

# Invasion impacts vary across the diel cycle: hemipterans supercharge ant ecosystem functions and restructure local invertebrate communities

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## Abstract

Biological invasions threaten biodiversity and ecosystem functions, often causing cascading effects across trophic levels. Yet how these impacts vary over the fundamental day–night cycle remains largely overlooked. On Barrow Island, a high-conservation-value reserve off northwestern Australia, we examined how infestations of the non-native scale insect *Saissetia miranda* (Hemiptera: Coccidae) on a dominant native tree species – the fig *Ficus brachypoda* – influenced native arboreal invertebrate communities and ant-mediated ecosystem functions in a natural matched-pairs experiment, comparing scale-infested versus uninfested sites. We observed the ecologically dominant native meat ant *Iridomyrmex sanguineus* collecting carbohydrate-rich honeydew from the non-native scale insects on fig trees. Although the meat ants are primarily diurnal, their nocturnal foraging activity increased by 164% at infested sites. Scale-infested sites supported higher abundances of arboreal invertebrates (~28% higher overall). In particular, scale infestations boosted the nighttime abundances of moth and cockroach species, which were observed feeding on excess honeydew, as well as spiders, possibly preying on these groups. Moreover, we uncovered knock-on effects on ant-mediated ecosystem functions in the brown and green food webs of infested sites. On the ground, the higher nocturnal ant activity contributed to a 207% increase in the mass of protein baits scavenged by ants during the night. In the canopy, there was a nearly fivefold increase in ant attacks on insect herbivores, as well as a 35% reduction in leaf damage across diel periods. Our findings show that non-native hemipterans can trigger carbohydrate-rich pulses in ecosystems, reshaping ant behaviour and amplifying both ant–herbivore interactions and higher-order effects on multitrophic communities. Critically, these impacts are not uniform across the diel cycle, but disproportionately affect nocturnal assemblages. Recognising the temporal dimension of invasion impacts is crucial for understanding and managing their full ecological consequences.

## 39 Introduction

40 A major concern with biological invasions is their far-reaching cascading effects on species  
41 and ecosystem functions, extending across multiple trophic levels (Ehrenfeld, 2010;  
42 Simberloff et al., 2013). The temporal dimensions of such impacts – exactly when they  
43 emerge, intensify and diminish – can have important implications for understanding and  
44 managing invasions. Although many species vary strongly in their behaviour, physiology, and  
45 interactions over the fundamental day–night (diel) cycle (Kronfeld-Schor & Dayan, 2003;  
46 Wong & Didham, 2024), a pervasive bias in ecological research towards daylight hours has  
47 left nocturnal communities and their contributions to ecosystem processes underexplored  
48 (Gaston, 2019). The consequences of invasive species for nocturnal biodiversity and  
49 functions may therefore be systematically underestimated. Given that many ecological  
50 interactions – such as herbivory, predation, and scavenging – vary with time of day (Cox &  
51 Gaston, 2024), invasion impacts are unlikely to be temporally uniform.

52 The sap-sucking insects of the order Hemiptera are among the most diverse and widespread  
53 invertebrate invaders, with at least 1,691 species known to have spread to non-native regions  
54 globally (Liebhold et al., 2024). Non-native hemipterans often exert disproportionate  
55 ecological impacts through mutualisms with ants which feed on the carbohydrate-rich  
56 honeydew they excrete (Styrsky & Eubanks, 2007). While such ant-hemipteran associations  
57 are indeed ‘novel’ in a coevolutionary sense (Simberloff & Von Holle, 1999), they can  
58 nevertheless have broad ecological effects. Many studies in managed systems show that such  
59 interactions not only alter ant behaviour but can also lead to cascading effects on ant–  
60 herbivore dynamics, plant health and the structure of multitrophic food webs (e.g. Powell,  
61 2009; Tena et al., 2013; Kulikowski, 2020; Anastasio et al., 2021; reviewed in Styrsky &  
62 Eubanks, 2007; Nelson & Mooney, 2022). These wide-ranging effects reflect the central role  
63 of ants in ecosystem functioning, as they regulate green food webs by dispersing seeds and  
64 suppressing herbivores, and brown food webs through scavenging and nesting activities that  
65 drive decomposition, nutrient cycling and soil turnover (Rocha et al., 2024).

66 Islands are global hotspots for biological invasions (Roy et al., 2023) and have long served as  
67 model systems for uncovering their ecological mechanisms (Simberloff & Wilson, 1969).  
68 One well-documented example of a hemipteran invasion in this context is the co-invasion of  
69 the scale insect (*Coccus celatus*) and yellow crazy ant (*Anoplolepis gracilipes*) on Christmas  
70 Island (O’Dowd & Lake, 2003). The carbohydrate subsidies provided by the honeydew of *C.*  
71 *celatus* supported exceptionally high densities of *A. gracilipes*, driving severe ecological  
72 impacts. On the forest floor, these included the decimation of native crab populations and  
73 consequent transformation of forest structure through the release of seedling recruitment and  
74 reduced litter breakdown. In the canopy, excess honeydew promoted sooty mould  
75 proliferation, extensive tree dieback, and large-scale alterations to forest composition  
76 (O’Dowd & Lake, 2003). Yet aside from this classic case study, the ecological consequences  
77 of non-native hemipterans on islands, and particularly their interactions with native ants,  
78 remain poorly understood. Conceivably, such associations may generate similarly strong but

79 context-dependent effects, especially when involving ecologically dominant native ant  
80 species.

81 Moreover, whether novel ant–hemipteran interactions disproportionately impact specific diel  
82 communities and ecosystem functions is poorly known. Many ant species show strong diel  
83 foraging patterns (Ohyama et al., 2024; Middendorp et al., 2025), which can strongly  
84 influence the nature and extent of their ecosystem functions (Houadria et al., 2016). In natural  
85 ant–hemipteran pairs, tending can modify ant diel activity, for instance leading to higher ant  
86 abundance around aphid colonies at night (Depa, 2024). While such mutualisms can shift ant  
87 diel activity, the effects of interactions with non-native hemipterans are unknown. Potentially,  
88 these novel interactions could alter the timing and magnitude of ant impacts on other species  
89 and ecosystem processes, but whether diurnal or nocturnal communities are  
90 disproportionately affected remains unclear. Understanding these temporal dynamics is  
91 crucial to assessing the ecological consequences of hemipteran invasions and their ant  
92 partners.

93 We investigated diel variation in the impacts of a recent hemipteran invasion on Barrow  
94 Island, a 23,500-ha nature reserve off the northwest coast of Australia. The island is of  
95 considerable ecological significance, supporting numerous endemic species and remaining  
96 one of the largest land masses globally without established non-indigenous vertebrates (Scott  
97 et al., 2017). The predominant habitat is spinifex grassland, with individuals of the dominant  
98 tree species – a native fig (*Ficus brachypoda*) – occurring patchily within this landscape (Fig.  
99 1). The discrete patches of fig trees are integral to the island’s biodiversity as they provide  
100 vital trophic and structural (i.e. refuge and nesting habitats) resources for numerous  
101 invertebrates and vertebrates (Lynch et al., 2019).

102 In a recent survey of 94 fig trees across Barrow Island in 2022, high densities of the non-  
103 native Mexican black scale *Saissetia miranda* (Hemiptera: Coccidae) were documented on 39  
104 trees and found to be associated with the occurrence of sooty mould (Thomas & Framenau,  
105 2023). The Mexican black scale is globally widespread, recorded from 40 countries spanning  
106 all major zoogeographic regions, including Africa, Asia, the Americas, and Europe, as well as  
107 oceanic islands such as Hawaii, Fiji, Galapagos and Cape Verde (García Morales et al.,  
108 2016). Although the ecological impacts of Mexican black scale remain poorly studied, there  
109 have been recent anecdotal reports of damage to *Ficus carica* in Iran (Moghaddam & Watson,  
110 2024). Moreover, the closely related congener *Saissetia oleae* is a major pest of citrus and  
111 olive that reduces growth and yield through still poorly understood physiological mechanisms  
112 (dos Santos et al., 2022). While no clear impacts of Mexican black scale and sooty mould on  
113 the health of fig trees on Barrow Island have yet emerged (Thomas & Framenau, 2023), the  
114 potential knock-on effects of interactions between the scale and native ants on the island’s  
115 biodiversity and ecosystem functions are unknown.

116 We focused our investigation on the effects of the scale infestation on local invertebrate  
117 communities as well as ant-mediated ecosystem functions occurring on and around the  
118 ecologically pivotal fig trees of Barrow Island. Specifically, we studied (i) the composition of  
119 the local ant community and their foraging activity, (ii) the composition of multitrophic

120 invertebrate communities in the canopy, (iii) the performance of two crucial ant-mediated  
121 ecosystem functions for the brown and green food webs: ant scavenging on insect biomass on  
122 the ground, and ant attacks on herbivores in the canopy, and (iv) leaf damage from herbivory  
123 in the canopy. We leveraged a natural matched-pairs experimental design, comparing the  
124 study parameters for scale-infested trees to adjacent uninfested controls. To test for diel  
125 variation in invasion impacts, we conducted observations during both the day and night. We  
126 hypothesised that the presence of non-native scale insects would increase nocturnal ant  
127 diversity and activity (Depa, 2024), thereby enhancing nighttime rates of ant scavenging and  
128 attacks on canopy herbivores, ultimately reducing overall leaf damage and altering the  
129 composition of nocturnal invertebrate communities.

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## 131 **Methods**

### 132 **Study period and sites**

133 The study was conducted between 20 August and 25 September 2024, coinciding with the  
134 peak activity period of Mexican black scale and associated sooty mould outbreaks (Thomas  
135 & Framenau, 2023). We employed a natural matched-pairs experimental design with the two  
136 geographic regions of Barrow Island (northern and southern) as a blocking factor. We  
137 sampled a total of 20 paired sites (40 sites in total) relatively evenly distributed across these  
138 blocks (nine site-pairs in the northern block, eleven site-pairs in the southern block; Fig. S1).  
139 We first selected an infested site by identifying a fig tree hosting feeding populations of  
140 Mexican black scale on at least one-third of its canopy. Such trees were usually  
141 straightforward to locate, as the sooty mould associated with Mexican black scale typically  
142 enveloped the canopy, giving the trees a distinctly dark and conspicuous appearance (Fig. 1).  
143 For each infested site, we then selected a paired control site comprising a fig tree of  
144 comparable stature located at least 50 m away, verifying that it had not been previously  
145 infested in 2022 or 2023 using data from Thomas and Framenau (2023). The 50 m separation  
146 was chosen to minimise overlap in local ant communities and the low overlap was confirmed  
147 during the ant survey (see below). Surveys of all study parameters (below) were conducted at  
148 all 20 site-pairs during the day (0900–1500 h), and at a subset of 10 site-pairs (five northern  
149 block, five southern block) at night (2000–0000 h). All measurements for a given focal  
150 parameter (listed below) were conducted by the same observer across the entire study, with  
151 paired infested and control sites measured within the same hour to ensure consistency and  
152 minimise observer bias.

153 Although a minimum distance of 50 m between infested and control sites was expected to  
154 minimise overlap in ant communities, colonies of the abundant meat ant *Iridomyrmex*  
155 *sanguineus* occasionally established satellite nests only 5–10 m apart. To ensure  
156 independence, we confirmed colony identity prior to baiting by reciprocally translocating  
157 meat ant workers between sites and testing for antagonistic interactions with the resident  
158 meat ants.

### 159 **Ant community composition and trophic preferences**

160 We assessed ant communities using baits representing two key trophic resources required for  
161 colony maintenance and reproduction: carbohydrates and insect protein (Feldhaar, 2014).  
162 Each bait type was presented as standardised cubes of agar jelly, with carbohydrate baits  
163 made from 15% sugar solution and protein baits from 15% cricket powder solution, each set  
164 in 0.8% agar. We used agar-based baits because they retain structural integrity under high  
165 temperatures and allow resources to be presented in consistent amounts and configurations,  
166 thereby facilitating accurate comparisons of ant trophic preferences and resource removal. We  
167 cut the baits into standardised cubes (2 mm on each side; ~0.06 g) that could be readily  
168 carried by ants.

169 At each focal tree we established five bait stations on the ground around the crown perimeter,  
170 spaced approximately 2–4 m apart. Each station consisted of two petri dishes with small side  
171 openings to permit ant entry: one dish containing 16 cubes of carbohydrate bait, the other 16  
172 cubes of protein bait (Fig. 1). Baits were exposed for 30 minutes, during which we recorded  
173 the species identities and abundances of all ants recruiting at 10-minute intervals to determine  
174 community composition as well as species' preferences for specific resources. We collected  
175 representative specimens of each species for identification using taxonomic keys (Heterick,  
176 2021) in the laboratory. Using an infrared thermometer (Matador M207), we recorded the  
177 ground surface temperature at each bait station during each survey.

#### 178 **Arboreal invertebrate community composition**

179 We sampled invertebrates in the canopy of the fig tree in each site only after the ant baiting  
180 and predation assays (below) were completed so as not to disrupt ant activity. Each site was  
181 surveyed using two complementary methods (Fig. 1). First, one researcher conducted a 10-  
182 minute manual search of the canopy, collecting invertebrates by hand into vials containing  
183 100% ethanol. Second, another researcher performed systematic sweep-netting of all  
184 branches. Specimens were bagged and frozen immediately, then stored in 100% ethanol. This  
185 combination of sampling methods ensured broad coverage of canopy invertebrates across  
186 body sizes, mobility, and conspicuousness. To maximise consistency, the same researchers  
187 performed their respective tasks at all sites. Specimens collected using both methods were  
188 combined into a single pooled sample per site for all subsequent processing and analyses. All  
189 specimens were bagged and frozen immediately, then sorted and stored in ethanol.  
190 Taxonomic identifications to order, family, species, or morphospecies level were conducted  
191 using the Barrow Island invertebrate reference collection currently housed at Murdoch  
192 University, Western Australia.

#### 193 **Ant ecosystem functions in brown and green food webs**

194 To quantify the scavenging of insect biomass by ants, we measured the mass of protein bait  
195 removed during the 30-minute period. Pilot trials confirmed that this duration was sufficient  
196 for ants to completely remove all protein baits during both day and night sampling.

197 To quantify ant predation pressure on canopy herbivores, we used model caterpillars  
198 fashioned from green moulding clay (after Roslin et al., 2017) (Fig. 1). Five caterpillars were  
199 attached to the terminal branches of each tree, spaced evenly around the canopy's

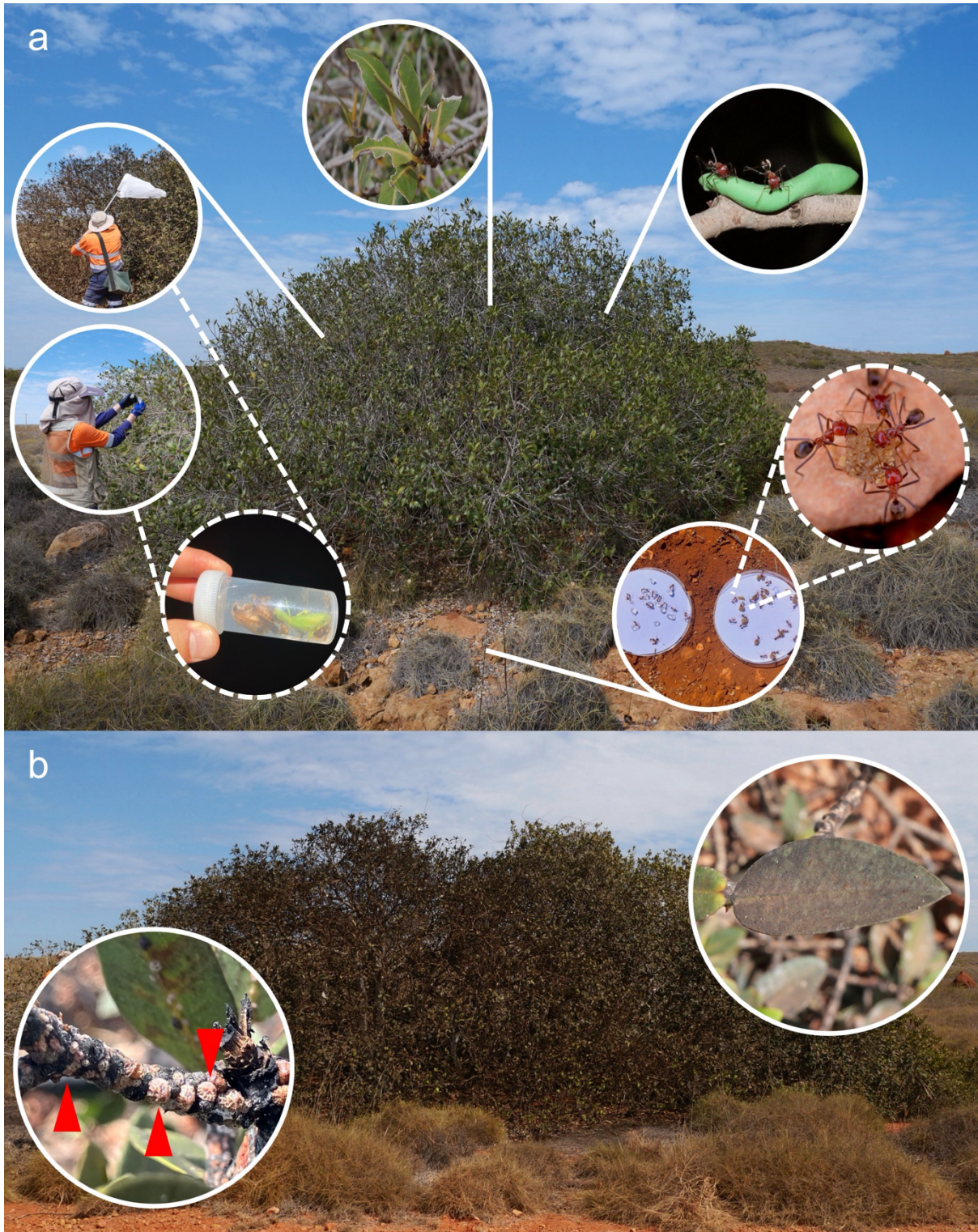
200 circumference and placed arbitrarily across both lower and upper canopy positions. After 30  
201 minutes of exposure, caterpillars were examined for signs of ant attack, typically conspicuous  
202 paired mandible impressions. We recorded the number of bite marks per caterpillar (attack  
203 severity) and the proportion of caterpillars attacked per site (attack incidence). We then  
204 multiplied the measure of incidence by the mean value of severity to obtain an incidence–  
205 severity measure of ant predation pressure at each site that was used for analysis.

## 206 **Foliar herbivory**

207 We assessed foliar herbivory on the fig tree at each site by visually surveying the entire  
208 canopy and estimating damage across leaves encountered throughout this inspection (Fig. 1).  
209 Specifically, we estimated (i) the incidence of herbivory, expressed as the percentage of  
210 leaves showing evidence of herbivore damage, and (ii) the severity of damage, quantified as  
211 the average proportion of leaf area removed on damaged leaves (after Smith et al., 2005).  
212 Rather than sampling a fixed number of leaves, the observer conducted a systematic visual  
213 scan of the canopy and integrated observations across all visible foliage to generate these  
214 whole-tree estimates. All assessments were made by the same observer across sites to ensure  
215 consistency. Unlike the other study parameters, which were assessed during both day and  
216 night, foliar herbivory was measured once per site, as this metric reflects cumulative leaf  
217 damage rather than instantaneous herbivore activity and was not expected to change  
218 appreciably within a 24-h period. We multiplied measures of incidence by severity to obtain  
219 an incidence–severity measure of foliar herbivory for analysis.

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**Figure 1.** Study system and sampling methods. (a) Illustration of sampling methods at each site, which centred on a native fig tree (photograph shows a control site with a fig tree lacking the non-native Mexican black scale). Sampling methods included (anti-clockwise from bottom right): baiting ant communities with carbohydrate and insect protein-enriched agar on the ground; assessing ant attacks on herbivore models in the canopy; assessments of foliar herbivory in the canopy; and sampling canopy invertebrates by sweep netting and direct sampling. (b) An infested site, comprising a fig tree infested with Mexican black scale (left inset: red arrows). Infested fig trees were typically enveloped in black sooty mould (right inset), giving their canopies a distinctly darker shade.

## 231 **Statistical analysis**

### 232 ***Ant community and foraging patterns of dominant ants***

233 We tested whether ant species richness (the total number of ant species recorded at baits in  
234 each site during a particular diel period) was significantly influenced by infestation status (the  
235 presence of Mexican black scale on the fig tree), the specific diel period, or an interaction  
236 between the two factors using a Poisson generalised linear mixed effects model (GLMM)  
237 with a logit link. The model included fixed effects for infestation status (infested vs. control),  
238 diel period (day vs. night), and their interaction. Random intercepts for site-pair (20 in total)  
239 and region (northern vs. southern) were included to account for repeated measures and spatial  
240 variation.

241 We also tested the effects of infestation status and diel period on the foraging activity and  
242 trophic preferences of the dominant ant species, the meat ant *I. sanguineus* using generalised  
243 linear mixed models. Foraging activity was quantified as bait occupancy by the meat ants. At  
244 each site and sampling period, five bait stations were deployed, and occupancy was modelled  
245 as a binomial response (number of baits occupied vs. unoccupied) with a logit link. Fixed  
246 effects included infestation status, diel period, and their interaction. Site-pair and region were  
247 included as random intercepts. Trophic preferences were assessed by analysing meat ant  
248 recruitment to carbohydrate versus protein resources. The response variable was the mean  
249 number of meat ant workers observed across all baits of a given resource type in each site  
250 during each baiting session. Recruitment was modelled using a Gamma error distribution  
251 with a log link. Fixed effects included infestation status, resource type, diel period, and all  
252 interactions. Site-pair and region were included as random intercepts.

### 253 ***Arboreal invertebrate community richness and composition***

254 We tested whether invertebrate richness (total number of orders, families, or species) and  
255 abundance recorded at each site during a given diel period was influenced by infestation  
256 status, diel period, or their interaction. Separate Poisson GLMMs were fitted for order-,  
257 family-, species-level richness, and abundance, with infestation status, diel period, and their  
258 interaction as fixed effects, and site-pair and region as random intercepts. We also tested  
259 whether invertebrate community composition differed between control and infested sites  
260 using permutational multivariate analysis of variance (PERMANOVA) based on Bray–Curtis  
261 dissimilarities calculated from site-level community matrices of taxonomic order abundances.  
262 Infestation status (control vs. infested), diel period (day vs. night), and their interaction were  
263 included as fixed factors, with site-pair included as a blocking factor in the model. A  
264 SIMPER analysis was used to identify the taxa that contributed most to compositional  
265 dissimilarity. Patterns in community composition were visualised using non-metric  
266 multidimensional scaling (NMDS) ordinations derived from the same dissimilarity matrix.

### 267 ***Ecosystem functions in brown and green food webs***

268 We tested the effects of infestation status and diel period on ant-mediated scavenging in the  
269 brown food web and on the incidence–severity of ant attacks on canopy herbivores in the

270 green food web. We also examined whether the incidence–severity of foliar herbivory on fig  
271 trees varied with infestation status.

272 Scavenging by ants was quantified as the proportion of protein baits removed at each site and  
273 analysed using a generalised linear mixed-effects model with a binomial error distribution  
274 and logit link. The response was specified as the number of protein baits removed versus not  
275 removed, with infestation status, diel period, and their interaction as fixed effects, and site-  
276 pair and region as random intercepts. Ant scavenging was quantified using protein bait  
277 removal because protein foraging by ants is more closely associated with predatory and  
278 scavenging behaviour, whereas carbohydrate foraging primarily reflects exploitation of sugar  
279 resources such as honeydew and would be strongly confounded by the presence of scale  
280 insects (Davidson, 1997). The incidence–severity of ant attacks on canopy herbivores was  
281 analysed using a generalised linear mixed model with a Tweedie error distribution and log  
282 link, with the same fixed and random effects. The incidence–severity of foliar herbivory was  
283 log transformed for normality and analysed using a linear mixed-effects model with  
284 infestation status as a fixed effect and site-pair and region as random intercepts.

285 All data were analysed in R version 4.3.0 (R Core Team, 2023). Linear and generalised linear  
286 mixed models were built using *lme4* (Bates et al., 2015) and *glmmTMB* (McGillucuddy et al.,  
287 2025), respectively. Model estimates were back-transformed to the original scale to aid  
288 interpretation and reported as multiplicative changes (i.e., percent differences) in the results.  
289 PERMANOVA and SIMPER analyses were conducted using *vegan* (Oksanen et al., 2013).

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## 291 **Results**

### 292 **Ant community and foraging patterns of dominant ants**

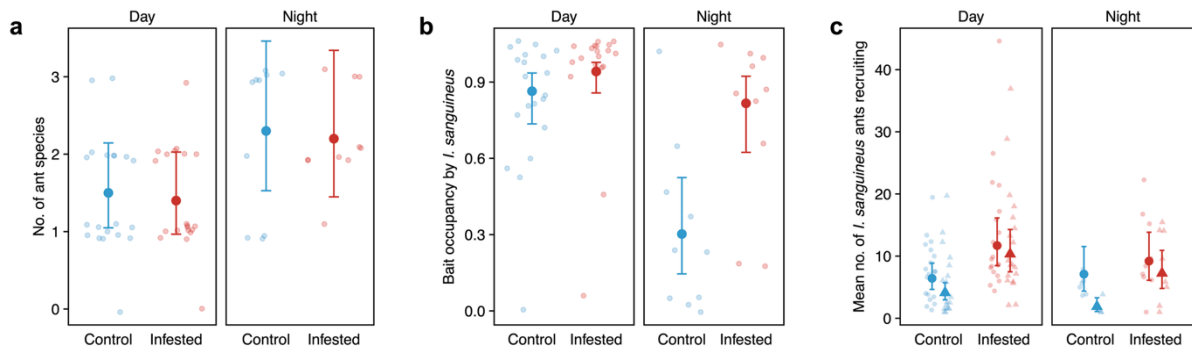
293 A total of 16 ant species representing five subfamilies were collected (Table S1). Ant species  
294 richness did not vary significantly between scale-infested and control sites, nor between diel  
295 periods (Fig. 2a; Table S2). Among the ant species, the meat ant *I. sanguineus* was  
296 numerically dominant, occurring at all sites and comprising a mean of 95.7% ( $\pm$  10.9 SD) of  
297 all individuals recorded at baits per site.

298 Foraging activity of the meat ant (measured as the proportion of baits occupied) was  
299 significantly influenced by diel period, declining sharply at night ( $\beta=-2.69$ ,  $P<0.001$ ).  
300 However, at scale-infested sites, this decline was mitigated, as indicated by a significant  
301 interaction between infestation status and diel period ( $\beta=1.40$ ,  $P=0.04$ ). At control sites, meat  
302 ant foraging declined by 66% from day (mean bait occupancy rate=0.83) to night  
303 (mean=0.28). In contrast, nocturnal foraging activity was higher at scale-infested sites  
304 (mean=0.74), representing a 164% increase relative to control sites (Fig. 2b).

305 The trophic preference of the meat ant was influenced by infestation status and diel period.  
306 Recruitment, measured as the mean number of workers recruiting over the baiting session,  
307 was significantly lower to protein resources than to carbohydrate resources ( $\beta=-0.44$ ,  
308  $P=0.002$ ), while scale infestation increased overall recruitment ( $\beta=0.60$ ,  $P=0.01$ ).

309 Recruitment to protein resources declined at night relative to the day, as indicated by a  
 310 negative interaction between resource type and diel period ( $\beta=-0.88$ ,  $P=0.008$ ). However, the  
 311 nocturnal reduction in recruitment to protein was moderated by infestation status, as reflected  
 312 by a positive three-way interaction between infestation status, resource type and diel period  
 313 ( $\beta=0.76$ ,  $P=0.06$ ) (Table S3). Specifically, at protein resources during the night, mean  
 314 recruitment increased from 1.8 workers at control sites to 8.2 workers at scale-infested sites  
 315 (Fig. 2c).

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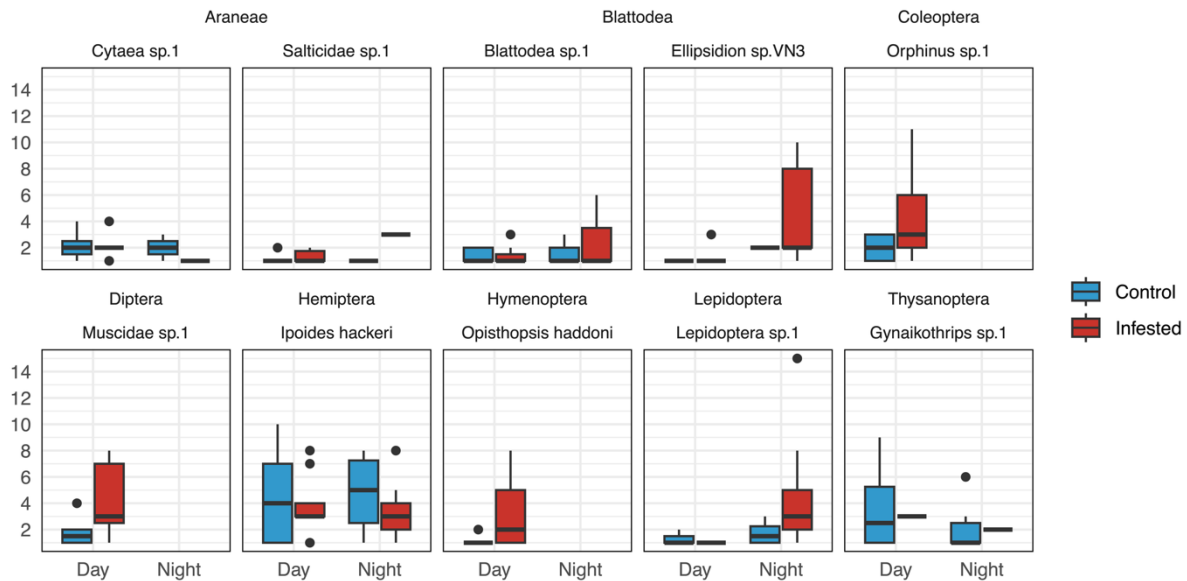
318 **Figure 2.** Effects of scale infestation and diel period on ants on Barrow Island. (a): Ant species richness neither  
 319 varied significantly with infestation status nor with diel period. (b): Bait occupancy of the dominant meat ant *I.*  
 320 *sanguineus* was significantly higher at infested sites at night. (c): The mean number of meat ant workers  
 321 recruiting to protein resources (triangles) was consistently lower than that to carbohydrate resources (circles),  
 322 but nocturnal recruitment to protein resources was higher at infested sites. In all plots, points indicate predicted  
 323 mean values and error bars indicate 95% confidence intervals (CIs).

324

### 325 *Arboreal invertebrate community richness and composition*

326 A total of 857 individual invertebrates were collected. These comprised 104 species from 70  
 327 families in 12 taxonomic orders, including numerous spiders (Araneae), cockroaches  
 328 (Blattodea), flies (Diptera), wasps (Hymenoptera), moths (Lepidoptera), crickets (Orthoptera)  
 329 and other invertebrates (Table S4). While there were no significant effects of infestation  
 330 status, diel period, or their interaction on the richness of invertebrate orders, families, or  
 331 species at a site (Table S2), a GLMM showed that invertebrate abundance was 28% higher at  
 332 scale-infested sites ( $B=0.25$ ,  $P=0.003$ ). Additionally, a PERMANOVA model showed that the  
 333 composition of the invertebrate community in a site (i.e. the relative abundances of  
 334 individuals of different species) was significantly influenced by infestation status ( $F_{1,56}=2.76$ ,  
 335  $R^2=0.04$ ,  $P=0.002$ ) and diel period ( $F_{1,56}=4.33$ ,  $R^2=0.07$ ,  $P=0.001$ ). Collectively, diel period  
 336 and infestation status explained 11% of the compositional dissimilarity among invertebrate  
 337 communities, as reflected by the overlap observed in the NMDS ordination (Fig. S2).  
 338 Compositional dissimilarity was primarily driven by changes in the abundances of a subset of  
 339 invertebrate species. As indicated by SIMPER analysis, 54% of the average dissimilarity  
 340 between scale-infested and control invertebrate communities was attributable to differences  
 341 in the abundances of 10 species representing eight orders (Fig. 3). In particular, one spider  
 342 (Araneae), two cockroaches (Blattodea), and one moth (Lepidoptera) species were more

343 abundant at scale-infested sites at night, whereas one fly (Diptera), ant (Hymenoptera), and  
 344 beetle (Coleoptera) species were more abundant at scale-infested sites during the day (Fig. 3).  
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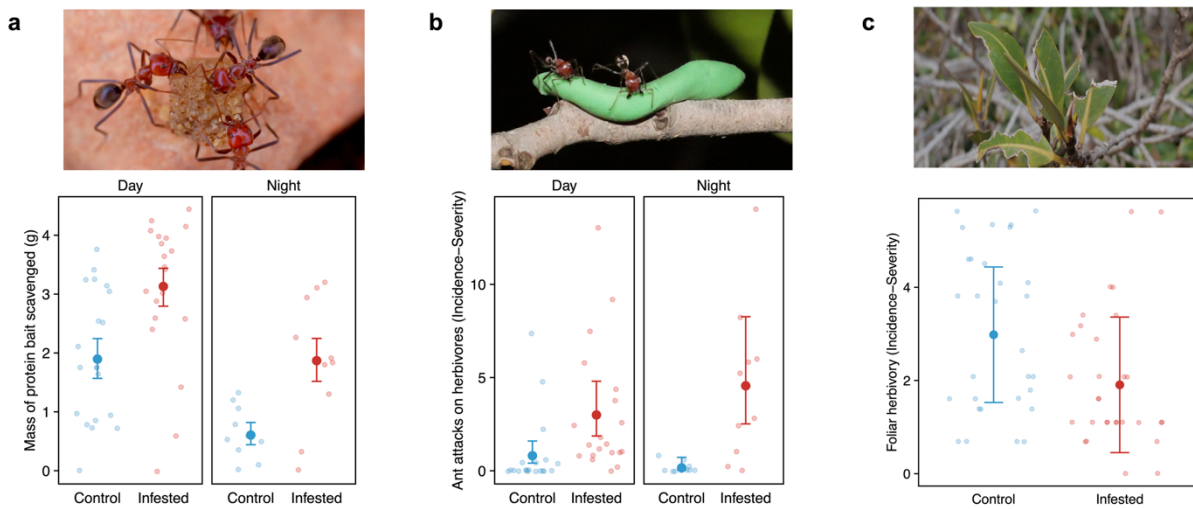
346  
 347 **Figure 3.** Plots showing sample-level species abundances, by scale-infestation status and diel period, for the 10  
 348 invertebrate species (from eight orders) that together accounted for approximately 54% of the average  
 349 dissimilarity between control and scale-infested sites based on SIMPER analysis. Black bars in boxplots indicate  
 350 median values and boxes represent interquartile range.

351  
 352 **Ecosystem functions in brown and green food webs**

353 Ant scavenging was significantly influenced by both diel period and infestation status.  
 354 Overall, a shift from day to night had a significant negative effect on scavenging ( $\beta=-1.51$ ,  $P$   
 355  $< 0.001$ ), while scale infestation had a significant positive effect ( $\beta=1.05$ ,  $P < 0.001$ ), with a  
 356 smaller positive interaction term ( $\beta=0.43$ ,  $P=0.004$ ) (Table S2). Consequently, scale  
 357 infestation increased the mean mass of protein bait removed by ants, with similar absolute  
 358 increases in mass removed during day and night (day: 3.13 g vs 1.90 g in controls; night: 1.87  
 359 g vs 0.61 g in controls), but substantially stronger proportional effects at night (day: +65%;  
 360 night: +207%).

361 Scale infestation also had a significant positive effect on the incidence–severity of ant attacks  
 362 on canopy herbivores ( $\beta=1.31$ ,  $P=0.002$ ). The incidence–severity of ant attacks in infested  
 363 sites was nearly fivefold higher (mean=3.52) than in control sites (mean=0.59). Attacks did  
 364 not vary significantly with diel period (Fig. 4b).

365 Scale infestation had a significant negative effect on the incidence–severity of foliar  
 366 herbivory on the fig trees ( $\beta=-1.08$ ,  $P=0.003$ ). The incidence–severity of foliar herbivory on  
 367 fig trees in infested sites (mean=2.0) was 35% lower than in control sites (mean=3.07) (Fig.  
 368 4c).



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371 **Figure 4.** Effects of *S. miranda* infestation status and diel period on ecosystem functions on Barrow Island. (a):  
 372 The mass of protein baits removed by scavenging ants was significantly higher during the day as well as at  
 373 infested sites. (b): The incidence–severity (proportion of model caterpillars attacked  $\times$  mean bite marks per  
 374 caterpillar) of ant attacks on canopy herbivores was significantly higher at infested sites. (c): The incidence–  
 375 severity (percentage of leaves damaged  $\times$  mean proportion of leaf area removed per damaged leaf) of herbivory  
 376 on leaves of fig trees was marginally lower at infested sites. In all plots, points indicate predicted mean values  
 377 and error bars indicate 95% confidence intervals (CIs).

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## 379 Discussion

380 Despite being widespread globally, the ecological consequences of hemipteran invasions  
 381 remain poorly resolved outside of a handful of case studies. Our findings demonstrate that  
 382 invasions by non-native hemipterans on islands can trigger short-term cascading effects that  
 383 influence multitrophic interactions and ecosystem processes across both brown and green  
 384 food webs. These effects were associated with the carbohydrate-rich honeydew produced by  
 385 the non-native Mexican black scale *S. miranda*, which supported strong nocturnal activity of  
 386 the dominant native meat ant *I. sanguineus* and altered its trophic foraging. The resulting  
 387 changes in ant activity had temporally variable impacts across the diel cycle, with  
 388 disproportionate consequences for nocturnal assemblages, including shifts in the abundance  
 389 of multiple canopy invertebrate taxa.

### 390 Carbohydrate subsidies amplify ant foraging and alter trophic preferences

391 The strongest effect of scale infestations was the creation of a carbohydrate subsidy in the  
 392 form of honeydew. Carbohydrates are critical for fuelling ant activity, and honeydew  
 393 provision is known to stimulate higher recruitment and prolonged foraging in both native and  
 394 invasive ants (Styrsky & Eubanks, 2007; Nelson & Mooney, 2022). In our study, honeydew  
 395 presence coincided with sustained nocturnal activity of the meat ant, a period when foraging  
 396 otherwise declined sharply at control sites.

397 Davidson (1997) proposed that surplus carbohydrates in ant colonies can be allocated to  
398 activities that enhance acquisition of limiting protein resources. Our results are consistent  
399 with this hypothesis: at infested sites, meat ant colonies allocated a greater share of foragers  
400 to protein resources at night. This suggests that carbohydrate subsidies did not only attract  
401 workers to honeydew itself, but also enhanced overall colony capacity for acquiring limiting  
402 nitrogenous nutrients necessary for brood development and colony growth (Feldhaar, 2014).  
403 In other words, access to abundant honeydew may have allowed colonies to redirect foraging  
404 effort toward protein resources, illustrating a mechanistic link between carbohydrate  
405 subsidies and altered trophic allocation.

#### 406 **Cascading effects onto ecosystem functions in brown and green food webs**

407 The behavioural changes of meat ants had clear consequences for ant-mediated ecosystem  
408 functions. In the brown food web, enhanced nocturnal activity translated into a 207%  
409 increase in ant scavenging of insect protein baits, offsetting the typical nocturnal decline in  
410 scavenging observed at control sites. This suggests that honeydew subsidies can not only  
411 increase total rates of nutrient recycling via scavenging but also redistribute them across diel  
412 periods. In the green food web, ant attacks on model herbivores were more than fourfold  
413 higher at infested sites, and leaf damage was reduced, consistent with strengthened top-down  
414 control by ants. While the herbivory reduction was not statistically significant, the direction  
415 of the effect aligns with previous studies demonstrating honeydew-fuelled increases in ant  
416 protection of plants (Tena et al., 2013; Anastasio et al., 2021). Taken together, these findings  
417 show that hemipteran–ant associations can amplify both brown and green food web  
418 processes, with consequences for plant–herbivore dynamics and energy flows through detrital  
419 pathways.

#### 420 **Effects on nocturnal multitrophic communities**

421 Beyond direct effects on ants and their functions, scale infestations also restructured arboreal  
422 invertebrate communities, but these changes were concentrated at night. Moths (Lepidoptera)  
423 and cockroaches (Blattodea) were more abundant on infested trees during nocturnal surveys,  
424 with spiders (Araneae) likewise increasing, possibly reflecting greater prey availability. One  
425 possible explanation is that reduced ant activity in the canopy at night freed up honeydew for  
426 other insects, or lowered the risk of ant aggression. However, our assays indicated that ant  
427 attack rates on model herbivores in the canopy remained consistently high during both day  
428 and night on infested trees (Fig. 4b). This suggested that ant pressure did not diminish  
429 sufficiently at night to account for the observed shifts.

430 Instead, the predominance of nocturnal responses is more likely to reflect the intersection of  
431 two processes: intrinsic diel activity patterns of canopy invertebrates and temporal variation  
432 in resource availability. Many insects are adapted to exploit cooler nighttime conditions and  
433 reduced exposure to vertebrate predators such as birds (Wong & Didham, 2024), which likely  
434 concentrates activity at night. At the same time, honeydew subsidies generated by scale  
435 insects may not be temporally constant. Plant physiological processes and phloem quality can  
436 vary across the diel cycle, potentially influencing sap-feeding activity and honeydew  
437 excretion rates. If honeydew availability is elevated at night, this would further amplify

438 nocturnal aggregation and feeding by canopy invertebrates. However, we did not quantify  
439 diel variation in honeydew production by *S. miranda*, and resolving whether subsidy supply  
440 itself varies over the diel cycle represents an important avenue for future work. Together,  
441 these mechanisms suggest that non-native hemipterans not only introduce novel carbohydrate  
442 subsidies into ecosystems, but also modulate when those subsidies are most strongly  
443 expressed and utilised.

#### 444 **Broader implications and future research**

445 One limitation of our study was the relatively brief observation window within a single  
446 season. Hemipteran populations often fluctuate with plant phenology, climatic conditions,  
447 and their own life cycles (Bodino et al., 2019; Santos et al., 2019), which may in turn  
448 modulate their levels of honeydew production and ecological impacts. Future research should  
449 therefore examine invasion impacts across multiple seasons and years, testing whether the  
450 strength and timing of honeydew-mediated subsidies vary not only over the diel cycle but  
451 also across seasonal and interannual scales. Such work would provide a more complete  
452 picture of how hemipteran invasions interact with temporal dynamics to shape communities  
453 and ecosystem functions. Another key direction for future research would be to quantify  
454 honeydew production and composition by the non-native scale insect and experimentally test  
455 its role in shaping ant–hemipteran associations and downstream community impacts, thereby  
456 establishing a mechanistic link between honeydew availability and observed trophic  
457 restructuring.

458 In summary, our results highlight that the ecological impacts of non-native hemipterans are  
459 strongly context-dependent, emerging not only from novel associations with ants but also  
460 from the temporal structuring of ecological processes. By disproportionately altering  
461 nocturnal interactions and communities, non-native scale insects reshaped an ecological  
462 domain typically underrepresented in invasion studies. Critically, if our surveys had been  
463 restricted to daylight hours, the strongest community-level effects would not have been  
464 detected. This underscores the need to integrate nocturnal perspectives into invasion ecology  
465 more broadly.

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