

1 **A global evidence synthesis of outcomes of urban bird conservation**
2 **interventions**

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28 **Data Availability:** All primary data and code is available at the Open Science Framework
29 repository for this project:

30 https://osf.io/v6rwz/overview?view_only=8b459b888409479b8077d5ed80cc4ec5

31

32 **ABSTRACT**

33 Avian populations face high concentrations of threats in urban areas. Understanding the
34 outcomes of urban conservation interventions to tackle these threats could inform more effective
35 evidenced-based approaches. Using an evidence synthesis and meta-analysis of peer-reviewed
36 literature on conservation interventions to tackle four leading causes of urban bird mortality in
37 urban areas (cat predation, bird-window collisions, road mortalities, light pollution, and bird-
38 vehicle collisions), we: 1) summarized the current state of the literature on urban threats and
39 conservation interventions for birds, and 2) evaluated the outcomes of interventions aimed at
40 reducing impacts of major urban threats on birds. Our evidence synthesis found that across all
41 threats, few studies (35 out of 522) explored outcomes of conservation interventions compared to
42 those that determined the establishment, magnitude, and drivers of threats. Our meta-analysis of
43 nine types of collar-based interventions to reduce urban cat predation found that rainbow-colored
44 Birdsbesafe™ collars most consistently demonstrated evidence of reduced numbers of dead birds
45 brought home by cats (pooled reduction of 54% compared to controls). However, these results

46 were drawn from a very small number of studies (n=13). To reduce mortality from bird-window
47 collisions, closely spaced vertical stripes and circular or square decals reduced collisions by 98%
48 (n=6) There were no studies that explored outcomes of interventions to reduce light pollution in
49 urban areas. Two studies explored the use of traffic control measures (speed bumps, chicanes and
50 barrier poles) to reduce bird-vehicle collisions, although we could not draw conclusions on their
51 outcomes. Overall findings suggest there is a need for greater research on the effectiveness of
52 conservation interventions for all urban threats.

53 *Keywords:* artificial light at night, bird mortality, bird-vehicle collision, bird-window collision,
54 cat predation, conservation interventions, evidence synthesis, urban birds

55 **LAY SUMMARY**

- 56 • We reviewed research on four key urban threats facing birds: bird-window collisions, cat
57 predation, light pollution, and bird-vehicle collisions
- 58 • We found that very little research is dedicated to understanding outcomes of interventions
59 compared to understanding the magnitude of threats
- 60 • Most studies on interventions to reduce cat predation looked at outcomes of cat collars,
61 with Rainbow Birdsbesafe™ collars demonstrating the greatest reduction in bird deaths
- 62 • Window treatments with high coverage, such as vertical stripes and circular or square
63 decals resulted in the greatest reduction of bird-window collisions
- 64 • More research is needed to understand the outcomes of bird conservation interventions,
65 particularly for light pollution and urban bird-vehicle collisions, as well as on outcomes
66 of feral cat removal and roaming bylaws for owned cats

67 INTRODUCTION

68 The rapid expansion of urban landscapes is a major contributor to biodiversity loss
69 worldwide (McKinney, 2006). Urban development typically reduces vegetation, simplifies
70 habitat, and concentrates threats, resulting in the intolerance of many bird species to urbanized
71 landscapes (Marjakangas et al., 2024). Many cities lie along migratory flyways, and habitat for
72 birds at risk of extinction overlap with highly urbanized landscapes (Jokimäki et al., 2018).
73 Given the critical roles birds play in ecosystems (Wenny et al., 2011) along with their value as
74 effective biodiversity indicators (Gregory & Strien, 2010), bird conservation in urban areas can
75 help cities foster broader biodiversity.

76 Among the most severe anthropogenic threats facing birds in urban areas are predation by
77 cats, collisions with windows, disorientation by light pollution, and bird-vehicle collisions
78 (Marzluff, 2001). It is estimated that feral and owned outdoor cats kill 70-200 million birds
79 annually in Canada alone, making them the leading source of mortality after habitat conversion
80 (Calvert et al., 2013). Collisions with windows kill another estimated 16-42 million birds
81 annually (Calvert et al., 2013). Birds cannot perceive vegetation in clear or reflective glass,
82 resulting in fatal collisions (Klem, 2009). Light pollution causes fewer direct mortalities but
83 influences birds' photoperiod cycles and disorients birds that migrate nocturnally, increasing the
84 risk of window collisions and other urban sources of mortality. As a result, large numbers of
85 birds have been found dead or injured near brightly lit buildings during migration seasons (Van
86 Doren et al., 2021). An estimated 96 million birds are killed by vehicles in the United States
87 (Rabie et al., 2024), where species that forage near roadsides are particularly vulnerable
88 (Orłowski, 2008).

89 Interventions to tackle these threats have been proposed and, in some cases, tested. For
90 example, the application of decals can result in birds being able to perceive and avoid colliding
91 with windows (Klem, 1990). A study in Delta, British Columbia, found that application of
92 Feather Friendly® dot decals at 2-inch intervals resulted in a 95% reduction in window-
93 collisions post-treatment (De Groot et al., 2022). Other examples include specialized cat collars
94 to reduce cat predation (Willson et al., 2015) and using downward-facing lighting designs and
95 different spectra of light to reduce the effects of light pollution (Poot et al., 2008). However, a
96 comprehensive understanding of the effectiveness of different interventions across different
97 contexts is lacking.

98 In general, knowledge of the effectiveness and outcomes of proposed conservation
99 interventions remains limited globally across taxa. For example, Binley et al. (2025) found a
100 scarcity of research on outcomes of interventions for species at risk of extinction in Canada.
101 Williams et al. (2020) found a similar pattern in a review of conservation science literature,
102 where nearly half of all studies described the state of nature, 10% linked a mechanism to the
103 driver of change, and very few tested the effectiveness of an intervention.

104 As global urban land cover is expected to increase threefold from 2000 to 2030 (Fragkias
105 et al., 2013) and continue expanding thereafter, it is essential to understand and implement
106 conservation interventions to reduce bird mortalities in cities. A number of programs address
107 bird conservation in cities, such as the American Bird Conservation and Environment for the
108 Americas' Bird City Network (American Bird Conservancy, 2025) and Nature Canada's Bird
109 Friendly Cities program. These programs encourage cities to implement policies and initiatives
110 that support bird conservation (Nature Canada, 2025). The objectives of these programs are to
111 reduce bird mortalities in cities, towns and villages and build bird conservation awareness by

112 advocating for nature-positive municipal policies, implementing restoration projects of local bird
113 habitat, addressing and mitigating the leading causes of urban bird mortalities, and hosting
114 events that connect and educate community members. Certification criteria are based on how
115 municipalities propose to tackle known major threats to urban birds (Nature Canada, 2025). An
116 evidence-based understanding of the outcomes of conservation interventions for birds could help
117 guide bird friendly city programs in promoting the most effective interventions.

118 To this end, we conducted an evidence synthesis to explore outcomes of interventions to
119 address the top causes of direct bird mortality of birds in cities: cat predation, bird-window
120 collisions, light pollution, and bird-vehicle collisions. The primary objective of the review was to
121 synthesize available evidence on the outcomes of bird conservation interventions that address
122 key threats in urban areas to provide evidence for bird-friendly city certification programs.
123 Specifically, we: 1) conducted a narrative synthesis of the state of the literature on key urban
124 threats and interventions for birds and 2) conducted a meta-analysis of outcomes of interventions
125 addressing each threat, including reductions in direct bird mortality, improvements to
126 reproductive success, changes in physiological or behavioral stress indicators, or any other bird-
127 related conservation outcomes. We discuss how an understanding of the effectiveness of
128 interventions can inform urban programs that aim to improve bird conservation.

129 **METHODS**

130 Our evidence synthesis followed, as closely as possible, the Collaboration for
131 Environmental Evidence's (CEE) standardized guidelines for systematic maps (CEE, 2022). All
132 full-text screening and data extraction were conducted by two reviewers following a double-
133 blind process (i.e., reviewers did not see each other's responses until a conflict arose), with
134 conflicts resolved by discussion and consultation with the principal investigator. We used

135 separate search strings and conducted independent screenings for each threat category: cat
136 predation, bird-window collisions, light pollution, and bird-vehicle collisions.

137 **Search for articles**

138 *Search sources*

139 The Bird Friendly Cities program National Advisory Panel (NAP), made up of 10 experts
140 in bird conservation from across Canada, were consulted early in the review process (Nature
141 Canada, 2025). The NAP helped to identify target studies (i.e., De Groot et al., 2022; Gordon et
142 al., 2010; Hall et al., 2015; Klem, 1990; Poot et al., 2008; Zuberogoitia et al., 2015) and provided
143 recommendations for keywords used in search strings. Following consultation with librarians at
144 Carleton University, we selected Web of Science Core Collections and Scopus as relevant
145 databases for our literature search, as they are more repeatable databases. Moreover, we ran a
146 preliminary test on 250 studies in Google Scholar and found that including Google Scholar
147 citations in the review search stage did not produce any additional relevant citations. Using
148 citations of excluded review articles (see “Eligibility Criteria”), we performed supplemental
149 searching to capture studies that did not come up in our initial search.

150 *Search strings*

151 For each threat category, we used search strings that targeted the title, abstract and
152 keywords within relevant studies (“Topic” in Web of Science, or “TITLE-ABS-KEY” in Scopus;
153 Supplementary Material Table S1). A list of target studies was used to develop and test the
154 comprehensiveness of the search strings (listed here: Barton et al., 2017; Geiger et al., 2022;
155 Groot et al., 2022; Poot et al., 2008; Willson et al., 2015; Zuberogoitia et al., 2015) and
156 respective population, intervention, comparator, and outcome (PE/ICO) elements (see
157 Supplementary Material Table S2 for full search strings). Search string elements were further

158 refined through multiple rounds of preliminary abstract screening under each threat category.
159 Search strings did not necessarily contain all PE/ICO elements to remain inclusive of all relevant
160 studies. For example, under the cat predation threat topic, it may have been relevant to include
161 the term “urban” or related keywords to keep the studies within an urban setting, or “predation”
162 or “mortality” to specify exposure or outcome. However, as the target studies did not include the
163 words “urban”, “predation”, or “mortality” within their title, abstract, or keywords, specifying
164 these terms in the search string would have excluded target studies. We instead used the terms
165 “domestic”, “pet”, “feral” and “*Felis catus*”. We also excluded the terms “influenza” or “virus”
166 or “disease” or “flu”, as these studies fell out of the scope of our objectives. Similarly, the term
167 “aircraft” was excluded from window collisions studies to focus on building collisions and
168 “poultry” was excluded from light pollution studies to focus on the effects of artificial light on
169 wild birds.

170 **Article screening**

171 For eligibility screening, studies were uploaded from Web of Science and Scopus onto
172 Covidence (Veritas Health Innovation, Melbourne, Australia), a web-based systematic review
173 software. Duplicates between databases were removed automatically by the software; missed
174 duplicates were removed manually during the screening process. Each of the four threat
175 categories was screened in Covidence as an independent review. For abstract screening, a kappa
176 test was performed in which both reviewers each completed the same 30% (a random subset) of
177 abstracts under each threat category and achieved a score representing a good strength of
178 agreement ($\kappa = 0.7$, where $\kappa > 0.6$ is considered good strength of agreement) (Sim & Wright,
179 2005). Following the kappa test, abstracts were screened independently. Each article was judged

180 for eligibility based on its abstract, and when reviewers were uncertain, articles were included for
181 full-text screening.

182 *Eligibility criteria*

183 Following abstract screening, the full-text screening stage was completed in full by two
184 review authors (double-blind screening). All conflicts were resolved by discussion, with
185 unsolved conflicts resolved by the principal investigator. During both screening stages (abstract
186 and full-text), studies were excluded if they were not peer reviewed, not available in English (as
187 it was beyond project capacity to review studies not in English), or not related to the (PE/ICO)
188 elements defined for each topic (Supplementary Material Table S2 & S3). No studies were
189 excluded based on the presence or absence of a comparator. Present, historical, and even
190 archaeological evidence (e.g., Onar et al., 2021) were all considered eligible, provided that they
191 presented new data or findings not previously published in peer-reviewed literature. Otherwise,
192 they were excluded as “non-primary research” and used for supplemental searching. These were
193 typically review papers and perspective pieces. There were no geographical or time-based limits
194 on publications for the review. We reached out to the corresponding author for studies we could
195 not access. Studies were excluded as “inaccessible” if we did not hear from the author within 14
196 days. See Supplementary Material Table S3 for detailed exclusion criteria.

197 *Eligible setting*

198 To date, there is no well-established global standard definition for the term “urban”.
199 Thus, we followed the definition for “urban” outlined by the country in which each study
200 occurred, using the most up-to-date list collated by the United Nations Statistics Division (United
201 Nations Statistics Division, 2005). University campuses were considered eligible even when
202 outside of strict urban boundaries, as they typically emulate an urban environment. We also

203 included experiments completed in controlled or non-urban settings if the goal of the study was
204 to understand outcomes in an urban environment (particularly important for studies on light
205 pollution). Studies that contained a gradient or interface between urban, rural, and natural
206 habitats were included.

207 **Data extraction**

208 *Defining threat and intervention categories*

209 Following full-text screening, we separated studies into threat and intervention studies.
210 Drawing from the framework created by (Williams et al., 2020), threat studies were assigned
211 three categories: a) studies that incidentally discover or establish the presence of a threat (i.e.,
212 studies in which the threat is discovered as a by-product of the study rather than its main goal,
213 such as reports on wildlife rehabilitation center admission records), b) studies whose primary
214 objective is to assess the magnitude of the threat, and c) studies that explore drivers that
215 influence how the threat impacts birds. We defined intervention studies as those that assess the
216 outcome of an implemented conservation intervention on bird populations. Studies that explored
217 the design and implementation of a proposed intervention, but did not explore outcomes, or
218 studies that included surgical or care-related treatments of birds, were not considered for further
219 data extraction.

220 *Data extraction from threat studies*

221 From both threat and intervention-based studies we extracted: bibliometric information,
222 geographic location of the study, bird species or community studied, and conservation status of
223 each species. We also collected information on whether the study targeted a specific individual
224 bird species. Furthermore, we collected information on whether the study was conducted in a
225 controlled setting (i.e., in a lab or field setting in which the environment and population were

226 strictly controlled) or was observational in a less strictly controlled, usually field-based, setting.
227 Data extraction was completed in full (i.e., double-blind data extraction and conflict resolution)
228 by two reviewers. The full data extraction template can be found in the Open Science Framework
229 page for this project (**link to be provided upon acceptance**).

230 *Data extraction for meta-analysis of intervention studies*

231 For intervention studies only, we collected additional information for quality assessment
232 and meta-analyses. This included information on the study design, type of intervention, sample
233 sizes and more specific data according to the type of effect size used (see information on meta-
234 analyses below). Due to the limited number of intervention studies in the light pollution and
235 vehicle collision threat categories, we only extracted further information for meta-analysis in the
236 cat predation and bird-window collisions categories. A list of all included studies and their data
237 can be found in the Open Science Framework page for this project. These data were extracted by
238 the lead author.

239 We assessed intervention studies for the quality of their methodological approach and
240 potential for risk of bias. One author (AK) assessed study quality based on modified CEE-based
241 study validity assessments of two previous systematic reviews (Harper et al., 2022; Rytwinski et
242 al., 2023). Our rubric (Supplementary Material Table S4) defined and assessed selection bias
243 (i.e., whether the study was appropriately controlled, whether there was true replication, and
244 whether replicates were well matched); assessment bias (i.e., consistency in sampling methods
245 and whether a quantitative outcome was provided); performance bias (i.e., whether the study was
246 free of extenuating circumstances and whether study settings were consistent); and attrition bias
247 (i.e., whether units of replication were removed from the study, leading to sampling differences
248 between the control and intervention groups – particularly applicable to cat predation studies).

249 To determine relevant confounding factors to assess selection and performance bias
250 (Supplementary Material Table S4), we used information from threat studies. The full quality
251 assessment rubric with results can be found in in the Open Science Framework page for this
252 project.

253 **Statistical analysis**

254 All analyses were conducted using R statistical software version 4.4.3 (R core team,
255 2024). To retrieve the conservation status of each species studied, we used the “rredlist” package
256 (Gearty & Chamberlain, 2025) to extract information directly from the International Union for
257 Conservation of Nature (IUCN) Red List of Threatened Species Version 2025-1 (API v4)
258 (International Union for Conservation of Nature, 2025).

259 ***Meta-analysis of intervention studies***

260 Meta-analysis was conducted using the “metafor” package (Viechtbauer, 2010). We built
261 a multi-level mixed-effects model using the *rma.mv* function with intervention type as a
262 moderator. Several studies included multiple interventions; in these cases, each intervention was
263 independently assessed for quality, and separate effect sizes were calculated. To account for
264 multiple study comparisons, study ID was included as a random factor in the model. Due to the
265 small number of studies, we opted to fit the model without an intercept (i.e., without a reference)
266 to summarize the average effects of each intervention rather than compare their relative
267 effectiveness (Viechtbauer, 2010).

268 For both meta-analyses, we assessed heterogeneity using tests for τ^2 (between-study
269 variance and within-study variance) and I^2 (proportion of variability due to true differences
270 among effect estimates) (Viechtbauer, 2010). Residual heterogeneity was evaluated using QE,
271 which tests whether unexplained variation remained after accounting for moderators (i.e.,

272 intervention type) and the random effects (i.e., study ID). Moderator effects were tested using
 273 QM, which evaluates whether intervention type significantly explains variation in effect
 274 estimates (results in Table S9, Table S10, Table S11, and Table S12).

275 Publication bias was also assessed by using two complementary approaches (Figure S3
 276 and Figure S4). First, funnel plot asymmetry was evaluated using Egger's regression test (Egger
 277 et al., 1997). However, these results should be interpreted with caution as this test suffers from
 278 poor power when the number of effect sizes is low (Simmonds, 2015). We also used
 279 Rosenberg's fail-safe number N (Rosenberg, 2005), which is commonly used for random effects
 280 models and estimates the number of additional studies with an average effect size of zero that
 281 would be required to make an outcome not significant. We considered the meta-analysis robust
 282 against publication bias if the failsafe number was greater than $5k + 10$, where k is the number of
 283 effect sizes in the analysis.

284 For cat predation studies, our meta-analysis evaluated the reduction in birds brought
 285 home by cats from collar-based interventions. We used the standardized mean difference (SMD)
 286 (otherwise known as Hedges' g , which includes a correction for smaller sample sizes) as our
 287 effect size measure, calculated manually using the steps in Koricheva et al., (2013):

$$288 \quad g = \frac{\bar{X}_{G1} - \bar{X}_{G2}}{S_{pooled}} \cdot J$$

289 where \bar{X}_{G1} is the mean of group 1, or the control group (i.e., the average number of birds brought
 290 home per cat with no intervention collar), and \bar{X}_{G2} is the mean of group 2, or the intervention
 291 group (i.e., the average number of birds brought home per cat wearing the intervention collar).
 292 S_{pooled} is the pooled standard deviation between the two groups. J is a correction applied to
 293 remove small sample bias:

294
$$J = \left[1 - \frac{3}{4(n_1 + n_2) - 1} \right]$$

295 where n_1 is the sample size of group 1 and n_2 is the sample size of group 2. A positive Hedges' g
 296 indicates that the average number of birds brought home per cat was lower when cats wore an
 297 intervention collar than cats with no intervention collars, while a negative Hedges' g corresponds
 298 with an increase in birds brought home by cats with intervention collars compared to controls.
 299 We also calculated sampling variance to build confidence intervals around the effect size,
 300 defined as:

301
$$v_i = \frac{n_1 + n_2}{n_1 n_2} + \frac{g^2}{2(n_1 + n_2)}$$

302 Mean, standard deviation, and sample size were collected from all studies where possible,
 303 where mean was the average number of birds killed per cat and the sample size was the number
 304 of cats. For some studies, we used PlotDigitizer (PORBITAL, 2025) to extract mean and
 305 standard error values from figures, which were then converted to Hedges' g and variance values.
 306 For one study reporting confidence intervals, we manually calculated standard error, assuming a
 307 normal distribution.

308 For bird-window collision studies, we examined log odds ratios for each intervention
 309 using the *escalc* function (Viechtbauer, 2010). The simplified equation (Koricheva et al., 2013)
 310 is:

311
$$\log OR = \ln \left(\frac{a_i \cdot d_i}{b_i \cdot c_i} \right)$$

312 where a_i is the number of times a bird collides with a treated window, b_i is the number of times a
 313 bird avoids a collision with a treated window, c_i is the number of times a bird collides with an
 314 untreated window, and d_i is the number of times a bird avoids a collision with an untreated

315 window. Negative log-odds ratios indicate a reduction in window collisions following the
316 intervention.

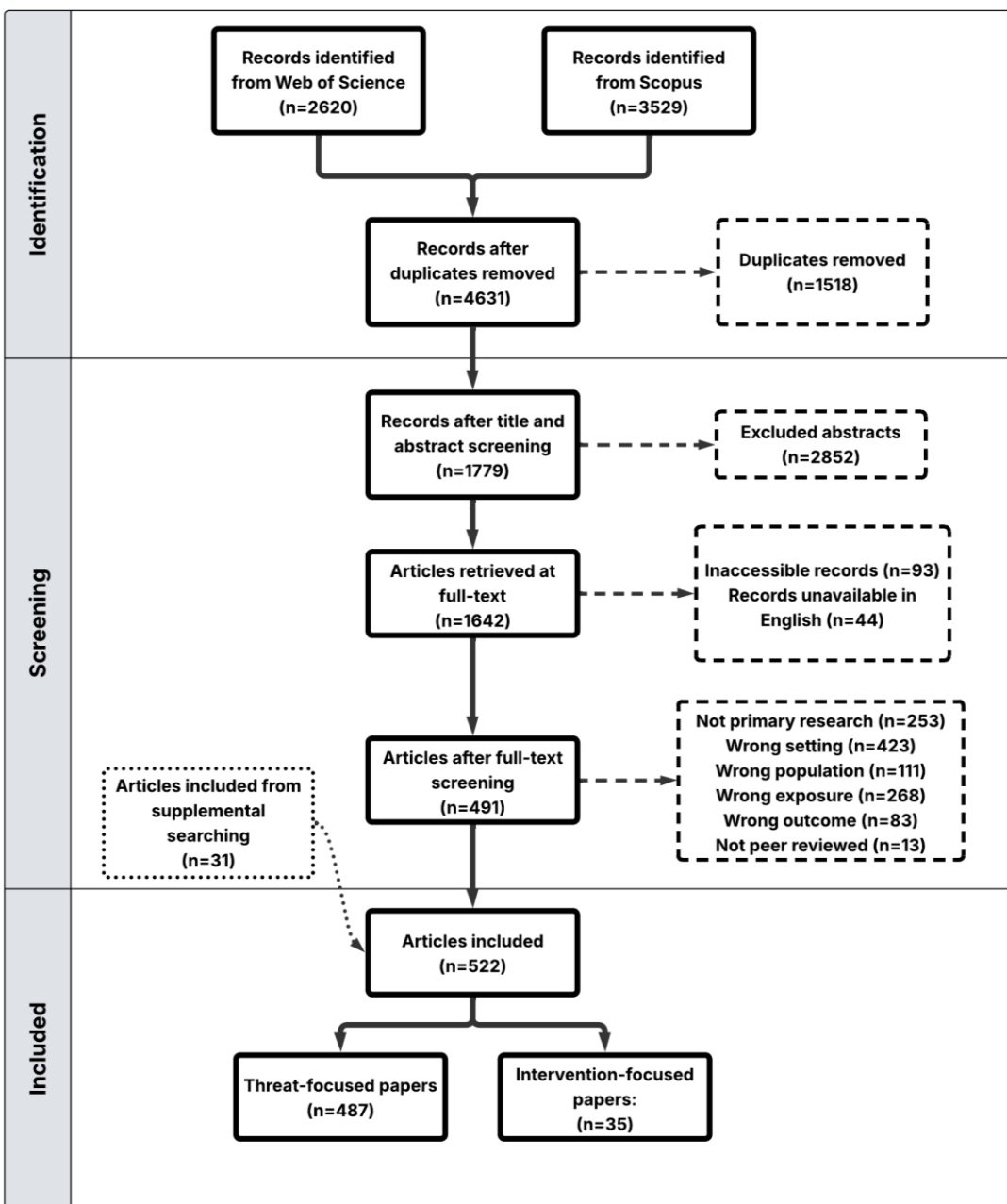
317 Separate effect sizes between the two meta-analyses were used due to the differences in
318 how outcomes were reported in studies for each threat category. For the cat predation studies,
319 outcomes were continuous, making Hedges' *g* an appropriate effect size to compare mean values
320 between groups (i.e., before vs. after intervention). In contrast, bird–window collision studies
321 reported binary outcomes (collision vs. no collision), making a log odds ratio more appropriate.

322 We could not extract and calculate effect estimates from all studies, as some lacked
323 replication, some did not maintain consistent controls, and some did not present the results in a
324 way that allowed effect sizes to be calculated. For this reason, we also conducted and compared
325 results from a vote-count for each threat category, in which the number of significant results for
326 each intervention was tallied across studies (Koricheva et al., 2013).

327 **RESULTS**

328 **Literature search and screening**

329 Our review retrieved a total of 522 articles (Figure 1), including 118 on cat predation of
330 birds, 93 on bird-window collisions, 221 on light pollution, and 101 on bird-vehicle collisions.
331 The United States of America was the most common country in which studies were located,
332 comprising 26.6% of all studies, followed by the United Kingdom (6.5%), Canada (6.1%) and
333 Australia (6.1%; Supplementary Material Figure S1).

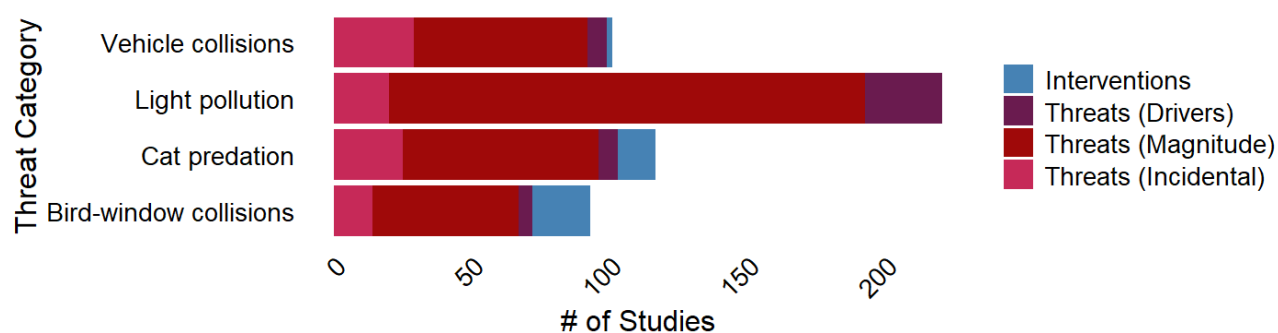


334

335 **Figure 1.** ROSES flow diagram (modified from Haddaway et al., 2018) showing literature
 336 sources and inclusion/exclusion process with criteria for combined evidence syntheses on cat
 337 predation, bird-window collisions, light pollution, and bird-vehicle collisions.

338 **Narrative synthesis of threats and intervention studies**

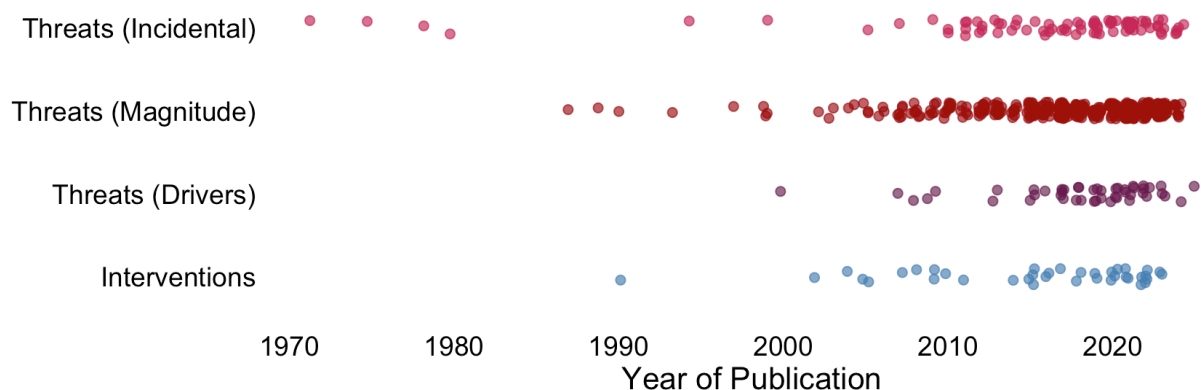
339 Across threat categories, we found 6.7% of studies (n=35 of 522) explored outcomes of
 340 conservation interventions (Figure 2). The majority (68.7%, n=359) of all studies were focused
 341 on understanding the magnitude of each threat. 16.5% of studies incidentally explored threats,
 342 typically examining and reporting on causes of admittance of birds to wildlife rehabilitation
 343 centers. 9% of studies specifically explored drivers that influence the effects of the threat on
 344 birds, such as studies exploring effects of different wavelengths of light on birds' sleep cycles
 345 (Liu et al., 2019). The bird-window collisions threat category had the greatest number (n=20)
 346 and proportion of intervention studies, where 21.5% of all window collisions studies explored
 347 outcomes of interventions, comprising over half (54%) of intervention studies across all
 348 categories. Cat predation had the second greatest number of intervention studies (n=13),
 349 comprising 11% of all studies in cat predation and 35.1% of all intervention studies. The light
 350 pollution category had no intervention studies, but it had the greatest proportions of studies
 351 assessing drivers of threats, comprising 12.7% of studies in the light pollution threat category
 352 and 59.3% of all studies on drivers. Only two studies were found on interventions for bird-
 353 vehicle collisions.



354
 355 **Figure 2.** The number of studies extracted for the narrative synthesis on each key urban threat to
 356 bird populations. Colors represent studies that incidentally establish the presence of a threat

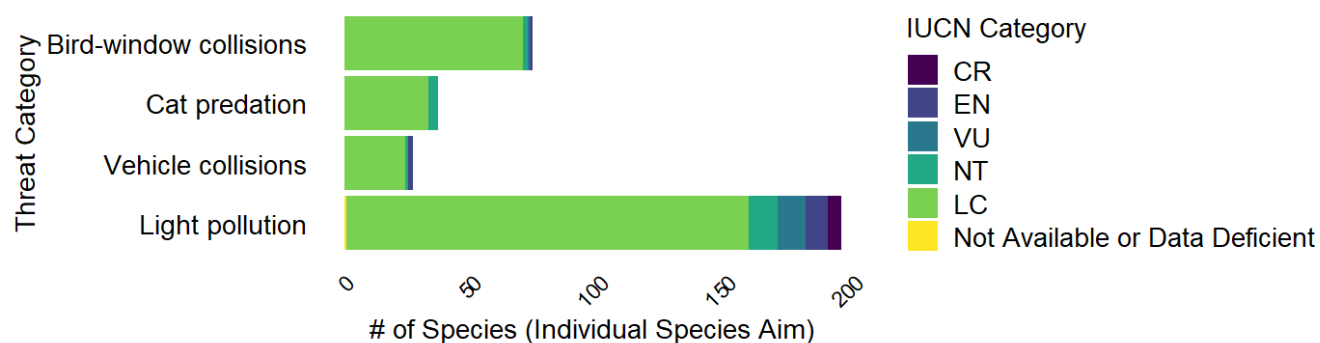
357 (incidental), assess the magnitude of a threat (magnitude), explore drivers of the threat (drivers),
 358 or assess outcomes of interventions addressing the threat (interventions).

359 Studies were published from 1971 to 2024 (Figure 3). 96.6% studies were published on
 360 or after the year 2000. The first paper in 1971 provided records of ringing recovery and
 361 established the first incidental findings of bird-window collisions, bird-vehicle collisions, and cat
 362 predation (Glue, 1971). Studies assessing the magnitude of threats were published almost three
 363 decades later in 1989, when experiments were conducted to determine whether birds can see
 364 glass (Klem, 1989). Intervention studies across all threat categories began to appear in the 1990s
 365 with experiments on window treatments that prevent bird-window collisions (Klem, 1990), but
 366 did not pick up in frequency until after 2000.



367
 368 **Figure 3.** Densities of studies on cat predation, bird-window collisions, light pollution and bird-
 369 vehicle collisions over time from an evidence synthesis aiming to explore the outcomes of
 370 conservation interventions for major urban threats. Each point represents one paper. Colors
 371 represent studies that incidentally establish the presence of a threat (incidental), assess the
 372 magnitude of a threat (magnitude), explore drivers of the threat (drivers), or assess outcomes of
 373 interventions addressing the threat (interventions).

374 Most studies (81.1%) were natural experiments which did not control the setting. Light
 375 pollution studies had the greatest number (n=62) and proportion (32.8%) of studies in controlled
 376 settings, the majority of which were indoor avian experiments to examine the effects of different
 377 wavelengths and intensities of artificial light on bird health and behaviour. A little under half of
 378 studies targeted individual taxa (47.9%) (Supplementary Material Figure S2). Light pollution had
 379 the highest number of such studies (n=170), comprising 66.7% of all studies targeting individual
 380 taxa. We were able to retrieve data on individual species from 60.4% of studies for a total of 646
 381 unique bird species from 97 families. Of these, 579 were Least Concern, 27 were Near
 382 Threatened, 22 were Vulnerable, 11 were Endangered, five were Critically Endangered, and two
 383 were Data Deficient. In most threat categories, very few birds targeted were red-listed birds
 384 (10.1%) (Figure 4). Light pollution was an exception, in which 20.6% of birds targeted for
 385 research were listed. Intervention studies almost exclusively reported on species listed as Least
 386 Concern except for a single Near Threatened species mentioned in an intervention study for bird-
 387 window collisions.



388
 389 **Figure 4.** Proportions of 260 species extracted from studies targeting individual species under
 390 each threat category. Bars are stacked by the International Union for Conservation of Nature's
 391 (IUCN) Red List category.

392 **Synthesis and meta-analysis of evidence from intervention studies**

393 ***Cat predation***

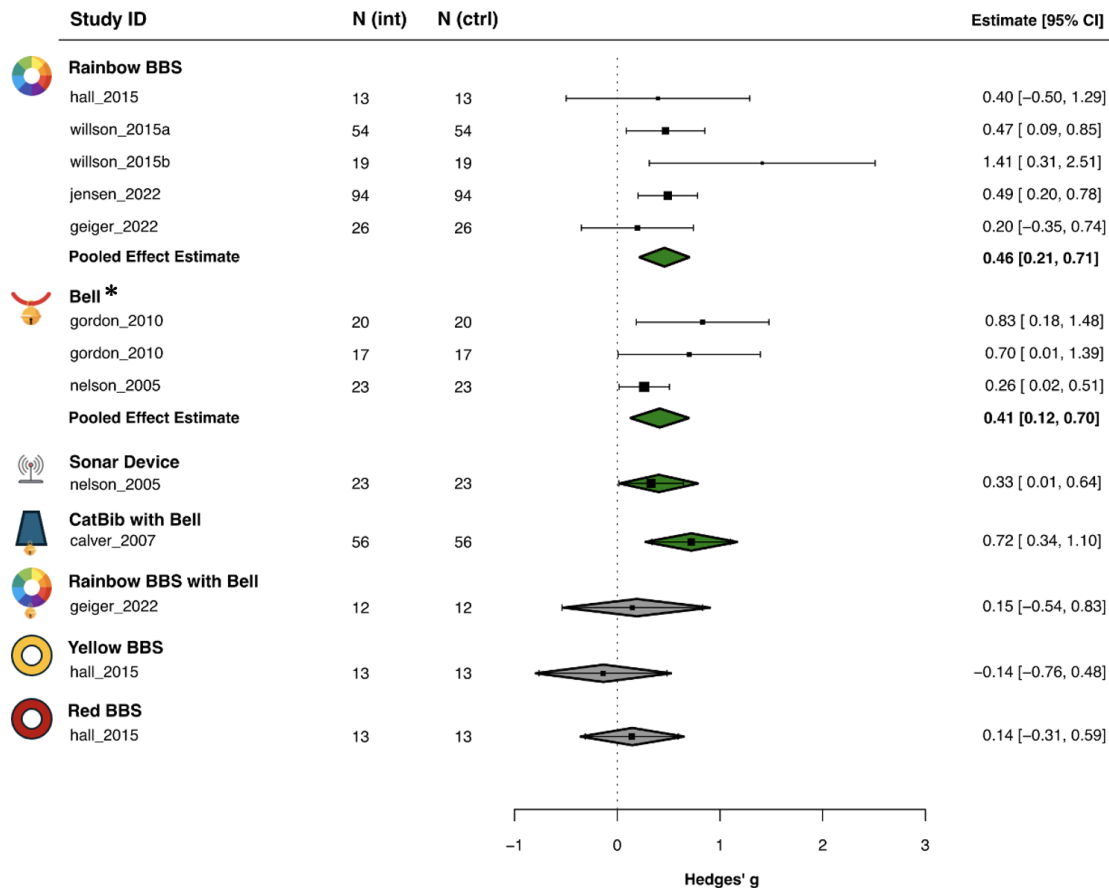
394 *Effectiveness of interventions*

395 Of the 13 intervention studies, we found eight studies on Birdsbesafe™ (BBS) cat
396 collars, five on belled collars, two on other sonar devices, two studies on CatBibs™ (Calver et
397 al., 2007), one study on feral cat removal (Meckstroth & Miles, 2005) and three studies on
398 behavioural or diet-based interventions including using a puzzle feeder, increased play with
399 wands and other toys, and a higher protein diet (Cecchetti et al., 2021) (vote count results are in
400 Supplementary Material Table S5). Three studies tested outcomes of combined interventions:
401 one tested outcomes of Rainbow-coloured BBS collars combined with bells (Geiger et al., 2022),
402 and two tested outcomes of the CatBib™ with a bell (Calver et al., 2007). For the one study on
403 outcomes of predator removal in an urban area we were unable to isolate the specific effects of
404 cat removal from the effects of removing all predators (Meckstroth & Miles, 2005).

405 For our multi-level mixed-effects model, we extracted 13 standardized mean differences
406 from seven different studies (Figure 5; Supplementary Material Table S6). Of the six studies we
407 could not include, three did not separate the bird-based outcomes from other taxa and three
408 lacked sufficient data to calculate a Hedges' *g* effect size. Consistent with our vote count results
409 (Supplementary Material Table S5), the Rainbow BBS collar had the greatest effect estimate
410 supported by the largest number of studies (0.46 ± 0.13 , $p < 0.05$; $n=4$). In magnitude, the
411 weighted mean percent reduction in birds brought home by cats with collars was 53%. The effect
412 for mounted sonar devices was also significantly positive (0.40 ± 0.20 , $p < 0.05$). We were able to
413 include three out of five studies on belled collars in our multi-level model, which had a
414 significant positive pooled effect (0.41 ± 0.15 , $p < 0.05$, $n = 3$). The two studies from which we
415 could not extract a Hedges' *g* effect size did not find a significant effect of bells; however, both

416 had low quality assessment scores for study design, selection bias and performance bias (see
417 quality assessment below). The combination of the CatBib™ with the bell had a significantly
418 positive effect (0.72 ± 0.23 , $p < 0.05$). However, the combination of the Rainbow BBS and bell
419 was not significant (0.19 ± 0.37 , $p = 0.61$) nor were red and yellow BBS collars (0.15 ± 0.26 ,
420 $p = 0.58$; -0.13 ± 0.34 , 0.69 , respectively).

421 Our test for heterogeneity found that intervention type significantly explained the
422 variation in effect estimates ($QM = 32.39$, $df = 7$, $p < 0.001$) (Table S9). Within-intervention
423 tests found moderate inconsistency in results among studies within the “bell” intervention
424 category ($I = 46.26\%$), and high consistency within the “rainbow BBS” category ($I^2 = 0.01\%$)
425 (Table S10). The Rosenberg’s fail safe number N was 130, suggesting that the meta-analysis is
426 robust against publication bias. The funnel plot and Egger’s test results for publication bias can
427 be found in the Supplementary Materials (Figure S3).



428

429 **Figure 5.** Standardized mean different effect size (Hedges' g) of collar-based interventions on
 430 rates of birds brought home by pet cats. The columns "N (int)" and "N (ctrl)" refer to the sample
 431 sizes of cats tested with the interventions on, compared to cats tested during the control.

432 Interventions are pooled into seven different types, with the diamond depicting pooled effects
 433 (i.e., for intervention types with >1 effect size comparison). Error bars indicate 95% confidence
 434 intervals. A positive mean value (right of dashed zero line) indicates that the number of birds
 435 brought home by cats with interventions was less than control groups (no intervention) (i.e., a
 436 significant positive reduction in birds brought home). Sizes of squares depict relative weight on
 437 pooled effect size. Multiple trials within the same study are included as separate effect sizes.

438 *We could only draw effect sizes from studies that found a significant positive effect for belled

439 collars. Two studies which could not be included did not find a significant positive effect for
440 belled collars.

441 *Quality assessment*

442 Intervention assessments from every article except for two had a repeated-measures
443 crossover design (the exceptions being a correlative study and a BACI design) and true
444 replication, resulting in a low risk of selection and assessment bias. However, none of the studies
445 recorded the relative lengths of time each cat was let out during the day, resulting in a moderate
446 risk of performance bias. Many studies were at moderate or high risk of attrition bias as they did
447 not specify exactly when cats were removed during the trials and how the removals affected their
448 final sample sizes and resulting analyses (see OSF data repository for full assessments).

449 ***Bird-window collisions***

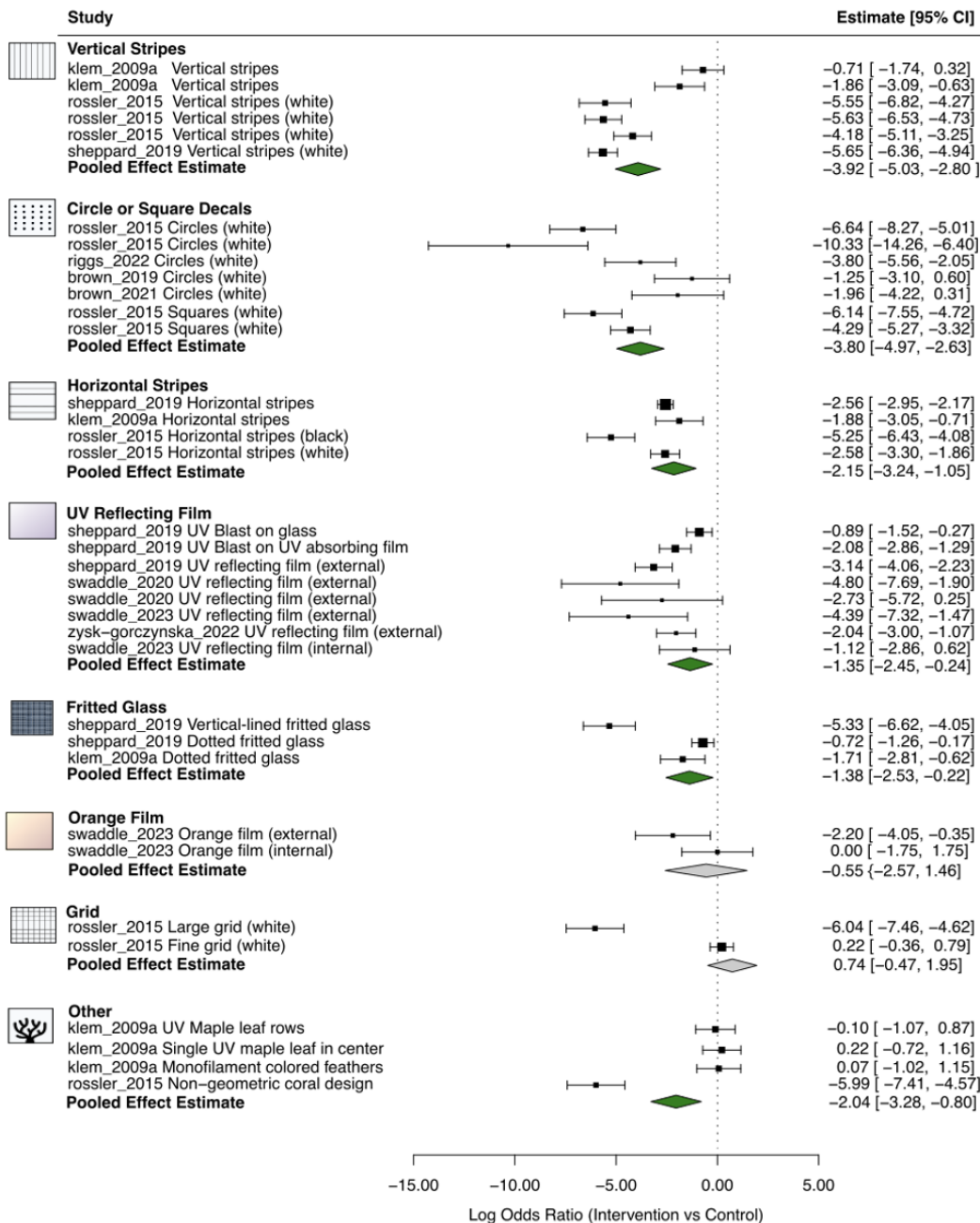
450 *Effectiveness of Interventions*

451 We included 8 out of 20 studies in our odds ratio analysis –four field studies and four
452 tunnel experiments – for a total of 36 effect sizes (Figure 6). Studies were included if they had an
453 adequate control, replication, and results presented in a way that would allow odds ratio values to
454 be extracted. Field studies applied interventions in real-world settings under natural conditions,
455 typically on pre-existing buildings. Tunnel experiments were randomized controlled trials in
456 which birds were released into an enclosed tunnel containing constructed window surfaces to
457 assess collision risk in response to different window treatments. As with our vote-count results
458 (Supplementary Material Table S7), our meta-analysis found that vertical stripes had the greatest
459 effect size (-3.92 ± 0.57 , $p < 0.05$), resulting in a pooled reduction in window collisions of 98%.
460 The only study that found vertical strips to be ineffective had the strips greater than 10 cm apart
461 (Klem, 2009). Circular or square decals had almost as large an effect size as vertical stripes (-

462 3.80 ± 0.60, $p < 0.05$), with circular decals also being the most frequently studied intervention
463 (n=9). UV reflective coatings or films were also widely studied and found to reduce the number
464 of birds colliding with windows (-1.35 ± 0.56 , $p < 0.05$), except where the film was applied to the
465 interior rather than the exterior surface of the window (Swaddle et al., 2023). Fritted glass (-1.38
466 ± 0.59 , $p < 0.05$) and horizontal stripes (-2.15 ± 0.56 , $p < 0.05$) also reduced collisions but had
467 fewer studies (n=3 and n=4, respectively). Pooled results for the “other” category (coral, maple
468 leaf, and feather designs) had a relatively large and statistically significant effect size ($-2.04 \pm$
469 0.63 , $p < 0.05$); however, of these studies, only the effect size of the coral design was significant.
470 Studies on bird-of-prey decals, which could not be included in the meta-analysis, were the only
471 intervention found to be consistently ineffective in preventing window collisions based on our
472 vote-count.

473 Our test for heterogeneity found that intervention type significantly explained the
474 variation in effect estimates ($QM = 255.77$, $df = 8$, $p < 0.001$) (Table S11). Tests for within-
475 intervention heterogeneity found that vertical stripes showed high heterogeneity ($QE = 87.82$, p
476 < 0.001 ; $\tau^2 = 4.341$; $I^2 = 94.31\%$) as well as circle or square decals ($QE = 37.70$, $p < 0.001$; $\tau^2 =$
477 6.141 ; $I^2 = 84.08\%$), and horizontal stripes ($QE = 20.58$, $p < 0.001$; $\tau^2 = 1.759$; $I^2 = 85.42\%$)
478 (Table S12). The Rosenberg’s fail safe number N was 8161, suggesting that the meta-analysis is
479 robust against publication bias. The funnel plot and Egger’s test results for publication bias can
480 be found in the Supplementary Materials (Figure S4).

481



482
 483 **Figure 6.** Log-odds ratio of 36 effect sizes for 21 specific interventions from 8 studies on
 484 outcomes of window treatments on bird-window collisions. Interventions are pooled into eight
 485 different intervention types, with the diamond depicting the pooled effects. Horizontal bars
 486 depict the effect size and 95% confidence intervals. Negative log-odds ratios indicate a reduction
 487 in window collisions following intervention. Effect sizes whose confidence intervals do not

488 overlap the dotted line of no effect are statistically significant. Sizes of squares depict the relative
489 weight on pooled effect size.

490 *Quality assessment*

491 Two BACI field studies had perfect quality scores (Riggs et al., 2023; Zyśk-Gorczyńska
492 & Źmihorski, 2022). Of the remaining field studies, 10 of 13 did not have true replication, eight
493 did not report potential environmental confounding factors, and only four adequately described
494 confounding factors between the units of replication (i.e., the window area, angling and specific
495 color/tinting, if applicable). All tunnel test experiments were randomized controlled trials in
496 strictly controlled settings which accounted for potential confounders; thus, almost all received
497 perfect scores through our quality assessment rubric. The exception was a tunnel test study
498 which was not transparent about its statistical methodology for determining significance (Klem,
499 1990).

500 *Bird-vehicle collisions*

501 Only two studies tested outcomes of interventions for bird-vehicle collisions. The first
502 paper comprised three experimental studies involving different bird taxa that found vertical pole
503 barriers were successful at altering the flight paths of birds, which could result in avoidance of
504 bird-vehicle collisions (Zuberogoitia et al., 2015). All three experiments received perfect quality
505 scores for assessment and performance biases but lacked true replication and independent control
506 sites.

507 The second paper compared road-kill rates before and after the installation of a series of
508 traffic calming devices, such as chicanes and speed bumps, finding that road mortality rates were
509 lower in forested areas where the road calming devices were installed (Jones et al., 2014). This
510 study lacked replication, had a high risk of performance bias as it did not account for related

511 confounding factors, and outcomes for birds and other taxa were pooled, making it impossible to
512 discern the effects of the intervention on birds independently.

513 **Discussion**

514 In synthesizing the literature on the effectiveness of conservation interventions to tackle
515 urban threats, we found a dearth of literature on outcomes for birds, supporting results from other
516 reviews in conservation science (e.g., Binley et al., 2025; Esler et al., 2010; Williams et al.,
517 2020). While research on threats has greatly increased in volume, those exploring outcomes of
518 interventions have not seen the same growth, remaining infrequent in urban bird research. Most
519 studies (93.3%) presented the prevalence, magnitude, or drivers of threats. Of these, studies
520 exploring drivers, which are critical for informing intervention research and practice, were also
521 proportionately few (9% of all studies).

522 Of the literature on outcomes of interventions, over half (57.1%) explored outcomes of
523 window treatments to reduce bird-window collisions. These studies found that most window
524 treatments, except for bird of prey decals and other films or decals with low area coverage or
525 placed on the inside of the window, reduced bird-window collisions by 98%. Out of collar-based
526 interventions to reduce the number of birds killed by cats, Rainbow Birdsbesafe™ collars
527 showed the greatest reduction in the number of birds cats brought home. However, cats wearing
528 these collars still brought home some birds, whereas hypothetically none would be brought home
529 if cats were kept indoors. The quality of intervention studies was found to be particularly poor
530 for bird-window collision field studies, requiring greater true replication and accounting for
531 confounders at the level of replication and between treatment and control environments.

532 For cat predation, existing intervention studies were high in quality but were primarily
533 focused on collar-based or behavioural interventions for owned cats that were still let outdoors,

534 rather than interventions that keep owned cats indoors or reduce feral cat populations. While we
535 found proportionately more studies for window collision interventions, many of the field tests
536 suffered from significant quality issues, making it difficult to conduct statistical analyses on most
537 studies. We found no urban research on the effectiveness of proposed conservation interventions
538 for light pollution. Conclusions from the two studies for bird-vehicle collisions were also limited
539 by quality.

540 A very small proportion of studies tested outcomes on or detected red-listed species
541 (International Union for Conservation of Nature, 2025). Species at risk of extinction may not be
542 detected in studies of urban threats as they may be more uncommon in urban areas. However, it
543 is likely that they are still vulnerable to these threats. For instance, many species at risk of
544 extinction are migratory, and light pollution is specifically more concentrated along migratory
545 hotspots (Cabrera-Cruz et al., 2018). Among the threat categories, light pollution had the greatest
546 proportion of research involving red-listed species (20.6%), possibly due to existing knowledge
547 on the impacts on migratory birds (Cabrera-Cruz et al., 2018; Van Doren et al., 2021).

548 **Research on interventions**

549 *Cat Predation*

550 We found that existing research on interventions addressing cat predation in cities
551 focused predominantly on collar-based devices that work to mitigate predation by owned outdoor
552 cats. The Rainbow Birdsbesafe™ collar, a brightly rainbow-colored fabric tube worn over a
553 breakaway collar which makes the cats highly visible to birds (Hall et al., 2015; Willson et al.,
554 2015), was studied most widely and had consistently demonstrated reductions in the number of
555 birds brought home, with a pooled reduction of 53%. Studies on belled collars had mixed results,
556 although the two studies that found bells did not significantly reduce bird catch had lower scores

557 for quality – one was a correlative study (Morgan et al., 2009), and the other had a medium risk
558 of selection and performance bias (Cecchetti et al., 2021). Bells are intended to deter birds by
559 making a sound, but it is possible they may be less effective in urban areas where the sound
560 could be masked by anthropogenic noise (Damsky & Gall, 2016). Other interventions, such as
561 the CatBib™ (another visibility-based collar) (Calver et al., 2007) and various mounted sonic
562 devices (Calver & Thomas, 2010; Nelson et al., 2005), were each found to significantly reduce
563 rates of birds brought home, but research remains too limited to draw definitive conclusions.
564 Cats are known to be particular and vary greatly in their individual tolerances of different collar
565 types (Harrod et al., 2016). Therefore, continued research would benefit from testing a variety of
566 potential interventions and cats' relative tolerance alongside the outcomes for birds. However, it
567 is important to note that even the Rainbow Birdsbesafe™ collar only reduced the number of
568 birds brought home by just over half (54%), whereas keeping cats indoors would reduce prey
569 brought home entirely.

570 With this in mind, we recommend continued prioritization of education and research on the
571 efficacy of other interventions for keeping cats indoors. We found there is still little research on
572 outcomes of other interventions for owned cats, such as restricting cat movement using leashes
573 and catios, or the introduction of roaming bylaws. One study found that areas where cats were
574 not allowed or cats faced strict curfews resulted in no significant outcomes for local small
575 mammal diversity and abundance, although factors such as vegetation characteristics may have
576 also played a role (Lilith et al., 2010). Another study that correlated cat density with bird
577 abundance and diversity within different neighbourhoods in the City of Ottawa, Canada, found
578 that the two variables were only weakly related (Perkins et al., 2021). Although threat studies
579 demonstrate that unowned cats have high predation rates of birds in urban areas (Loss et al.,

580 2013), we found limited research on potential interventions for unowned cats. While many
581 studies exist on outcomes of cat eradication for island birds (Nogales et al., 2004), we found only
582 one study that explored the outcomes of predator removal on birds in San Francisco (Meckstroth
583 & Miles, 2005). However, cats were only one predator removed, and the presence of other more
584 common predator species confounded results (Meckstroth & Miles, 2005). Several studies
585 suggest that ongoing and comprehensive trap-and-neuter programs are the most effective way to
586 lower urban cat populations (Luzardo et al., 2025; Spehar & Wolf, 2018); however, more
587 research is needed on whether these programs result in reduced predation rates of birds in cities.

588 A common source of bias across most cat intervention studies was attrition bias, which
589 can be particularly impactful where sample sizes are already quite small. Clearer reporting of
590 how this bias may have affected the final treatment group sizes would improve transparency and
591 allow for more accurate interpretation of results. Additionally, none of the studies recorded the
592 relative lengths of time each cat was let out during the day, and whether these lengths may have
593 been different between the treatment and control periods. A significant driver of predation rates
594 is the amount of time a cat has spent outside, since this correlates with the amount of territory
595 they cover and gives them more opportunity to catch prey (Cordonnier et al., 2022).

596 Additionally, a single cat may exhibit significant variation in the time it spends outside on a
597 given day, depending on season or weather (Dunford et al., 2024; Jensen et al., 2022).

598 Cat predation is a major threat faced by many other taxa, and conservation interventions
599 applied to reduce harm to birds may have variable outcomes for other animals. For example,
600 Cecchetti et al. (2021) discovered that puzzle feeders made cats bring home proportionally fewer
601 birds but greater amounts of prey overall. The same study also found that playing with toys at
602 home made cats bring home less prey overall, but proportionally more birds.

603 *Bird-window collisions*

604 Interventions that most consistently reduced bird-window collisions were circular decals
605 and vertical stripes on windows spaced less than 10 cm apart. Other interventions were found to
606 reduce collisions but were only tested in single studies, such as white sheets, square decals, UV-
607 reflective films, or full-window white films. Interventions such as bird-of-prey decals or artwork
608 that only covered small parts of a window did not significantly reduce bird-window collisions.
609 Our findings are comparable to meta-analyses found in grey literature from recent years
610 (Colbaugh, 2023; Samuels, 2025), which have similarly emphasized the importance of window
611 treatments with minimal gaps for the most effective outcomes.

612 There were significant quality issues with the field studies in bird-window collisions.
613 Many studies did not have true replication or did not control for or report confounders which are
614 known to influence collision risk such as the sizes or angles of the windows, nearby vegetation,
615 or placement of bird feeders (Brown et al., 2019; Klem et al., 2004). While field studies can be
616 limited in their ability to control for covariates and confounders, some confounders such as
617 window area can be statistically accounted for post hoc (De Groot et al., 2022). With this in
618 mind, we recommend that future studies aim to improve replication and at least report
619 confounders where possible.

620 All tunnel test studies received perfect scores in all categories of the quality assessment,
621 except for one that did not clearly describe the statistical tests used to determine significance
622 (Klem, 1990). Tunnel tests can be highly useful as they can control for all potential confounders.
623 However, this also limits their applicability in real-world settings (Samuels, 2025). In addition,
624 tunnel test studies use specific bird taxa, and responses to window treatments may differ across
625 different species (Klem, 1989; Ocampo-Peñuela et al., 2016); thus, results from tests may not be

626 representative of responses by bird communities in the real world. Therefore, we recommend
627 prioritizing well-designed field studies that can complement tunnel experiments and better assess
628 outcomes of proposed interventions in real-world settings.

629 *Light pollution*

630 We did not find any studies on outcomes of interventions to reduce light pollution on
631 birds in urban areas. Proposed interventions include lights-out events, switching to warm
632 lighting, and employing downward-facing lighting (e.g., FLAP Canada, 2023; Audubon, 2025).
633 However, due to the spatially pervasive nature of light pollution (e.g., skyglow or diffuse
634 luminance of the night sky by artificial lights) compared to other threats in our review, it can be
635 difficult to clearly assess spatially controlled outcomes of these interventions (Seymoure et al.,
636 2025).

637 While we did not find any intervention studies, light pollution had the greatest proportion
638 of studies on drivers of threats. These studies were typically experiments exploring the effects of
639 different wavelengths and intensities of light. High intensity white and broad spectrum light had
640 the largest effect on sleep quality and immune functioning of birds (Kernbach et al., 2020;
641 Ouyang et al., 2017). However, based on the species and setting, sleep quality could also be
642 affected by green light (Ulgezen et al., 2019), orange light (Liu et al., 2019), and violet light
643 (Ren et al., 2021). Additionally, blue and green light were found to be more likely to result in
644 fatal light attraction (Rebke et al., 2019; Zhao et al., 2020). However, one intervention study,
645 which was excluded because it was not in an urban environment, found that green lighting was
646 the least disorienting for birds, whereas red and white lights were the most disorienting (Poot et
647 al., 2008). We recommend future research testing outcomes of interventions that reduce outdoor
648 light intensity and scatter in real-world settings.

649 Light pollution also influences bird-window collisions. An intervention study we
650 included in our the bird-window collisions threat category found that lowered interior lighting
651 during midday, when reflectivity is the highest, significantly reduced bird-window collisions
652 (Emerson et al., 2022). Although most window collisions occur during the day, modelled
653 estimates suggest that reducing light pollution can reduce collisions at night during peak
654 migration periods by 53-59% (Van Doren et al., 2021). However, models have not been
655 empirically tested, so actual outcomes of these programs remain unknown.

656 *Vehicle collisions*

657 We found very little research on interventions for urban bird-vehicle collisions. This is
658 surprising given that road mortalities are a widely studied phenomenon globally and many
659 interventions have been proposed for other taxa including mammals, reptiles, and amphibians,
660 such as overpasses, underpasses, and road barriers, which have been reviewed elsewhere (Glista
661 et al., 2009; Rytwinski et al., 2016). The two intervention studies we found had positive results.
662 The study on pole barriers found that they were effective in circumventing flight paths for tamed
663 raptors, waterbirds, and griffon vultures. However, all experiments in this paper lacked true
664 replication and were conducted in open fields rather than the roads upon which they would be
665 applied (Zuberogitia et al., 2015). The study on traffic calming devices found reductions in road
666 mortalities in areas where the devices were installed. However, this study was at high risk of
667 selection and performance bias. In addition, while results were provided on specifically bird-
668 related mortality accidents prior to the upgrade, results following the upgrade and installation of
669 calming devices were pooled for all wildlife, so the effectiveness on birds alone could not be
670 determined.

671 A few threat-based studies in this category (7.1%) explored drivers of bird-vehicle
672 mortalities and could be used as a starting point for future intervention research. For example,
673 Husby (2016) found that species that were lighter in weight and were attracted to roads for
674 foraging opportunities were more likely to be hit by vehicles. Other studies have found that the
675 height of the embankment and presence of hedgerows or other vegetation nearby also influence
676 bird-vehicle collisions (Husby, 2016; Orłowski, 2008; Pons, 2000). Based on these findings,
677 future interventions could target known hotspots and employ strategies such as removing food
678 attractants like roadkill, applying structural deterrents such as fences, or moving vegetation
679 farther away from the road (Orłowski, 2008).

680 **Future directions and conclusions**

681 The available evidence base for conservation intervention studies for urban birds was
682 relatively small and unevenly distributed across intervention types, which limited the strength of
683 conclusions drawn from the meta-analyses, including our ability to assess heterogeneity and
684 publication bias. While some interventions, such as the Rainbow Birdsbesafe™ collars, showed
685 consistent effects across studies, the overall number of studies was still low (e.g., n=4). In
686 addition, several intervention studies in the bird-window collisions threat category lacked the
687 appropriate controls and replication to be included in the meta-analysis. Furthermore, as is the
688 case for many global reviews, a major limitation of our study was that studies had to be restricted
689 to English, potentially missing key studies published in other languages which is reported to bias
690 outcomes of ecological meta-analyses (Konno et al., 2020). Our review also only examined peer-
691 reviewed studies, thus missing key results from grey literature (Jacobson, 2022; Samuels, 2025).
692 Finally, while the focus of our study was on urban settings, it is important to consider that studies
693 occurring in non-urban settings can still influence our knowledge on overall threats and

694 interventions. Future meta-analyses could expand their scope to include non-English and grey
695 literature to add strength to their conclusions.

696 While our review explored outcomes of interventions for bird populations once applied,
697 the success of the application of these interventions relies on the awareness and uptake by
698 people. Some of the intervention studies for cat predation also included questionnaires for
699 owners to understand owner preferences and collar uptake by the cats (Calver et al., 2007; Hall et
700 al., 2015). These questionnaires found that people's preferences for aesthetics and cats' attitudes
701 around wearing the collar influenced whether an owner was likely to keep the collar on the cat.
702 For application of window collision interventions, a study by Lysyk & Khan et al. (2025) found
703 that barriers such as time, cost, and aesthetic preferences may prevent people from applying
704 window collision stickers to individual homes. In summary, understanding the biological
705 outcomes of bird conservation interventions is important, and understanding the potential social,
706 economic, and political barriers to uptake is equally important for successful implementation.

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1006 **A global evidence synthesis of outcomes of urban bird conservation interventions:**

1007 **Supplementary materials**

1008 **Supplementary Material Table S1.** List of search terms for each threat category in ISI Web of Science Core Collections and Scopus,
 1009 along with the number of articles retrieved from each database, the total number of articles, and the date of the search performed using
 1010 Carleton University's institutional subscription.

| Topic | Search Terms (Web of Science) | Search Terms (Scopus) | # Articles (Web of Science) | # Articles (Scopus) | Total | Date of search |
|--------------|---|--|------------------------------------|----------------------------|--------------|-----------------------|
| Cat | (*bird* OR avian OR aves) | (*bird* OR avian OR aves) | 886 | 1127 | 1640 | April 29, 2024 |
| Predation | AND (domestic cat\$ OR pet cat\$ OR feral cat\$ OR feline OR "Felis catus" OR "Felis domesticus") NOT | AND ((domestic AND cat) OR (pet AND cat) OR (feral AND cat) OR feline OR "Felis catus" OR "Felis domesticus") AND NOT (influenza OR virus OR | | | | |

| | | | | | | |
|------------|--|--|-----|-----|------|----------------|
| | (influenza OR virus OR disease OR flu OR disease OR flu OR toxoplasma*) | disease OR flu OR toxoplasma*) | | | | |
| Window | (*bird* OR avian OR aves) | (*bird* OR avian OR aves) | 312 | 458 | 577 | April 29, 2024 |
| Collisions | AND (glass OR window\$) AND (collision\$ OR strik* OR mortalit* OR death\$) NOT (aircraft\$) | AND (glass OR window) AND (collision OR strik* OR mortalit* OR death) AND NOT (aircraft) | | | | |
| Light | (*bird* OR avian OR aves) | (*bird* OR avian OR aves) | 655 | 889 | 1125 | April 28, 2024 |
| Pollution | AND (((pollution OR artificial OR anthropogenic) NEAR | AND (((artificial OR anthropogenic) PRE/1 (light*)) OR "light pollution" OR streetlight | | | | |

(light*)) OR streetlight\$ OR "ALAN") AND NOT
 (poultry)
 NOT
 (poultry)

| | | | | | | |
|------------|--|---|-----|------|------|-------------|
| Collisions | (*bird* OR avian OR aves) | (*bird* OR avian OR aves) | 767 | 1141 | 1470 | May 2, 2024 |
| with | AND | AND | | | | |
| Vehicles | (car OR cars OR vehic* OR road\$ OR street\$ OR windshield\$ OR traffic) | (car OR cars OR vehic* OR road OR street OR windshield OR traffic) AND (collision OR strik* OR mortalit* OR death) AND NOT (poultry OR aircraft\$ OR virus) | | | | |

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1013 **Supplementary Material Table S2.** PE/ICO and setting criteria for abstract and full-text screening. Note, no studies were excluded
 1014 based on the presence or absence of a comparator.

| | |
|------------|---|
| Population | <p>WILD BIRDS. This included terms like songbirds, raptors, waterbirds, shorebirds, ground birds, “aerial insectivores”, and specific species names of non-captive birds.</p> <p>*Included if an experiment was done on birds in captivity to understand effects on wild birds</p> <p>*Included historical/archaeological evidence</p> <p>*Anything involving chickens was excluded</p> |
|------------|---|

| | |
|---------|---|
| Setting | <p>URBAN ENVIRONMENT. We followed the definition for “urban” outlined by the country in which each study takes place</p> <ul style="list-style-type: none"> • List of definitions for “urban” under each country: https://unstats.un.org/unsd/demographic/sconcerns/densurb/defintion_of%20urban.pdf <p>*Included University campuses</p> <p>*Included if an experiment was done in a non-urban environment if the study specifically stated that its goal was to better understand what the outcomes would be in an urban environment (particularly important for ALAN studies).</p> <p>*Excluded if study was conducted explicitly in “agricultural” or “farmland” or “rural” habitats</p> |
|---------|---|

| Threat | Bird-window collisions | Light pollution | Bird-vehicle collisions | Cat predation |
|-----------------|--|--|--|--|
| Category | | | | |
| Exposure | WINDOWS , including but not limited to “windows”, “buildings” (with windows), “glass” | ARTIFICIAL LIGHTS , including but not limited to “ALAN”, “artificial light at night”, “City lights”, “Streetlights” | VEHICLES ON ROADS , including but not limited to “cars”, “trucks”, “roads”, “streets”, “traffic”, including urban trains and railways | DOMESTIC OR FERAL CATS , including but not limited to “pet cats”, “ <i>Felis catus</i> ”, and “ <i>Felis domesticus</i> ” |
| Intervention | Any action that seeks to prevent a bird-window collision. | Any action that seeks to reduce the effects of artificial light pollution on birds. | Any action that seeks to reduce bird-vehicle collisions. | Any action that seeks to reduce the effects of cat predation on birds. |
| Outcome | Any sort of collision or collision-associated behaviour (e.g., study on | Any sort of effect from artificial light pollution on birds (e.g., collisions, | Any sort of collision or collision-associated behaviour (e.g., stress | Any sort of predation by cats on birds or behavior by birds in response to the |

| | | | |
|--|--|--|---|
| birds trying to avoid collisions), which has an effect on the population (bird). | migratory cycle disruption, changes to sleeping patterns or vocalization, changes to abundance). | response or avoidance behaviour). Alternatively, any sort of outcome caused by the intervention on birds. | presence of cats (e.g., mortality, various stress responses, abundance/diversity counts, etc) |
| Alternatively, any sort of outcome caused by the intervention on birds. | Alternatively, any sort of outcome caused by the intervention on birds. | the intervention on birds. | |

1015

1016 **Supplementary Material Table S3.** Definitions for exclusions other than PI/ECO and setting criteria for full-text screening.

| Exclusion Term | Definition |
|------------------------|--|
| Duplicate | Papers that were not already screened out by Covidence as “duplicates”. Errata and corrections were checked for information relevant to data extraction, then placed in this category. |
| Full-text inaccessible | Papers that could not be accessed via Carleton University’s institutional journal access, SciHub, ResearchGate or any other miscellaneous tools available at the review team’s disposal. |

| | |
|---|---|
| | Where possible, corresponding authors of inaccessible papers were emailed. Papers were excluded if corresponding authors could not be reached within 14 days. |
| Full-text unavailable in English | Full-text articles which could be accessed but were written in a language besides English. Papers were excluded if the full-text was unavailable in English even when the abstract was available in English. |
| Not primary research | Any research that primarily used data that was previously published in a peer-reviewed setting. This includes reviews, perspectives pieces, or response papers that only used previously peer-reviewed and published research. Studies that would be considered “primary research” include original experiments and field collections, as well as “reviews”/meta-analyses/modelling papers that use broadly available citizen science data, data provided by the government or other forms of gray literature. |
| Intervention-based paper with wrong population or setting | Any paper that investigated an intervention to one of our threats but was missing the population (birds) or setting (urban) [later sorted into wrong population, setting or outcome] |
| Not peer-reviewed | Papers that were not peer-reviewed (e.g., magazine articles or theses and dissertations). |
| Conference paper | Meeting abstracts or papers in conference journals that were not peer-reviewed. |

1018 **Supplementary Material Table S4.** Quality assessment rubric for intervention-based studies on cat predation of birds, bird-window
 1019 collisions, and bird-vehicle collisions.

| Pre-assessment filter | | | | |
|---|--|--|---|--|
| Question/Criterion | Response to Question/Criterion | Cat Predation | Bird Window Collisions | Vehicle Collisions |
| A. Is the methodological description sufficient to allow study quality assessment? | YES or NO <i>If NO, study is excluded from further assessment.</i> | Must include clear description of: - Study design (e.g., repeated measures, BACI) - Sampling units (cats, households) - Data collection methods (owner logs, cameras) | Must include clear description of: - Collision monitoring methods (daily searches, etc..) - Sampling units (windows, buildings) - Timing and duration of study | Must include clear description of: - Collision monitoring methods (daily searches, etc..) - Sampling units (road section, birds) |

| | | | | |
|---|--|--|---|---|
| | | - Timing and duration of study | | - Timing and duration of study |
| B. Is the intervention description sufficient to allow study quality assessment? (<i>i.e., is the intervention clearly related to cats and well described/easy to interpret?</i>) | YES or NO <i>If NO, study is excluded from further assessment.</i> | Must clearly describe intervention details, e.g.: - Type of collar (bell, bib, etc.) - Duration and consistency of use - Whether other cat behaviors or conditions were changed | Must clearly describe intervention details, e.g.: - Type of window treatment (film, decals) - Placement (inside/outside glass) - Installation timing relative to migration | Must clearly describe intervention details, e.g.: - Type of traffic barrier/mitigation |
| C. Is the outcome data sufficient to allow study quality assessment? | YES or NO <i>If NO, study is excluded from further assessment.</i> | Bird-specific outcomes must be reported separately | Bird collision or avoidance data must be quantitative and specific | Bird-specific outcomes must be reported |

(i.e., does the study clearly provide a direct empirical link between the intervention and outcome for birds, specifically?)

(e.g., number of birds caught, species). to intervention and control windows.

Study validity assessment

(risk of bias)

Selection Bias (study design)

| | | | | |
|---|---|---|---|---|
| 1.0. Is the study design well controlled? | HIGH (1) – BACI OR repeated-measures crossover design | A good cat study is BACI or repeated-measures crossover (cats act as own controls). | A good collision study is BACI, or RCT-like tunnel tests if applicable. | A good bird-vehicle collisions study would be BACI. Medium score for BA or CI |
| | MEDIUM (0.5) – BA or CI | | | |
| | LOW (0) – Correlative or anecdotal | | | |

| | | | | |
|---|--|--|--|--|
| | | | | Low score for correlative or anecdotal. |
| 1.1. Is there true replication in the study at the level of the intervention? <i>(i.e., pseudo-replication would be "NO")</i> | YES or NO <i>Where "YES" gives the study a score of "1" and "NO" gives the study a score of "0"</i> | True replication = multiple cats independently tested. | True replication = multiple windows or tunnels independently tested. For tunnel tests, replication at bird level. | Replication = multiple roads or multiple settings in case of trial tests Pseudoreplication = multiple road segments |
| 1.2. Are the comparator and intervention replicates/samples well-matched? | YES or NO or UNCLEAR | Matching or control for: - Outdoor access time | Matching or control for: - Window orientation and characteristics | Matching or control for: |

| | | | | |
|---|---|---|---|---|
| <i>(i.e., at the comparator and treatment samples similar at baseline with regards to confounding factors; otherwise, are these factors appropriately controlled for or addressed?)</i> | If UNCLEAR , treat as <i>NO</i> . Where “YES” gives the study a score of “1” and “NO” gives the study a score of “0” | - Cat age - Previous hunting experience Score 0 if factors reported but not randomized, controlled or statistically dealt with post-hoc. Score 0 if not reported/controlled. | including color and size/area Score 0 if reported but not controlled. Score 0 if unreported/uncontrolled. | Road setting (adjacent habitat type, surrounding landscape) |
| Assessment Bias | | | | |
| 2.0 Is the study consistent in its sampling method(s) between the intervention and comparator(s)? | YES or NO or UNCLEAR <i>Where “YES” gives the study a score of “1”;</i> | For example, prey measured similarly for both conditions (e.g., owner logs both | Collision data collected consistently for all windows (same observers, frequency). | Collision data collected consistently for all roads/road |

| | | | | |
|---|---|--|---|---|
| | <i>Where UNCLEAR, treat as NO; and “NO” gives the study a score of “0”</i> | intervention and control cats). | | segments (same observers, frequency). |
| 2.1 Does the study provide a measured quantitative outcome between the intervention and comparator? | YES or NO <i>In case of a qualitative outcome being provided, score of 0 is given 0.5 if a quantitative approximation is given only</i> | Quantitative counts of birds caught per cat or per time period. Qualitative or indirect estimates get 0. Score 0 if for whatever reason not all birds are included in the final outcome. | Quantitative counts of collisions or avoidance. | Quantitative counts of collisions or avoidance. |
| Performance Bias | | | | |
| 3.0 Is the study free of extenuating circumstances | YES or NO or UNCLEAR | | | |

| | | | | |
|---|---|--|---|--|
| following initiation of the study, or otherwise appropriately controlled or addressed? <i>(i.e., severe confounding factors such as weather events, unplanned human alterations; after study initiation)</i> | <i>If UNCLEAR, treat as NO</i> <i>If NO, describe extenuating circumstance (e.g., severe weather event, demolition of building shortening timespan of intervention tested).</i> <i>Where “YES” gives the study a score of “1” and “NO” gives the study a score of “0”</i> | | | |
| 3.1 Do the study settings of the intervention and comparator(s) match? | YES or NO or UNCLEAR | Intervention and control cats should have similar housing, | Intervention and control windows/buildings should be similar in | Intervention and control road settings should be |

| | | | | |
|--|--|--|--|--|
| <i>(i.e., do the comparator and intervention sites have similar characteristics; otherwise, were important confounding variables controlled by using randomization?)</i> (risk of confounding biases) | <i>Where “YES” gives the study a score of “1”,</i> <i>Where UNCLEAR, treat as NO; and “NO” gives the study a score of “0”.</i> | neighborhood, and outdoor conditions; differences should be dealt with via randomization or statistically. | exposure, surroundings, and season; differences should be dealt with via randomization or statistically. | similar (adjacent habitat type, surrounding landscape) |
|--|--|--|--|--|

Attrition Bias

| | | | | |
|---|---|---|----|----|
| 4.1. (For cat predation only) Were all cats originally enrolled in the study accounted for in the outcome data, or were missing data properly addressed? | YES or NO or UNCLEAR <i>Where “YES” gives the study a score of “1”,</i> <i>Where UNCLEAR, treat as NO; and “NO” gives the study a score of “0”.</i> | Studies should describe whether cats had to be withdrawn during trials and what the resulting final sample sizes were for treatment and control | NA | NA |
|---|---|---|----|----|

groups. Studies should
have appropriate
statistical adjustments
to deal with missing
data.

1020

1021

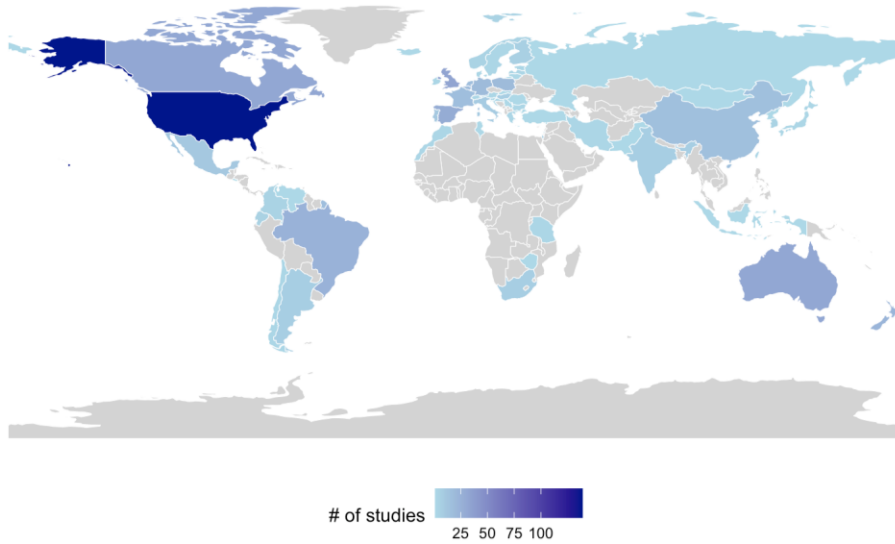
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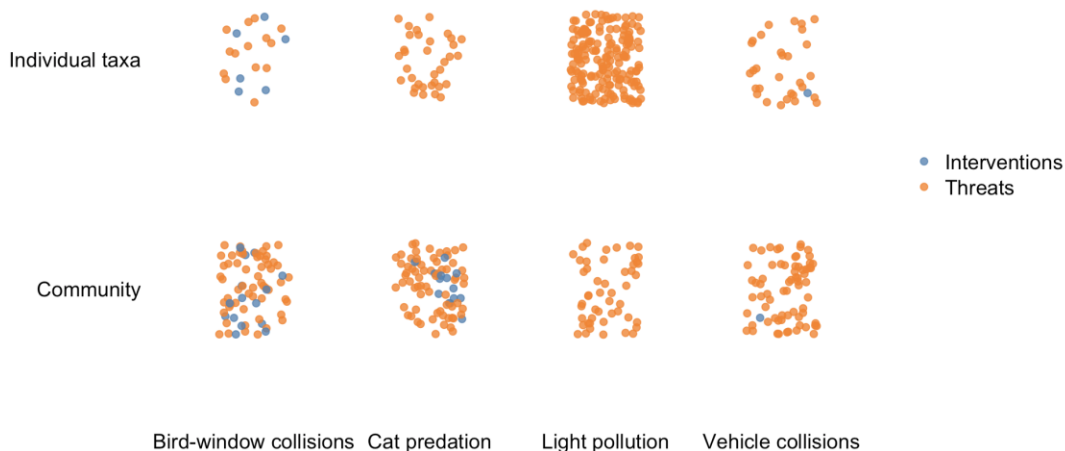


1027

1028 **Supplementary Material Figure S1.** Choropleth map showing the geographic distribution of
1029 533 studies in a global evidence synthesis of peer-reviewed literature on key urban threats to bird
1030 populations and their interventions. Darker blue shading represents regions with a higher density
1031 of studies.

1032

1033



1034

1035 **Supplementary Material Figure S2.** Distribution of 533 papers based on whether research
 1036 targeted individual taxa or bird communities. Each point represents an individual paper and is
 1037 colored based on whether the study was categorized as a threat-based paper or an intervention-
 1038 based paper.

1039

1040 **Supplementary Material Table S5.** Summary of vote-count of 11 types of cat predation
 1041 interventions across 13 papers showing the percentage of studies reporting statistically
 1042 significant positive results (i.e., a significant reduction in birds brought home by cats following
 1043 intervention), non-significant results, and cases where statistical significance is not provided.
 1044 Statistical significance is considered at $p < 0.05$.

| Type of Intervention | # of Papers | # of Studies | Statistically significant positive results (%) | Statistically non-significant results (%) | Statistical significance not provided (%) |
|----------------------|-------------|--------------|--|---|---|
|----------------------|-------------|--------------|--|---|---|

| | | | | | |
|----------------|---|---|-------|-----|-------|
| Rainbow | 5 | 6 | 100 | 0 | 0 |
| Birdsbesafe | | | | | |
| Mixed | 2 | 3 | 66.67 | 0 | 33.33 |
| Interventions | | | | | |
| CatBib | 1 | 2 | 100 | 0 | 0 |
| Collar with | 2 | 2 | 100 | 0 | 0 |
| Sonar Device | | | | | |
| High-protein | 1 | 1 | 100 | 0 | 0 |
| food | | | | | |
| Collar with | 5 | 5 | 60 | 40 | 0 |
| Bell | | | | | |
| Red Birdbesafe | 1 | 1 | 0 | 100 | 0 |
| Yellow | 1 | 1 | 0 | 100 | 0 |
| Birdbesafe | | | | | |
| Puzzle feeder | 1 | 1 | 0 | 100 | 0 |
| Object play | 1 | 1 | 0 | 100 | 0 |
| Predator | 1 | 1 | 0 | 100 | 0 |
| Removal | | | | | |

1045

1046 **Supplementary Material Table S6.** Pooled effect estimates from multi-level mixed effects
1047 model with Hedges' g effect of outcomes of collar-based interventions to reduce cat predation
1048 from 13 studies with intervention type as a moderator. The number of asterisks (*) indicates the

1049 degree of significance (i.e., “****” corresponds to $p < 0.001$, “***” corresponds to $p < 0.01$, and
 1050 “**” corresponds to $p < 0.05$).

| Intervention Type | Estimate ± SE | p-value |
|--------------------------|----------------------|----------------|
| Rainbow BBS | 0.46 ± 0.13 | <0.05**** |
| Catbib with Bell | 0.72 ± 0.23 | <0.05** |
| Bell | 0.41 ± 0.15 | <0.05** |
| Sonar Device | 0.40 ± 0.20 | <0.05* |
| Ranbow BBS with Bell | 0.19 ± 0.37 | 0.61 |
| Red BBS | 0.15 ± 0.26 | 0.58 |
| Yellow BBS | -0.13 ± 0.34 | 0.69 |

1051

1052 **Supplementary Material Table S7.** Summary vote-count of effectiveness of 15 types of bird-
 1053 window collision interventions across 20 papers showing the percentage of studies reporting
 1054 statistically significant positive results (i.e., a significant reduction in recorded bird-window
 1055 collisions following intervention), non-significant results, and cases where statistical significance
 1056 is not provided. Statistical significance is considered at $p < 0.05$.

| Type of Intervention | # of Papers | # of Field Studies | # of Tunnel Test Studies | Statistically Significant Positive Results (%) | Statistically Non-significant Results (%) | Significance Not Provided (%) |
|-----------------------------|--------------------|---------------------------|---------------------------------|---|--|--------------------------------------|
| Circular decals | 9 | 8 | 2 | 90 | 0 | 10 |

| | | | | | | |
|---|---|----|---|-------|-------|-------|
| UV reflective coating or film | 8 | 7 | 5 | 91.67 | 8.33 | 0 |
| Horizontal stripes | 4 | 1 | 4 | 100 | 0 | 0 |
| Vertical stripes | 4 | 1 | 5 | 83.33 | 16.67 | 0 |
| Miscellaneous window artwork and artifacts | 4 | 3 | 4 | 57.14 | 42.86 | 0 |
| Fritted glass (circular) | 3 | 3 | 0 | 66.67 | 33.33 | 0 |
| Grid | 3 | 0 | 8 | 37.5 | 62.5 | 0 |
| Bird-of-prey decals | 3 | 3 | 2 | | 80 | 20 |
| Vertical shades | 1 | 1 | | 100 | 0 | 0 |
| Square decals | 1 | 0 | 2 | 100 | | 0 |
| Orange film | 1 | | 2 | 50 | 50 | 0 |
| Miscellaneous objects placed near window | 1 | 3 | 3 | 16.67 | 83.33 | 0 |
| UV absorbing film | 1 | 12 | | 16.67 | | 83.33 |

| | | | | | | |
|-----------------------------|---|---|---|----|----|-----|
| Fritted glass (vertical) | 1 | 1 | 0 | 0 | 0 | 100 |
| Lighting conditions | 1 | | 2 | 50 | 50 | 0 |

1057

1058 **Supplementary Material Table S8.** Pooled effect estimates from multi-level mixed effects
 1059 model with odds ratio effect on outcomes of window treatments on bird-window collisions from
 1060 36 studies with intervention type as a moderator. “Other” refers to non-geometric coral design,
 1061 single UV maple leaf in center, UV maple leaf rows, monofilament covered feathers.

| Intervention Type | Estimate ± SE | p-value |
|----------------------------|--------------------------|----------------|
| Vertical stripes | -3.92 ± 0.57 | <.0001 |
| Circle or square decals | -3.80 ± 0.60 | <.0001 |
| Horizontal stripes | -2.15 ± 0.56 | 0.0001 |
| Other | -2.04 ± 0.63 | 0.0013 |
| UV reflecting film | -1.35 ± 0.56 | 0.0166 |
| Fritted glass | -1.38 ± 0.59 | 0.0199 |
| Orange film | -0.55 ± 1.03 | 0.5895 |
| Grid | 0.74 ± 0.62 | 0.2322 |

1062

1063

1064 **Supplementary Material Table S9.** Results from tests of overall heterogeneity from 13 effect
 1065 estimates across seven studies in a meta-analysis of interventions to reduce predation of birds by
 1066 cats.

| Test | Statistic |
|-----------------------------------|---------------------------|
| QM (df, p- value) | 32.39 (df = 7), p < 0.001 |
| QE (df, p-value) | 7.41 (df = 6), p = 0.284 |
| τ^2 (between-study variance) | 0.0156 |
| τ^2 (within-study variance) | 0.000 |
| I ² | 13.93% |

1067

1068 **Supplementary Material Table S10.** Results from tests of heterogeneity within each
 1069 intervention type from meta-analysis of interventions to reduce predation of birds by cats.
 1070 Heterogeneity statistics are only reported for intervention types with more than one effect
 1071 estimate.

| Intervention type | # of effect estimates | QE | τ^2 | I ² (%) |
|-------------------|-----------------------|------------------|----------|--------------------|
| Bell | 3 | 3.58 (p = 0.167) | 0.059 | 46.26 |
| Rainbow | 5 | 3.84 (p = 0.429) | 0.000 | 0.01 |

Birdbesafe™

1072

1073 **Supplementary Material Table S11.** Results from tests of overall heterogeneity from 36 effect
 1074 estimates across eight studies in a meta-analysis of interventions to reduce bird-window
 1075 collisions.

| Test | Statistic |
|------|-----------|
|------|-----------|

| | |
|-----------------------------------|-----------------------------|
| QM | 255.77 (df = 8), p < 0.001 |
| QE | 340.59 (df = 28), p < 0.001 |
| τ^2 (between-study variance) | 1.876 |
| τ^2 (within-study variance) | 2.721 |
| I ² | 74.96% |

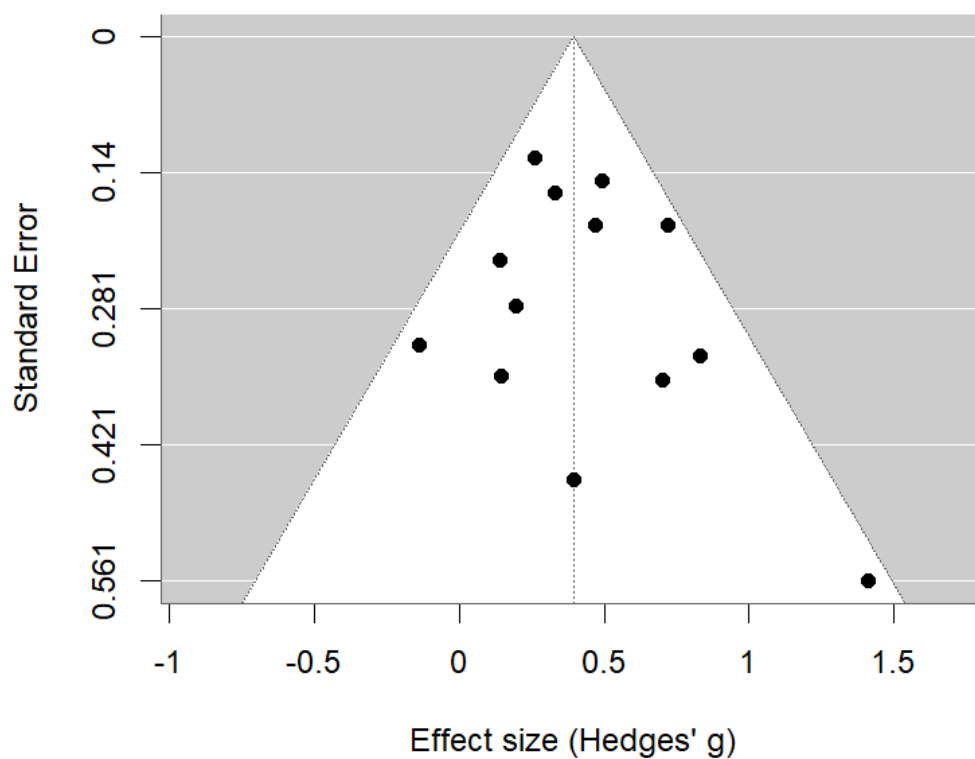
1076

1077 **Supplementary Material Table S12.** Results from heterogeneity tests within each of the eight

1078 intervention types from meta-analysis of interventions to reduce bird-window collisions.

| Intervention Type | # of effect estimates | QE | τ^2 | I² (%) |
|--------------------------|------------------------------|-------------------|----------------------------|--------------------------|
| Vertical Stripes | 6 | 87.82 (p < 0.001) | 4.341 | 94.31 |
| Circle or Square Decals | 7 | 37.70 (p < 0.001) | 6.141 | 84.08 |
| Horizontal Stripes | 4 | 20.58 (p < 0.001) | 1.759 | 85.42 |
| UV Reflecting Film | 8 | 25.06 (p < 0.001) | 0.806 | 72.07 |
| Fritted Glass | 3 | 42.12 (p < 0.001) | 5.529 | 95.25 |
| Orange Film | 2 | 2.87 (p = 0.090) | 1.576 | 65.13 |
| Grid | 2 | 64.14 (p < 0.001) | 19.288 | 98.44 |
| Other | 4 | 60.31 (< 0.001) | 8.585 | 95.02 |

1079



1080

1081 **Supplementary Material Figure S3.** Relationship between effect estimates and standard error

1082 to assess publication bias in meta-analysis of interventions to reduce cat predation of birds.

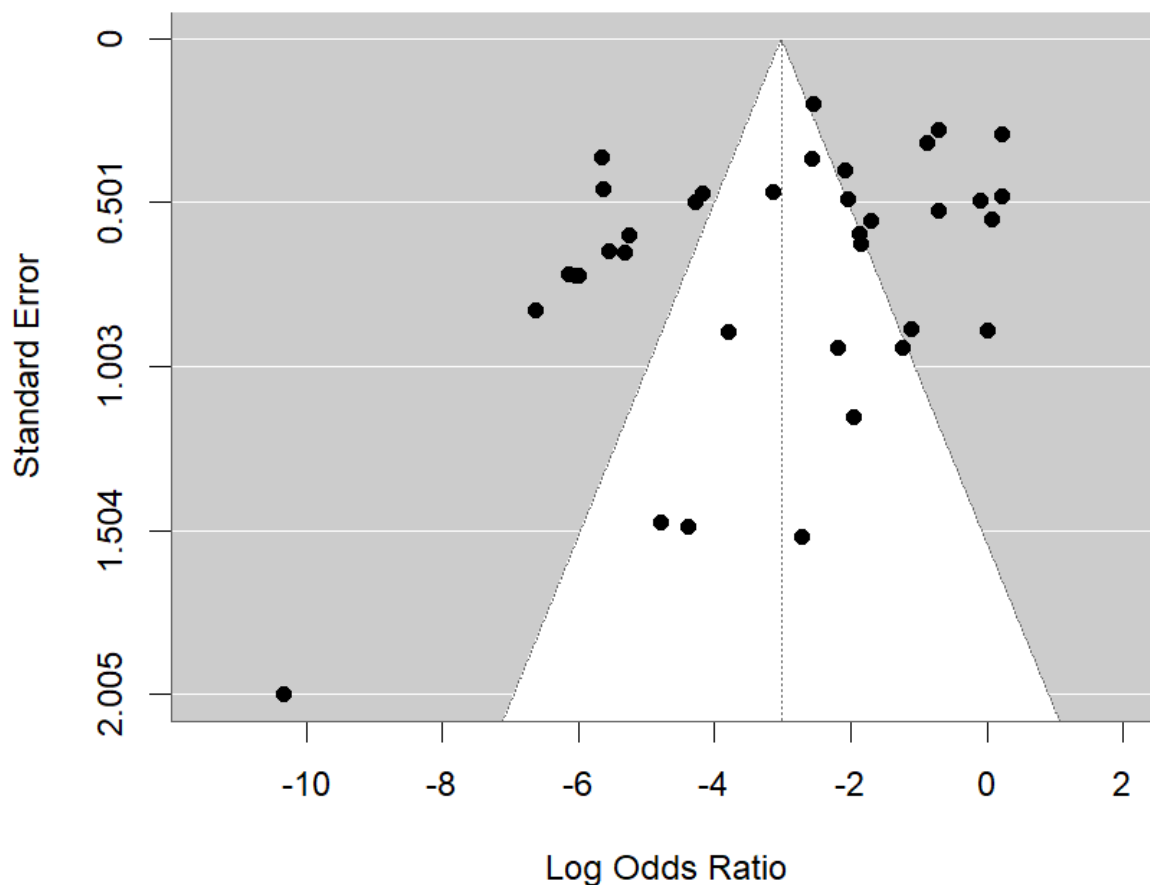
1083 Dashed line is the summary mean-weighted effect estimate from meta-analysis across 13 effect

1084 estimates from seven studies. Egger's test result was significant ($z = 0.83$, $p = 0.41$), suggesting

1085 there is no significant funnel plot asymmetry or publication bias. Results should be interpreted

1086 with caution due to the small sample size.

1087



1088
 1089 **Supplementary Material Figure S4.** Relationship between effect estimates and standard error
 1090 to assess publication bias in meta-analysis of interventions to reduce bird-window collisions.
 1091 Dashed line is the summary mean-weighted effect size from meta-analysis across 36 effect
 1092 estimates from eight studies. Egger's test result was significant ($z = -2.27$, $p = 0.02$), suggesting
 1093 there is significant asymmetry. Specifically, studies with higher standard errors tended to report
 1094 more negative effect estimates (i.e., greater effectiveness of an intervention). Results should be
 1095 interpreted with caution due to the small sample size.
 1096