

1 **Title: Comparative urban behaviour of two sympatric columbids: *Columba palumbus* and**  
2 ***Columba livia***

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11

12 **Abstract**

13 *Comparative urban behaviour of two sympatric columbids: Columba palumbus and Columba*  
14 *livia*. Urban environments favour species tolerant of human disturbance. We compared time  
15 budgets and anti-predator behaviour of the common wood pigeon (*Columba palumbus*) and rock  
16 pigeon (*Columba livia*) in five urban parks in León (Spain) using video recordings and predator  
17 approach tests. The common wood pigeon devoted most time to foraging (81.1%) with moderate  
18 vigilance (9.9%), while the rock pigeon showed higher movement (35.25%) and near-absent  
19 vigilance (2.95%). It escaped mainly by flying (84%), with greater flight initiation distances (9.47  
20 vs. 4.59 m) and distances to refuge (4.27 vs. 1.89 m). Group size reduced vigilance in the common  
21 wood pigeon but only movement in the rock pigeon. These findings show that the rock pigeon  
22 exhibits a more habituated, low-vigilance behavioural profile, whereas the common wood pigeon  
23 retains a more cautious anti-predator strategy, indicating contrasting responses to urban  
24 environments.

25 **Keywords:** common wood pigeon, rock pigeon, urban ecology, behavioural ecology,  
26 antipredatory behaviour, flight initiation distance.

27

## 28 **Introduction**

29       Urbanisation is one of the most pervasive and rapidly expanding forms of land-use change  
30 globally, fundamentally transforming natural habitats into highly modified environments  
31 characterised by increased human presence, altered resource availability, novel predation  
32 pressures, and chronic anthropogenic disturbance (McKinney, 2002; Grimm et al., 2008). These  
33 conditions impose strong selective pressures on wildlife, acting as an ecological filter that favours  
34 individuals capable of tolerating or exploiting the urban matrix while excluding species with  
35 narrower behavioural or ecological tolerances (Blair, 1996; Sol et al., 2005). As a consequence,  
36 urban animal communities tend to be dominated by a subset of behaviourally flexible species —  
37 the so-called urban adapters — while a large proportion of taxa retreat to less disturbed habitats or  
38 experience local extinction (McKinney, 2002; Fontaine & Martin, 2006).

39       Among vertebrates, birds have emerged as a model system for studying the behavioural  
40 consequences of urbanisation, given their diversity, long-standing scientific interest, and the  
41 extensive comparative data available across urban gradients (Marzluff, 2001; Chace & Walsh,  
42 2006). A central topic in urban avian ecology is the degree to which species modify their behaviour  
43 in response to the predictable, non-lethal nature of human disturbance, a process broadly referred  
44 to as habituation (Fernández-Juricic & Jokimäki, 2001). This process is especially evident in anti-  
45 predator behaviour: urban birds consistently exhibit reduced escape distances, lower investment  
46 in vigilance, and greater tolerance of human approach compared to their rural conspecifics,  
47 reflecting a recalibration of perceived predation risk in environments where direct predation threat

48 is reduced (Møller, 2008; Vincze et al., 2016). This reduction in fear responses is one of the clearest  
49 examples of urban behavioural convergence, operating both within species across populations and  
50 among phylogenetically distant species sharing the same urban environment, and is now  
51 recognised as part of a broader process of behavioural homogenisation driven by urbanisation  
52 (Mikula et al., 2026). However, these adjustments may also carry costs, as reduced vigilance and  
53 altered escape strategies may increase vulnerability in contexts where predation pressure remains  
54 high, such as in less urbanised or natural environments. (Lima & Dill, 1990).

55         Beyond anti-predator behaviour, urbanisation also reshapes time budgets and foraging  
56 strategies. The abundance of anthropogenic food subsidies, the spatial predictability of resources,  
57 and the reduction in interspecific competition in urban environments can profoundly alter the  
58 proportion of time individuals devote to foraging, movement, and social interactions (Shochat et  
59 al., 2006). These shifts in behavioural allocation are not uniform across species. Instead, they are  
60 expected to vary according to each species' evolutionary history and degree of association with  
61 human-modified environments, often described along the urban avoider–adapter continuum (Blair,  
62 1996; Croci et al., 2008). From the perspective of behavioural homogenisation, species that  
63 successfully colonise urban environments tend to converge towards similar behavioural profiles,  
64 characterised by increased boldness, reduced wariness, and opportunistic foraging strategies,  
65 regardless of their taxonomic identity or evolutionary background (Mikula et al., 2026). This  
66 convergence implies that sympatric species differing in their degree of urban association may  
67 occupy functionally distinct behavioural niches within the same habitat, reflecting their positions  
68 along the trajectory of urban behavioural adaptation.

69         To date, most studies on urban bird behaviour have focused on single species or on  
70 comparisons between urban and rural populations of the same taxon (e.g. Møller, 2008; Díaz et

71 al., 2013; Samia et al., 2015). Comparisons between sympatric species differing in their degree of  
72 urban association — and thus potentially occupying functionally distinct behavioural niches within  
73 the same habitat — remain comparatively scarce (Chace & Walsh, 2006; Croci et al., 2008). Such  
74 an approach offers the advantage of controlling for local environmental conditions while  
75 highlighting the role of evolutionary and ecological history in shaping behavioural responses to  
76 urbanisation (Møller, 2010; Díaz et al., 2013). Here, we present a comparative study of the  
77 behaviour of the common wood pigeon (*Columba palumbus*) and the rock pigeon (*Columba livia*)  
78 in urban parks of the city of León (Spain), examining both their time budgets and anti-predator  
79 strategies. These two species differ markedly in their history of association with humans: the rock  
80 pigeon has been domesticated and closely linked to human settlements for centuries (Johnston &  
81 Janiga, 1995; Giunchi et al., 2012), whereas the common wood pigeon has colonised urban  
82 environments more recently and retains a lower degree of synanthropy (Luna et al., 2024).  
83 Specifically, we aim to characterise and compare the behavioural profiles of the two species under  
84 comparable environmental conditions, assess interspecific differences in escape behaviour and  
85 flight initiation distance, and explore the relationships between movement, foraging, vigilance,  
86 and group size within each species. We predict that the rock pigeon, as the more synanthropic  
87 species, will invest less time in vigilance, exhibit shorter escape distances, and show a greater  
88 reliance on running as an escape strategy (Bocelli et al., 2022; Cooper & Blumstein, 2015),  
89 whereas the common wood pigeon will exhibit a more cautious behavioural profile consistent with  
90 a species at an earlier stage of urban adaptation (Blair, 1996; Møller, 2010).

91

## 92 **Material and methods**

### 93 **Study species**

94           The rock pigeon (*Columba livia*; Gmelin, 1789) is native to the rocky cliff habitats of  
95 Europe, North Africa, and South Asia. The Iberian Peninsula constitutes part of its original  
96 distribution range, where wild populations associated with coastal cliffs and inland gorges still  
97 persist, though their genetic integrity is increasingly compromised by introgression with feral  
98 individuals (Fernández-García 2022). Feral populations, derived from centuries of domestication  
99 and subsequent re-feralization, are now ubiquitous in Spanish cities, and León is no exception:  
100 feral rock pigeons are a permanent fixture of the city’s historic centre, plazas, and urban parks,  
101 exploiting the ledges and recesses of buildings as functional equivalents of their ancestral cliff  
102 habitats (Johnston & Janiga, 1995). Their long commensal history with humans has resulted in a  
103 fully synanthropic lifestyle, characterised by year-round breeding, opportunistic foraging on  
104 anthropogenic food subsidies, and a strong association with human infrastructure (Giunchi et al.,  
105 2012). The common wood pigeon (*Columba palumbus*; Linnaeus, 1758), by contrast, is  
106 traditionally a forest and agricultural species that has colonised urban parks and gardens across  
107 Spanish cities only in recent decades (Luna et al., 2024). In León, common wood pigeons are now  
108 a regular presence in the city’s main parks, where they forage and breed year-round, though their  
109 urban use remains largely restricted to green spaces with sufficient tree cover. Unlike the rock  
110 pigeon, the wood pigeon’s foraging strategy relies primarily on tree-derived resources rather than  
111 human food waste, and its urban colonisation in Spain is recent enough that it retains ecological  
112 traits more typical of woodland populations (Díaz et al., 2013; Luna et al., 2024). This contrast in  
113 urban history and degree of association with human settlements makes the two species a  
114 particularly suitable pair for comparative behavioural research in urban environments.

115

## 116 **Study area**

117 The study was conducted in León, a medium-sized city in north-western Spain founded as  
118 a Roman settlement and currently inhabited by just over 120,000 people, with a continental  
119 Mediterranean climate characterised by cold winters and warm, relatively dry summers. Our study  
120 took place in five urban parks located in the city of León, Spain: Parque de la Granja (42°35'56"N  
121 5°34'01"W), Parque Virgen del Camino (42°35'31.521"N 5°33'29.802"W), Jardín del Chantre  
122 (42°35'21"N 5°34'14"W), Parque Dos Hermanas (42°35'20.953"N 5°33'59.556"W), and Jardín de  
123 San Francisco (42°35'36"N 5°34'18"W) (Figure 1). All study sites showed heterogeneous habitat  
124 composition, including different substrate types, such as stone pathways, sand, grass, and bare soil.  
125 This heterogeneity favours the presence of the target species, which are commonly found in these  
126 areas.

127

### 128 **Sampling protocol**

129 Data were collected between August 2024 and February 2026, and all sampling sessions  
130 were conducted between 10:00 and 13:00 h, coinciding with the main period of activity of both  
131 species. A total of 25 common wood pigeons (*Columba palumbus*) and 40 rock pigeons (*C. livia*)  
132 were sampled. Surveys were performed through active searching by a single observer throughout  
133 the study period. At the beginning of each sampling day, we recorded four variables to control for  
134 their potential effects on pigeon behaviour: (1) human activity, quantified over a 10-minute period  
135 using a mobile application and subsequently converted to persons per minute, (2) temperature (°C),  
136 (3) relative humidity (%), and (4) wind speed (m/s). The temperature was measured by a laser  
137 thermometer whereas relative humidity and wind speed were measured using an anemometer.

138 Upon locating an individual, a photograph was taken for identification purposes and to  
139 minimise pseudoreplication. Following this, focal observations were conducted using 6 min video

140 recordings, or until the individual fled or the recording was otherwise interrupted, for subsequent  
141 behavioural analysis (ethogram, Table 1). We used BORIS software (Friard & Gamba, 2016) to  
142 obtain the duration of each behaviour, in seconds. Once the video recording had ended, a predator  
143 approach test was performed to assess the flight initiation distance (hereafter, FID): the observer  
144 walked in a straight line towards the pigeon at approximately 0.5 m/s while maintaining direct eye  
145 contact (Blumstein, 2006). In addition, the starting distance — defined as the distance at which the  
146 observer initiated the approach, a variable known to influence FID (Blumstein, 2006) — was  
147 recorded. The escape mode was also noted, as urban birds may choose between running and flying  
148 as alternative escape strategies depending on the perceived level of threat (Bocelli et al., 2022).  
149 Furthermore, the distance to the nearest refuge was measured, as proximity to shelter may  
150 influence both general behaviour and anti-predator responses in these species.

151

## 152 **Statistical analyses**

153 Statistical analyses were performed using R software (version 4.3.2; R Core Team, 2023).  
154 All statistical tests were conducted at a significance level of  $\alpha = 0.05$ . When the data did not meet  
155 the assumptions of normality and homogeneity of variances, non-parametric tests were applied.  
156 Interspecific differences were assessed using the Mann-Whitney U test, both for general behaviour  
157 (e.g., time investment across the recorded behavioural categories) and anti-predator behaviour (e.g.  
158 starting distance, FID, distance to the nearest refuge). A Chi-square test ( $\chi^2$ ) was used to assess  
159 interspecific differences in escape strategy (flying vs. running). Linear regressions were used to  
160 examine relationships between continuous behavioural and environmental variables.

161

## 162 **Results**

163 No significant interspecific differences were found in any of the environmental variables recorded  
164 during behavioural observations, including human presence ( $p = 0.761$ ), temperature ( $p = 0.716$ ),  
165 relative humidity ( $p = 0.299$ ), and wind speed ( $p = 0.695$ ).

166 The behavioural time budgets of the two species differed markedly (Figure 2). The  
167 common wood pigeon devoted most of its active time to foraging (median = 81.10%), followed  
168 by vigilance (median = 9.90%) and movement (median = 5.00%). In contrast, the rock pigeon  
169 allocated a considerably higher proportion of time to movement (median = 35.25%) and a lower  
170 proportion to foraging (median = 42.80%) and vigilance (median = 2.95%). Additional behaviours  
171 such as preening, resting, aggression, reproduction, and short escape flights were recorded in both  
172 species, though they collectively represented a minor fraction of the total time budget (Figure 2).  
173 Interspecific differences were statistically significant for all three main behavioural categories:  
174 movement (Mann-Whitney U test,  $U = 225.5$ ,  $p < 0.001$ ), foraging ( $U = 801.5$ ,  $p < 0.001$ ), and  
175 vigilance ( $U = 708$ ,  $p = 0.005$ ).

176 Regarding escape behaviour, the two species differed significantly in their choice of escape  
177 strategy ( $\chi^2 = 17.5$ ,  $p < 0.001$ ; Figure 3). The common wood pigeon predominantly escaped by  
178 flying (84%,  $n = 21$ ), with only 16% of individuals ( $n = 4$ ) choosing to run. The pattern was  
179 reversed in the rock pigeon, which escaped primarily by running (72%,  $n = 29$ ), while only 28%  
180 of individuals ( $n = 11$ ) took flight. Within the common wood pigeon, individuals that escaped by  
181 running were encountered under significantly higher levels of human presence than those that  
182 escaped by flying (median running = 10.2 pers/min vs. median flying = 3.6 pers/min; Mann-  
183 Whitney U test,  $U = 11$ ,  $p = 0.023$ ), and showed lower vigilance prior to escape (median running =  
184 2.8% vs. median flying = 12.8%; Mann-Whitney U test,  $U = 70$ ,  $p = 0.041$ ). No significant  
185 differences in FID were found between escape types in this species (Mann-Whitney U test,  $U = 63$ ,

186  $p = 0.132$ ). Within the rock pigeon, individuals that escaped by flying showed significantly greater  
187 FID than those that ran (median flying = 7.1 m vs. median running = 4.15 m; Mann-Whitney U  
188 test,  $U=253$ ,  $p = 0.005$ ), suggesting a higher perceived threat in individuals that chose to fly. No  
189 significant differences in human presence or vigilance were detected between escape types in the  
190 rock pigeon (Mann-Whitney U test,  $U=188$ ,  $p = 0.396$  and  $U=135.5$ ,  $p = 0.476$ , respectively).

191 With respect to anti-predator escape distances, the common wood pigeon exhibited  
192 significantly greater values than the rock pigeon across all three measured variables (Figure 4).  
193 Starting distance was significantly higher in the common wood pigeon (median = 15.77 m) than  
194 in the rock pigeon (median = 12.08 m; Welch's t-test,  $t=2.78$ ,  $p = 0.009$ ). FID also differed  
195 markedly between species (common wood pigeon: median = 9.47 m vs. rock Pigeon: median =  
196 4.59 m; Mann-Whitney U test,  $U=796$ ,  $p < 0.001$ ), as did the distance to the nearest refuge at the  
197 moment of escape (common wood pigeon: median = 4.27 m vs. rock pigeon: median = 1.89 m;  
198 Mann-Whitney U test,  $U=745$ ,  $p = 0.001$ ).

199 In both species, movement was negatively associated with foraging time (Figure 5A):  
200 common wood pigeon ( $\rho = -0.541$ ,  $R^2 = 0.651$ ,  $p = 0.005$ ) and rock Pigeon ( $\rho = -0.551$ ,  $R^2 =$   
201  $0.362$ ,  $p < 0.001$ ). The relationship between movement and vigilance, however, differed between  
202 species (Figure 5B): no significant association was found in the common wood pigeon ( $\rho = -0.019$ ,  
203  $p = 0.929$ ), whereas in the rock Pigeon a weak but significant negative relationship was detected  
204 ( $\rho = -0.359$ ,  $R^2 = 0.070$ ,  $p = 0.023$ ).

205 Group size had contrasting effects on the behavioural budget of the two species. In the  
206 common wood pigeon, larger group size was positively associated with foraging time ( $\rho = 0.556$ ,  
207  $R^2 = 0.192$ ,  $p = 0.004$ ) and negatively associated with vigilance ( $\rho = -0.502$ ,  $R^2 = 0.156$ ,  $p = 0.011$ ).  
208 No significant relationship between group size and movement was found in this species ( $p =$

209 0.182). In the rock pigeon, group size was negatively associated with movement ( $\rho = -0.363$ ,  $R^2$   
210  $= 0.093$ ,  $p = 0.021$ ), while no significant effects on foraging or vigilance were detected ( $p = 0.120$   
211 and  $p = 0.199$ , respectively).

212

## 213 **Discussion**

214 This study reveals marked behavioural divergence between two sympatric columbids  
215 occupying the same urban parks in León (Spain). Since they were sampled under similar human  
216 presence and environmental conditions, these differences are likely due to their contrasting degrees  
217 of synanthropy.

218 With respect to the time budget, the common wood pigeon devoted the majority of its  
219 active time to foraging, with moderate time for vigilance and limited time for locomotion, which  
220 is consistent with a stationary, resource-exploitation strategy typical of species still undergoing  
221 urban adaptation (Whittingham & Evans, 2004). The rock pigeon showed the opposite profile,  
222 characterised by high mobility, reduced foraging, and near-absent vigilance, consistent with a fully  
223 synanthropic species that has effectively decoupled alertness from predation risk assessment  
224 (Johnston & Janiga, 1995). This pattern has been previously documented in other urban-tolerant  
225 passerines such as *Passer domesticus* and *Sturnus vulgaris* (Fernández-Juricic et al., 2002; Lima  
226 & Dill, 1990). Together, these differences suggest that the two species occupy distinct behavioural  
227 positions along the urban adaptation continuum.

228 Regarding escape strategy, the two species diverged significantly in their choice of  
229 response: the common wood pigeon escaped predominantly by flight, whereas the rock pigeon  
230 relied mainly on running, reflecting differences in habitat use and perceived threat level consistent  
231 with optimal escape theory, which predicts that prey should select the escape response that

232 maximises survival while minimising the costs of fleeing (Ydenberg & Dill, 1986; Tryjanowski et  
233 al., 2016). Escape distances further corroborated interspecific differences in perceived risk:  
234 common wood pigeons exhibited significantly greater starting distances (15.77 vs. 12.08 m), FID  
235 (9.47 vs. 4.59 m), and distances to the nearest refuge at the moment of escape (4.27 vs. 1.89 m),  
236 consistent with the economic escape model, according to which escape decisions depend on the  
237 balance between the costs of fleeing and the perceived risk of predation (Blumstein, 2006; Cooper  
238 & Blumstein, 2015) and indicative of a higher baseline threat threshold. By contrast, the markedly  
239 shorter FIDs of the rock pigeon reflect a substantially lowered threat threshold following long-  
240 term habituation to human disturbance, as widely reported across urban bird communities (Møller,  
241 2008; Vincze et al., 2016). Overall, these patterns indicate that the common wood pigeon maintains  
242 a more cautious anti-predator strategy, whereas the rock pigeon exhibits a more relaxed response  
243 to potential threats in urban environments.

244         Regression analyses between movement and behavioural categories revealed the obvious  
245 trade-off between locomotion and foraging in both species, reflecting the incompatibility of  
246 simultaneous movement and feeding. However, movement was negatively associated with  
247 vigilance only in the rock pigeon, suggesting a trade-off between locomotion and alertness in this  
248 species. By contrast, the common wood pigeon appears to maintain vigilance independently of  
249 movement, possibly through intermittent scanning during resource searching (Beauchamp, 2004).

250         Finally, group size effects diverged markedly between species: in the common wood  
251 pigeon, larger groups were associated with increased foraging and reduced vigilance. This aligns  
252 with the many-eyes hypothesis and the anti-predator dilution effect, whereby individuals in larger  
253 groups benefit from increased collective vigilance and a reduced probability of being targeted by  
254 predators (Pulliam, 1973; Caraco et al., 1980), as expected in a species that relies on group-based

255 risk reduction. In the rock pigeon, group size reduced movement but had no effect on foraging or  
256 vigilance, suggesting that aggregations in this species reflect resource exploitation dynamics rather  
257 than anti-predator coordination (Johnston & Janiga, 1995), which is consistent with a species that  
258 has largely decoupled social grouping from predation risk management. Collectively, these results  
259 indicate that group formation plays different functional roles in the two species, depending on their  
260 degree of adaptation to urban environments.

261       Taken as a whole, these findings support the urban adapter–avoider framework (Blair,  
262 1996; McKinney, 2002) and reinforce the view that urbanisation acts as a selective filter on  
263 behaviour, favouring species capable of adjusting foraging and escape strategies to the chronic,  
264 predictable nature of human disturbance in cities (Sol et al., 2005). More specifically, our results  
265 show that even closely related species coexisting in the same urban habitat may differ markedly in  
266 their behavioural responses according to their degree of synanthropy and history of association  
267 with humans. This study therefore highlights the value of comparative within-habitat approaches  
268 for understanding how urban environments shape behavioural divergence and adaptation in  
269 wildlife. Extending this approach across species and urban contexts will help clarify how  
270 differences in synanthropy shape behavioural responses to urbanisation and ultimately determine  
271 species success in cities.

272

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279

### 280 **CRedit authorship contribution statement**

281 Iñigo Navidad: Data acquisition, Data curation, Formal analysis, Investigation, Visualization,  
282 Writing – original draft. Zaida Ortega: Conceptualization, Formal analysis, Supervision, Writing  
283 – review & editing.

284

### 285 **Declaration of Generative AI used in the writing process**

286 During the preparation of this manuscript, the authors used ChatGPT and Claude to improve the  
287 English language and grammar. The authors reviewed and edited the output as necessary and take  
288 full responsibility for the final version of the manuscript.

289

### 290 **References**

291 Beauchamp G, 2004. Reduced flocking by birds on islands with relaxed predation: a meta-analysis.  
292 *Proceedings of the Royal Society B: Biological Sciences* 271, 1039-1042.

293 Blair RB, 1996. Land use and avian species diversity along an urban gradient. *Ecological*  
294 *Applications* 6, 506-519.

295 Blumstein DT, 2006. Developing an evolutionary ecology of fear: How life history and natural  
296 history traits affect disturbance tolerance in birds. *Animal Behaviour* 71, 389-399.

297 Bocelli ML, Morelli F, Benedetti Y, Leveau L, 2022. Escape strategies of birds in urban  
298 environments. *El Hornero* 37, 75-85.

299 Caraco T, Martindale S, Pulliam HR, 1980. Avian flocking in the presence of a predator. *Nature*  
300 285, 400-401. DOI: 10.1038/285400a0

301 Chace JF, Walsh JJ, 2006. Urban effects on native avifauna: a review. *Landscape and Urban*  
302 *Planning* 74, 46-69.

303 Cooper WE, Blumstein DT (Eds.), 2015. *Escaping from Predators: An Integrative View of Escape*  
304 *Decisions*. Cambridge University Press, Cambridge.

305 Croci S, Butet A, Clergeau P, 2008. Does urbanization filter birds on the basis of their biological  
306 traits? *The Condor* 110, 223-240.

307 Díaz M, Møller AP, Flensted-Jensen E, Grim T, Ibáñez-Álamo JD, Jokimäki J, Markó G,  
308 Tryjanowski P, 2013. The geography of fear: a latitudinal gradient in anti-predator escape  
309 distances of birds across Europe. PLoS ONE 8, e64634.

310 Fernández-García JM, 2022. Paloma torcaz *Columba palumbus*. A: *III Atlas de las aves en época*  
311 *de reproducción en España* (B Molina, A Nebreda, AR Muñoz, J Seoane, R Real, J Bustamante,  
312 JC del Moral, Eds). SEO/BirdLife, Madrid.

313 Fernández-Juricic E, Jokimäki J, 2001. A habitat island approach to conserving birds in urban  
314 landscapes: cases and considerations. *Biodiversity and Conservation* 10, 841-861.

315 Fernández-Juricic E, Jimenez MD, Lucas E, 2002. Factors affecting intra- and inter-specific  
316 variations in the difference between alert distance and flight initiation distance for birds in  
317 forested habitats. *Canadian Journal of Zoology* 80, 1212-1220.

318 Fontaine JJ, Martin TE, 2006. Parent birds assess nest predation risk and adjust offspring  
319 development and fledging behaviour accordingly. *Proceedings of the Royal Society B* 273,  
320 2551-2557.

321 Friard O, Gamba M, 2016. BORIS: A free, versatile open-source event-logging software for  
322 video/audio coding and live observations. *Methods in Ecology and Evolution* 7, 1325-1330.

323 Giunchi D, Albores-Barajas YV, Baldaccini NE, Vanni L, Soldatini C, 2012. Feral pigeons:  
324 problems, dynamics and control methods. A: *Integrated Pest Management and Pest Control —*  
325 *Current and Future Tactics* (S Soloneski, Ed). InTech. DOI: 10.5772/32995

326 Grimm NB, Faeth SH, Golubiewski NE, Redman CL, Wu J, Bai X, Briggs JM, 2008. Global  
327 change and the ecology of cities. *Science* 319, 756-760.

328 Iannacone J, Villegas W, Calderón M, Huamán J, Silva-Santiesteban M, Alvariño L, 2012. Diurnal  
329 behavior patterns of the Peruvian Thick-knee *Hesperoburhinus superciliaris* in modified  
330 habitats of the central coast of Peru. *Acta Zoológica Mexicana (n.s.)* 28, 507-524.

331 Johnston RF, Janiga M, 1995. *Feral Pigeons*. Oxford University Press, Oxford.

332 Lima SL, Dill LM, 1990. Behavioral decisions made under the risk of predation: a review and  
333 prospectus. *Canadian Journal of Zoology* 68, 619-640.

334 Luna A, Pomedá-Gutiérrez F, Galán Díaz J, 2024. Feeding ecology of the common wood pigeon  
335 (*Columba palumbus*) in a major European city. *Royal Society Open Science* 11, 1231721. DOI:  
336 10.1098/rsos.231721

337 Marzluff JM, 2001. Worldwide urbanization and its effects on birds. A: *Avian Ecology and*  
338 *Conservation in an Urbanizing World* (JM Marzluff et al., Eds). Kluwer Academic Publishers,  
339 Boston.

340 McKinney ML, 2002. Urbanization, biodiversity, and conservation. *BioScience* 52, 883-890.

341 Mikula P, Blumstein DT, Tryjanowski P (in press). Behavioral convergence under urbanization:  
342 An overlooked dimension of biotic homogenization. *PLoS Biology*.

343 Møller AP, 2008. Flight distance of urban birds, predation, and selection for urban life.  
344 *Behavioural Ecology and Sociobiology* 63, 63-75.

345 Møller AP, 2010. Interspecific variation in fear responses predicts urbanization in birds.  
346 *Behavioural Ecology* 21, 365-371.

347 Pulliam HR, 1973. On the advantages of flocking. *Journal of Theoretical Biology* 38, 419-422.

348 R Core Team, 2023. *R: A language and environment for statistical computing*. R Foundation for  
349 Statistical Computing, Vienna.

350 Samia DSM, Møller AP, Weston MA, Guay PJ, Blumstein DT, Bessa-Gomes C, 2015.  
351 Urbanization and the drivers of avian escape distances across three continents. *PLoS ONE* 10,  
352 e0138929.

353 Shochat E, Warren PS, Faeth SH, McIntyre NE, Hope D, 2006. From patterns to emerging  
354 processes in mechanistic urban ecology. *Trends in Ecology & Evolution* 21, 186-191.

355 Sol D, Duncan RP, Blackburn TM, Cassey P, Lefebvre L, 2005. Big brains, enhanced cognition,  
356 and response of birds to novel environments. *Proceedings of the National Academy of Sciences*  
357 102, 5460-5465.

358 Tryjanowski P, Møller AP, Morelli F, Biaduń W, Brauze T, Ciach M, Czuchra M, 2016.  
359 Urbanization affects neophilia and risk-taking at bird feeders. *Scientific Reports* 6, 28575.

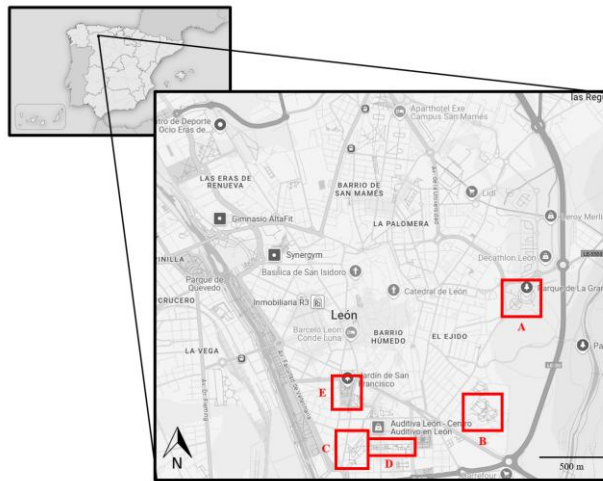
360 Vincze E, Papp S, Preiszner B, Seress G, Liker A, Bókony V, 2016. Habituation to human  
361 disturbance is faster in urban than rural house sparrows. *Behavioral Ecology* 27, 1304-1313.

362 Whittingham MJ, Evans KL, 2004. The effects of habitat structure on predation risk of birds in  
363 agricultural landscapes. *Ibis* 146, 210-220.

364 Ydenberg RC, Dill LM, 1986. The economics of fleeing from predators. *Advances in the Study of*  
365 *Behavior* 16, 229-249.

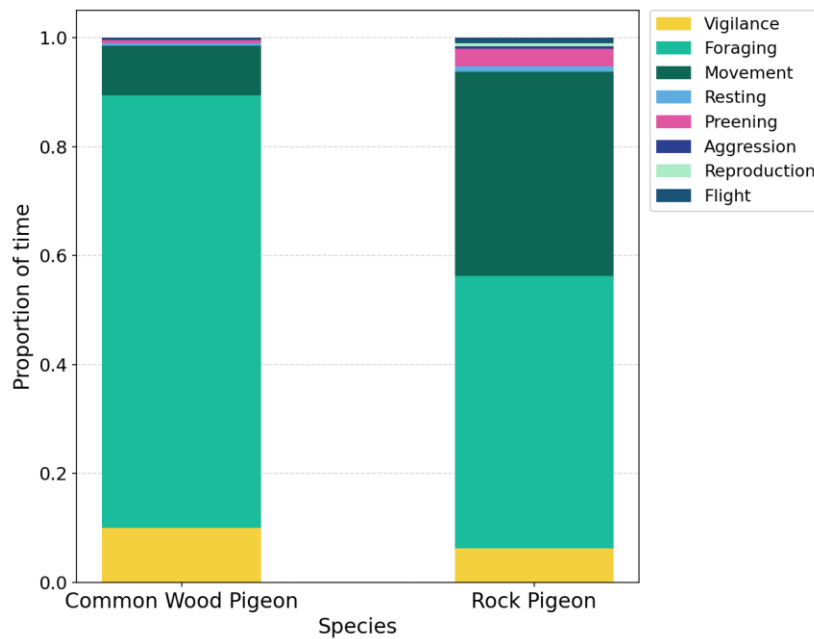
366 **Figures and tables**

367 **Figure 1:** Map of the sampling locations (León, Spain): (A) Parque de la Granja, (B) Parque  
368 Virgen del Camino, (C) Jardín del Chantre, (D) Parque Dos Hermanas and (E) Parque San  
369 Francisco.



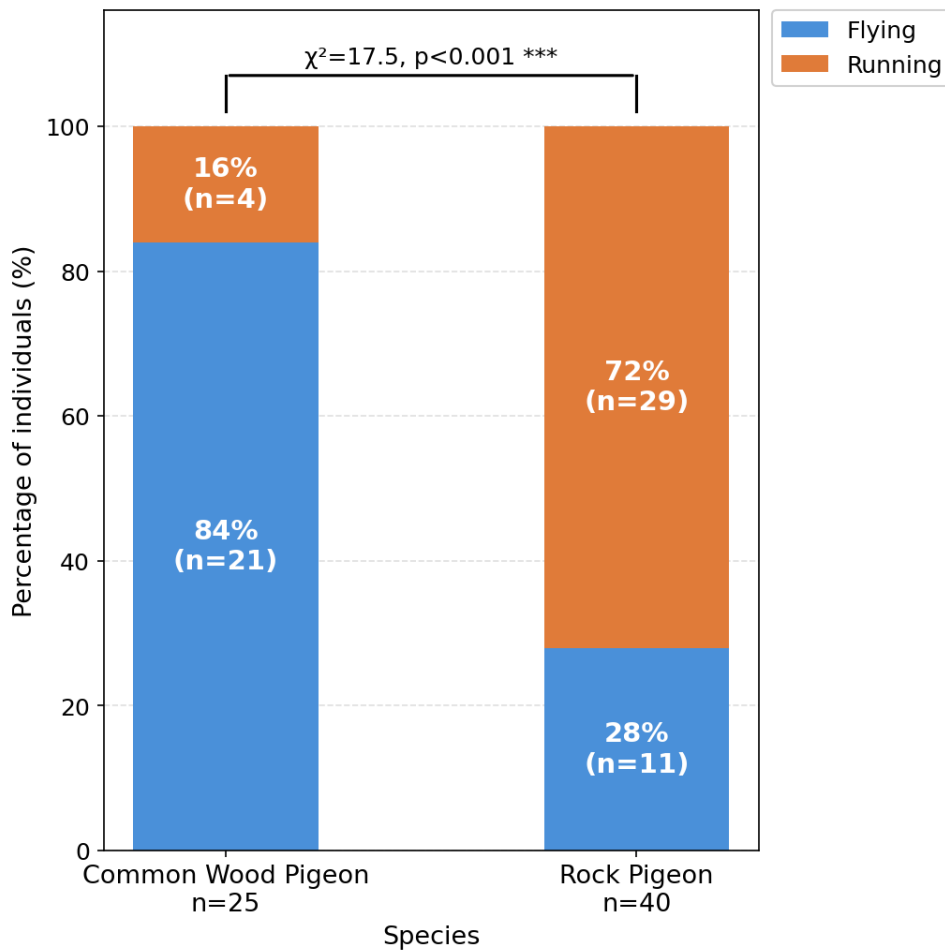
370

371 **Figure 2:** Proportion of time spent by each of the two pigeon species (*Columba palumbus* and *C.*  
372 *livia*) studied in green areas of León (Spain) performing each of the different behaviours (see  
373 definitions in the ethogram, Table 1).

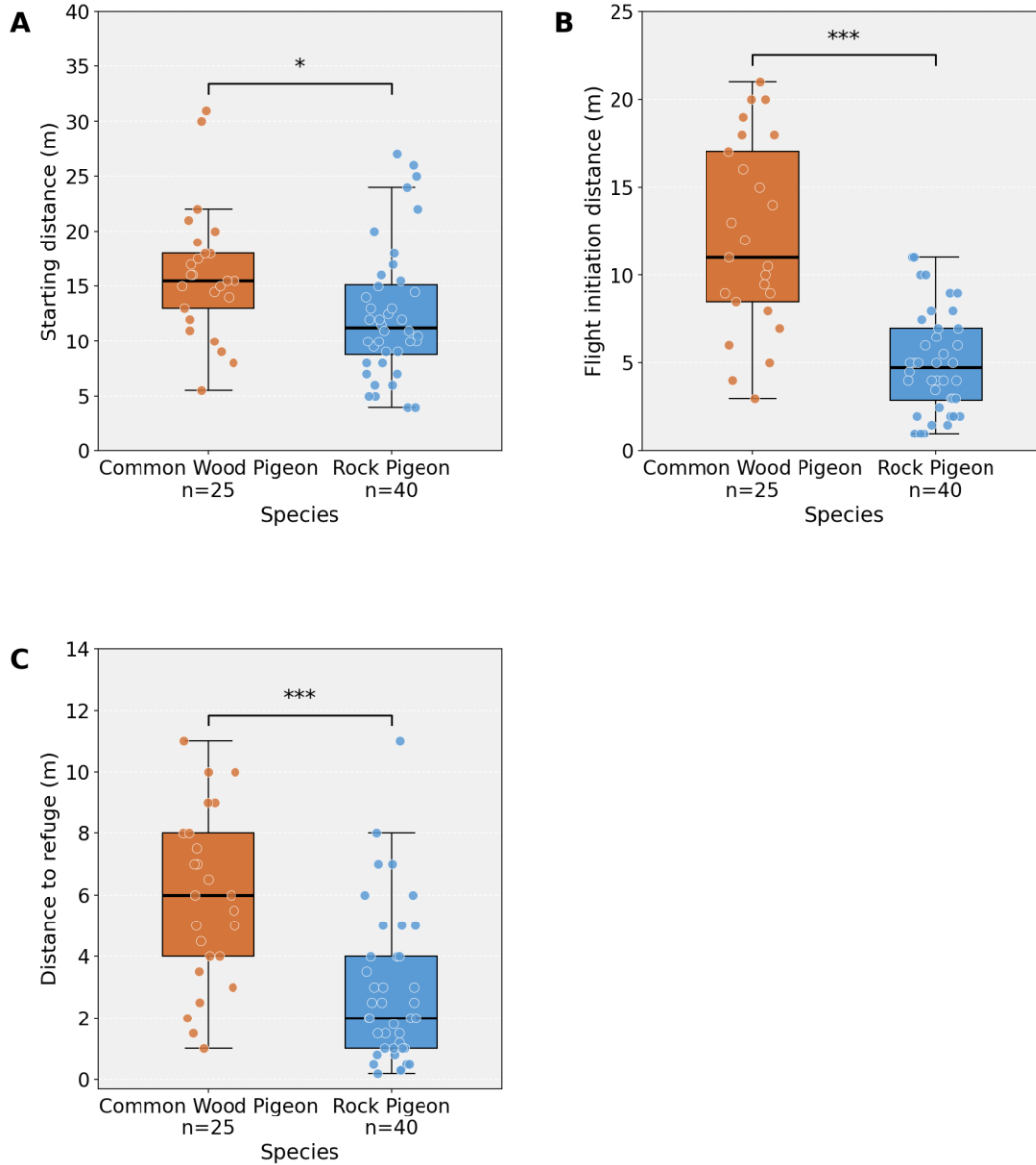


374

375 **Figure 3:** Escape strategy used by the common wood pigeon (*Columba palumbus*, n=25) and the  
 376 rock pigeon (*C. livia*, n=40) in urban green areas of León (Spain) expressed as the percentage of  
 377 individuals that chose flying (blue) or running (orange) as their escape response. Statistically  
 378 significant differences in escape strategy between species were detected using a Chi-square test  
 379 ( $\chi^2=17.5$ ,  $p<0.001$ ).



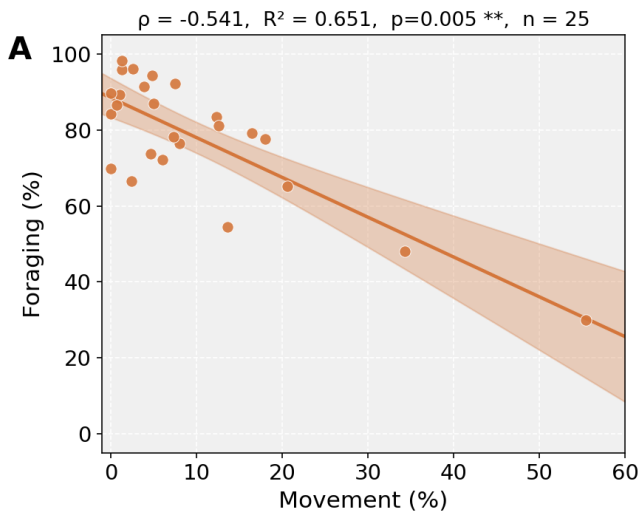
380  
 381 **Figure 4:** Comparison of escape distances between the common wood pigeon (*Columba*  
 382 *palumbus*, n=25) and the rock pigeon (*C. livia*, n=40) in urban green areas of León (Spain): starting  
 383 distance (A), flight initiation distance (FID; B), and distance to the nearest refuge (C). Significant  
 384 differences between species were assessed using the Mann-Whitney U test (\*  $p<0.05$ , \*\*  $p<0.01$ ,  
 385 \*\*\*  $p<0.001$ ).



386

387 **Figure 5:** Linear regression analyses between movement (%) and foraging (%) (A), and between  
 388 movement (%) and vigilance (%) (B), for the common wood pigeon (*Columba palumbus*, n=25,  
 389 orange) and the rock pigeon (*C. livia*, n=40, blue) in urban green areas of León, Spain. The solid  
 390 line represents the regression line and the shaded area indicates the 95% confidence interval.  
 391 Pearson's correlation coefficient ( $\rho$ ), coefficient of determination ( $R^2$ ), and significance level are  
 392 shown above each panel (n.s. = not significant, \*  $p < 0.05$ , \*\*  $p < 0.01$ , \*\*\*  $p < 0.001$ ).

Common Wood Pigeon



Rock Pigeon

