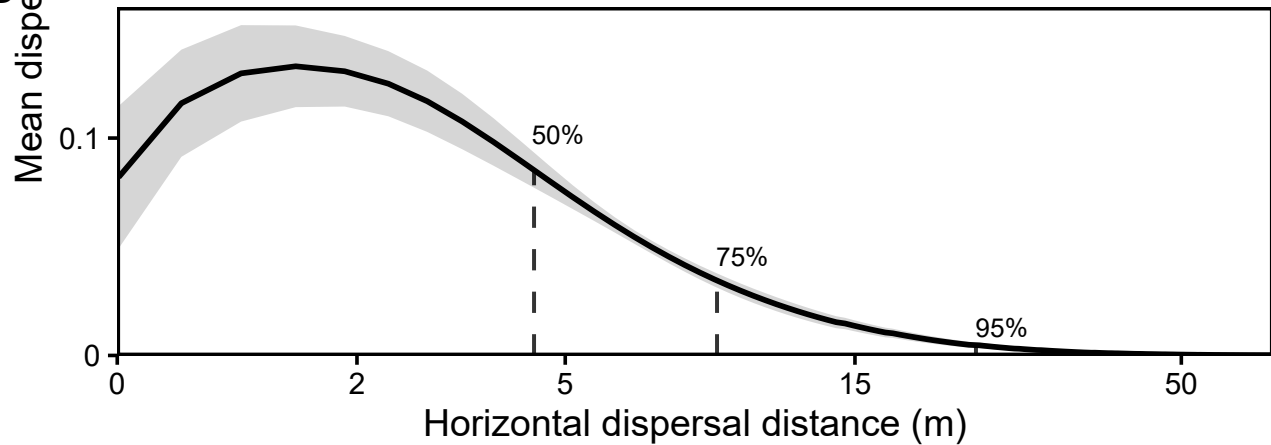
A

Terminal velocity (m s^{-1}) 0.01 0.1 0.3 0.5 1 — day ■ night

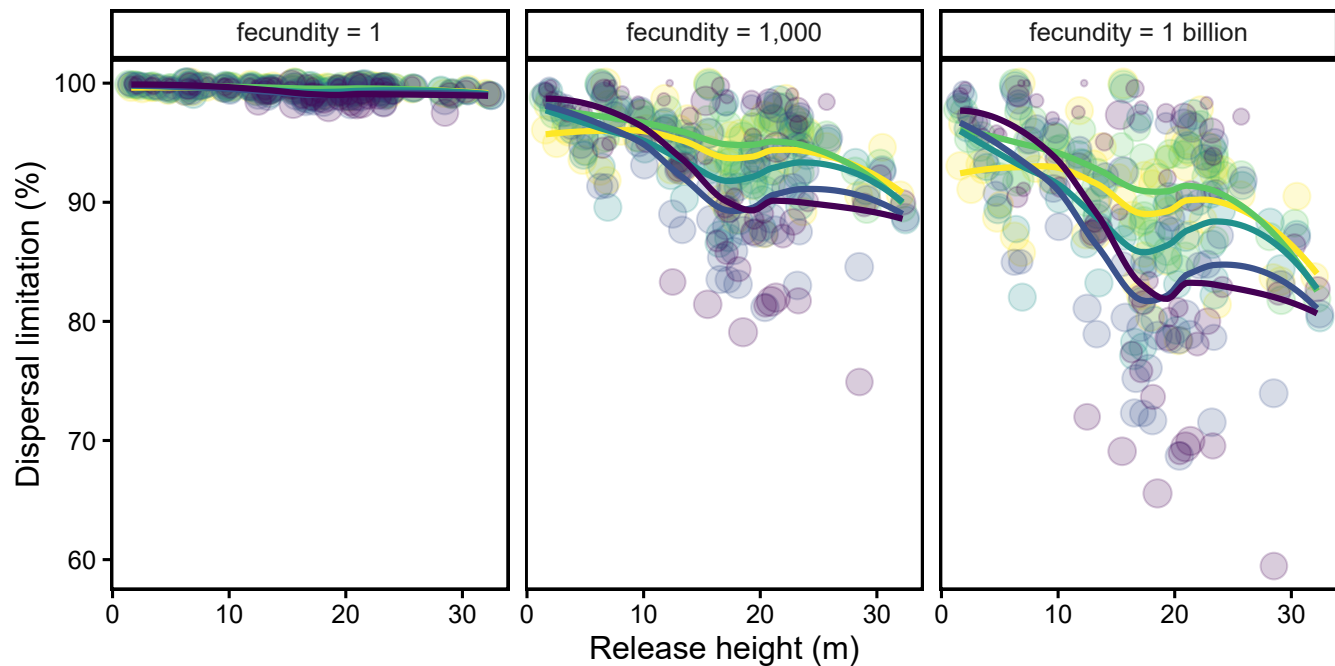


B



A**B**

Terminal velocity (m s^{-1}) 0.01 0.1 0.3 0.5 1
Bark interception probability (%) 0 10 20



Ecological Monographs

Appendix 1: Materials, Methods, Concepts

Patterns of wind dispersal within a forest canopy: An epiphyte perspective

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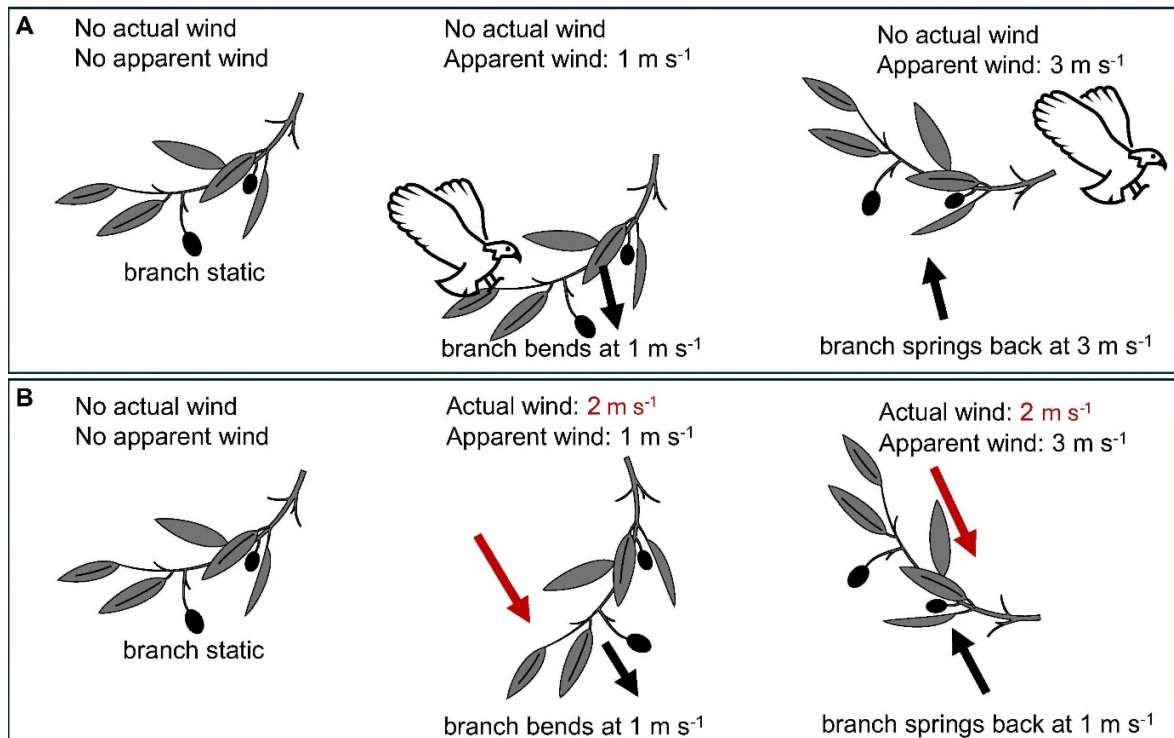


Figure S1. Conceptual diagram of branch movements and their apparent winds that diverge from the actual winds due to **(A)** mechanical disturbance and **(B)** branch sway due to wind loading. Note that other factors such as the direction of movement relative to the direction of the actual wind, inertia, resistance, or drag also play a role in the magnitude and the direction of the apparent wind. The icons used in this figure were sourced from the built-in icon library in Microsoft PowerPoint (Microsoft, 2024) and adapted by MH.

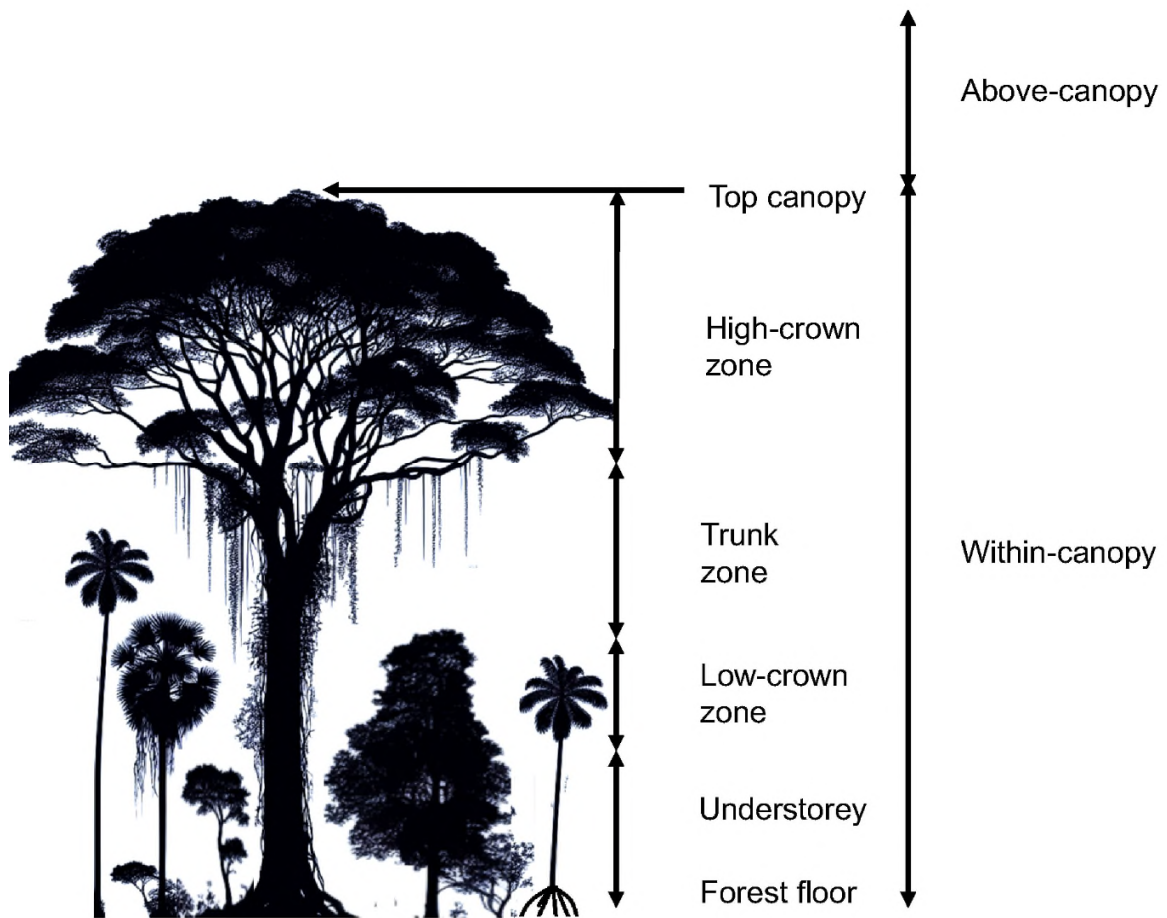


Figure S2. The different forest zones this study refers to and the terminology used. Note that the heights and vegetation densities of the zones may vary within and between forests, and that they are not necessarily representative of the San Lorenzo Forest studied here. The silhouette was adapted by MH from a graphic generated with Microsoft Copilot (Microsoft, 2024).

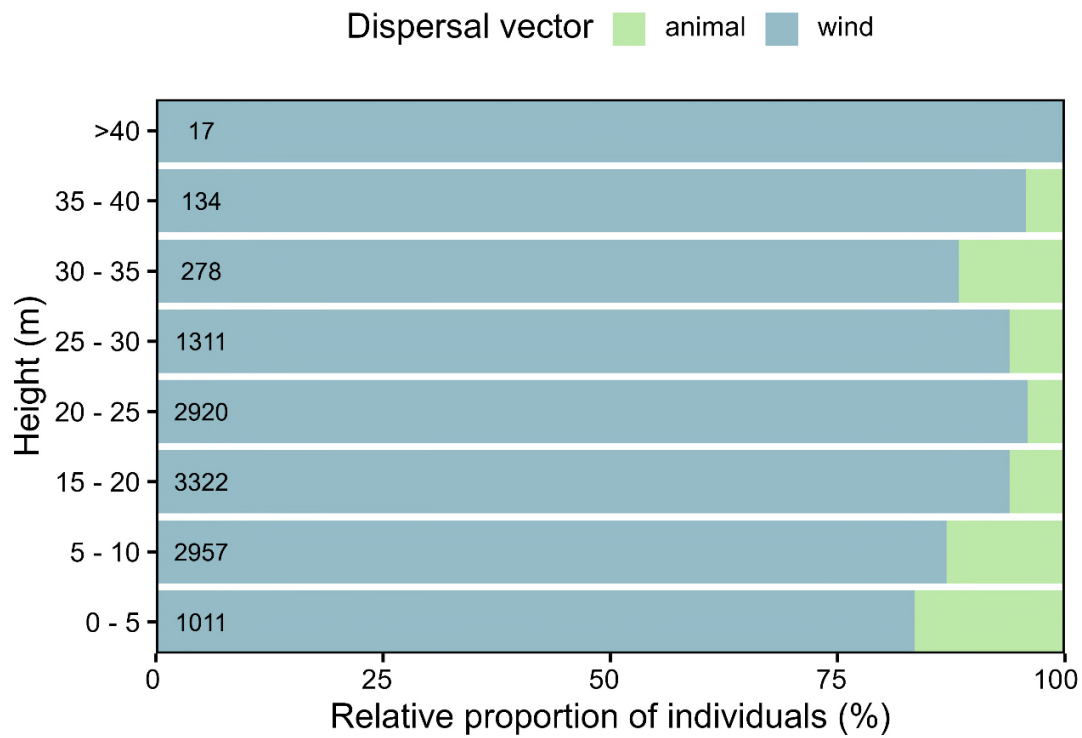


Figure S3. Percent of wind-dispersed vs. animal-dispersed epiphyte individuals at the San Lorenzo 0.4 ha forest plot with absolute numbers of wind-dispersed individuals given inside the columns. The data are based on the census by Zotz & Schultz (2008), encompassing 11,950 epiphyte individuals of 126 species.



Figure S4. Assessing wind dynamics along the forest height gradient. Sonic anemometers (**A**), screwed to an aluminium frame at two positions (in = near the stem, out = 1.35 m away from the stem), were strapped to stems or branches of trees at four heights to measure winds, gusts, and wind direction for 3- to 10-second intervals (0.3 to 0.1 Hz) for one to five consecutive days. Loggers (**B**) logged the measurements. Every anemometer was additionally coupled with a temperature sensor (**C**). The images were taken by MH.

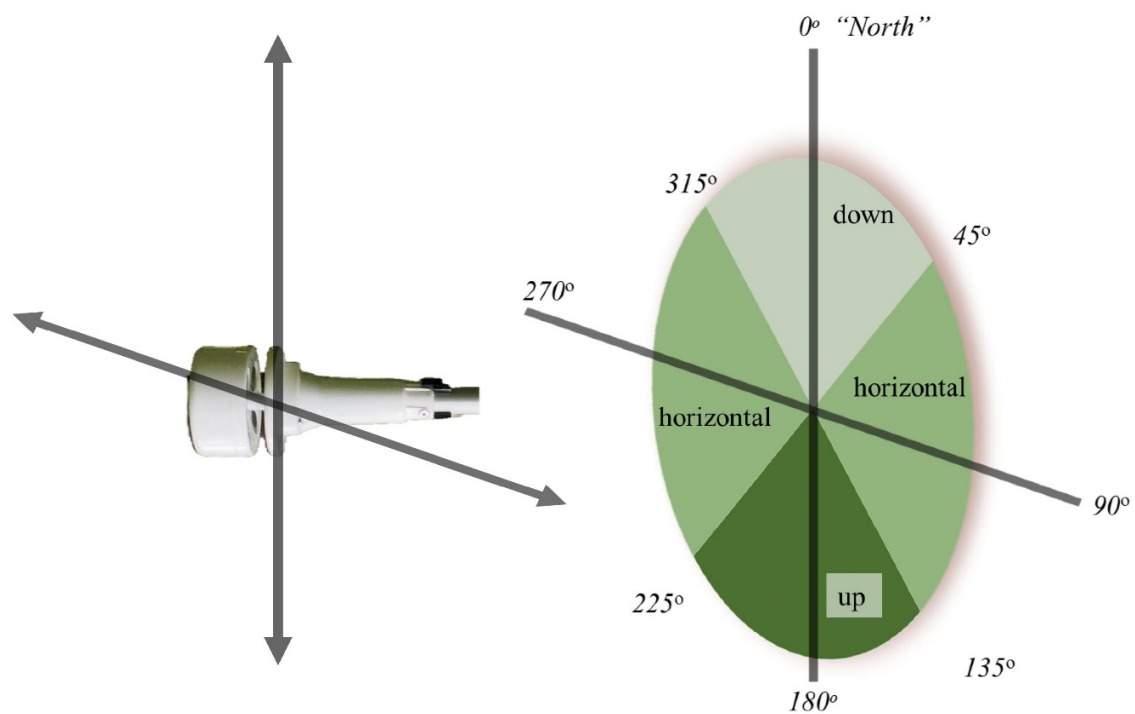


Figure S5. The plane of wind measurements during the study of the wind dynamics at San Lorenzo, Panama, showing measurement directions in cursive and geographic directions in bold. The measurement plane was divided into four equally sized portions, which denote the respective measurement direction categories (up, down, or horizontal [two times]). The geographic direction was derived with the help of the facing direction (see equations in Figure). However, because this horizontal alignment of the anemometers gave only two readings of horizontal direction, i.e., left (90°) and right (270°), horizontal winds/gusts were largely ignored during the analyses. The image of the anemometer was taken and modified by MH.

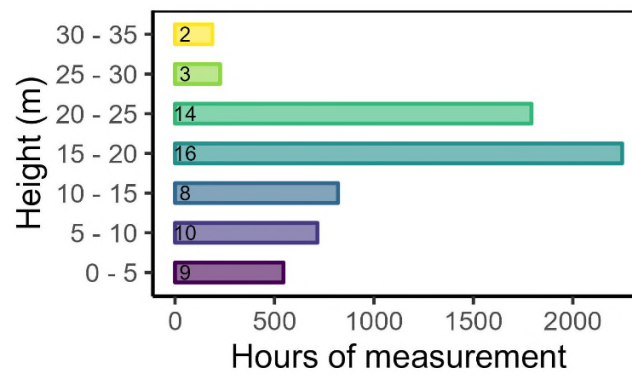


Figure S6. Hours of measurement per height category with numbers of replicates inside the bars. Note that this is for a single anemometer (i.e., the “in” treatment: 0.3 m from the stem); the hours of measurements duplicate when the second position (“out”: 1.3 m from the stem) is added.

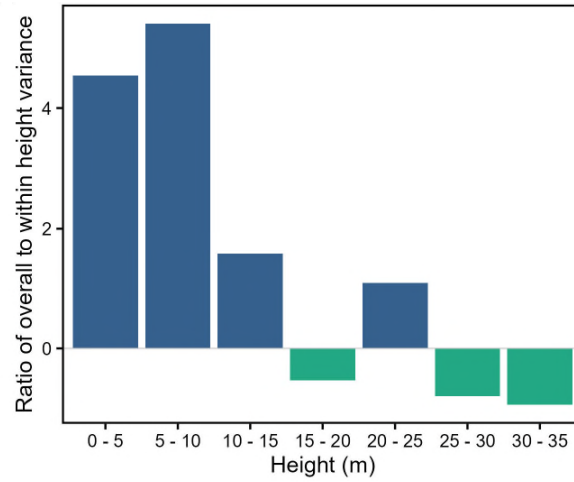


Figure S7. Ratio of inter-height to intra-height variance of gust speeds. A positive value indicates that the variance is greater between height categories than within a height category. Note that the axis is shifted down 1 unit to highlight the differences, i.e., a shown value of 0 corresponds to a ratio of 1, i.e., equal variances.

References

- Microsoft. 2024. "Microsoft 365 tools [software]". <https://microsoft.com/microsoft-365>.
- Zotz, G., and S. Schultz. 2008. "The Vascular Epiphytes of a Lowland Forest in Panama—Species Composition and Spatial Structure." *Plant Ecology* 195: 131–41.
<https://doi.org/10.1007/s11258-007-9310-0>.

Ecological Monographs

Appendix 2: ShinyApp

Patterns of wind dispersal within a forest canopy: An epiphyte perspective

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Figure S1. Screenshot of the interactive shinyApp, available at:

<https://mariehoensbroech.shinyapps.io/epi-wind-dispersal/>. Users can explore the study site as a 3D lidar scene, inspect vertical profiles and anemometer wind sensor positions, and animate above-canopy wind and gust conditions measured at 50 m from the top of the canopy access crane, and discover how below-canopy gust speeds change with above-canopy wind, temperature, or vegetation density. The data and R scripts associated with the ShinyApp will be made available in the study's data repository such that the app can be run locally as well.

Reference List of R Libraries used in the ShinyApp

- Barrett, T., M. Dowle, A. Srinivasan, J. Gorecki, M. Chirico, T. Hocking, B. Schwendinger, and I. Krylov. 2025. “data.Table: Extension of ‘data.Frame’.” R Package Version 1.17.8. <https://CRAN.R-project.org/package=data.table>.
- Bürkner, P.-C. 2017. “brms: An R Package for Bayesian Multilevel Models Using Stan.” *Journal of Statistical Software* 80:1–28. <https://doi.org/10.18637/jss.v080.i01>.
- Chang, W., J. Cheng, J. Allaire, C. Sievert, B. Schloerke, Y. Xie, J. Allen, J. McPherson, A. Dipert, and B. Borges. 2026. “shiny: Web Application Framework for R.” R Package Version 1.11.1. <https://CRAN.R-project.org/package=shiny>.
- Cheng, J., B. Schloerke, B. Karambelkar, Y. Xie, and G. Aden-Buie. 2025. “leaflet: Create Interactive Web Maps with the JavaScript ‘leaflet’ Library.” R Package Version 2.2.3. <https://CRAN.R-project.org/package=leaflet>.
- Hester, J., and J. Bryan. 2026. “glue: Interpreted String Literals.” R Package Version 1.8.0. <https://CRAN.R-project.org/package=glue>.
- Hijmans, R. J. 2025. “terra: Spatial Data Analysis.” R Package Version 1.8-80. <https://CRAN.R-project.org/package=terra>.
- North, A. 2026. “rdeck: Deck.gl Widget.” R Package Version 0.5.2. <https://github.com/qfes/rdeck>.
- Ooms, J. 2014. “The jsonlite Package: A Practical and Consistent Mapping Between JSON Data and R Objects.” arXiv:1403.2805. <https://arxiv.org/abs/1403.2805>.
- Pebesma, E. 2018. “Simple Features for R: Standardized Support for Spatial Vector Data.” *The R Journal* 10: 439–46. <https://doi.org/10.32614/RJ-2018-009>.

- R Core Team. 2023. *R: A Language and Environment for Statistical Computing*. Vienna: R Foundation for Statistical Computing. <https://www.R-project.org/>.
- Roussel, J.-R., D. Auty, N. C. Coops, P. Tompalski, T. R. H. Goodbody, A. Sánchez Meador, J.-F. Bourdon, F. de Boissieu, and A. Achim. 2020. "lidR: An R Package for Analysis of Airborne Laser Scanning (ALS) data." *Remote Sensing of Environment* 251: 112061. <https://doi.org/10.1016/j.rse.2020.112061>.
- Vaidyanathan, R., Y. Xie, J. J. Allaire, J. Cheng, C. Sievert, and K. Russell. 2023. "htmlwidgets: HTML Widgets for R." R Package Version 1.6.4. <https://CRAN.R-project.org/package=htmlwidgets>.
- Wickham, H., M. Averick, J. Bryan, W. Chang, L. D'Agostino McGowan, R. François, G. Grolemund, et al. 2019. "Welcome to the tidyverse." *Journal of Open Source Software* 4(43): 1686. <https://doi.org/10.21105/joss.01686>.

References List of External Data Sources

- Janner, L. E., and M. Fernández-Lucero, unpublished dataset.
- Mendieta-Leiva, G., K. Wagner, and G. Zotz. 2021. "Vascular Epiphyte Census Data at the San Lorenzo Crane Plot [dataset]." *Zenodo*. <https://doi.org/10.5281/zenodo.5645775>.
- Paton, S. 2019a. "Galeta–Wind Direction [dataset]." *Smithsonian Tropical Research Institute*. <https://doi.org/10.25573/data.10042640.v25>.
- Paton, S. 2019b. "Galeta–Wind Speed [dataset]." *Smithsonian Tropical Research Institute*. <https://doi.org/10.25573/data.10042658.v23>.
- Paton, S. 2019c. "San Lorenzo Crane–Wind Speed [dataset]." *Smithsonian Tropical Research Institute*. <https://doi.org/10.25573/data.10042688.v23>.

- Paton, S. 2019d. "San Lorenzo Crane—Wind direction [dataset]." *Smithsonian Tropical Research Institute*. <https://doi.org/10.25573/data.10042718.v23>.
- Paton, S. 2019e. "San Lorenzo Crane—Air Temperature [dataset]." Smithsonian Tropical Research Institute. <https://doi.org/10.25573/data.10042694.v23>.
- Paton, S. 2019f. "San Lorenzo Crane—Precipitation [dataset]." Smithsonian Tropical Research Institute. <https://doi.org/10.25573/data.10042679.v23>.
- Ramos, P., P. Villareal, and H. Muller-Landau. 2024. "Tree Height Allometry Data for 5109 Trees in Twenty-Four Small ForestGEO Plots in Panama, 2011–2020 [dataset]." Smithsonian Tropical Research Institute. <https://doi.org/10.25573/data.24954204.v1>.
- Vasquez V., M. Garcia, M. Hernandez, ForestGEO Smithsonian, and H. Muller-Landau. 2024. "Smithsonian ForestGEO San Lorenzo and Panama Small Plots Aerial Photogrammetry Orthomosaics, Digital Surface Models, Point Clouds and Raw Images for 2015–2024." Smithsonian Research Data Repository. <https://doi.org/10.60635/C33W2K>.
- Zotz, G., and S. Schultz. 2008. "The Vascular Epiphytes of a Lowland Forest in Panama—Species Composition and Spatial Structure." *Plant Ecology* 195: 131–41. <https://doi.org/10.1007/s11258-007-9310-0>.

Ecological Monographs

Appendix 3: Flowchart of the data wrangling and dispersal model for:

- **Patterns of wind dispersal within a forest canopy: An epiphyte perspective**

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Joining below-canopy wind data with the tree data

speeds are in m/s,

temperature in °C,

heights in m

Below-canopy

wind data

Replicate #

Position

Sensor ID

Datetime	Wind speed 11.3 in 1234567	Gust speed 11.3 in 1234567	Wind direction 11.3 in 1234567
2023-04-10 15:43:32	0	0.33	340
2023-04-10 15:43:35	0	0	10



Datetime	Replicate #	Position	Sensor ID	Wind speed	Gust speed	Wind direction
2023-04-10 15:43:32	11.3	in	1234567	0	0.33	340
2023-04-10 15:43:35	11.3	in	1234567	0	0	10

Replicate #	Measurement	Sensor ID	Position	Tree species	Sensor height	Sensor direction	Tree height	Canopy density (four readings)
11.3	Wind	1234567	in	<i>Brosimum</i>	17.4	60	36.1	24, 40, 5, 37
11.3	Temp	23456789	out	<i>Brosimum</i>	17.4	60	36.1	24, 40, 5, 37

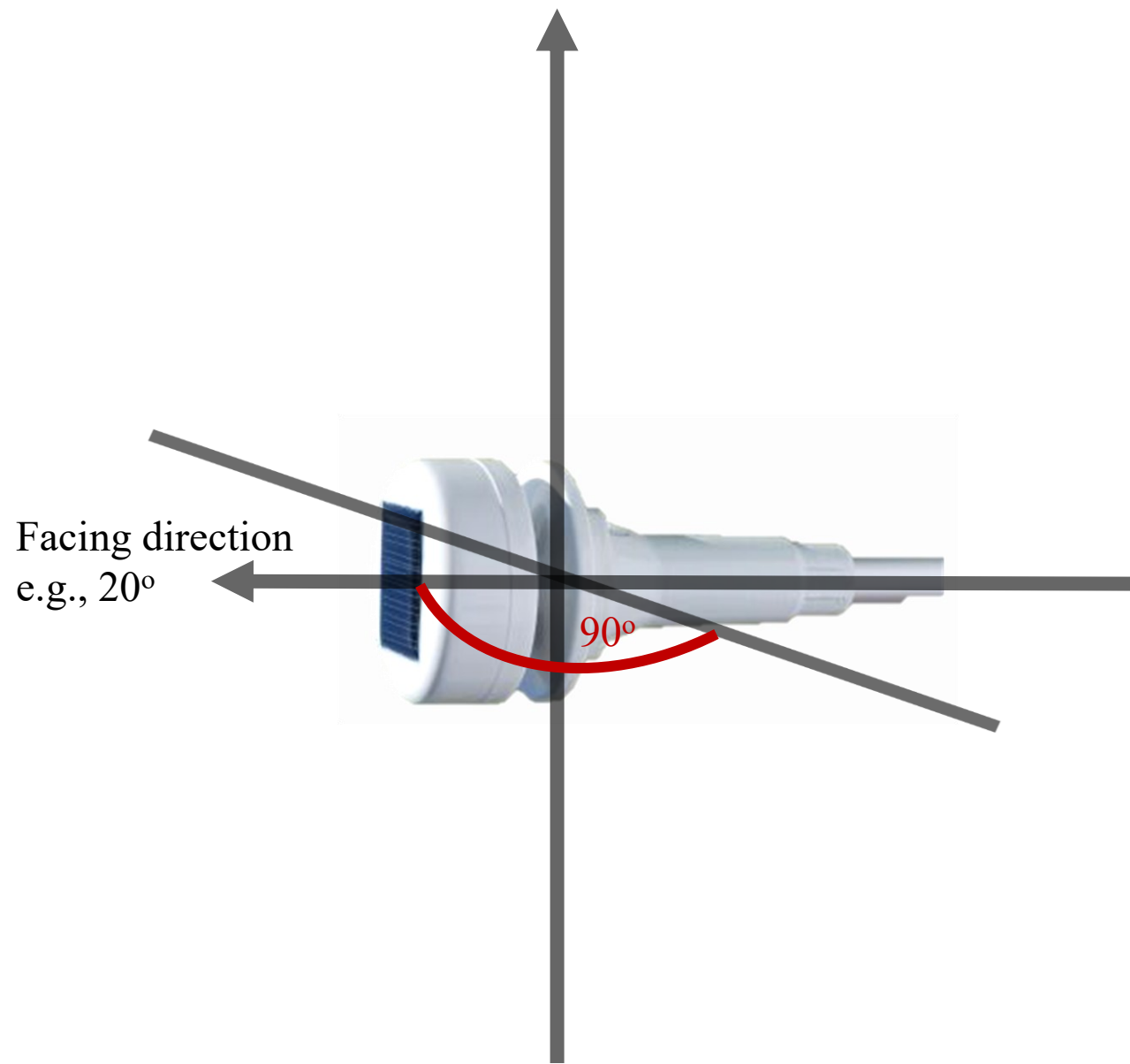
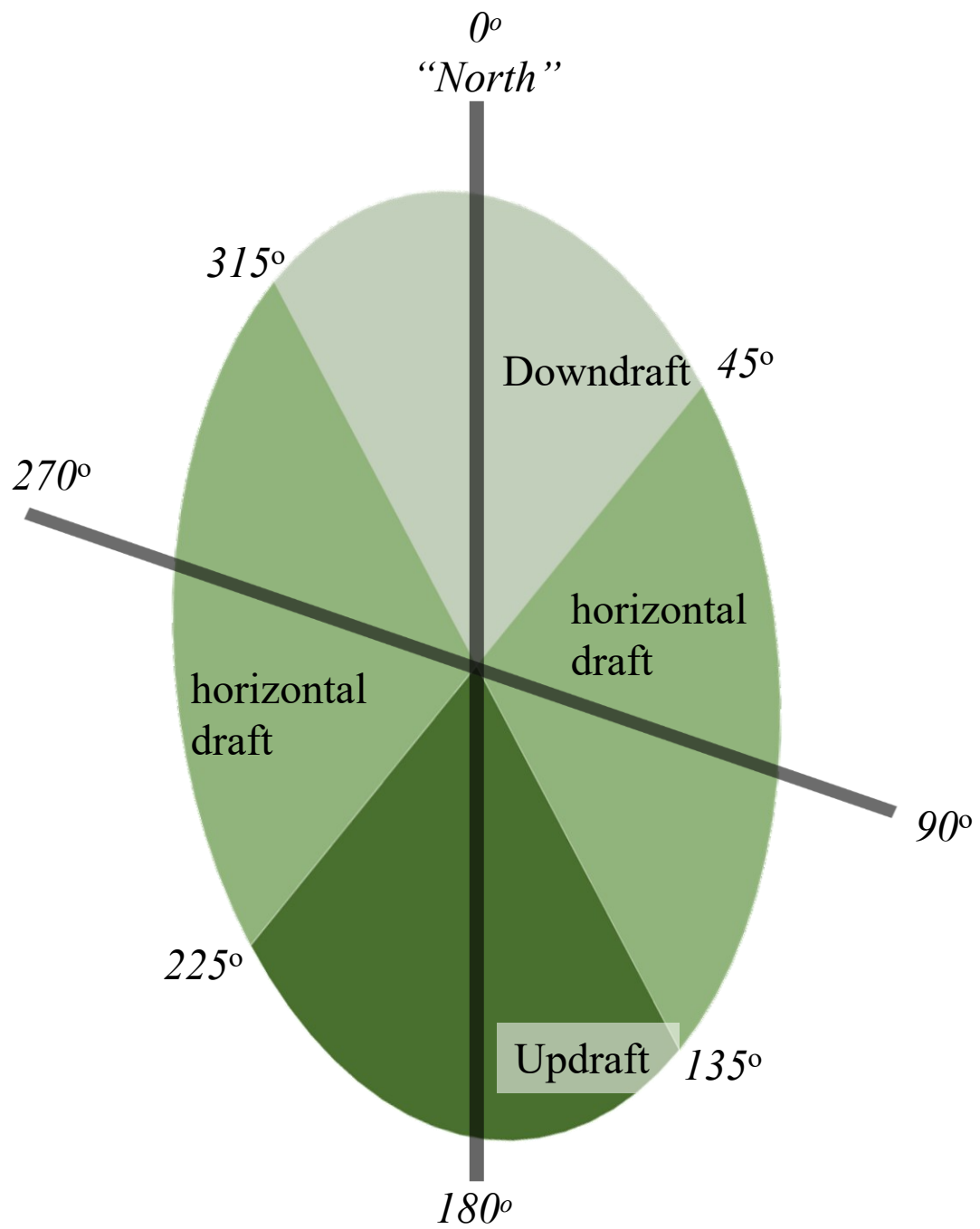


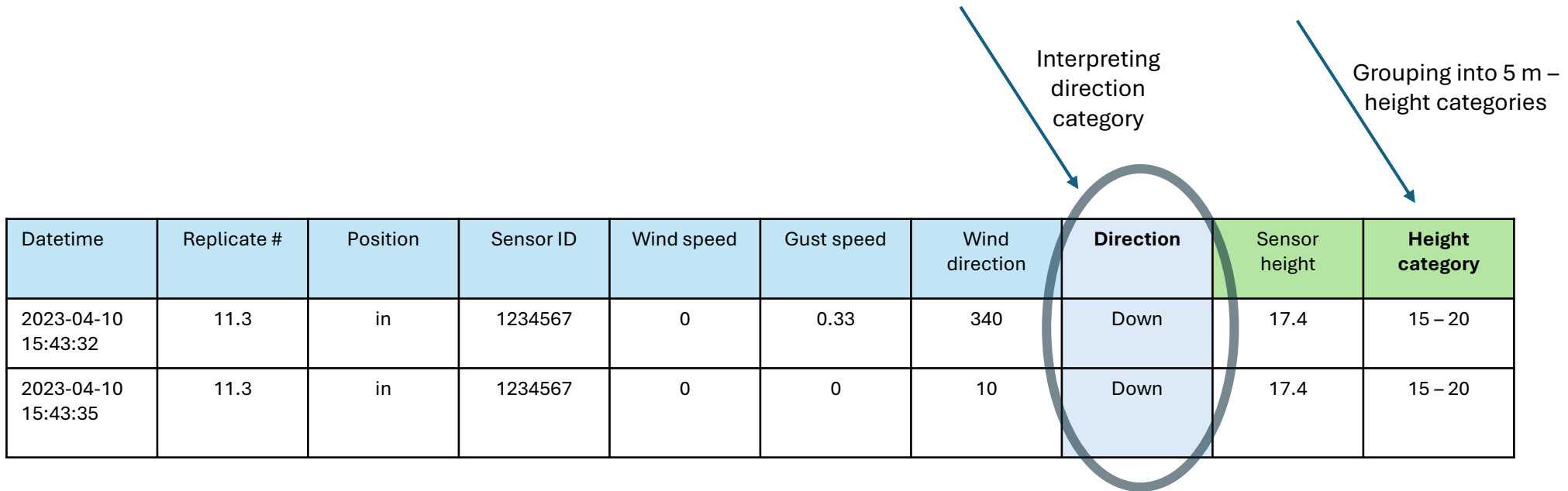
Datetime	Replicate #	Position	Sensor ID	Wind speed	Gust speed	Wind direction
2023-04-10 15:43:32	11.3	in	1234567	0	0.33	340
2023-04-10 15:43:35	11.3	in	1234567	0	0	10



Datetime	Replicate #	Position	Sensor ID	Wind speed	Gust speed	Wind direction	Tree species	Sensor height	Sensor direction	Tree height	Canopy density
2023-04-10 15:43:32	11.3	in	1234567	0	0.33	340	<i>Brosiimum</i>	17.4	60	36.1	24, 40, 5, 37
2023-04-10 15:43:35	11.3	in	1234567	0	0	10	<i>Brosimum</i>	17.4	60	26.1	24, 40, 5, 37

Interpreting direction





Datetime	Replicate #	Position	Sensor ID	Wind speed	Gust speed	Wind direction	Direction	Sensor height	Height category
2023-04-10 15:43:32	11.3	in	1234567	0	0.33	340	Down	17.4	15 – 20
2023-04-10 15:43:35	11.3	in	1234567	0	0	10	Down	17.4	15 – 20

Datetime	Replicate #	Position	Sensor ID	Wind speed	Gust speed	Wind direction	Direction	Sensor height	
2023-04-10 15:43:32	11.3	in	1234567	0	0.33	340	Down	17.4	
2023-04-10 15:43:35	11.3	in	1234567	0	0	10	Down	17.4	

lubridate::ceiling_date() -> 1 minute

mean

max

circular mean

per group: replicate # and position

Datetime	Replicate #	Position	Sensor ID	Wind speed	Gust speed	Wind direction	Direction	Sensor height
2023-04-10 15:44:00	11.3	in	1234567	0	0.33	350	Down	17.4
2023-04-10 15:45:00	11.3	in	1234567	0.12	0.66	345	Down	17.4
2023-04-10 15:46:00	11.3	in	1234567	0.22	1	270	Horizontal	Sensor height
2023-04-10 15:47:00	11.3	In	1234567	0.1	0.33	170	Up	17.4

3 second data



1 minute data

Datetime	Replicate #	Position	Sensor ID	Wind speed	Gust speed	Wind direction	Direction	Sensor height
2023-04-10 15:44:00	11.3	in	1234567	0	0.33	350	Down	17.4
2023-04-10 15:45:00	11.3	in	1234567	0.12	0.66	345	Down	17.4
2023-04-10 15:46:00	11.3	in	1234567	0.22	1	270	Horizontal	17.4
2023-04-10 15:47:00	11.3	In	1234567	0.1	0.33	170	Up	17.4

lubridate::ceiling_date() -> 5 minutes

mean

max

circular mean

per group: replicate # and position **and direction**

Datetime	Replicate #	Position	Sensor ID	Wind speed	Gust speed	Wind direction	Direction	Sensor height
2023-04-10 15:45:00	11.3	in	1234567	0.06	0.66	347.5	Down	17.4
2023-04-10 15:50:00	11.3	in	1234567	0.22	1	270	Horizontal	17.4
2023-04-10 15:50:00	11.3	In	1234567	0.1	0.33	170	Up	17.4

1 minute data



5 minute data

Joining below-canopy data with above-canopy data

Datetime	Mean wind	Max wind (gust)	Wind direction	Mean temperature	Rain
2023-04-10 15:45:00	5.4	7.7	0	29.5	10
2023-04-10 16:00:00	5.7	7.5	10	29.4	0

dplyr::na_spline()

Datetime	Mean wind	Gust	Wind direction	Mean temperature	Rain
2023-04-10 15:45:00	5.4	7.7	0	29.5	10
2023-04-10 15:50:00	5.5	7.64		29.46	
2023-04-10 15:55:00	5.6	7.57		29.42	
2023-04-10 16:00:00	5.7	7.5	10	29.4	0

Copy down

Divide by 3

Datetime	Mean wind	Gust	Wind direction	Mean temperature	Rain
2023-04-10 15:45:00	5.4	7.7	0	29.5	3.33
2023-04-10 15:50:00	5.5	7.64	0	29.46	3.33
2023-04-10 15:55:00	5.6	7.57	0	29.42	3.33
2023-04-10 16:00:00	5.7	7.5	10	29.4	0

Datetime	Replicate #	Position	Sensor ID	Wind speed	Gust speed	Wind direction	Direction	Sensor height	Mean wind	Gust	Wind direction	Mean temperature	Rain
2023-04-10 15:45:00	11.3	in	1234567	0.06	0.66	347.5	Down	17.4	5.4	7.7	0	29.5	3.33
2023-04-10 15:50:00	11.3	in	1234567	0.22	1	270	Horizontal	17.4	5.5	7.64	0	29.46	3.33
2023-04-10 15:50:00	11.3	In	1234567	0.1	0.33	170	Up	17.4	5.5	7.64	0	29.46	3.33

Calculating up-gust proportion

Datetime	Replicate #	Position	Sensor ID	Wind speed	Gust speed	Wind direction	Direction	Sensor height	
2023-04-10 15:43:32	11.3	in	1234567	0	0.33	340	Down	17.4	
2023-04-10 15:43:35	11.3	in	1234567	0	0	10	Down	17.4	

3 second data

Filter gusts above 0.33 m/s (the anemometers’ threshold)

Datetime	Replicate #	Position	Sensor ID	Wind speed	Gust speed	Wind direction	Direction	Sensor height
2023-04-10 15:43:32	11.3	in	1234567	0	0.33	340	Down	17.4
2023-04-10 15:43:43	11.3	in	1234567	0	0	10	Up	17.4

Calculate total number of gusts per hour, the number of gusts in each direction category, and the relative proportion

Datehour	Datetime	Replicate #	Position	Sensor ID	Wind speed	Gust speed	Wind direction	Direction	Sensor height	Total number of gusts	Number of gusts per group	Proportion
2023-04-10 16:00:00	2023-04-10 15:43:32	11.3	in	1234567	0	0.33	340	Down	17.4	2500	100	0.04
2023-04-10 16:00:00	2023-04-10 15:43:43	11.3	in	1234567	0	0	10	Up	17.4	2500	323	0.13

EpiDisp 1.0 – the trajectory model

Mean wind speed for every minute per height.

Vertical winds and gusts received a negative sign when downward flow, positive when upward flow.

Negative when cardinal wind direction between 90.1° and 270°

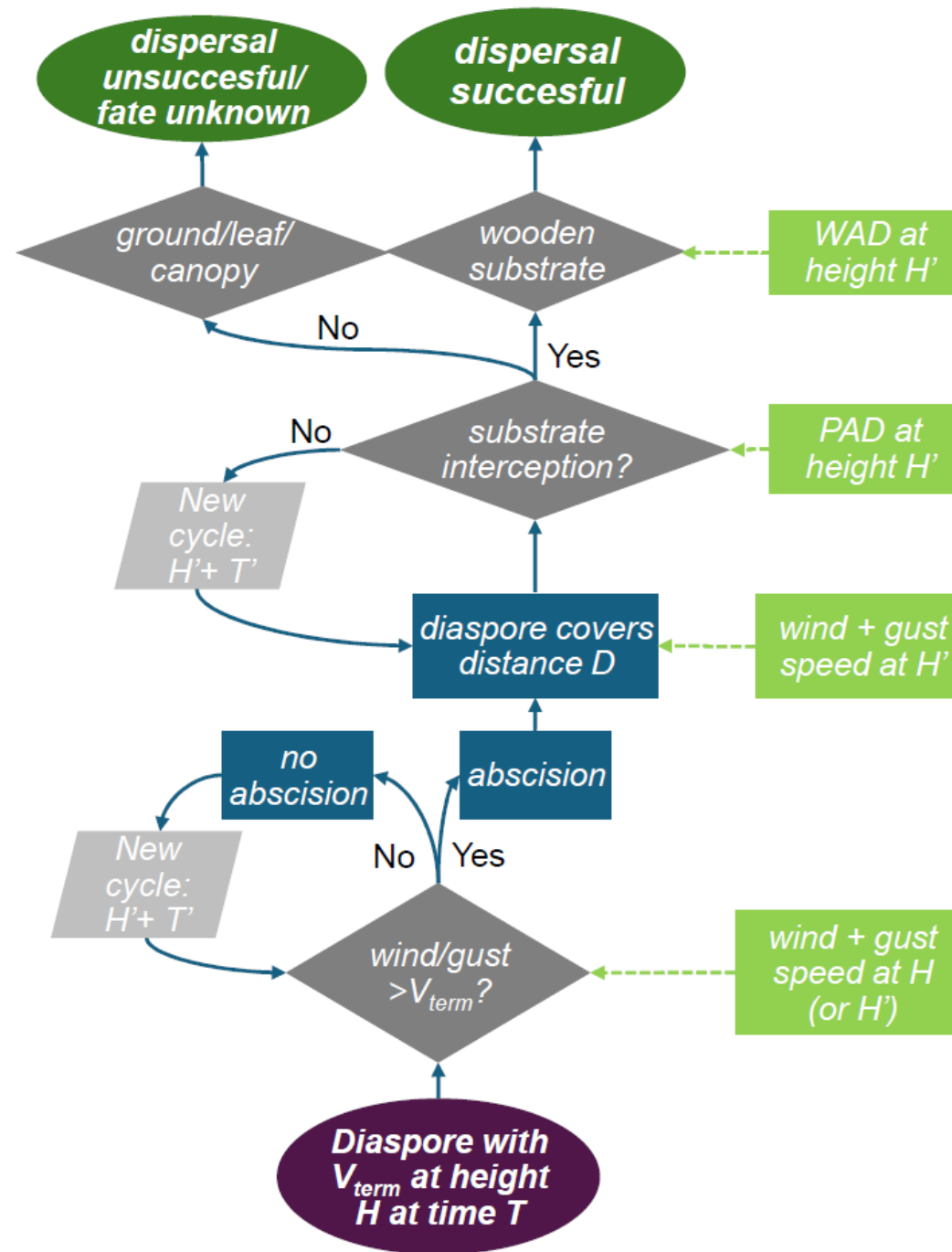
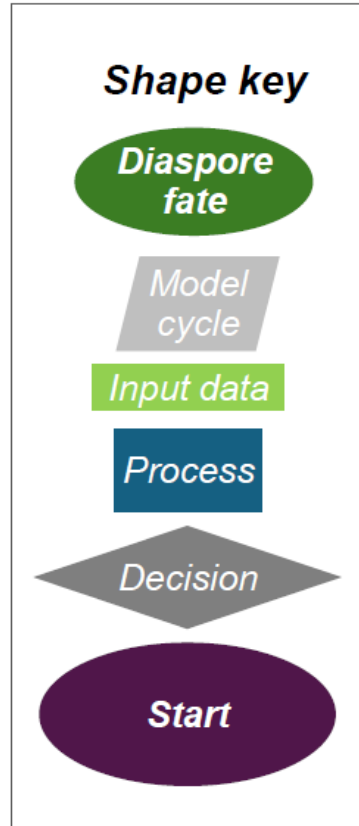
Datetime	Daytime	Replicate #	Position	Vertical wind speed	Vertical gust speed	Horizontal wind speed	Horizontal gust speed	Sensor height	Tree height
2023-04-10 15:43:32	Late day	11.3	In	0.1	0.33	1.6	-2.3	10.2	36.1
2023-04-10 15:43:32	Late day	11.3	In	0	0	0	0	13.2	36.1
2023-04-10 15:43:32	Late day	11.3	In	0.1	-0.67	1.6	0	17.4	36.1
2023-04-10 15:43:32	Late day	11.3	In	0.1	0	1.6	1.2	21.1	36.1
2023-04-10 15:43:33	Late day	11.3	In	0.1	0	1.6	1.5	13.2	36.1

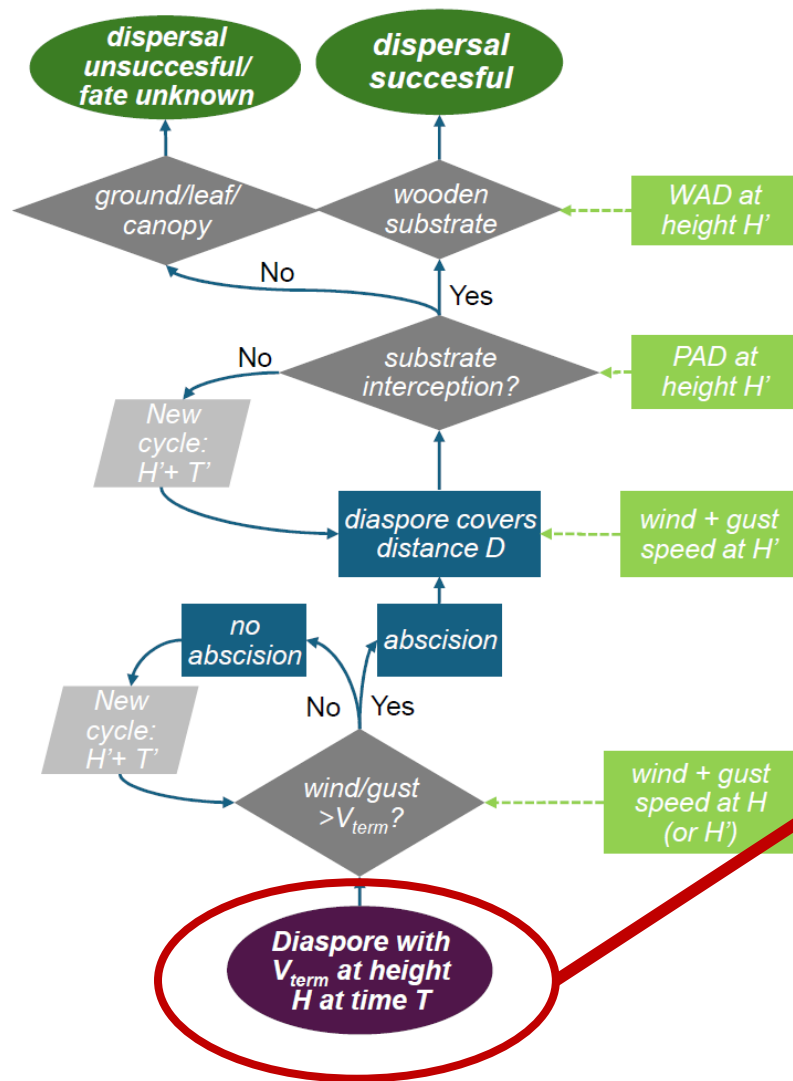
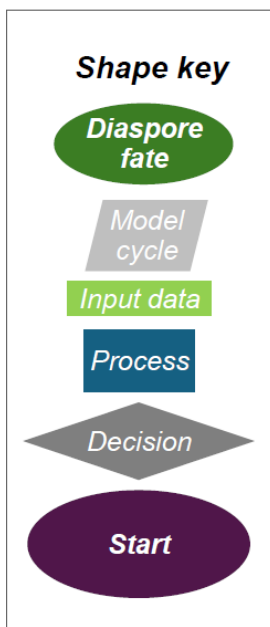
When at a certain time, at a given height, no measurements were taken, they received 0 as value

Dataframes were split by replicate #, position, and daytime

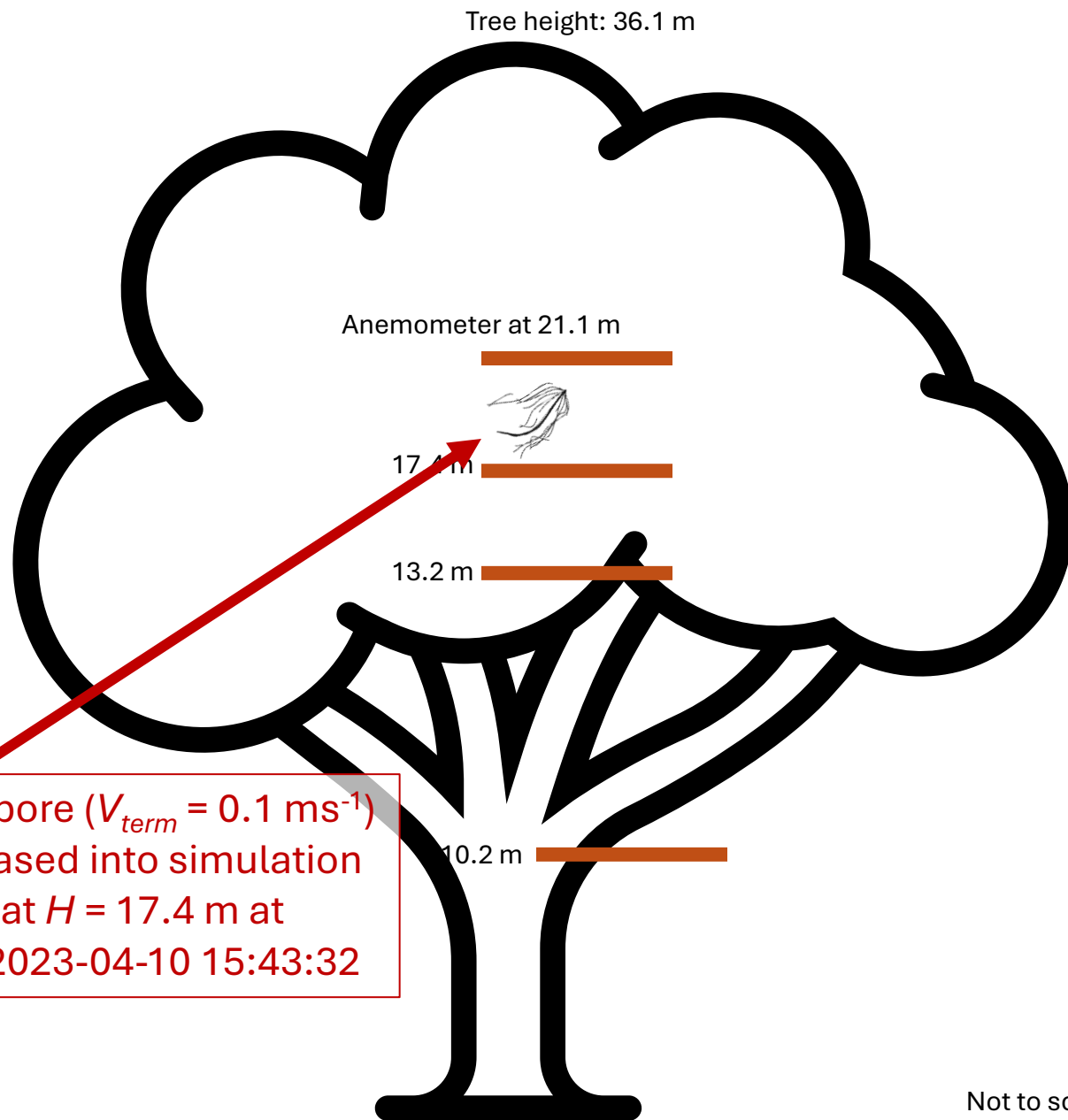
A ca. 30-min period (~1000 rows) were randomly chosen

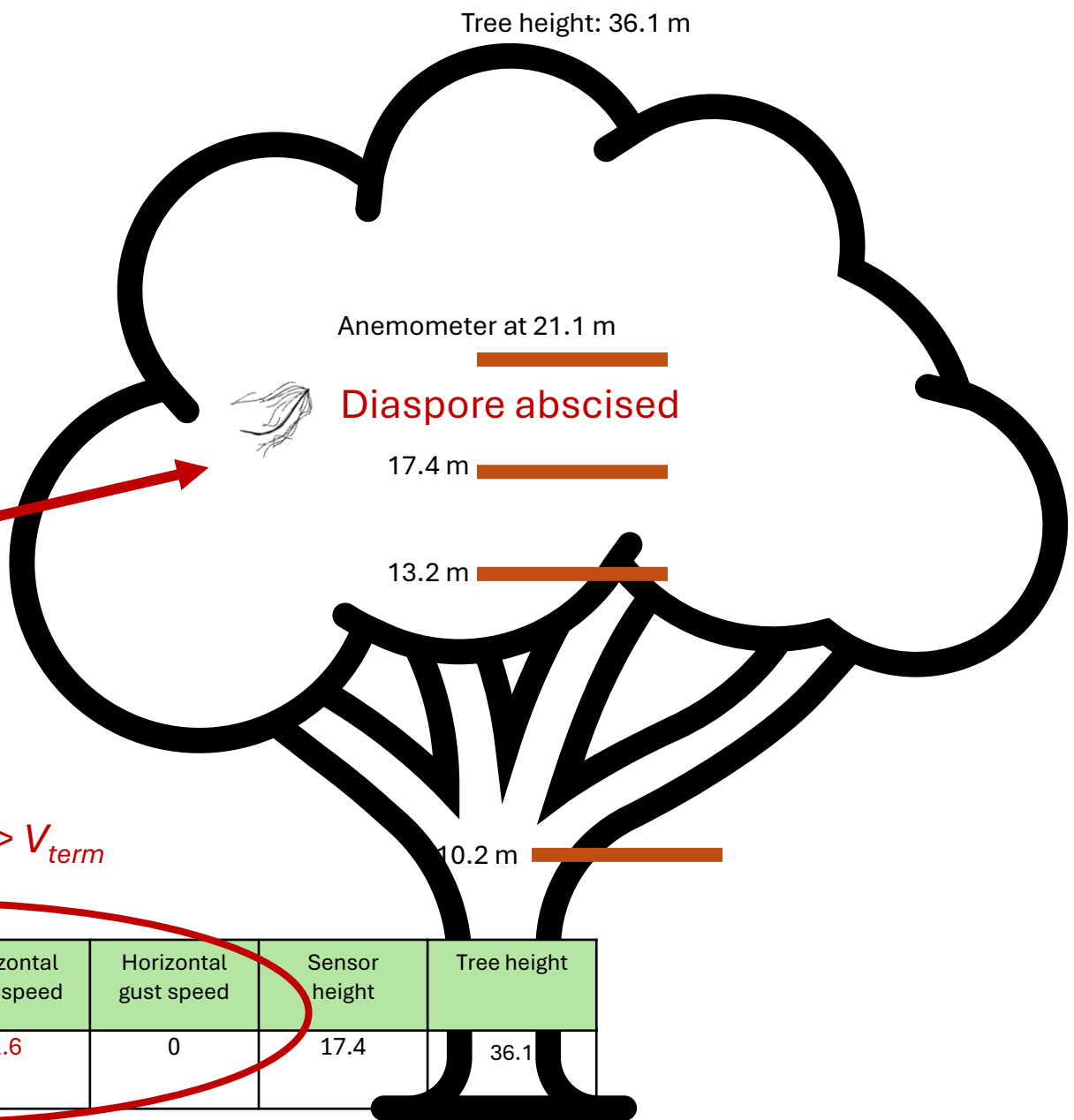
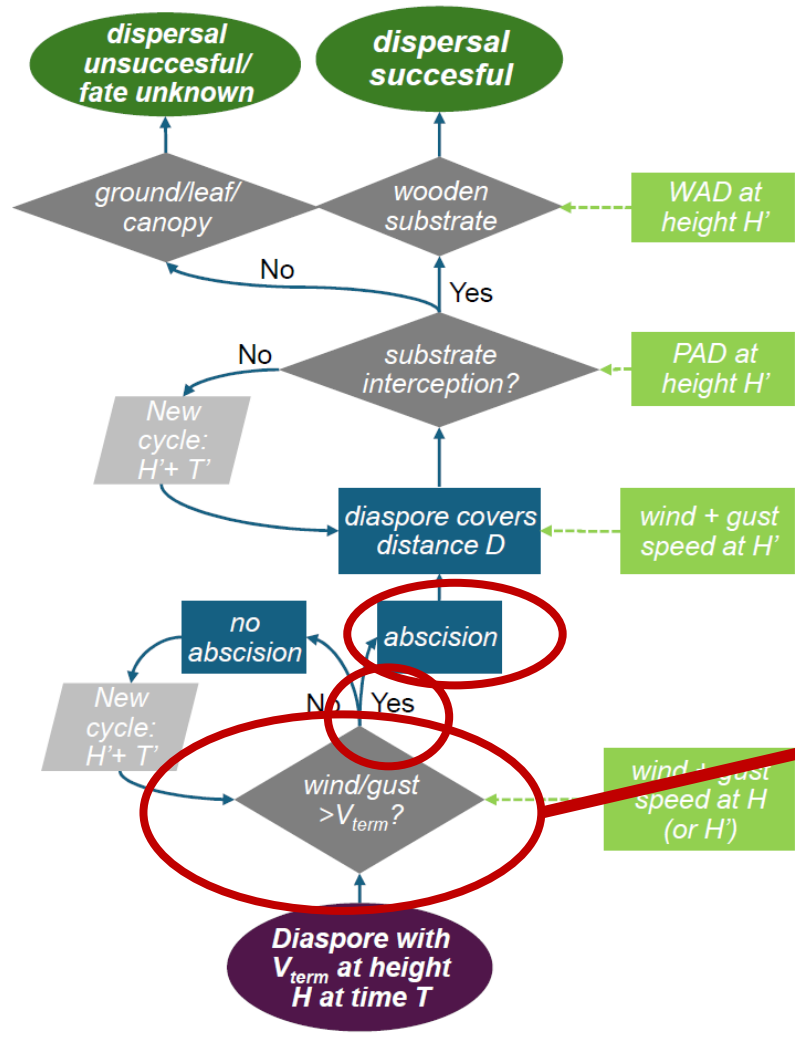
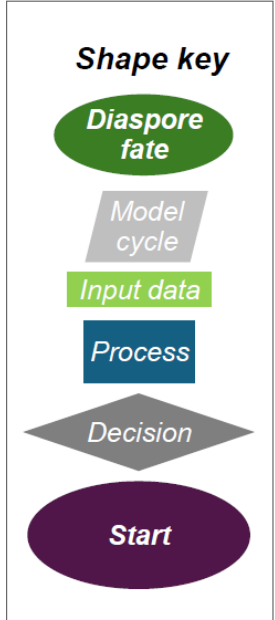
Datetime	Daytime	Replicate #	Position	Vertical wind speed	Vertical gust speed	Horizontal wind speed	Horizontal gust speed	Sensor height	Tree height
2023-04-10 21:43:32	Early night	11.3	Out	0.1	0.33	1.6	-2.3	10.2	36.1
2023-04-10 21:43:32	Early night	11.3	Out	0	0	0	0	13.2	36.1
2023-04-10 21:43:32	Early night	11.3	Out	0.1	-0.67	1.6	0	17.4	36.1
2023-04-10 21:43:32	Early night	11.3	Out	0.1	0	1.6	1.2	21.1	36.1
2023-04-10 21:43:33	Early night	11.3	Out	0.1	0	1.6	1.5	13.2	36.1





Diaspore ($V_{term} = 0.1 \text{ ms}^{-1}$)
released into simulation
at $H = 17.4 \text{ m}$ at
 $T = 2023-04-10 \ 15:43:32$





$|gust| \text{ and } |wind| > V_{term}$

Datetime	Daytime	Replicate #	Position	Vertical wind speed	Vertical gust speed	Horizontal wind speed	Horizontal gust speed	Sensor height	Tree height
2023-04-10 15:43:32	Late day	11.3	In	0.1	-0.67	1.6	0	17.4	36.1

Additional columns are added now

When a seed was released, at what height, and at what position

True, if a vertical or horizontal gust speed was greater than Vterm
terminal velocity (m/s)

Only filled when abscision = T and Time of abscision = NA

Datetime	Seed ID	Time interval (seconds)	Vertical wind speed	Vert. gust speed	Horizontal wind speed	Hor. gust speed	Sensor height = Original release height	Tree height	Vterm	Abscision	Time of abs.
2023-04-10 21:43:32	2023-04-10 21:43:32 __ 10.2 m __ Out	3	0.1	0.33	1.6	-2.3	10.2	36.1	0.1	T	2023-04-10 21:43:32
2023-04-10 21:43:32	2023-04-10 21:43:32 __ 10.2 m __ Out	3	0	0	0	0	13.2	36.1	0.1	F	NA
2023-04-10 21:43:32	2023-04-10 21:43:32 __ 10.2 m __ Out	3	0.1	-0.67	1.6	0	17.4	36.1	0.1	T	2023-04-10 21:43:32
2023-04-10 21:43:32	2023-04-10 21:43:32 __ 10.2 m __ Out	3	0.1	0	1.6	1.2	21.1	36.1	0.1	T	2023-04-10 21:43:32

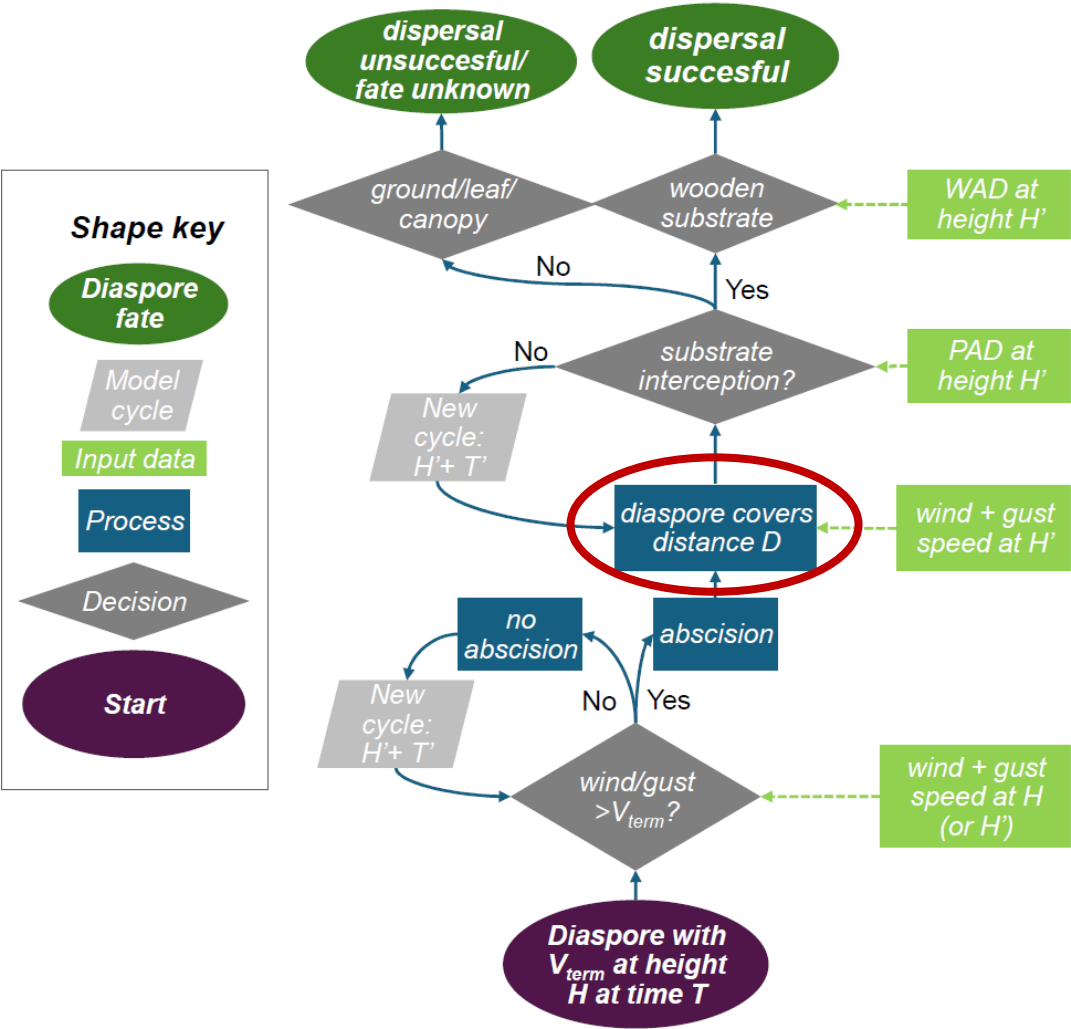
Calculating vertical and horizontal distances:

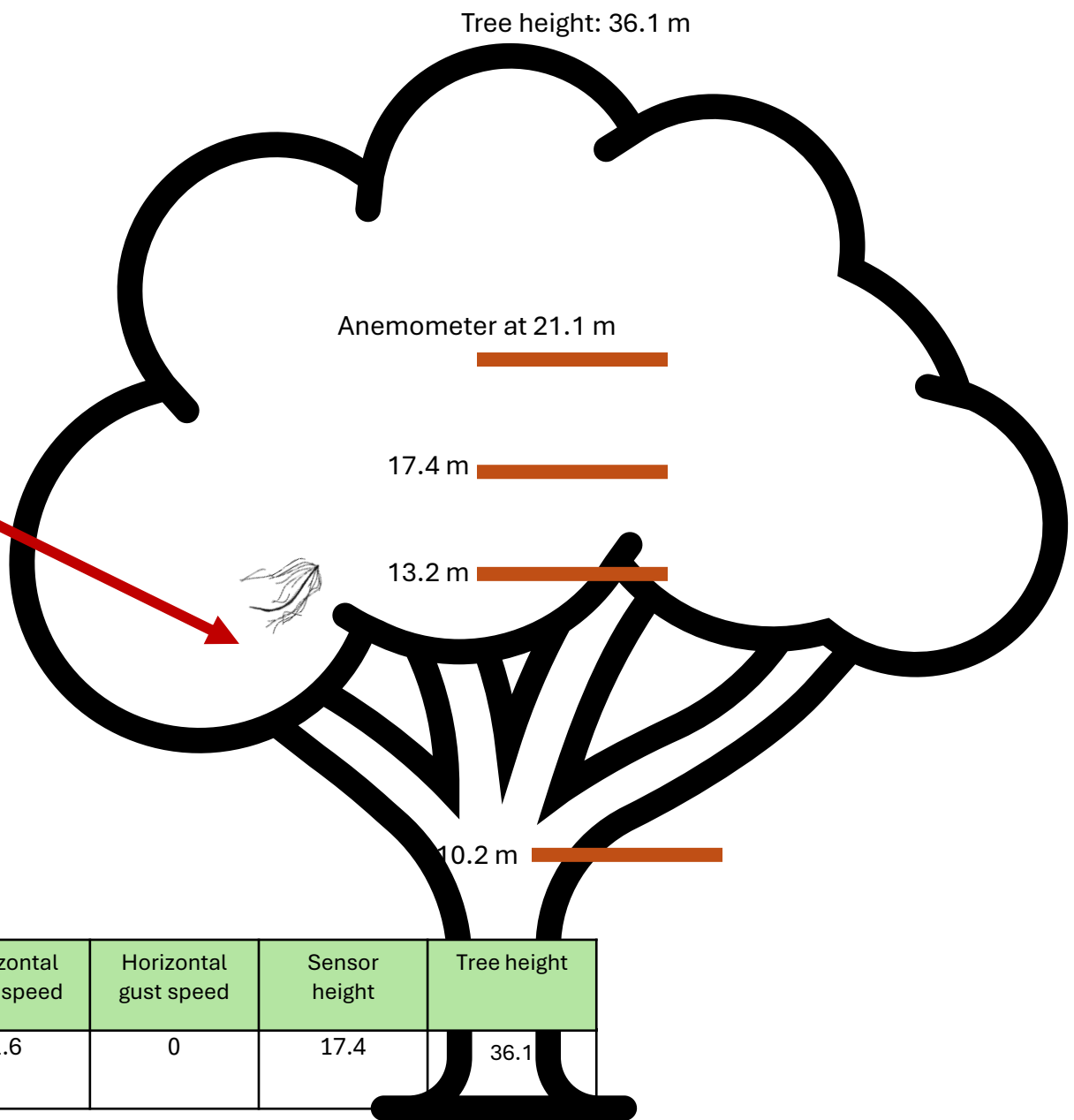
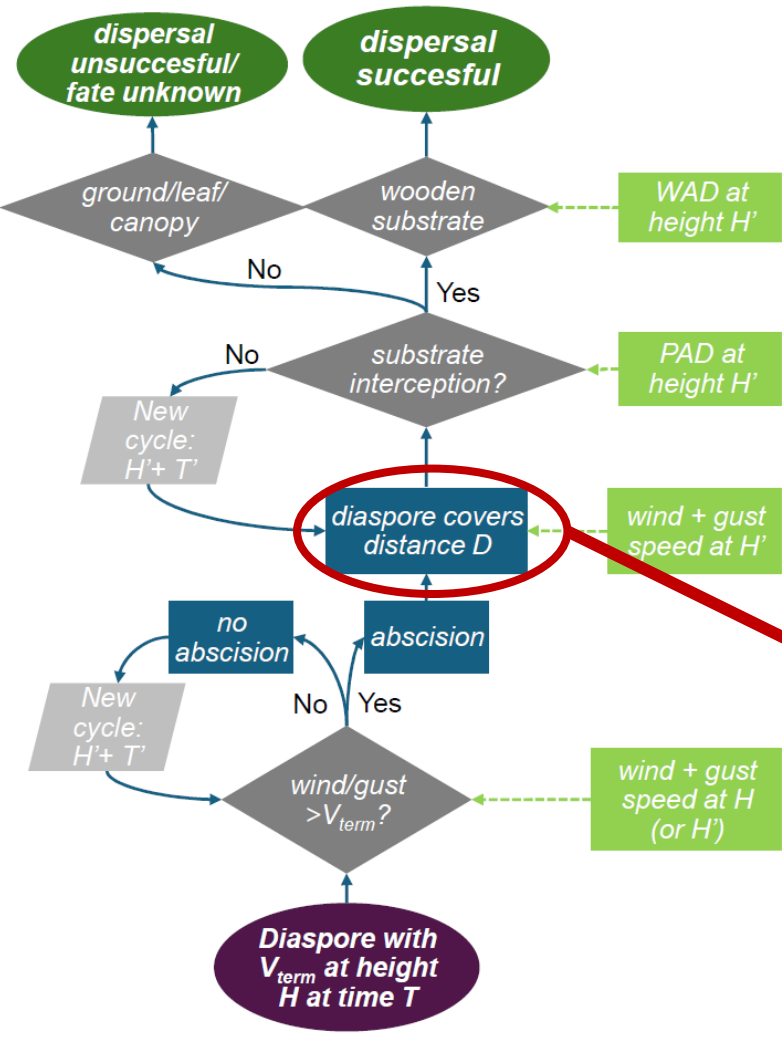
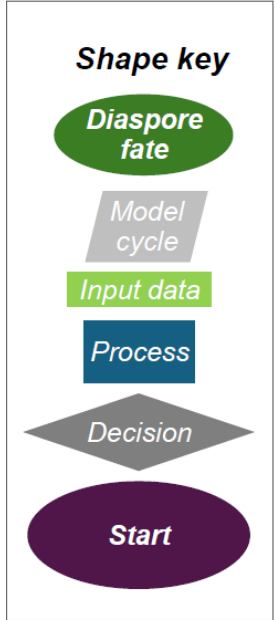
$$D_w = \sum \begin{cases} \frac{w' + 2\bar{w}}{3} - V_{term} , & \text{if } \frac{w' + 2\bar{w}}{3} > V_{term} \\ \frac{w' + 2\bar{w}}{3} , & \text{if } \frac{w' + 2\bar{w}}{3} < V_{term} \end{cases}$$

and $D_u = \sum \left(\frac{1}{3} u' + \frac{2}{3} \bar{u} \right) t$

(Equations 1a and 1b),

where D_w and D_u are the sum of the vertical and horizontal dispersal distances, w' and u' the vertical and horizontal gust speeds, $\bar{w}$ and $\bar{u}$ the 1-minute vertical and horizontal wind averages, t is the time interval and V_{term} the terminal velocity of a seed. The vertical distance was corrected for the apparent speed, meaning that the falling velocity did not affect the downward motion if winds were equal to or more negative than the falling velocity. For example, a seed with a terminal velocity of 0.1 m/s in a down-gust of 0.2 m/s does not fall at 0.3 m/s but at 0.2 m/s; conversely, a seed with a falling velocity of 0.5 m/s falls at 0.5 m/s. Wind vectors of the same direction cancel out; hence, the more negative value determines the downward distance.





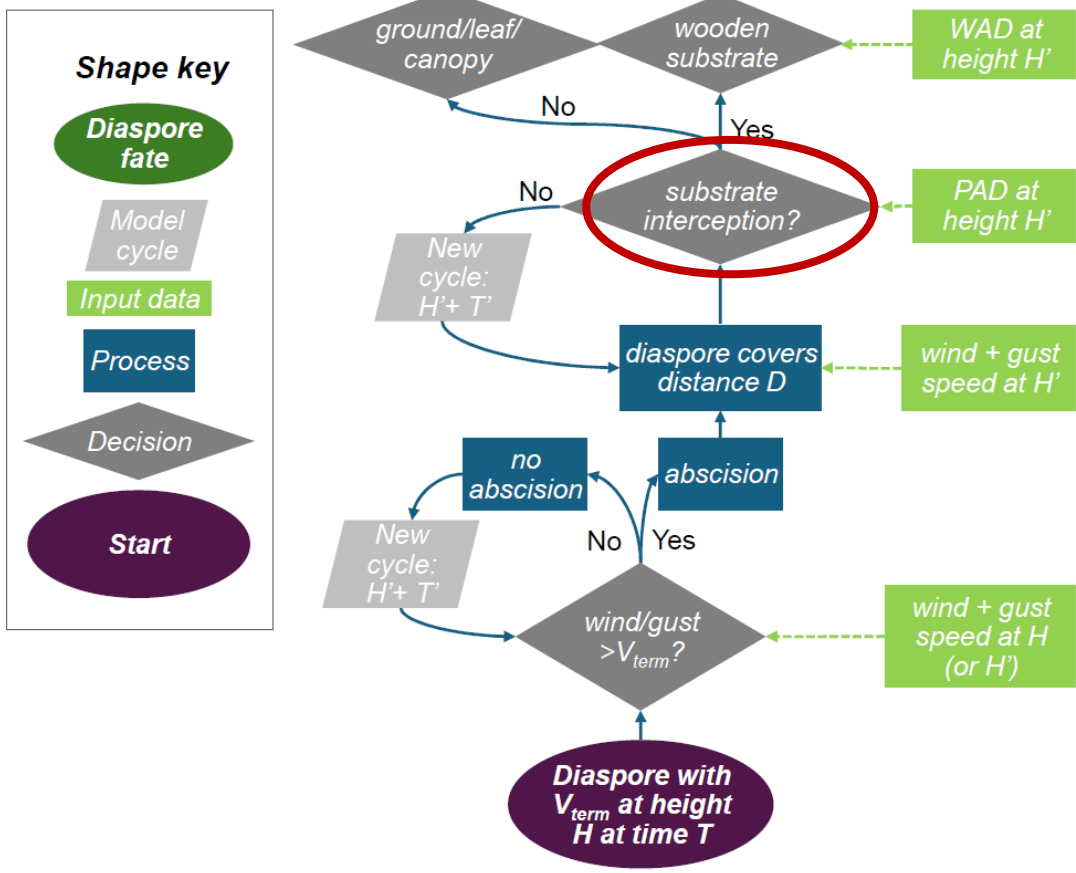
Datetime	Daytime	Replicate #	Position	Vertical wind speed	Vertical gust speed	Horizontal wind speed	Horizontal gust speed	Sensor height	Tree height
2023-04-10 15:43:32	Late day	11.3	In	0.1	-0.67	1.6	0	17.4	36.1

Calculating substrate interception proability

$$I_{Pw} = 1 - \exp(-PAD |D_w|) \qquad \text{or}$$

$$I_{Pu} = 1 - \exp(-PAD |D_u|)$$

(Equations 2a and 2b)



Where I_{Pw} and I_{Pv} are the probabilities that a diaspore is intercepted during the horizontal (P_u) or vertical (P_v) trajectory, ($|D_u|$ or $|D_v|$) the absolute horizontal or vertical distance it travels, and PAD is the plant area density (PAD) at height H .

The model generates a random number (R) between 0 and 1. When $R < P_u$ or $R < P_v$, interception occurs. Specifically, when $R < \text{wood area density (WAD)}$, it was intercepted by wood and considered **successful**.

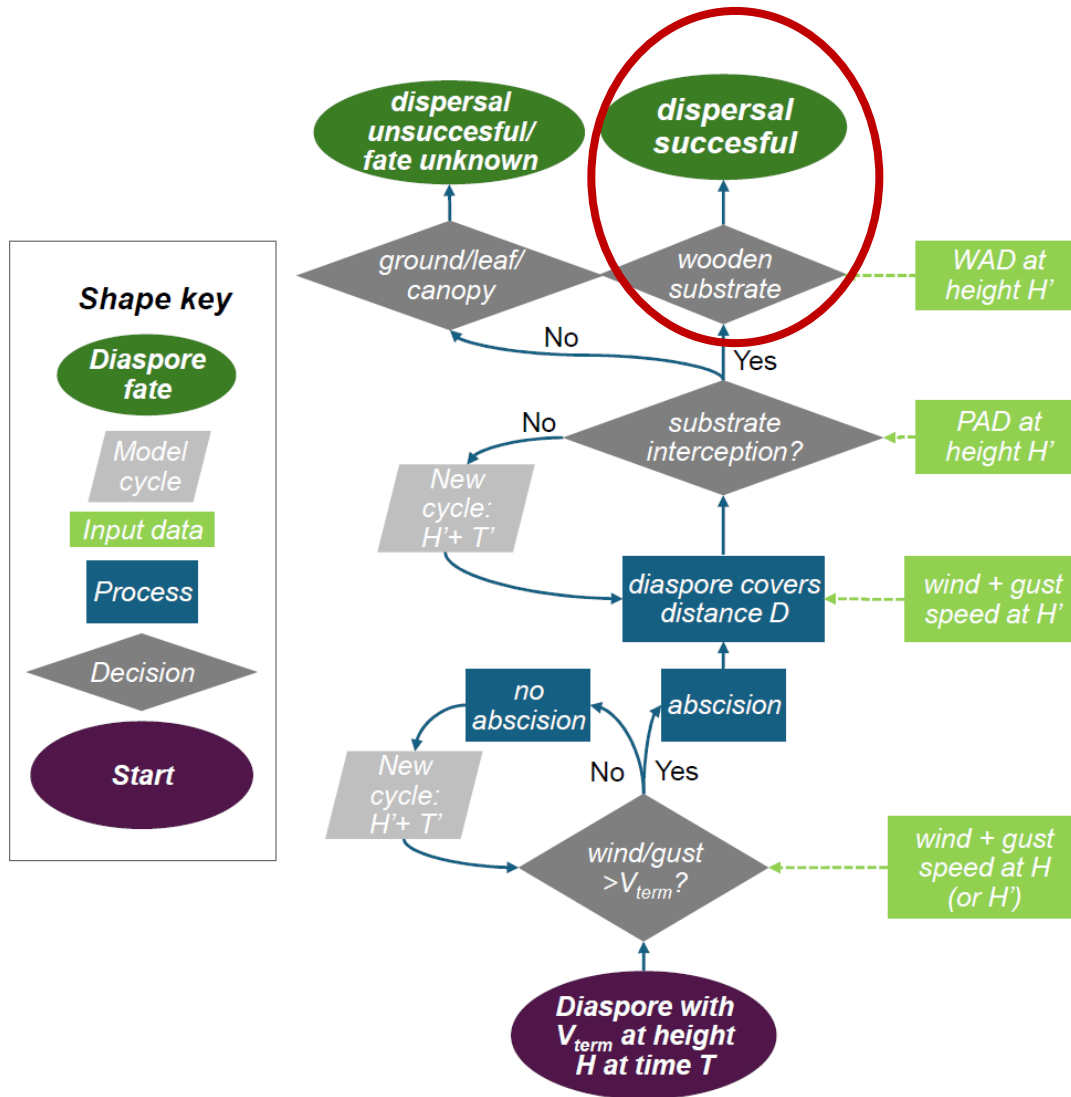
Calculating substrate interception probability

$$I_{Pw} = 1 - \exp(-PAD |D_w|) \quad \text{or}$$

$$I_{Pu} = 1 - \exp(-PAD |D_u|)$$

(Equations 2a and 2b)

When random number $R < \text{wood area density (WAD)}$, it was intercepted by wood and considered **successfully dispersed**



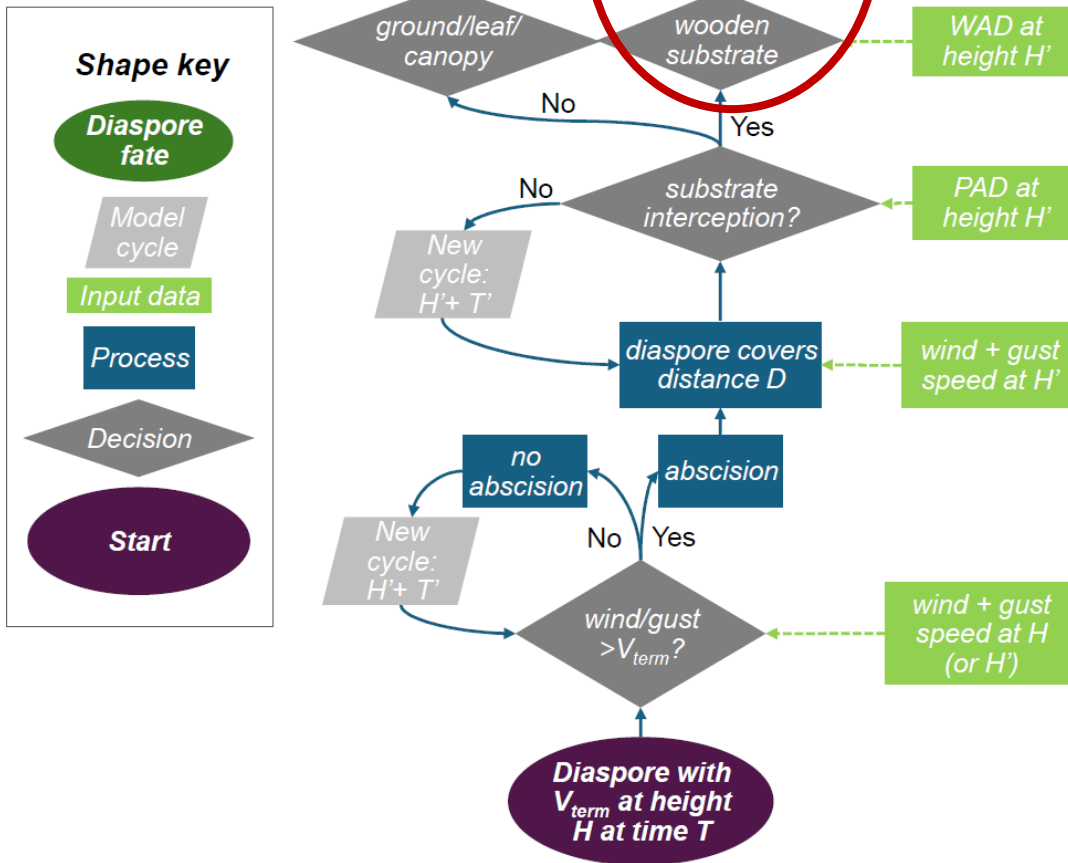
Calculating substrate interception probability

$$I_{Pw} = 1 - \exp(-PAD |D_w|) \quad \text{or}$$

$$I_{Pu} = 1 - \exp(-PAD |D_u|)$$

(Equations 2a and 2b)

When random number $R > \text{wood area density (WAD)}$ but $< PAD$ it was intercepted by a leaf and considered **unsuccessfully dispersed**.

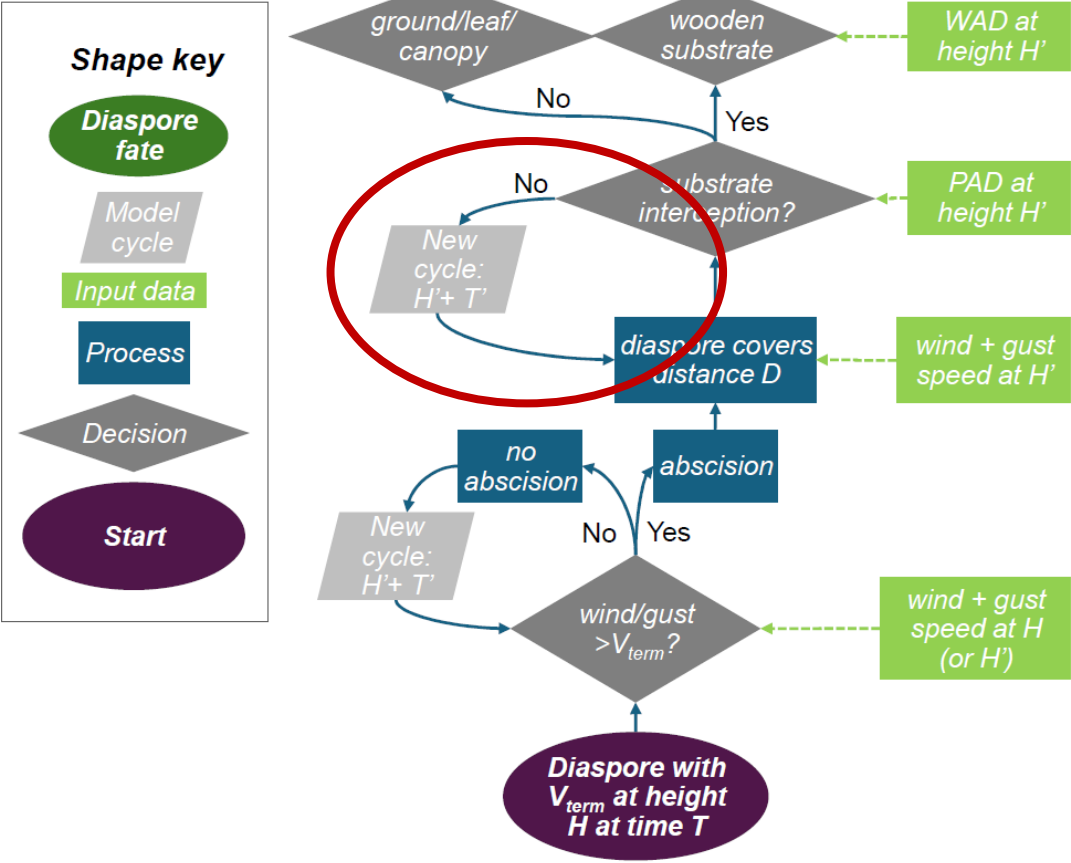


Calculating substrate interception proability

$$I_{Pw} = 1 - \exp(-PAD |D_w|) \qquad \text{or}$$

$$I_{Pu} = 1 - \exp(-PAD |D_u|)$$

(Equations 2a and 2b)



When random number $R > PAD$, it was not intercepted and the simulation continued

Additional columns are added now

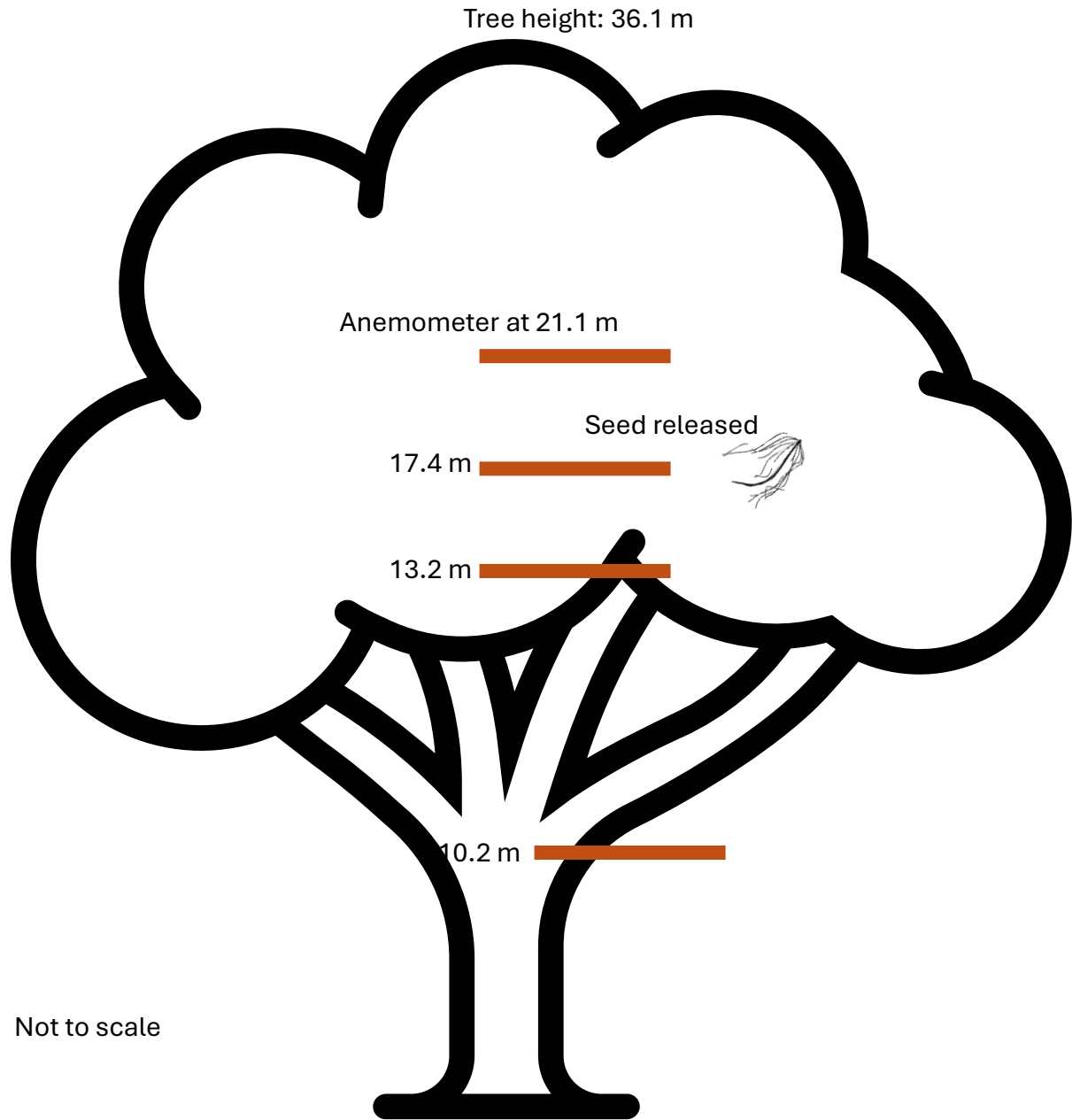
Calculated using
equations 1 and 2

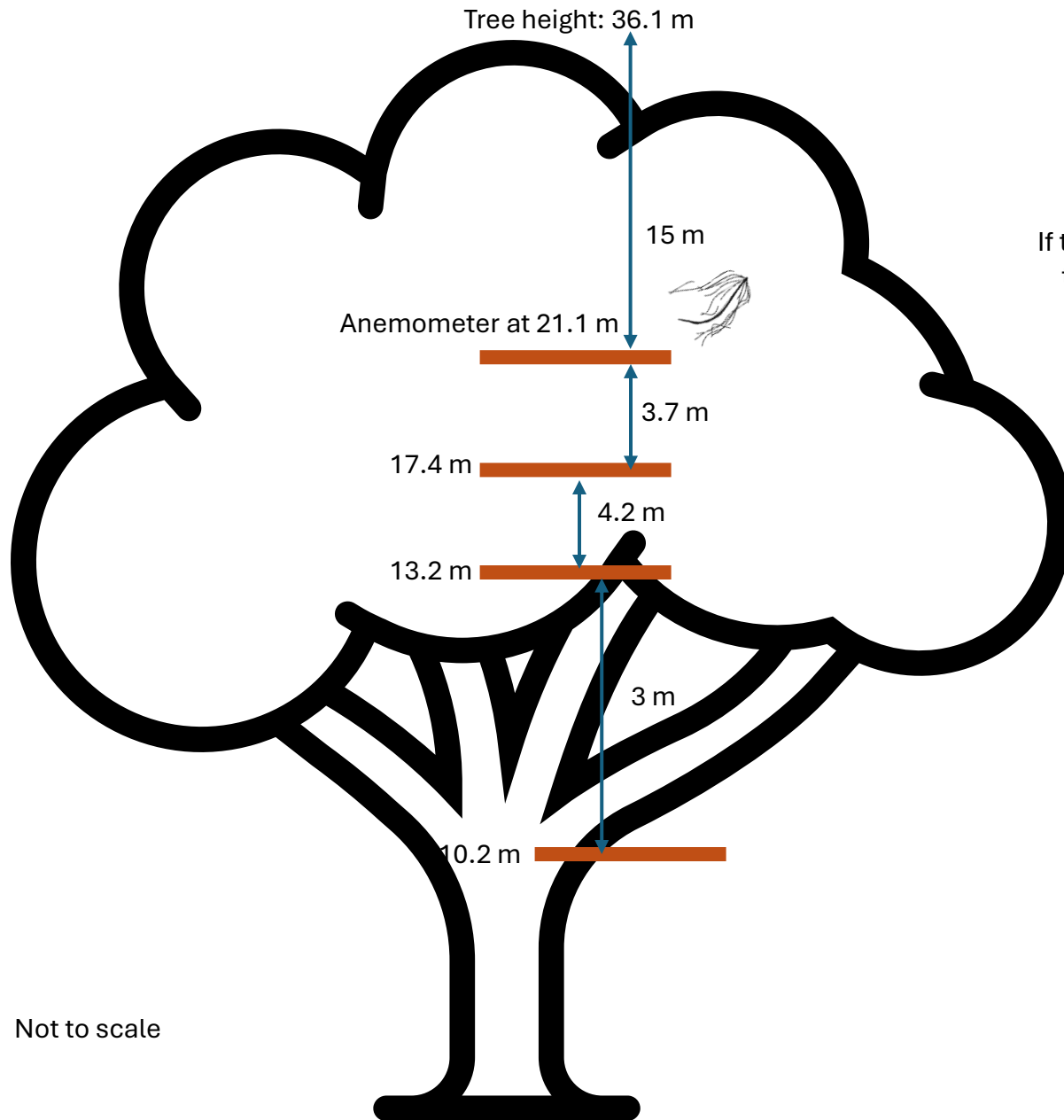
The new distances were
added to the value
present here. First
iteration = 0

*Explained in the
next slide*

Date time	Seed ID	Time interval (s)	Vert. wind speed	Vert. gust speed	Hor. wind speed	Hor. gust speed	Sensor height = Original release height	Tree height	Vterm	Ab-scission	Time of abs.	New vert. distance	New hor. distance	Sum vert. distance	Sum hor. distance	New release height
2023-04-10 21:43:32 2	2023-04-10 21:43:32 — 10.2 m — Out	3	0.1	0.33	1.6	-2.3	10.2	36.1	0.1	T	2023-04-10 21:43:32	0.23	2.43	0 + 0.21 = 0.21	0 + 2.43 = 2.43	10.2
2023-04-10 21:43:32 2	2023-04-10 21:43:32 — 10.2 m — Out	3	0	0	0	0	13.2	36.1	0.1	F	NA	0	0	0	0	13.2
2023-04-10 21:43:32 2	2023-04-10 21:43:32 — 10.2 m — Out	3	0.1	-0.67	1.6	0	17.4	36.1	0.1	T	2023-04-10 21:43:32	-0.67	3.2	0 + 0.32 = 0.32	0 + 3.2 = 3.2	17.4
2023-04-10 21:43:32 2	2023-04-10 21:43:32 — 10.2 m — Out	3	0.1	0	1.6	1.2	21.1	36.1	0.1	T	2023-04-10 21:43:32	0	3.6	0	0 + 3.6 = 3.6	21.1

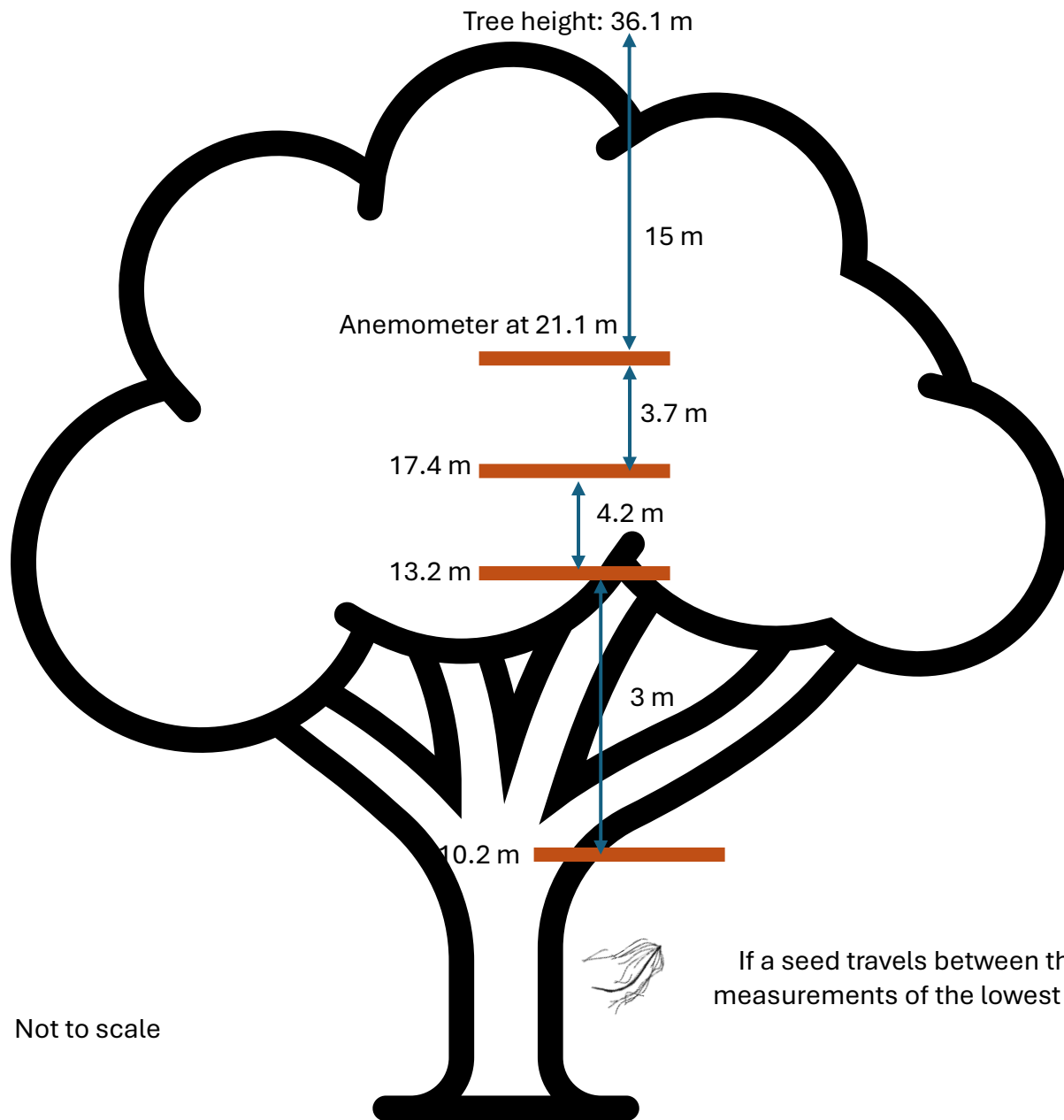
If the total vertical distance travelled by a seed exceeded a higher or lower sensor, that sensor was used for the next wind calculations. For example:





If the vertical distance travelled is equal to or greater than 3.7, the winds/gusts measured at 21.1 m height will affect the seeds trajectory in the next loop

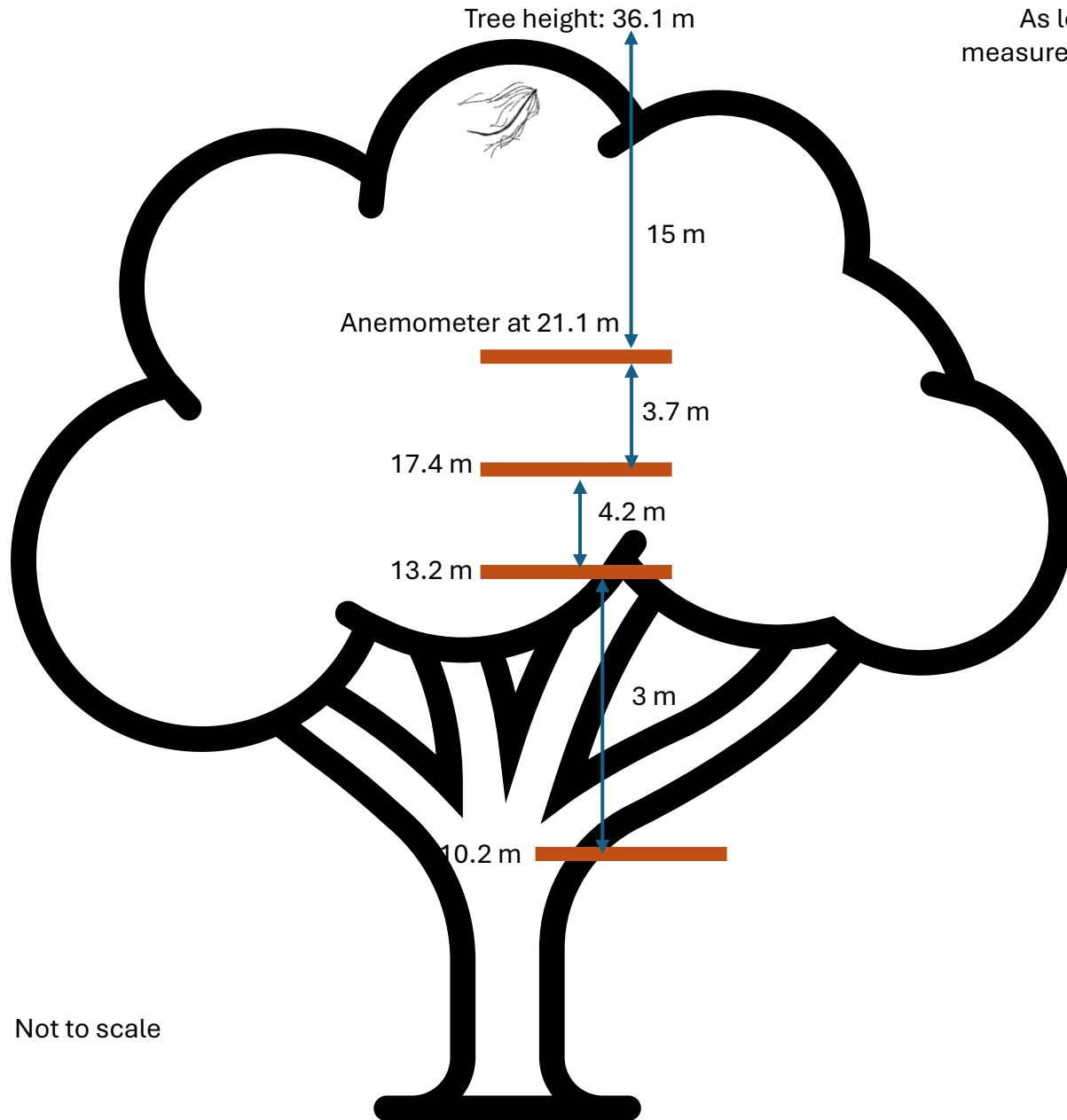
Not to scale



Not to scale

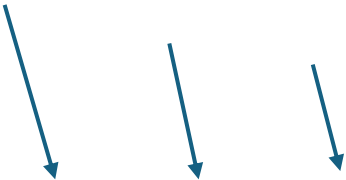


If a seed travels between the ground and the lowest anemometer, the measurements of the lowest anemometer will affect the seed's trajectory



As long as the seed is below or at tree height, the measurements of highest anemometer affect its trajectory

Once the diaspore is intercepted, new columns are generated:



Date time	Seed ID	Time interval (s)	Vert. wind speed	Vert. gust speed	Hor. wind speed	Hor. gust speed	Sensor height = Original release height	Tree height	Vterm	Ab-scission	Time of abs.	New vert. distance	New hor. distance	Sum vert. distance	New release height	Inter-ception	Inter-ception by wood	Height at interception
2023-04-10 21:43:32 2	2023-04-10 21:43:32 — 10.2 m — Out	3	0.1	0.33	1.6	-2.3	10.2	36.1	0.1	T	2023-04-10 21:43:32	0.23	2.43	0 + 0.21 = 0.21	10.2	T	T	10.2 + 0.23 = 10.53
2023-04-10 21:43:32 2	2023-04-10 21:43:32 — 10.2 m — Out	3	0	0	0	0	13.2	36.1	0.1	F	NA	0	0	0	13.2	NA	NA	NA
2023-04-10 21:43:32 2	2023-04-10 21:43:32 — 10.2 m — Out	3	0.1	-0.67	1.6	0	17.4	36.1	0.1	T	2023-04-10 21:43:32	-0.67	3.2	0 + 0.32 = 0.32	17.4	F	NA	NA
2023-04-10 21:43:32 2	2023-04-10 21:43:32 — 10.2 m — Out	3	0.1	0	1.6	1.2	21.1	36.1	0.1	T	2023-04-10 21:43:32	0	3.6	0	21.1	T	F	21.1 + 0 = 21.1

If seeds remain in the loop, results from previous iteration are added to the new

iteration New seeds																
Datetime	Seed ID	Time interval (seconds)	Vertical wind speed	Vert. gust speed	Horizontal wind speed	Hor. gust speed	Sensor height = Original release height	Tree height	Vterm	Abscission	Time of abs.	New vert. distance	New hor. Distance	Sum vert. distance	Sum hor. distance	New release height
2023-04-10 21:43:33	2023-04-10 21:43:33 __ 10.2 m __ Out	3	0	0	0	0	10.2	36.1	0.1	F	NA	0	0	0	0	10.2
2023-04-10 21:43:33	2023-04-10 21:43:33 __ 10.2 m __ Out	3	0.1	0.3	-1	-0.7	13.2	36.1	0.1	T	2023-04-10 21:43:33	0.2	-2.23	0 + 0.2 = 0.2	0 – 2.23 = -2.23	13.2
2023-04-10 21:43:33	2023-04-10 21:43:33__ 10.2 m __ Out	3	0	0	0	0	17.4	36.1	0.1	F	NA	0	0	0	0	17.4
2023-04-10 21:43:33	2023-04-10 21:43:33__ 10.2 m __ Out	3	0	0	0	0	21.1	36.1	0.1	F	NA	0	0	0	0	21.1
Old seeds with old ID, but with new datetime and new wind/gust measurements.																
2023-04-10 21:43:33	2023-04-10 21:43:32 __ 10.2 m __ Out	3	0	0	0	0	10.2	36.1	0.1	F	2023-04-10 21:43:32	0	0	0.21 + 0 = 0.21	2.43 + 0 = 2.43	10.2
2023-04-10 21:43:33	2023-04-10 21:43:32__ 10.2 m __ Out	3	0.1	0.3	-1	-0.7	13.2	36.1	0.1	T	2023-04-10 21:43:33	0.2	-2.23	0.2 + 0 = 0.2	0 - 2.23 = -2.33	13.2
2023-04-10 21:43:33	2023-04-10 21:43:32__ 10.2 m __ Out	3	0	0	0	0	17.4	36.1	0.1	F	2023-04-10 21:43:32	0	0	0 + 0.32 = 0.32	3.2 + 0 = 3.2	17.4
2023-04-10 21:43:33	2023-04-10 21:43:32__ 10.2 m __ Out	3	0	0	0	0	21.1	36.1	0.1	F	2023-04-10 21:43:32	0	0	0	3.6 + 0 = 3.6	21.1

Example: a seed moves into a new category

Datetime	Seed ID	Time interval (seconds)	Vertical wind speed	Vert. gust speed	Horizontal wind speed	Hor. gust speed	Sensor height = Original release height	Tree height	Vterm	Abscission	Time of abs.	New vert. distance	New hor. Distance	Sum vert. distance	Sum hor. distance	New release height
2023-04-10 21:50:30	2023-04-10 21:43:33 __ 10.2 m __ Out	3	0	0	0	0	10.2	36.1	0.1	F	2023-04-10 21:43:35	0	0	2.5	0	10.2

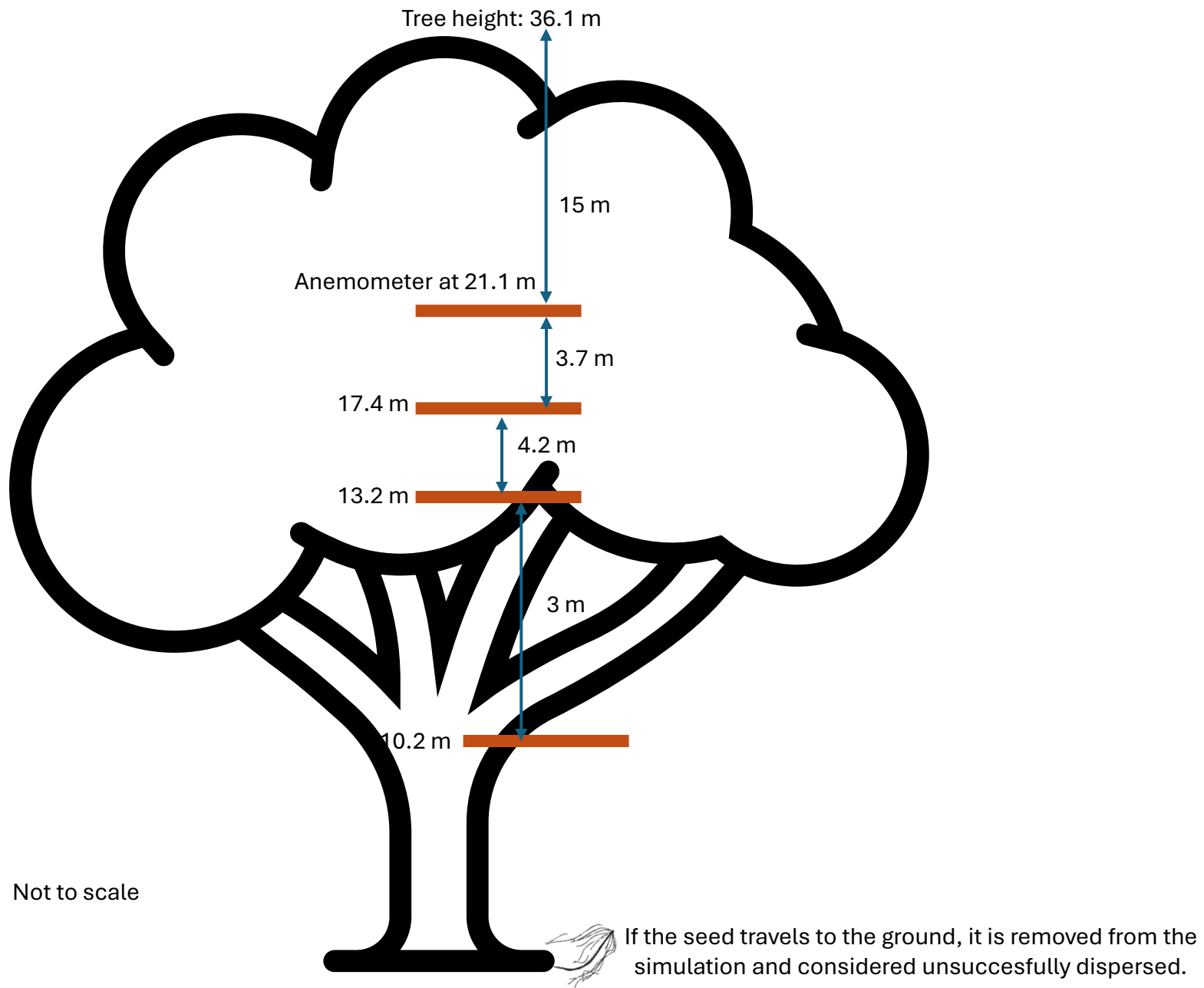
Datetime	Seed ID	Time interval (seconds)	Vertical wind speed	Vert. gust speed	Horizontal wind speed	Hor. gust speed	Sensor height = Original release height	Tree height	Vterm	Abscission	Time of abs.	New vert. distance	New hor. Distance	Sum vert. distance	Sum hor. distance	New release height
2023-04-10 21:50:33	2023-04-10 21:43:33 __ 10.2 m __ Out	3	1	2	0	0	10.2	36.1	0.1	F	2023-04-10 21:43:35	2.67	0	2.5 + 2.67 = 5.17	0	13.2

An upward dispersal of 5.17 m, in this case, exceeds the height of the next highest anemometer

The measurements of this anemometer are now used to calculate the further trajectory

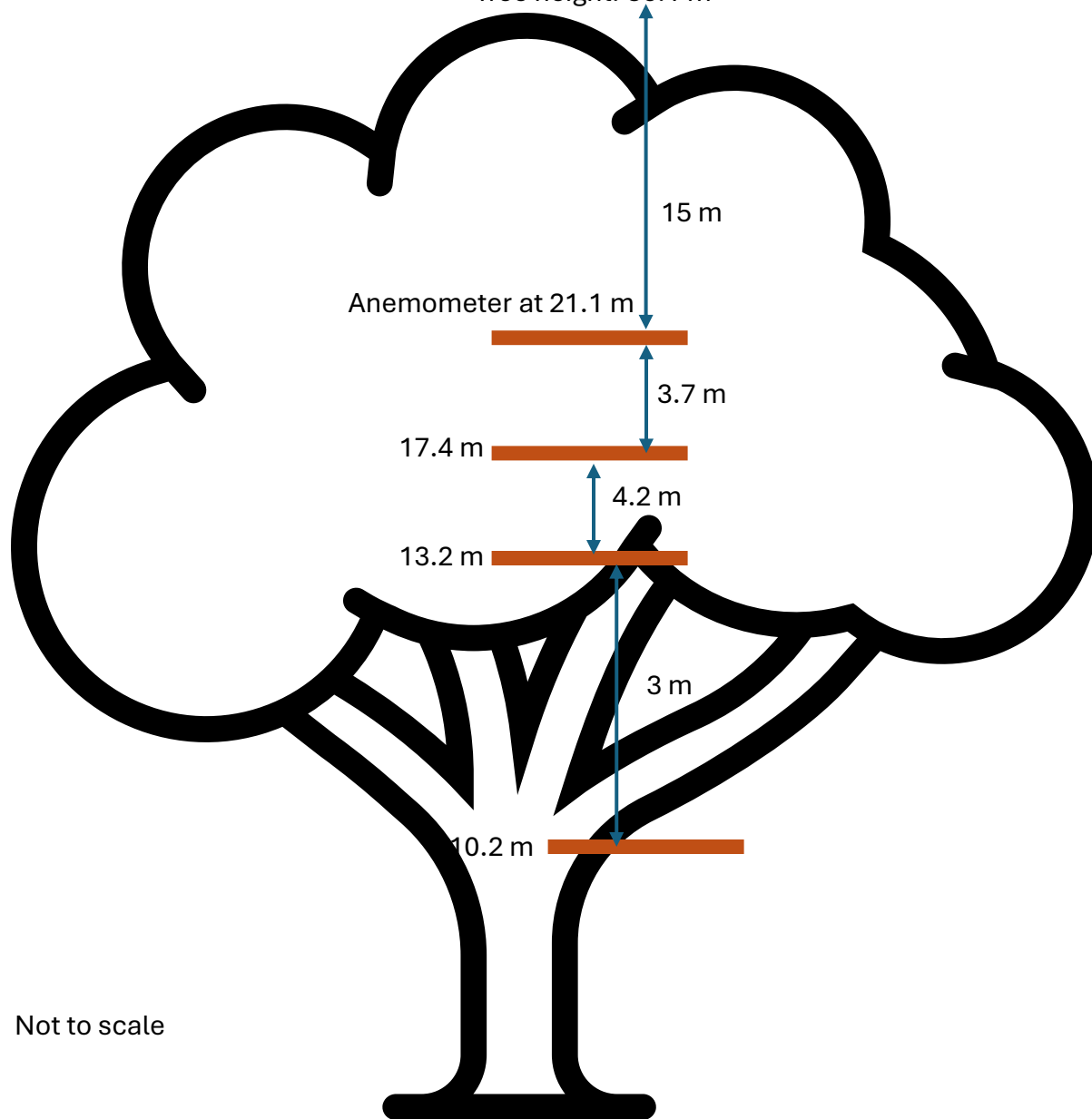
The iterations continue until the last datetime in a data frame is reached, and then begin anew with a different terminal velocity. All seeds dispersed beyond the tree height or to the ground are stored in a separate dataframe. All seeds still unabsorbed or travelling within the tree column remain.

Dispersal outcomes

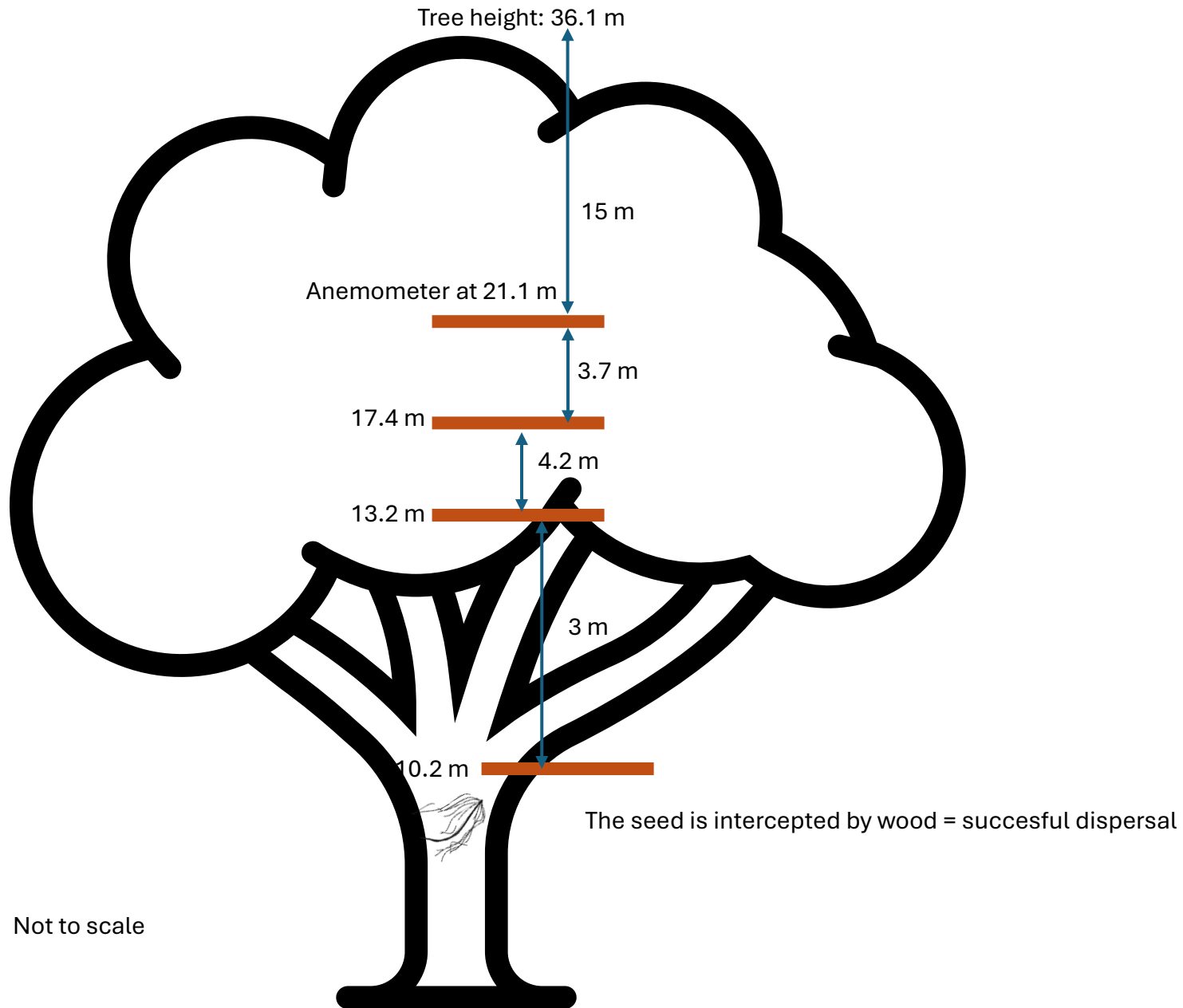


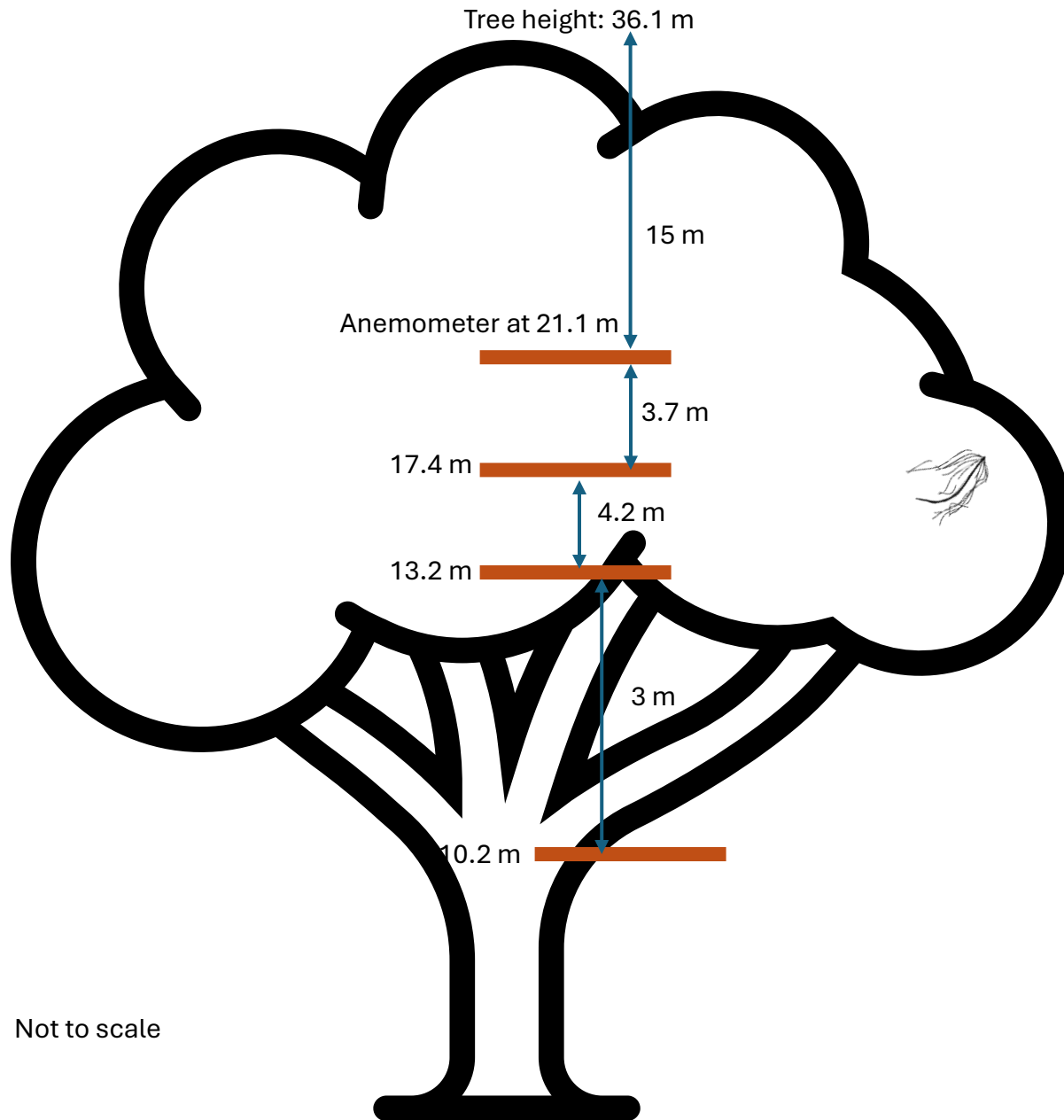


Once the seed has travelled beyond the tree height, it is removed from the simulation. *It now has the chance for long-distance dispersal.*



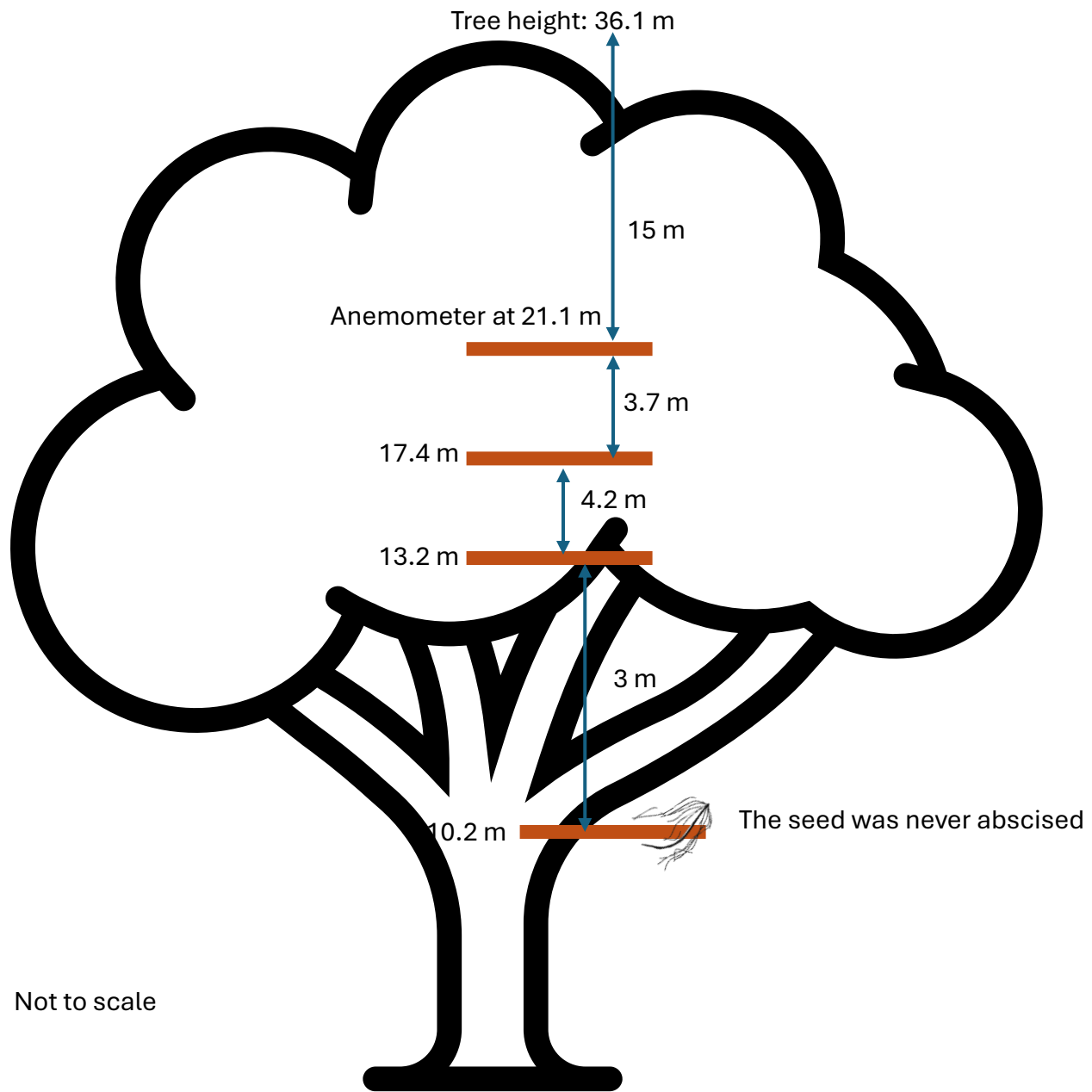
Not to scale



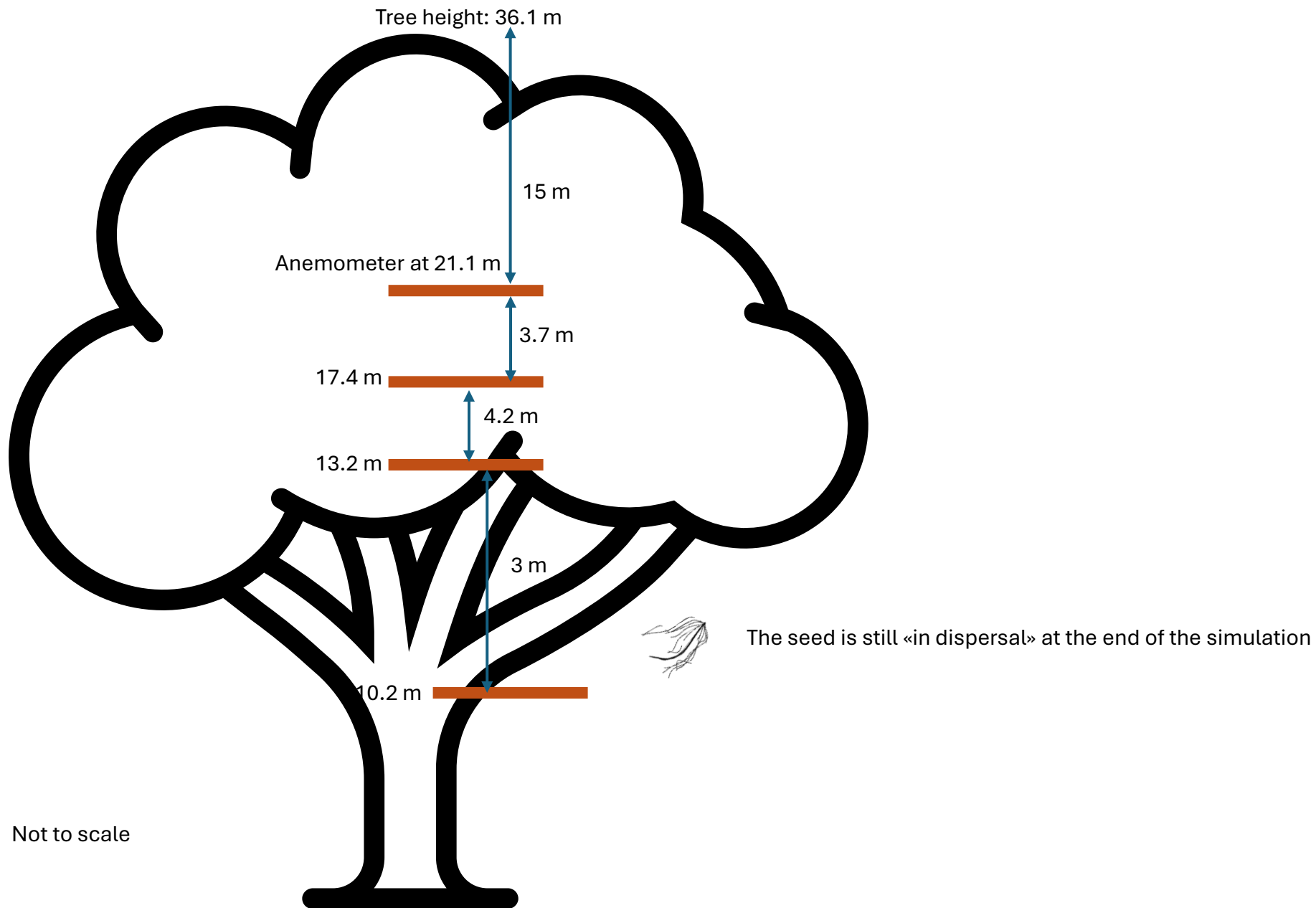


The seed is intercepted by a leaf = unsuccessful dispersal

Not to scale



Not to scale



Ecological Monographs

Appendix 4: Additional Results for:

Patterns of wind dispersal within a forest canopy: An epiphyte perspective

Marie Hoensbroech¹, Juliano Sarmiento Cabral², Gerhard Zotz^{1,3}

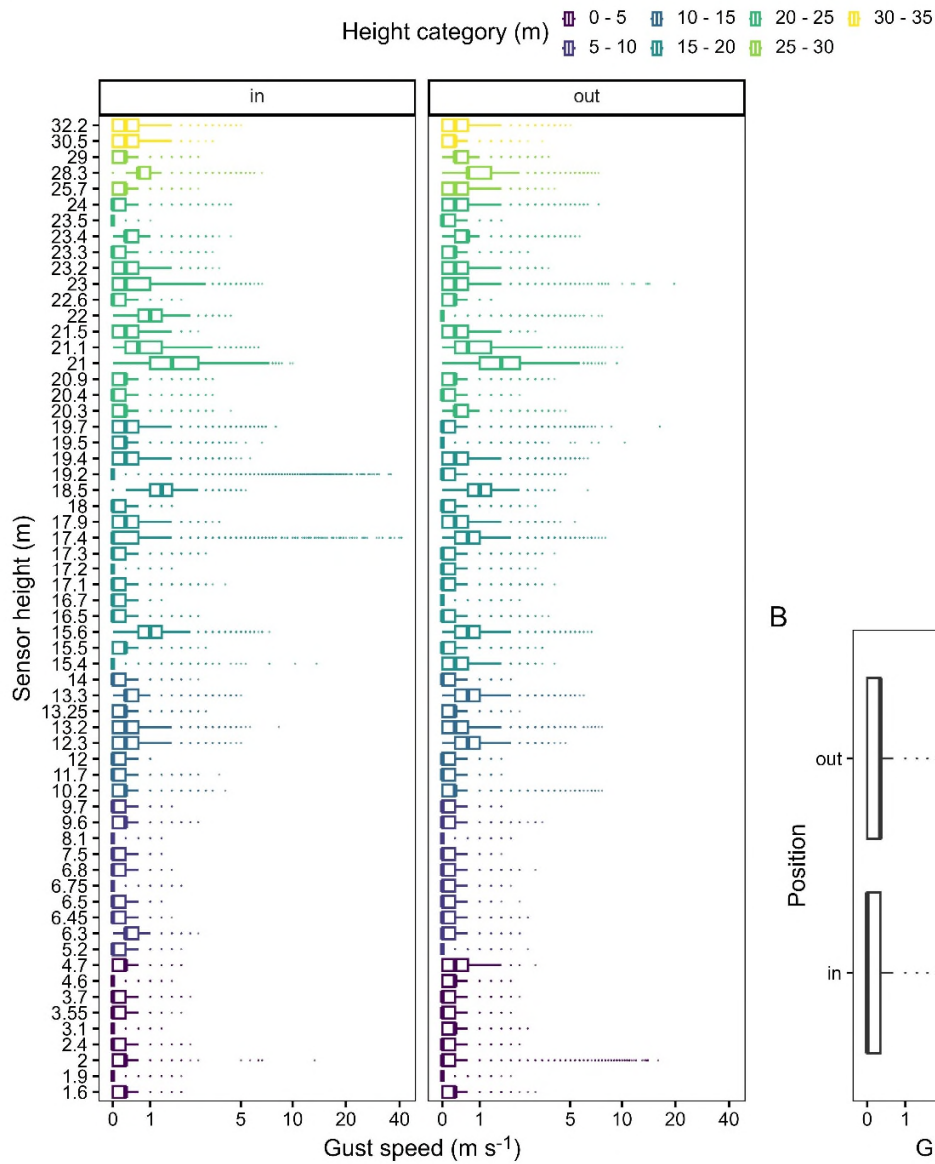
1: Institute of Biology and Environmental Sciences, Carl von Ossietzky Universität Oldenburg,
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A



B

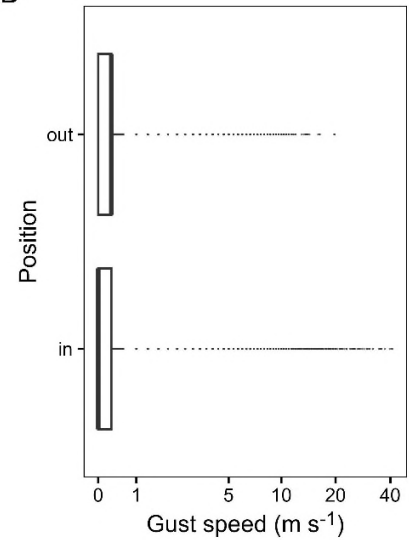
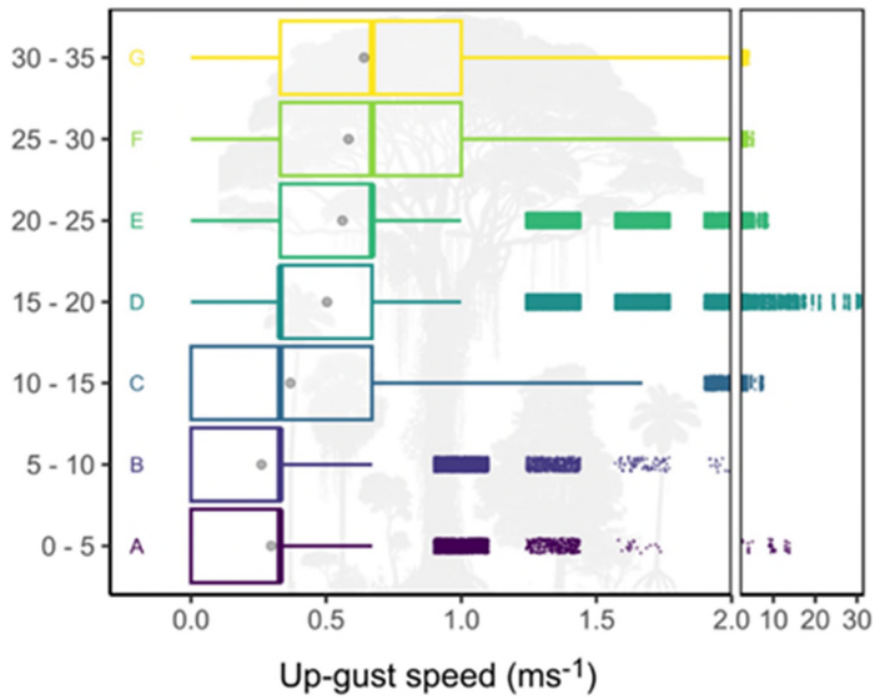


Figure S1. Variation in gust speed (all directions) across anemometer positions and measurement heights. **(A)** Differences in gust speed among replicates across height categories and anemometer positions (in = 0.35 m from the stem; out = 1.35 m from the stem). Boxplots show the median (vertical line), the interquartile range (box), values within $1.5 \times$ the interquartile range (whiskers), and outliers (points). The x-axis is truncated at 1.5 to emphasize the denser range of

the data. **(B)** Differences in gust speed between the in and out anemometer positions, shown as standard boxplots. Although Dunn's test detected a significant difference between positions (compact letter display above the x-axis), the two distributions differ mainly in the frequency and extent of outliers.



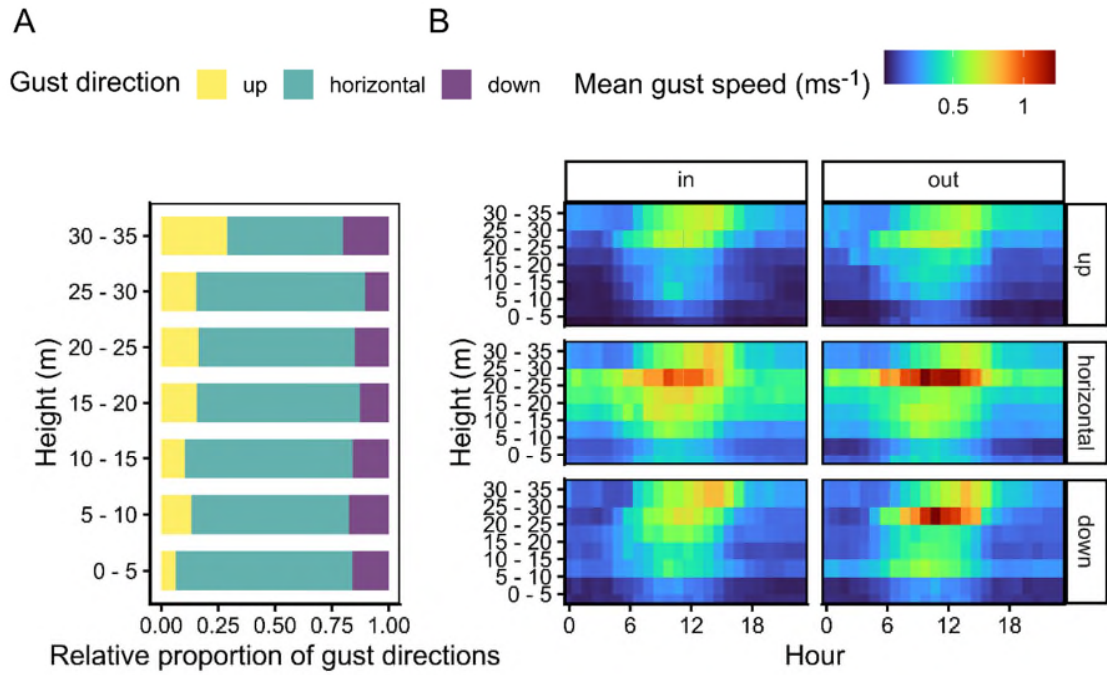


Figure S3. (A) Relative proportion of gust directions (gusts $\geq 0.33 \text{ ms}^{-1}$) per height category. **(B)** Mean gust speeds per hour of the day, based on their moduli (absolute values), separated by distance from stem (in = 0.35 cm, out = 1.35 cm) and gust direction.

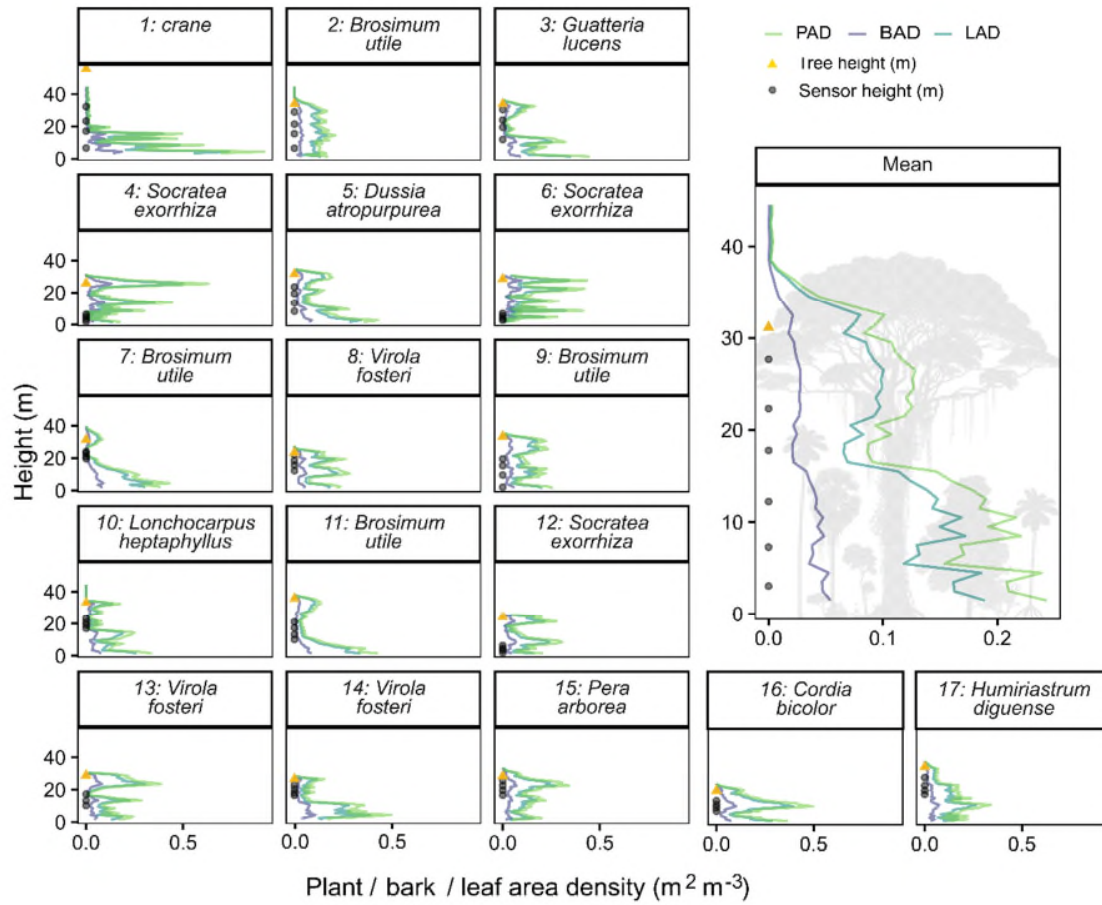


Figure S4. Bark area density (BAD), leaf area density (LAD), and their combination, the plant area density (PAD) along the individual vertical profiles and averaged across all, with the (mean) position of anemometers (grey dots) and tree tops (yellow triangle) shown. Refer to the Materials and Methods section in the main text for their derivation. The silhouette in the mean panel illustrates forest structure and was adapted by MH from a graphic generated with Microsoft Copilot (Microsoft, 2024). PAD was calculated from lidar data provided by Vasquez et al. (2024); the derivation of BAD and LAD leaned heavily on previous censuses at San Lorenzo (Janner & Fernández-Lucero, unpublished data; Ramos et al., 2024).

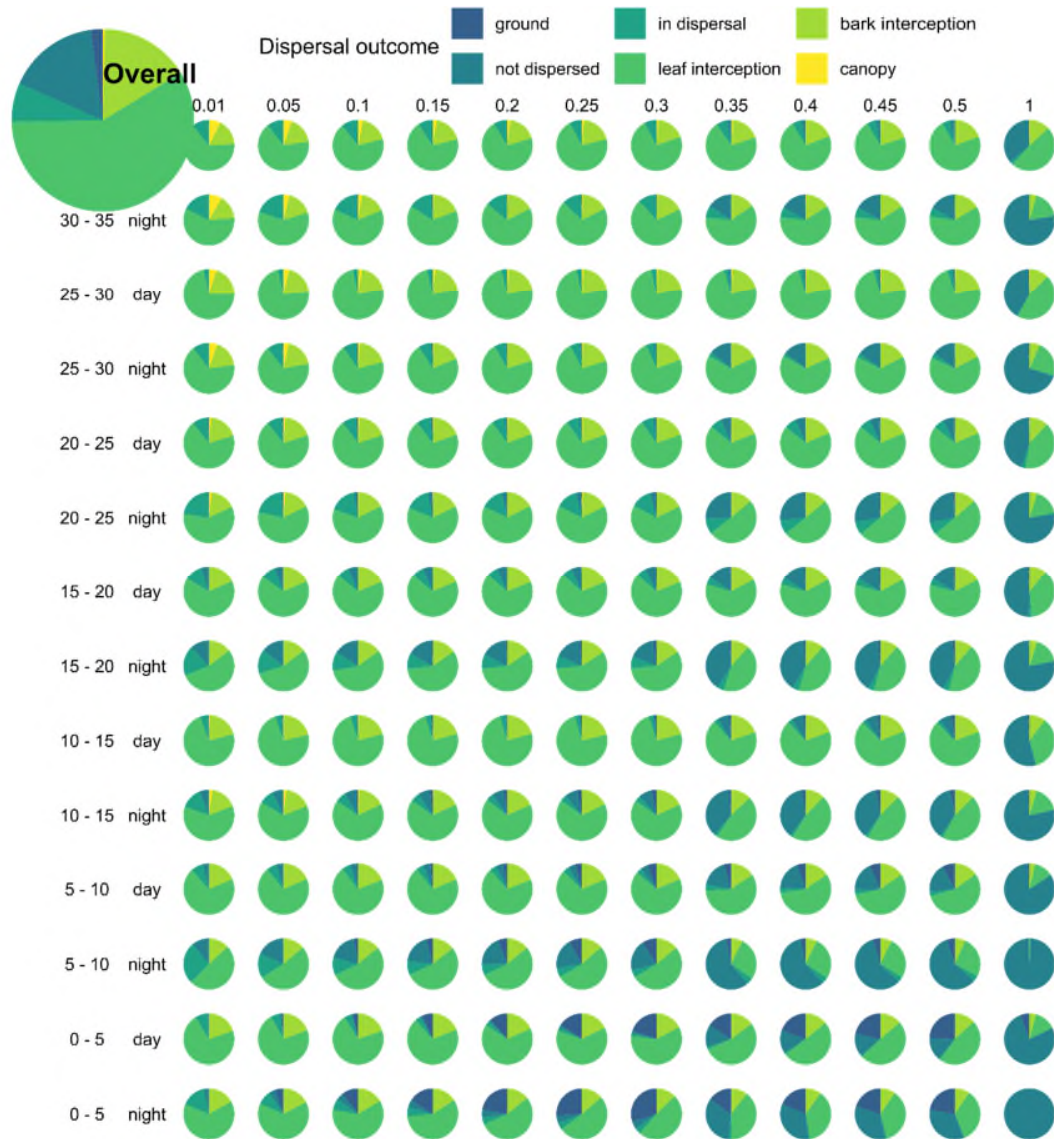


Figure S5. Relative proportion of dispersal outcomes of the hypothetical diaspores released by the trajectory model EpiDisp, separated by release height, time of day (rows), and terminal velocity (V_{term} , columns) and overall proportion (inset). Ground = diaspore was dispersed to the ground; unabscised = diaspore was not abscised from the mother plant; dispersal = diaspore was still in the air at the end of the simulation period; leaf interception = diaspore intercepted a leaf

within the forest; bark interception = diaspore intercepted by bark within the forest; canopy = diaspore was dispersed above the canopy, where the simulation stopped.

Table S1. The number of simulated days per V_{term} and release height category, the absolute numbers of diaspores released (N_{total}), the absolute number and percentage (in parentheses) of diaspores per dispersal outcome. The last column reports diaspores that settled on bark after dispersing more than 5 m vertically (first number) or horizontally (second number), with their proportion of all successfully dispersed diaspores (bark interception) shown in parentheses.

V_{term}	Height	Days	N_{total}	Un- abscised	In disper- sal	Ground	Leaf inter- ception	Bark Inter- ception	Above- canopy	>5m
0.01	0 - 5	3.4	30521	516 (1.7%)	3585 (11.7%)	68 (0.2%)	20731 (67.9%)	5621 (18.4%)	0	48 (<0.01%) / 1702 (0.3%)
	5 - 10	8.6	40808	2804 (6.9%)	6644 (16.3%)	7 (<0.01%)	24694 (60.5%)	6605 (16.2%)	54 (0.1%)	123 (<0.01%) / 1849 (0.3%)
	10 - 15	7.2	59345	1698 (2.9%)	5499 (9.3%)	76 (0.1%)	40175 (67.7%)	10985 (18.5%)	912 (1.5%)	357 (<0.01%) / 4006 (0.4%)
	15 - 20	13.7	10464 6	8690 (8.3%)	15598 (14.9%)	26 (<0.01%)	63007 (60.2%)	17106 (16.3%)	219 (0.2%)	907 (<0.01%) / 7851 (0.5%)
	20 - 25	11.9	90651	893 (1%)	13814 (15.2%)	0	58500 (64.5%)	15880 (17.5%)	1564 (1.7%)	871 (<0.01%) / 7133 (0.4%)
	25 - 30	3.8	20332	19 (<0.01%)	1360 (6.7%)	0	14095 (69.3%)	3809 (18.7%)	1049 (5.2%)	354 (<0.01%) / 2459 (0.6%)
	30 - 35	2.6	16784	2 (<0.01%)	2354 (14%)	0	10401 (62%)	2698 (16.1%)	1329 (7.9%)	615 (0.2%) / 1233 (0.5%)
0.1	0 - 5	3.4	30521	882 (2.9%)	1773 (5.8%)	2049 (6.7%)	20268 (66.4%)	5549 (18.2%)	0	3 (<0.01%) / 1665 (0.3%)
	5 - 10	8.6	38473	4073 (10.6%)	3047 (7.9%)	533 (1.4%)	24281 (63.1%)	6529 (17%)	10 (<0.01%)	167 (<0.01%) / 1754 (0.3%)
	10 - 15	7.1	57066	2247 (3.9%)	3012 (5.3%)	330 (0.6%)	40357 (70.7%)	10779 (18.9%)	341 (0.6%)	468 (<0.01%) / 3900 (0.4%)
	15 - 20	13.7	10025 6	10730 (10.7%)	9803 (9.8%)	83 (<0.01%)	62601 (62.4%)	16955 (16.9%)	84 (<0.01%)	1005 (<0.01%) / 7820 (0.5%)
	20 - 25	11.9	88450	1696 (1.9%)	11240 (12.7%)	5 (<0.01%)	58769 (66.4%)	15963 (18%)	777 (0.9%)	914 (<0.01%) / 7424 (0.5%)
	25 - 30	3.8	19787	80 (0.4%)	1191 (6%)	0	14218 (71.9%)	3925 (19.8%)	373 (1.9%)	236 (<0.01%) / 2518 (0.6%)
	30 - 35	2.6	16525	45 (0.3%)	2326 (14.1%)	0	10871 (65.8%)	2813 (17%)	470 (2.8%)	1005 (0.4%) / 1270 (0.5%)
0.3	0 - 5	3.4	30520	978 (3.2%)	855 (2.8%)	7543 (24.7%)	16550 (54.2%)	4594 (15.1%)	0	1 (<0.01%) / 1019 (0.2%)
	5 - 10	8.6	37043	4181 (11.3%)	1488 (4%)	2665 (7.2%)	22532 (60.8%)	6177 (16.7%)	0	576 (<0.01%) / 1347 (0.2%)
	10 - 15	7.1	54759	2338 (4.3%)	1850 (3.4%)	1099 (2%)	38765 (70.8%)	10690 (19.5%)	17 (<0.01%)	1175 (0.1%) / 3645 (0.3%)

	15 - 20	13.7	95489	10867 (11.4%)	6611 (6.9%)	1197 (1.3%)	60334 (63.2%)	16459 (17.2%)	21 (<0.01%)	2639 (0.2%) / 7139 (0.4%)
	20 - 25	11.9	81985	1896 (2.3%)	8473 (10.3%)	405 (0.5%)	55736 (68%)	15173 (18.5%)	302 (0.4%)	2408 (0.2%) / 6612 (0.4%)
	25 - 30	3.8	17430	100 (0.6%)	737 (4.2%)	1 (<0.01%)	12966 (74.4%)	3529 (20.2%)	97 (0.6%)	353 (0.1%) / 2197 (0.6%)
	30 - 35	2.6	15202	60 (0.4%)	1392 (9.2%)	0	10875 (71.5%)	2753 (18.1%)	122 (0.8%)	1595 (0.6%) / 1021 (0.4%)
0.5	0 - 5	3.3	30521	7220 (23.7%)	293 (1%)	7137 (23.4%)	12395 (40.6%)	3476 (11.4%)	0	6 (<0.01%) / 585 (0.2%)
	5 - 10	7.6	37732	13996 (37.1%)	1162 (3.1%)	2483 (6.6%)	15841 (42%)	4250 (11.3%)	0	482 (0.1%) / 999 (0.2%)
	10 - 15	6.6	56435	12473 (22.1%)	1145 (2%)	1431 (2.5%)	32301 (57.2%)	9074 (16.1%)	11 (<0.01%)	1309 (0.1%) / 3437 (0.4%)
	15 - 20	13.3	97532	26851 (27.5%)	4030 (4.1%)	1445 (1.5%)	51115 (52.4%)	14079 (14.4%)	12 (<0.01%)	2849 (0.2%) / 6657 (0.5%)
	20 - 25	11.7	81373	13296 (16.3%)	5954 (7.3%)	596 (0.7%)	48274 (59.3%)	13091 (16.1%)	162 (0.2%)	2763 (0.2%) / 5899 (0.5%)
	25 - 30	3.8	16771	1343 (8%)	505 (3%)	14 (<0.01%)	11619 (69.3%)	3224 (19.2%)	66 (0.4%)	637 (0.2%) / 1926 (0.6%)
	30 - 35	2.6	14641	1392 (9.5%)	813 (5.6%)	0	9802 (66.9%)	2591 (17.7%)	43 (0.3%)	1637 (0.6%) / 860 (0.3%)
1	0 - 5	3.1	30521	26847 (88%)	0	902 (3%)	2119 (6.9%)	653 (2.1%)	0	0 / 219 (0.3%)
	5 - 10	7.0	38740	35055 (90.5%)	109 (0.3%)	288 (0.7%)	2527 (6.5%)	761 (2%)	0	84 (0.1%) / 241 (0.3%)
	10 - 15	6.0	58314	37551 (64.4%)	240 (0.4%)	684 (1.2%)	15403 (26.4%)	4436 (7.6%)	0	753 (0.2%) / 2115 (0.5%)
	15 - 20	12.1	10257 9	63952 (62.3%)	1159 (1.1%)	837 (0.8%)	28403 (27.7%)	8223 (8%)	5 (<0.01%)	1865 (0.2%) / 4870 (0.6%)
	20 - 25	10.4	85380	51218 (60%)	1581 (1.9%)	376 (0.4%)	24858 (29.1%)	7309 (8.6%)	38 (<0.01%)	1675 (0.2%) / 4082 (0.6%)
	25 - 30	3.5	19835	10874 (54.8%)	182 (0.9%)	129 (0.7%)	6746 (34%)	1873 (9.4%)	31 (0.2%)	519 (0.3%) / 1494 (0.8%)
	30 - 35	2.3	15527	8796 (56.6%)	192 (1.2%)	0	5156 (33.2%)	1354 (8.7%)	29 (0.2%)	916 (0.7%) / 564 (0.4%)

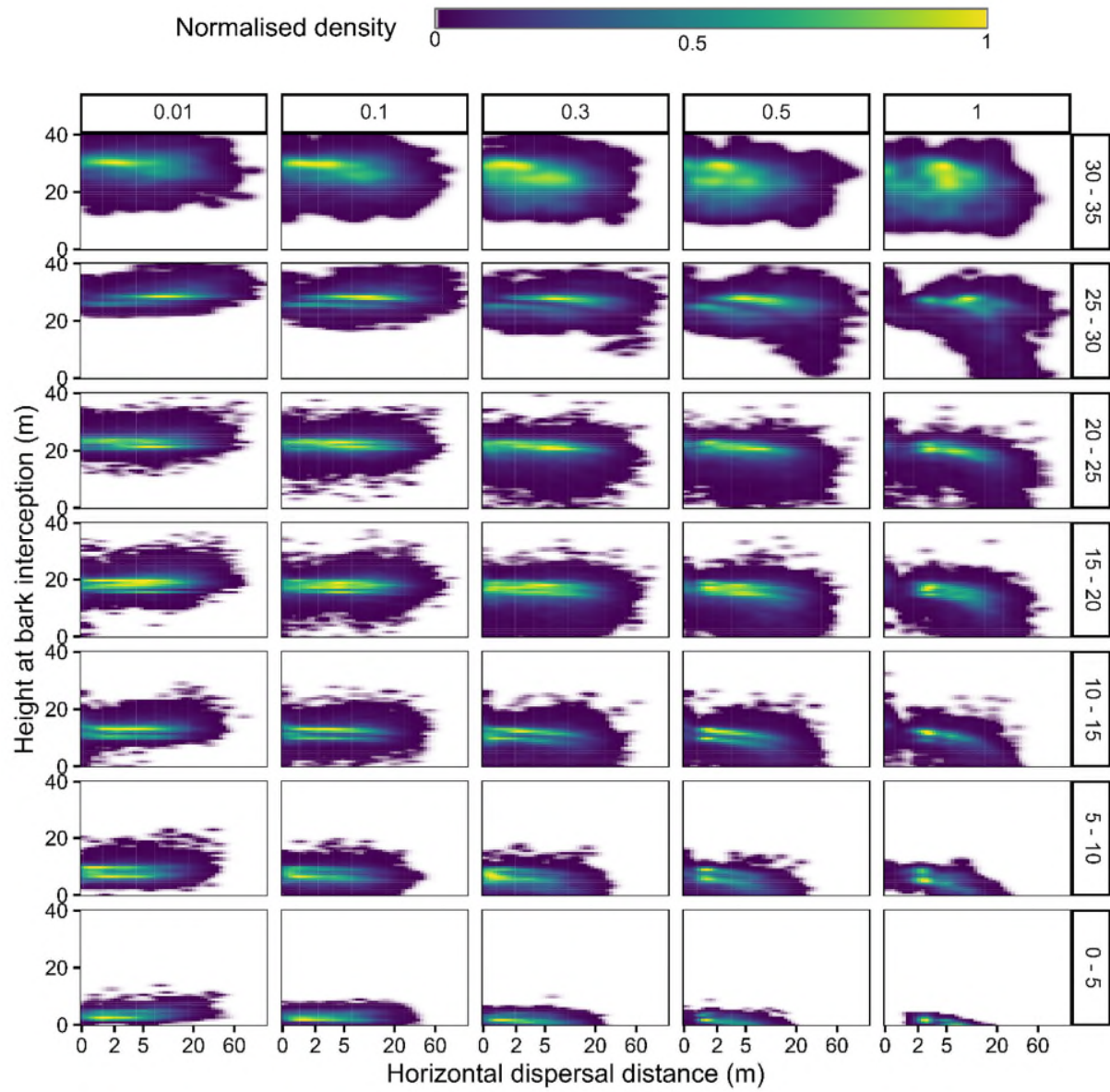


Figure S6. Density distribution of bark-intercepted diaspores. Normalized two-dimensional densities of successful dispersal outcomes within the forest, based on horizontal dispersal distance and height after bark interception, shown across terminal velocity classes (columns) and release-height classes (rows). Zero densities are shown as white gradient fill.

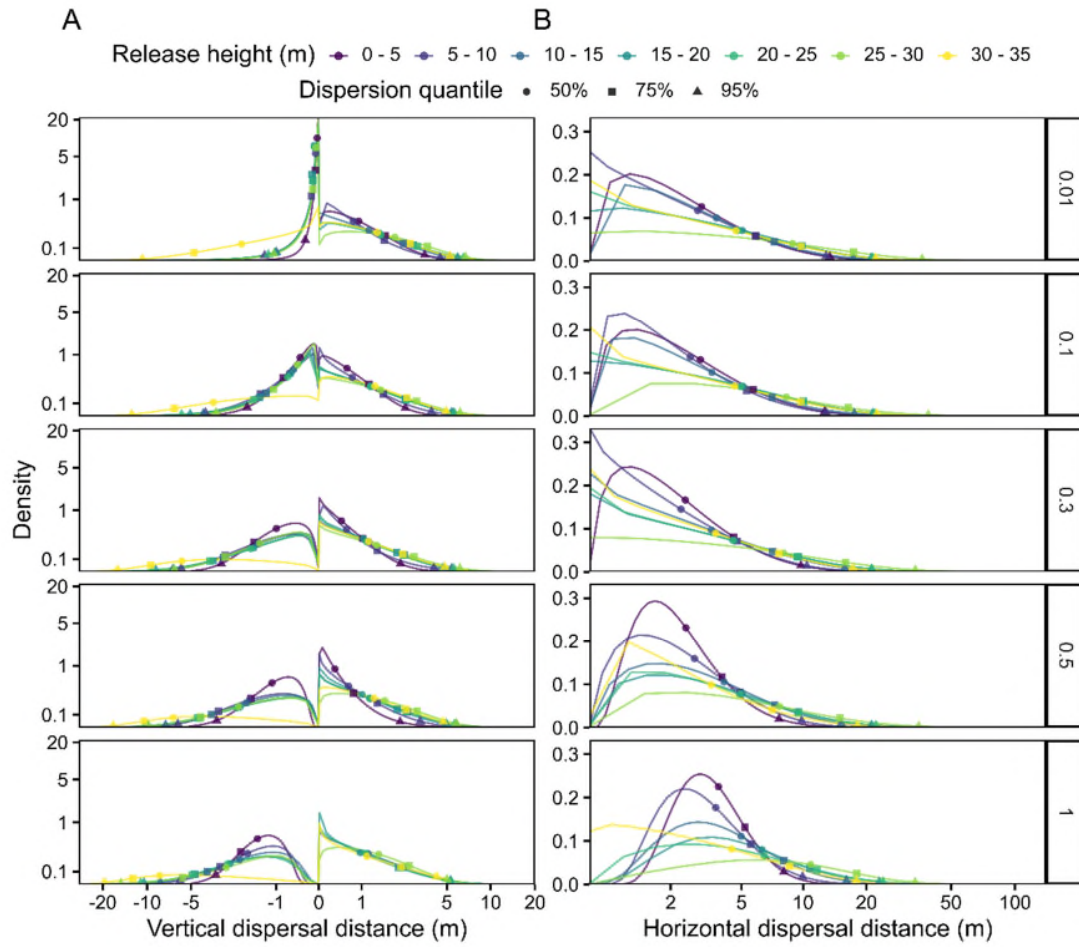


Figure S7. Fitted dispersal kernels for vertical (A) and horizontal (B) dispersal distances across release heights (colours) and terminal velocities (rows) for diaspores intercepted by bark.

Markers show at which distance the cumulative dispersed population reaches a given percentage.

The axes have been pseudo-log transformed; note the different scales.

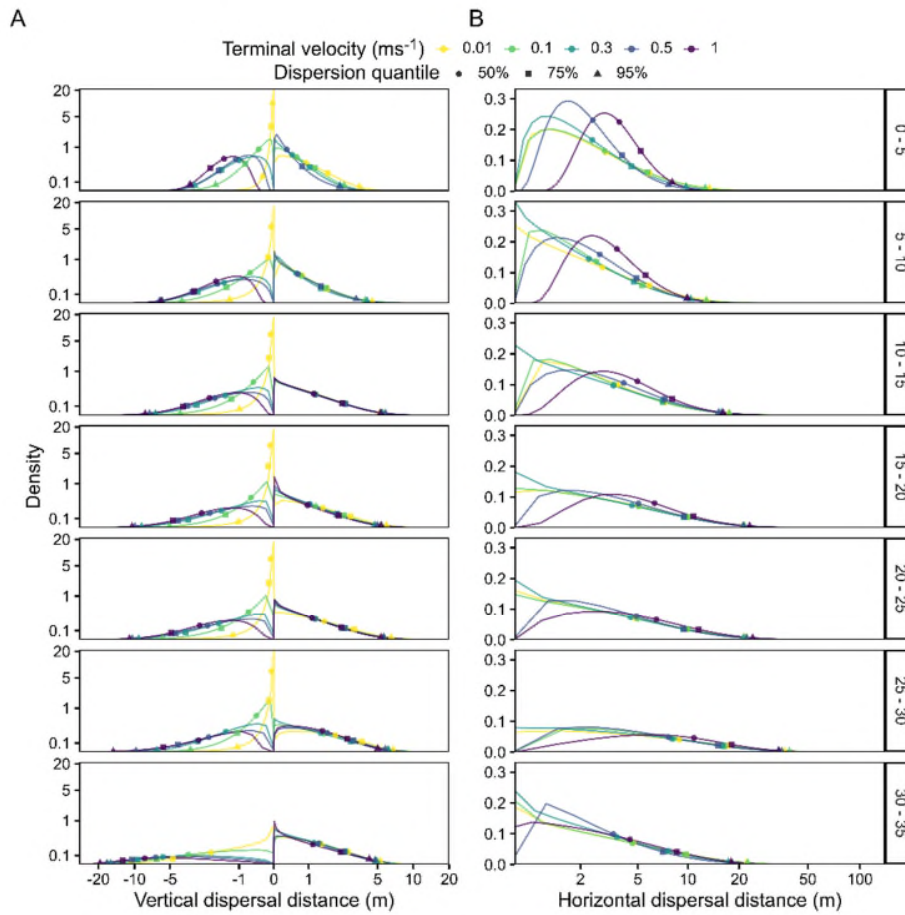


Figure S8. Fitted dispersal kernels for vertical (**A**) and horizontal (**B**) dispersal distances across terminal velocities (colours) and release height (rows) for diaspores intercepted by bark. Markers show at which distance the cumulative dispersed population reaches a given percentage. The axes have been pseudo-log transformed; note the different scales.

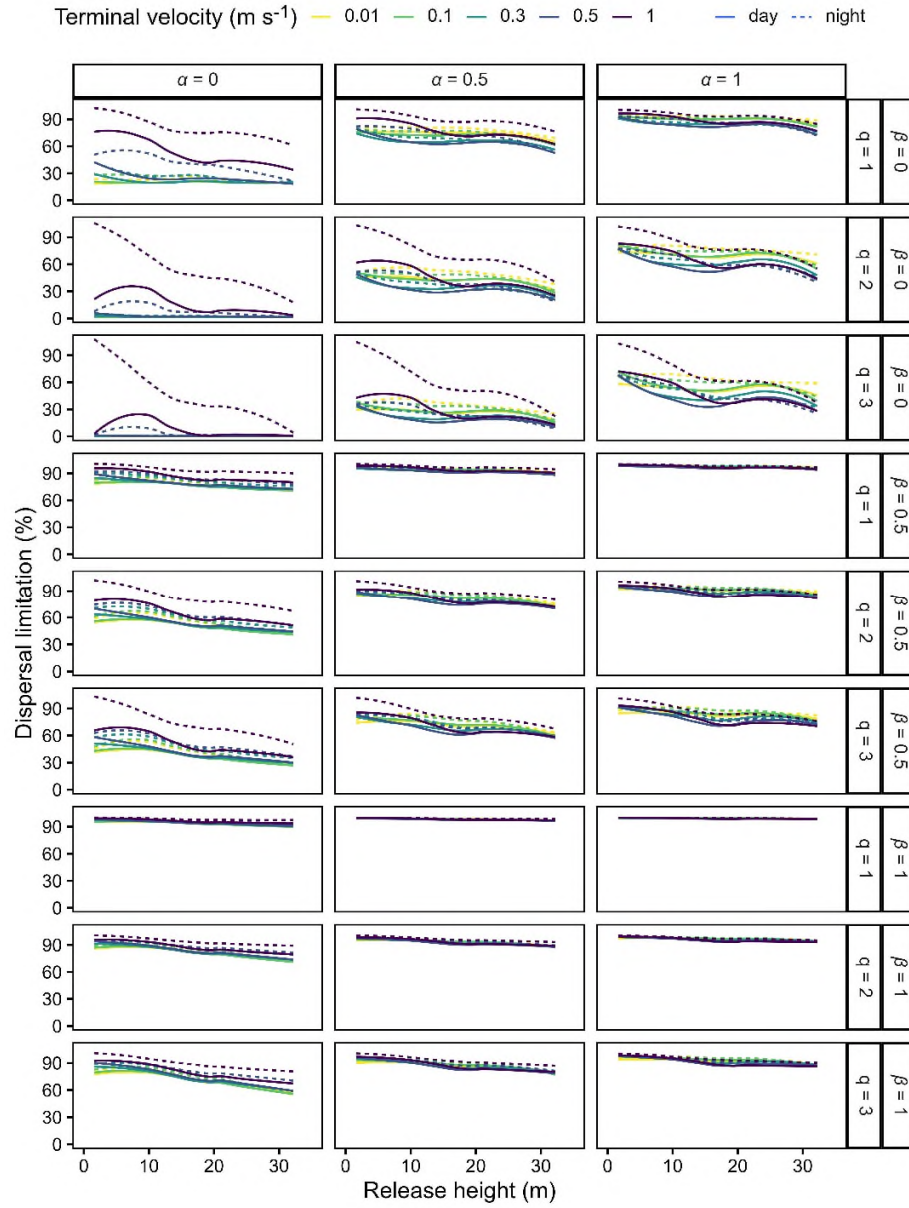


Figure S9. Sensitivity of dispersal limitation to vertical and horizontal weighting and aggregation of dispersal contributions. Dispersal limitation (%) is shown as a function of release height for combinations of vertical ($\alpha = 0, 0.5, 1$) and horizontal ($\beta = 0, 0.5, 1$) weighting, and power-mean exponent ($q = 1, 2, 3$). Colours indicate terminal velocity (m s⁻¹), and solid versus dashed lines represent daytime and nighttime conditions, respectively.

References

- Mendieta-Leiva, G., K. Wagner, and G. Zotz. 2021. "Vascular Epiphyte Census Data at the San Lorenzo Crane Plot [dataset]." *Zenodo*. <https://doi.org/10.5281/zenodo.5645775>.
- Microsoft. 2024. "Microsoft 365 tools [software]". <https://microsoft.com/microsoft-365>.
- Ramos, P., P. Villareal, and H. Muller-Landau. 2024. "Tree Height Allometry Data for 5109 Trees in Twenty-Four Small ForestGEO Plots in Panama, 2011–2020 [dataset]." *Smithsonian Tropical Research Institute*. <https://doi.org/10.25573/data.24954204.v1>.
- Vasquez V., M. Garcia, M. Hernandez, ForestGEO Smithsonian, and H. Muller-Landau. 2024. "Smithsonian ForestGEO San Lorenzo and Panama Small Plots Aerial Photogrammetry Orthomosaics, Digital Surface Models, Point Clouds and Raw Images for 2015–2024." *Smithsonian Research Data Repository*. <https://doi.org/10.60635/C33W2K>.

Ecological Monographs

Video S1 Metadata: Branch Sway for:

Patterns of wind dispersal within a forest canopy: An epiphyte perspective

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Video S1. Branch of *Dussia atropurpurea* gently swaying in the wind. Branch motion can be substantially more pronounced than shown here and may increase the apparent wind experienced by epiphytes relative to the actual wind. Even relatively weak branch sway can generate turbulence that may contribute to diaspore release, yet this remains a poorly investigated process. The video was taken by Marie Hoensbroech close to the canopy top of the San Lorenzo permanent forest plot (9.28°, -79.97°), approximately 30 m above ground.