

1 **When does sampling uncertainty**
2 **matter in matrix population models?**
3 **Evidence from published projection**
4 **matrices**

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13

14 Abstract

- 15 1. The collation of thousands of population projection matrices in the COM(P)ADRE
16 Matrix Databases has enabled large-scale comparative analyses in ecology, evolution, and
17 demography. A persistent challenge is that transition rates are estimated from finite
18 samples, yet the resulting sampling uncertainty is rarely reported and typically ignored in
19 downstream analyses. Although sampling uncertainty is well recognised, it remains
20 unclear when it affects demographic inference.
- 21 2. Using a subset of studies from the COMPADRE Plant Matrix Database for which
22 transition-specific sample sizes could be obtained, we reconstructed sampling
23 distributions for transition rates and derived demographic parameters, and conducted
24 three related analyses: a comparative analysis, a within-species time-series analysis, and a
25 multi-site extension of the climate-demography analysis, each fit with and without
26 explicit sampling uncertainty.
- 27 3. The effects of sampling uncertainty were strongly trait-dependent. It had little influence
28 on integrated demographic metrics such as population growth rate, but accounted for a
29 substantial fraction of the total variation in traits related to mortality trajectories,
30 particularly the shape of the mortality distribution. In these cases, point estimates were
31 often systematically displaced from their underlying sampling distributions due to
32 boundary estimates of stage-specific survival (i.e. survival probabilities near 0 or 1),
33 which are common in published matrices. By contrast, inference in a within-species
34 analysis of weather effects on vital rates was largely unaffected.
- 35 4. Together, these analyses show that ignoring sampling uncertainty does not universally
36 compromise inference from published projection matrices, but can yield fragile or biased
37 conclusions for certain demographic traits. We identify structural features of MPMs and

38 analyses that may indicate heightened sensitivity to sampling uncertainty, and discuss
39 implications for researchers using compiled demographic databases.

40 Keywords: age-from-stage methods; parametric bootstrapping; estimation bias; parameter
41 uncertainty; stage-structured models; sensitivity analysis; plant demography; life history traits

42

43 **Introduction**

44 The collation of published demographic data in databases such as PanTHERIA (Jones et
45 al. 2009), AnAge (de Magalhães and Costa 2009), Myhrvold et al.'s (2015) amniote life-history
46 database, and COM(P)ADRE (Salguero-Gómez et al. 2015, Salguero-Gómez et al. 2016a) has
47 enabled fascinating comparative analyses in ecology, evolution, and demography (e.g. Garratt et
48 al. 2013, Adler et al. 2014, Healy et al. 2014, Fristoe et al. 2017). By leveraging open-source
49 databases that aggregate data from hundreds of individual studies, researchers can test
50 hypotheses of interest across taxonomic, spatial, and temporal scales that would otherwise be
51 impossible. However, analyses based on published demographic results rather than underlying
52 raw data also present their own challenges.

53 One challenge in working with large collections of published data is understanding and
54 modelling the sources of underlying uncertainty. For instance, the demographic parameters
55 reported in the databases noted above have almost always been estimated from a sample of
56 individuals drawn from a larger population of interest, and therefore carry inherent sampling
57 uncertainty. However, parameter-specific sample sizes and standard errors are rarely available
58 and are often difficult to obtain even from the original data sources. These original studies,
59 which produce the parameter estimates later compiled in demographic databases, usually account
60 for sampling uncertainty in their own analyses, but uncertainty is often not reported in a form
61 that can be propagated from the published matrices. Therefore, studies that subsequently analyse
62 parameter estimates compiled in demographic databases rarely incorporate this uncertainty,
63 raising questions about the robustness of resulting inferences (Caswell 2001, Ch. 12; Biggs et
64 al. 2009).

65 Despite widespread recognition that sampling uncertainty is inherent in demographic estimates
66 (Caswell 2001), a key unresolved question concerns when this uncertainty materially affects
67 inference from compiled databases. Comparative analyses using projection matrices rely on a

68 diverse set of derived demographic traits, ranging from highly integrated metrics such as
69 population growth rate to traits that depend on long-term survivorship or tail behaviour, such as
70 life expectancy or the shape of mortality trajectories. Whether these different classes of traits are
71 equally sensitive to sampling uncertainty under typical data regimes remains poorly understood.

72 Recent work has reviewed the importance of uncertainty propagation in matrix population
73 models and outlined best-practice guidelines for addressing it (Simmonds and Jones 2024). That
74 paper focused on methodology and characterised the sources of uncertainty, the available
75 propagation techniques, and recommendations for practice, but did not empirically quantify the
76 consequences of ignoring uncertainty across a range of real analyses and demographic traits.

77 Here, we address that complementary question directly: not how uncertainty should be handled,
78 but when handling it actually changes the conclusions researchers draw from published
79 projection matrices.

80 We investigate sampling uncertainty in the COMPADRE Plant Matrix Database (Salguero-
81 Gómez et al. 2015; www.compadre-db.org), which currently contains over 9,000 matrix
82 population models (i.e. projection matrices, hereafter MPMs) drawn from more than 650 studies.

83 Over the course of its development, researchers have used COMPADRE (and the sister animal
84 matrix database COMADRE; Salguero-Gómez et al. 2016a) to address diverse topics in ecology,
85 evolution, and demography, including studies of the structuring of life-history variation (Franco
86 and Silvertown 1996, Salguero-Gómez et al. 2016b), comparative analyses of ageing (Baudisch et
87 al. 2013, Jones et al. 2014), tests of the demographic buffering hypothesis (McDonald et
88 al. 2017), and analyses of the impact of climate on population dynamics (Coutts et al. 2016,
89 Csergő et al. 2017). In practice, these analyses involve deriving parameters from a set of MPMs
90 in COMPADRE (e.g. equilibrium population growth rate λ , elasticities of λ , reproductive values,
91 mean vital rates, generation time, life expectancy), and incorporating these parameter estimates,
92 or further derivations, into a statistical model.

93 Our aim is to examine how sampling uncertainty in published projection matrices propagates to
94 specific MPM-derived quantities and to explore when point estimates are robust or fragile. We
95 expect sensitivity to vary systematically across traits, reflecting a spectrum from highly integrated
96 metrics — such as population growth rate, which averages over many transitions and should be
97 buffered against uncertainty in any single element — to traits that depend on long-term
98 survivorship inferred indirectly via age-from-stage methods, such as life expectancy and mortality
99 shape, where small errors in late-stage survival can amplify into large errors in derived quantities.
100 We structure our three analyses to bracket this spectrum deliberately. The focal quantities are
101 mature life expectancy, the shape of the mortality trajectory, log population growth rate, and
102 recruitment-climate regression coefficients. Because COMPADRE does not routinely include
103 transition-level sample sizes, our analysis is restricted to the subset for which these data could be
104 assembled (695 MPMs from 31 studies; details in Methods). We illustrate trait-dependent
105 uncertainty effects using a multi-species comparative analysis of life expectancy-shape
106 relationships (analysis 1), a single-species time-series analysis of weather effects on recruitment
107 (analysis 2), and a multi-site extension of the same climate-demography framework (analysis 3),
108 and use these to identify diagnostic features that help researchers assess robustness in their own
109 work.

110 **Methods**

111 *Background on the COMPADRE Plant Matrix Database*

112 The COMPADRE database includes MPMs and associated metadata from published studies
113 dating back to 1966 (Salguero-Gómez et al. 2015). Where possible, MPMs are decomposed into
114 component processes of survival, growth, and retrogression (the U matrix), sexual reproduction
115 (F), and clonal reproduction (C). The database provides a range of metadata for each MPM,
116 including the relevant time period, geographic coordinates of the study site, population or site

117 identifiers, and treatment conditions. From a given source study, COMPADRE often contains
118 multiple MPMs reflecting different sites, time periods, and/or treatment conditions, as well as
119 one or more “mean MPM” calculated as an element-by-element arithmetic mean across multiple
120 years and/or sites.

121 *Target studies*

122 We used an earlier version of COMPADRE to identify an initial group of 86 studies from which
123 we sought data, applying the following criteria: studies published from 2010 onwards; species of
124 common perennial growth forms (herbaceous, shrub, tree, palm, and succulent); wild
125 populations not subjected to experimental treatments; and MPMs with three or more stage
126 classes and a one-year projection interval. Data collection and analysis were subsequently
127 conducted using COMPADRE version 6.26.3.0 (created Mar_11_2026). Of the 86 initially
128 identified studies, we obtained the data required to model sampling uncertainty for 26: most
129 exclusions reflected insufficient reporting of transition-specific sample sizes, and a small number
130 used estimation methods incompatible with our bootstrapping approach (e.g. mark-recapture or
131 multi-stage vital rate products). We cannot rule out selection bias, but have no strong reason to
132 expect systematic differences between included and excluded studies. The 26 studies yielded 580
133 MPMs for 116 populations of 25 species. We supplemented this set with 5 additional studies
134 published before 2010 for which we had previously obtained data (targeted based on a criterion
135 of at least 4 years of temporal replication), giving a final set of 31 studies representing 695 MPMs
136 for 132 populations of 30 species.

137 *Typical construction of matrix population models*

138 A matrix population model A is a square projection matrix whose elements A_{ij} represent
139 estimated transition rates from stage class j to stage class i over one projection interval (typically
140 one year). Transitions may reflect the passage of individuals from one stage to another through

141 survival, growth, or retrogression (the U component), or the production of new individuals via
142 reproduction (the F component); $A = U + F$. Sometimes the reproductive transitions are
143 further split into sexual and clonal contributions.

144 The methodology for estimating transition rates varies across studies. In plant MPM studies,
145 survival-related transitions are often estimated from observed stage transitions of marked
146 individuals between censuses, as in Eq. (1); in other systems, especially mobile animals, transition
147 and survival estimation more commonly uses models that explicitly account for imperfect
148 detection.

$$149 \quad U_{ij,t} = \frac{\text{number transitioned from stage } j \text{ in year } t - 1 \text{ to stage } i \text{ in year } t}{\text{number in stage } j \text{ in year } t - 1}. \quad (1)$$

150 Reproductive transitions are sometimes estimated directly from counts of new recruits at time t
151 per reproductive individual at time $t - 1$, as in Eq. (2),

$$152 \quad F_{sr,t} = \frac{\text{number in recruit stage } s \text{ in year } t}{\text{number in reproductive stage } r \text{ in year } t - 1}. \quad (2)$$

153 The demographic interpretation of F depends on census timing and stage definitions (e.g. pre-
154 breeding vs post-breeding formulations). Here, Eq. (2) is a count-based estimator of the recruit
155 transition across one projection interval and we use the published matrix entries as represented
156 in COMPADRE rather than reparameterising each study to a single census convention.

157 Alternatively, reproductive transitions may be split into multiple lower-level vital rates
158 (e.g. numbers of flowers per plant, fruits per flower, and seeds per fruit, seed survival probability,
159 and seed germination probability), which are typically estimated from separate samples, and then
160 multiplied to derive the overall transition rate. In some studies, transition rates are estimated
161 using more nuanced statistical models that partially pool estimates across stage classes, sites,
162 and/or years, or incorporate additional variables external to the MPM.

163 *Sampling distribution for a single MPM*

164 Techniques for estimating sampling uncertainty in transition rates and derived parameters are
165 reviewed in Caswell (2001, Ch. 12). The method we use here is parametric bootstrapping, in
166 which we generate a sampling distribution for each MPM by taking 2,000 independent random
167 draws from the sampling distribution of each underlying transition rate. For example, if U_{ij} is a
168 survival-related transition from stage j to i , estimated from the observation that k out of the n
169 individuals in stage j survived and transitioned to stage i (and the remaining $n - k$ died), we can
170 model the U_{ij} transition as a binomial sample with k successes in n trials. We could equivalently
171 use a multinomial likelihood here (given observed counts x_m of each possible outcome m),
172 which flexibly scales to instances with more than two possible outcomes (e.g. an individual from
173 stage j might die, survive and remain in stage j , or survive and grow to stage $j + 1$). For
174 reproductive transitions, we use a Poisson likelihood to model the number of recruits k
175 produced by a given number of reproductive individuals n . We distinguish structural zeros
176 (biologically impossible transitions) from sampling zeros (possible but unobserved transitions):
177 structural zeros are fixed at 0 in all bootstrap draws, whereas sampling zeros are modelled with
178 the same likelihood-based uncertainty framework as non-zero transitions.

179 *Analyses*

180 We chose parametric bootstrapping because the sampling distributions of survival and
181 reproductive transitions follow well-characterised probability models (multinomial and Poisson,
182 respectively) whose parameters can be directly estimated from the transition frequencies
183 reported in source studies. This makes parametric bootstrapping more tractable than MCMC for
184 our scale of analysis and more principled than the empirical bootstrap, which cannot generate
185 variation when all individuals in a stage class experience the same outcome (e.g. all survive), a
186 situation common in published matrices. When a stage-specific sample size was recovered as

187 zero, we retained the published point-estimate transition column unchanged in all bootstrap
 188 draws rather than treating it as a structural zero. In such cases, the matrix remains usable in the
 189 same way it would be in a comparative database analysis, but no additional sampling variation is
 190 propagated for that unsupported column.

191 *Illustrative example*

192 We illustrate with an MPM for the perennial forb *Agrimonia eupatoria* (Rosaceae) from a study
 193 carried out in southern Sweden (Kiviniemi 2002) (Population ‘B’; projection interval 1997-1998).

194 The full matrix is

$$195 \quad A = \begin{bmatrix} 0 & 0 & 0 & 0.19 \\ 0.86 & 0.78 & 0 & 0 \\ 0.06 & 0.67 & 0.28 & 0 \\ 0 & 0 & 0.22 & 0.70 \end{bmatrix}.$$

196 Decomposed into survival/growth (U) and reproduction (F) components:

$$197 \quad U = \begin{bmatrix} 0 & 0 & 0 & 0 \\ 0.86 & 0.78 & 0 & 0 \\ 0.06 & 0.67 & 0.28 & 0 \\ 0 & 0 & 0.22 & 0.70 \end{bmatrix}, \quad F = \begin{bmatrix} 0 & 0 & 0 & 0.19 \\ 0 & 0 & 0 & 0 \\ 0 & 0 & 0 & 0 \\ 0 & 0 & 0 & 0 \end{bmatrix}.$$

198 In that study, ~80 plants were monitored each year in a population estimated at ~1000
 199 individuals. Mirroring the $A = U + F$ decomposition above, the observed transition frequencies
 200 can be written as:

$$201 \quad N_U = \begin{bmatrix} 0 & 0 & 0 & 0 \\ 12/14 & 14/18 & 0 & 0 \\ 0 & 1/18 & 6/9 & 13/47 \\ 0 & 0 & 2/9 & 33/47 \\ 2/14 & 3/18 & 1/9 & 1/47 \end{bmatrix},$$

$$202 \quad N_F = \begin{bmatrix} 0 & 0 & 0 & 9/47 \\ 0 & 0 & 0 & 0 \\ 0 & 0 & 0 & 0 \\ 0 & 0 & 0 & 0 \end{bmatrix}.$$

203 Here, columns index origin stage j ; rows 1-4 in N_{ij} index destination stages, and the extra row d
 204 contains stage-specific death frequencies. The N_F matrix remains 4×4 and contains only
 205 reproductive transitions. From these counts, the sampling distributions of the underlying
 206 transition rates are:

207 $(12,2) \sim \text{Multinomial}(U_{21}, U_{d1}),$

208 $(14,1,3) \sim \text{Multinomial}(U_{22}, U_{32}, U_{d2}),$

209 $(6,2,1) \sim \text{Multinomial}(U_{33}, U_{43}, U_{d3}),$

210 $(13,33,1) \sim \text{Multinomial}(U_{34}, U_{44}, U_{d4}),$

211 $9 \sim \text{Poisson}(\exp(F_{14} + \log 47)).$

212 To create a sampling distribution for the focal MPM, we take 2,000 independent draws from the
 213 sampling distributions implied above (see Supplementary Methods section S1.1 for a full
 214 workflow description). At each iteration, we arrange the drawn transition rates into the form of
 215 our MPM and then use that MPM draw to calculate our derived parameter(s) of interest. After
 216 2,000 iterations, we therefore have 2,000 independent draws from the sampling distribution of
 217 each transition rate and derived parameter.

218 The modest sample sizes underpinning this MPM (9-47 plants per stage class) result in relatively
 219 high sampling uncertainty in the transition rates (Fig. 1). For instance, the sample estimate for
 220 the vegetative-reproductive growth transition is 0.22 (2/9), but the upper bound of the 95%
 221 bootstrap interval suggests this transition rate could plausibly be as high as 0.45 in the population
 222 at large. Correspondingly, there is substantial sampling uncertainty in some derived parameters
 223 (Fig. 2). Based on the 95% bootstrap interval bounds, the equilibrium population growth rate
 224 could be as low as -17% per year or as high as $+1\%$ per year; life expectancy could be as low as

225 3.5 years or as high as 18.2 years; and the reproductive value of the vegetative stage class could
226 be 22% lower than that of the seedling stage or 1460% higher.

227 We chose three analyses to illuminate contrasting contexts in which sampling uncertainty may be
228 relevant. Analysis 1 is a comparative analysis across many species and populations, in which
229 derived traits depend heavily on long-term survivorship inferred indirectly from stage-structured
230 matrices, precisely the setting in which sampling uncertainty might compound. Analysis 2, by
231 contrast, is a within-species time-series analysis in which the response variable is a directly
232 observed count, the statistical model is closely tied to the raw data, and the question concerns an
233 environmental driver rather than a demographic trait per se. Analysis 3 extends analysis 2 to
234 multiple populations of the same species to assess whether inference robustness is consistent
235 across sites. Together, the three analyses bracket a range of typical uses of published MPMs and
236 allow us to ask whether the consequences of sampling uncertainty are consistent across contexts
237 or depend on the nature of the analysis.

238 *Analysis 1: Comparative analysis of mortality trajectories*

239 The analysis we focus on here tests the relationship between mature life expectancy and the
240 “shape” of the mortality distribution (sensu Baudisch 2011), where shape denotes the extent to
241 which mortality rates increase with age after reproductive maturity. Evolutionary biologists have
242 long been interested in variation in mortality trajectories within evolutionary theories of
243 senescence, but previous metrics used to compare rates of senescence may have conflated
244 variation in the shape of mortality trajectories with variation in the average level of mortality
245 (i.e. the inverse of life expectancy). Measures of shape proposed by Baudisch and colleagues
246 (Baudisch 2011, Baudisch et al. 2013, Wrycza et al. 2015) are mathematically independent of life
247 expectancy and therefore well-suited for comparative analyses. Whether shape is biologically
248 independent of life expectancy is a separate question, and an analysis by Baudisch et al. (2013)
249 (based on MPMs for 290 angiosperm species) found a positive relationship between life

250 expectancy and the shape of mortality trajectories. In addition to that analysis, several other
251 studies have used estimates of life expectancy and shape derived from MPMs in COM(P)ADRE,
252 primarily in studies of life-history variation (e.g. Salguero-Gómez et al. 2016b; Healy et al. 2019).

253 To estimate mature life expectancy (L) and shape (S), we first used age-from-stage methods
254 (Cochran and Ellner 1992, Caswell 2001 Sec. 5.3) to derive a survivorship trajectory from each
255 MPM, starting from the first stage class(es) with positive reproduction (see Supplementary
256 Methods section S1.2 for details). We then estimated mature life expectancy by summing over
257 the survivorship trajectory, as in Eq. (3),

$$258 \quad L = \sum_x l_x, \quad (3)$$

259 where l_x is the proportion of a cohort surviving from reproductive maturity to age x . To
260 estimate shape, we used a metric proposed by Baudisch and Stott (2019), which compares the
261 area under a survivorship trajectory to that under a reference trajectory assuming constant
262 mortality with age. The metric can take any continuous value between -0.5 (maximal negative
263 senescence; all individuals die at the age of maturity) and $+0.5$ (maximal senescence; all
264 individuals die at the age of maximum longevity), with a value of 0 indicating a constant
265 mortality rate (negligible senescence). Values of shape greater than, equal to, or less than zero
266 correspond to Type I, II, and III survivorship curves, respectively. When deriving a survivorship
267 trajectory from an MPM, the apparent mortality rate is guaranteed to approach a plateau as the
268 stage distribution approaches its equilibrium (Horvitz and Tuljapurkar 2008). To avoid bias from
269 artefactual mortality plateaus, we calculated shape from truncated survivorship trajectories,
270 limited to the period before a cohort reaches its quasi-stable distribution (based on a distance
271 threshold of 1%; see Supplementary Methods section S1.2).

272 We calculated life expectancy and shape at the population level, using one population-level MPM
273 per population. For populations with temporal replication, we used a pooled population MPM if

274 available; otherwise, we used a population mean MPM; if neither was available, we constructed a
 275 population-level matrix by taking the element-wise mean of the available annual individual
 276 MPMs. When database records included both site-specific populations and aggregate population
 277 labels (e.g. combined multi-site entries), we retained the site-specific populations and excluded
 278 the aggregate rows to avoid double counting. To model sampling uncertainty for pooled or mean
 279 population-level MPMs, we pooled stage-specific sample sizes across years (see Supplementary
 280 Methods section S1.3). This approach ignores uncertainty arising from temporal variation in
 281 transition rates and therefore underestimates overall uncertainty in the pooled estimates. The
 282 degree of underestimation depends on how much transition rates vary across years relative to
 283 within-year sampling error; for species with high interannual variability, the bias could be
 284 substantial. Results should therefore be interpreted conservatively with respect to the influence
 285 of sampling uncertainty.

286 To model the relationship between point estimates of mature life expectancy and shape, we used
 287 a linear model with random intercepts at the species level (to account for non-independence
 288 among populations of the same species), as in Eq. (4),

$$289 \quad y_{ij} \sim \text{Normal}(\alpha_j + \beta x_{ij}, \sigma^2), \quad (4)$$

290 where x_{ij} and y_{ij} represent point estimates of life expectancy and shape, respectively, for
 291 population i and species j , α_j represent species-specific intercepts, β is the slope of the
 292 relationship between S and L , and σ^2 is residual variance. Species-level intercept variation is
 293 modelled as $\alpha_j \sim \text{Normal}(\mu_\alpha, \tau^2)$, where τ^2 is the among-species intercept variance (i.e.,
 294 variance of random intercepts; distinct from residual variance σ^2). To incorporate sampling
 295 uncertainty, we used a measurement error model (based on McElreath 2020, sec. 14.1.2) where
 296 latent x_{ij} and y_{ij} are linked to sampling-distribution means (\bar{x}_{ij} and \bar{y}_{ij}) and variances ($\sigma_{x,ij}^2$ and
 297 $\sigma_{y,ij}^2$) derived as described in the sampling-distribution section above, as in Eqs. (5), (6), and (7).

298 The structural form of the model is unchanged when sampling uncertainty is incorporated; Eqs.
299 (5)-(7) restate the outcome model and add explicit sub-models for the true latent values of x_{ij}
300 and y_{ij} , which are now treated as unknown quantities estimated from their bootstrap-derived
301 sampling distributions rather than as fixed plug-in values:

$$302 \quad y_{ij} \sim \text{Normal}(\alpha_j + \beta x_{ij}, \sigma^2), \quad (5)$$

$$303 \quad \bar{x}_{ij} \sim \text{Normal}(x_{ij}, \sigma_{x,ij}^2), \quad (6)$$

$$304 \quad \bar{y}_{ij} \sim \text{Normal}(y_{ij}, \sigma_{y,ij}^2). \quad (7)$$

305 For these models, we used priors and settings as described in Supplementary Methods section
306 S1.4.

307 In supplementary analyses, we estimated the variance components of life expectancy, shape, and
308 other life history traits using a sampling-to-other variance ratio: the proportion of total observed
309 variation in a demographic parameter attributable to transition-rate sampling uncertainty relative
310 to variation from all other sources (principally biological variation among populations and
311 species). For each parameter we partitioned total variance into a sampling component, estimated
312 as the mean of the within-MPM sampling variances derived from the bootstrap distributions,
313 and a residual component representing the remainder. A ratio greater than 1 indicates that
314 sampling uncertainty exceeds biological variation; a ratio less than 1 indicates that biological
315 variation dominates. Full details are given in Supplementary Methods section S1.5 and
316 Supplementary Table S3.

317 *Analysis 2: The impact of weather on demographic traits*

318 We focus on a 12-year MPM time-series for *Silene spaldingii* (Lesica and Crone 2007), a perennial
319 herb found in semi-arid grasslands of western North America. We selected this case because it
320 provides a rare combination of multi-year temporal replication, recoverable transition-level

321 sample sizes, and matched local climate covariates, making it well suited for testing uncertainty
 322 propagation in a climate-response context. In that study, transition rates were estimated using
 323 the methods described above, as ratios of two integers (Eqs. (1) and (2)), and the corresponding
 324 stage-specific sample sizes are reported in Ellis et al. (2012). Eight of the MPMs in the 12-year
 325 time-series were non-ergodic due to missing estimates for the “rosette” stage class (i.e. no
 326 rosettes were observed in the sample plots). Instead of examining variation in population growth
 327 rates (which assumes ergodicity), we focus only on the reproductive transition (hereafter
 328 ‘recruitment’, the number of seedlings in year t per reproductive plant in year $t - 1$). We note
 329 that recruitment estimated in this way is a relatively aggregated transition whose behaviour under
 330 sampling uncertainty may differ substantially from that of survival transitions, which are
 331 estimated from smaller stage-specific samples and are more prone to boundary estimates. The
 332 robustness we observe in this analysis therefore reflects properties of aggregate reproductive
 333 transitions under moderate sample sizes, and should not be taken as evidence that within-species
 334 analyses are always insensitive to sampling uncertainty.

335 We examined the effect of monthly temperature deviations, derived from PRISM monthly
 336 temperature rasters extracted at the study population coordinates (Daly et al. 2008; PRISM
 337 Group, Oregon State University 2026; Supplementary Methods section S1.6), on annual
 338 recruitment using a similar approach to Teller et al. (2016), in which we model separate slope
 339 coefficients for a series of K monthly lags (i.e., going back in time from the end of a given
 340 transition interval) in a way that allows smoothing of coefficients across nearby lags. The
 341 smoothing model we use here is a Gaussian process with an exponential covariance function, as
 342 in Eq. (8),

$$343 \quad \log(y_t) \sim \text{Normal} \left(\alpha + \sum_k b_k x_{tk}, \sigma^2 \right) \quad (8)$$

344 with the monthly-coefficient prior and covariance structure given in Eqs. (9) and (10),

345
$$\mathbf{b} \sim \text{MVN}(\boldsymbol{\mu}_b, \mathbf{K}) \quad (9)$$

346 and

347
$$K_{i,j} = \eta^2 \exp(-0.5 \rho^2 (j - i)^2) \quad (10)$$

348 respectively.

349 where \mathbf{y}_t is annual recruitment in year t , and x_{tk} is the temperature deviation for year t at
 350 monthly lag k . Model parameters include the intercept (α), residual variance (σ^2), vector of
 351 monthly temperature coefficients (\mathbf{b}), mean of the distribution of \mathbf{b} ($\boldsymbol{\mu}_b$), covariance matrix (\mathbf{K}),
 352 maximum covariance for the \mathbf{b} distribution (η^2), and the rate of decay in covariance with time
 353 (ρ^2).

354 In the model above, \mathbf{y}_t represents a point estimate of annual recruitment derived from the
 355 relevant MPM. Recall that these values were initially calculated as the number of seedlings in year
 356 t divided by the number of reproductive adults in year $t - 1$. To propagate sampling
 357 uncertainty, we model the raw seedling counts directly with a Poisson likelihood (Eq. (11)) and
 358 include $\log(N_{t-1})$ as an offset via the year-specific intercept model in Eq. (12), so expected
 359 counts scale with the number of reproductive plants sampled. Note that the two formulations
 360 also differ in model structure: the point-estimate model uses a Gaussian likelihood on log per-
 361 capita recruitment, whereas the uncertainty-aware model uses a Poisson likelihood on raw counts
 362 with a log-offset. Observed differences between formulations therefore reflect both structural
 363 change and uncertainty propagation, though in practice the two are virtually identical.

364
$$Y_t \sim \text{Poisson} \left(\exp \left(\alpha_t + \sum_k b_k x_{tk} \right) \right) \quad (11)$$

365 year-specific intercepts

366
$$\alpha_t \sim \text{Normal}(\mu_\alpha + \log(N_{t-1}), \sigma_\alpha^2) \quad (12)$$

367 This implies that uncertainty in per-capita recruitment declines with sample size: if $\hat{r}_t = Y_t/N_{t-1}$,
 368 then $\text{Var}(\hat{r}_t)$ scales as $1/N_{t-1}$ for fixed underlying recruitment intensity.

369 with the same coefficient prior and covariance structure (Eqs. (13) and (14)):

370
$$b \sim \text{MVN}(\mu_b, \mathbf{K}) \quad (13)$$

371
$$K_{i,j} = \eta^2 \exp(-0.5 \rho^2 (j - i)^2) \quad (14)$$

372 where the response variable Y_t is the number of seedlings in year t , and N_{t-1} is the sample size
 373 of reproductive adults in year $t - 1$. The model is otherwise similar to the point estimates
 374 model, except that we use year-specific intercepts (α_t) to incorporate the Poisson offset (N_{t-1}),
 375 rather than a single model intercept as before. We fit the above-described models again, with
 376 priors and settings as described in Supplementary Methods section S1.4.

377 For the stage-specific survival comparison shown in the climate figures, we derived annual stage-
 378 specific survival probabilities as column sums of U (survive and transition to any living stage).
 379 We then estimated temperature effects for each stage under two matched formulations on the
 380 logit scale: an SSD-weighted quasibinomial GLM for point estimates ($\text{surv} \sim \text{tmp}$, with weights
 381 from the stable stage distribution of each annual U matrix) and a binomial count model for
 382 uncertainty propagation ($\text{cbind}(\mathbf{n}_{\text{surv}}, \mathbf{n}_{\text{fail}}) \sim \text{tmp}$). Full implementation details are provided in
 383 Supplementary Methods sections S1.6 and S1.9.

384 *Analysis 3: Multi-site extension of climate-demography models*

385 To test whether the analysis 2 conclusions generalise across populations, we applied the same
 386 paired modelling framework to three *Astragalus scaphoides* populations (Haynes Creek, McDevitt
 387 Creek, and Sheep Corral Gulch) (Crone & Lesica, 2004) . For each site, we fit the same two

388 recruitment-temperature formulations used in analysis 2: a point-estimate model and a sampling-
389 uncertainty model, for both (i) monthly-lag coefficients in the Gaussian-process framework and
390 (ii) a simplified spring-temperature model.

391 We also fit the same paired stage-specific survival-temperature comparison used in analysis 2,
392 deriving annual stage-specific survival from U -column sums and comparing an SSD-weighted
393 quasibinomial point-estimate model with a binomial count model that propagates sampling
394 uncertainty. This preserves comparable linear predictors across formulations while varying only
395 how uncertainty enters the response model. Full implementation details are given in
396 Supplementary Methods section S1.9.

397 All analyses were run in R (4.5.3) using RStan (2.32.7), with data handling and plotting in
398 tidyverse packages.

399 **Results**

400 *Analysis 1: Comparative analysis of mortality trajectories*

401 We first summarise the overall contribution of sampling uncertainty to observed trait variation
402 using variance-component ratios (Table 1), then ask whether this variation is large enough to
403 alter inference in a specific comparative analysis.

404 For our parameters of interest, life expectancy and shape, the estimated sampling-to-other
405 variance ratios across population-level MPM entries in analysis 1 were 1.93 and 3.31, respectively
406 (Table 1). Point estimates of life expectancy and shape often deviated from their respective
407 sampling distributions (Fig. 3). For shape, the most extreme positive and negative point
408 estimates tended to have sampling distributions shifted much closer to zero, whereas for mature
409 life expectancy, many point estimates were skewed upward from the median of the sampling
410 distribution. The mismatches between point estimates and sampling distributions usually

411 coincided with a boundary estimate of stage-specific survival (i.e. a point estimate approaching
412 either 0% or 100%). Supplementary analyses suggest that these boundary estimates can skew a
413 point estimate of life expectancy or shape away from the majority of its sampling distribution
414 (see Supplementary Figure S1 and Supplementary Figure S2).

415 Despite the mismatch between point estimates and sampling distributions, accounting for
416 sampling uncertainty in our test of the relationship between life expectancy and shape only
417 moderately affected the overall result (Fig. 4). For the slope parameter β , the posterior
418 distribution from the point-estimate model was centred slightly to the left of zero ($P[\beta > 0] =$
419 0.146), whereas the corresponding posterior from the sampling-uncertainty model was shifted
420 somewhat to the right ($P[\beta > 0] = 0.770$), indicating stronger evidence for a positive relationship
421 between life expectancy and shape.

422 Counterintuitively, there was less posterior uncertainty in β in the sampling-uncertainty model
423 than in the point-estimate model. To see why, consider that a regression slope is harder to
424 estimate precisely when the data are very spread out, because many different slopes could fit the
425 data reasonably well. The point estimates of shape are artificially spread out because boundary-
426 survival estimates drive some of them to extreme values. When those extreme point estimates
427 are replaced by their sampling distributions, which, as Figure 4A shows, tend to cluster much
428 closer to zero, the data are effectively less dispersed, and the slope becomes easier to pin down.

429 *Analysis 2: The impact of weather on demographic traits*

430 Despite the relatively small sample sizes underlying some recruitment estimates (the three
431 smallest sample sizes of reproductive individuals were 8, 12, and 13, respectively), results from
432 the point-estimate model and the sampling-uncertainty model were virtually identical (Fig. 5).
433 Both models identified temperature deviations in monthly lags 7-9 (February-April, prior to the
434 start of the transition interval in July) as having a positive effect on recruitment. If we simplify
435 our model and consider only the relationship between recruitment and mean spring temperature

436 deviations (see Supplementary Methods section S1.6 and Supplementary Table S5), we again find
437 that the point-estimate model is similar to the model incorporating sampling uncertainty (Fig. 5).
438 The estimate of the temperature coefficient β was 8.0% higher, and the 95% interval width for
439 that estimate was 10.9% higher, in the model incorporating sampling uncertainty compared with
440 the point-estimate model. A difference of this magnitude is unlikely to qualitatively change any
441 inferences drawn from a model. In the stage-specific survival panel (Fig. 5C), temperature
442 coefficients were estimable for stages 2 and 3; for stage 1 we show a descriptive open-circle
443 marker at mean survival across years, and for stage 4 (all years at 100% survival) we show a
444 descriptive open-circle marker at survival = 1, both without interval bars. This is analogous to
445 the pattern in Figure 4: once sampling variation in the annual survival estimates is modelled
446 explicitly, part of the apparent between-year spread is reallocated from biological signal to
447 measurement noise, which can yield more precise coefficient estimates.

448 *Analysis 3: Multi-site extension of climate-demography models*

449 Extending the same framework to three *Astragalus scaphoides* populations yielded the same overall
450 pattern: point-estimate and sampling-uncertainty models were closely aligned within site, while
451 differences among sites were more pronounced (Fig. 6). Across sites, we had 9-11 years of
452 annual demographic records, with 8-10 years retained for model fitting after filtering (0-3 years
453 excluded; Supplementary Table S7).

454 For spring-temperature models, posterior medians differed only slightly between formulations
455 (maximum absolute within-site difference: 0.026 on the linear predictor scale), and 95% interval
456 widths changed by only -5.3% to 4.7% across sites (Supplementary Table S8). Monthly-lag
457 coefficient profiles were likewise highly concordant between formulations, with within-site
458 correlations of posterior medians ranging from 0.992 to 0.997 and a maximum lag-specific
459 median difference of 0.042 (Fig. 6). As in analysis 2, these results indicate limited sensitivity of
460 climate-effect inference to transition-rate sampling uncertainty in this setting.

461 **Discussion**

462 *When does transition-rate sampling uncertainty materially affect inference from published MPMs?*

463 Sampling uncertainty in the transition rates underlying matrix population models can, in
464 principle, propagate into substantial uncertainty in the demographic traits derived from those
465 models. However, our results show that the impact of sampling uncertainty on inference from
466 published MPMs is strongly trait-dependent. In our analyses, accounting for transition-rate
467 sampling uncertainty moderately affected inference in a comparative analysis of mortality
468 trajectories, but had little impact on a within-species analysis of weather effects on recruitment.
469 It is important to note that the within-species result pertains specifically to an aggregate
470 reproductive transition (seedlings per reproductive adult) rather than to stage-specific survival
471 transitions, which are more commonly affected by boundary estimates. The contrast between
472 analyses 1 and 2 therefore reflects both the type of trait analysed and the type of transition
473 underlying it.

474 We use “materially affects inference” to mean a difference large enough to change a qualitative
475 conclusion, for instance, by shifting the direction of an estimated effect, moving a credible
476 interval from largely overlapping to largely excluding zero, or substantially altering the rank order
477 of populations. This is a judgment call rather than a formal threshold, and readers with different
478 tolerances for uncertainty may draw the lines differently. This contrast illustrates that ignoring
479 sampling uncertainty does not universally compromise inference from published projection
480 matrices, but that its consequences depend on the traits being analysed and the structure of the
481 underlying models.

482 For several commonly used integrated demographic metrics, such as population growth rate (λ),
483 generation time, and individual growth rate, sampling uncertainty accounted for only a small
484 fraction of the total variation observed among populations. By contrast, traits related to mortality

485 trajectories, particularly the shape of the mortality distribution, were far more sensitive: nearly
486 half of the total variation in shape was attributable to sampling uncertainty. In these cases, much
487 of the apparent variation among published MPMs is not distinguishable from statistical noise.
488 This result helps explain why incorporating sampling uncertainty shifted inference in the
489 senescence analysis but had little effect in the climate-recruitment analysis, where the response
490 variable was more directly linked to observed counts and less dependent on long-term
491 survivorship.

492 *Why are some MPM-derived traits particularly sensitive to sampling uncertainty, and what are the practical*
493 *implications?*

494 The sensitivity of mortality-trajectory traits has a single underlying cause: life expectancy and
495 shape are inferred via age-from-stage methods, which project a cohort forward through repeated
496 matrix multiplication. Errors in late-stage survival are therefore not simply carried forward but
497 compounded across projection steps, amplifying small transition-rate uncertainties into large
498 uncertainties in long-term survivorship. Integrated metrics like population growth rate are
499 insulated from this because they reflect a weighted average over all transitions simultaneously,
500 where errors tend to partially cancel.

501 Several structural features of published matrices exacerbate this. In our analysis 1 set ($n = 68$
502 population-level matrices), 47% (32/68) contain only a single reproductive stage class, 69%
503 (47/68) reach their quasi-stable distribution in less than one life expectancy, and 46% (31/68)
504 contain at least one stage-specific survival estimate of exactly 100%. These are not rare edge
505 cases; they are routine features of published matrices that systematically displace derived trait
506 estimates from the centre of their sampling distributions.

507 *Practical implications for users of compiled demographic databases*

508 These structural constraints have practical implications. Analyses based on highly aggregated
509 demographic metrics, such as population growth rate, mean vital rates, or generation time, are
510 often robust to sampling uncertainty and may not require explicit uncertainty modelling.
511 Analyses of ageing, longevity, or mortality trajectories are considerably more sensitive and should
512 be treated with caution unless sampling uncertainty is explicitly incorporated.

513 Analyses warrant caution when: any stage has survival >0.95 or <0.05 , or is estimated from
514 fewer than 20 individuals; the matrix has only a single reproductive stage class; or the time to
515 quasi-stable distribution is less than one life expectancy. Conversely, analyses are likely robust
516 when survival estimates are well away from boundaries, key transitions are estimated from >50
517 individuals, and the focal metric is highly integrated (e.g. population growth rate, generation
518 time). The absence of a detectable uncertainty effect (as in our climate-recruitment analysis)
519 should not be taken as general evidence that propagation is unnecessary; robustness is trait- and
520 context-specific.

521 *Sampling uncertainty versus demographic stochasticity*

522 It is important to emphasise that our analysis of sampling uncertainty is distinct from
523 demographic stochasticity, a distinction that is sometimes blurred in the literature (Kendall
524 1998). In the current study, we assume that the parameter of interest is the population mean of a
525 demographic trait and focus exclusively on uncertainty arising from finite sampling of that mean.
526 In addition, individuals within a population are likely to exhibit genuine variation around the
527 population mean in demographic rates, giving rise to demographic stochasticity. Both sources of
528 uncertainty can have important implications for applications such as population viability
529 analyses, where failing to account for either demographic stochasticity or sampling uncertainty
530 can lead to biased estimates of extinction risk (Kendall and Fox 2002, McGowan et al. 2011).

531 Distinguishing between these sources of uncertainty is therefore essential for appropriate
532 interpretation.

533 *Data limitations and reporting considerations*

534 Incorporating transition-rate sampling uncertainty into analyses of published MPMs remains
535 challenging, primarily because the data required to reconstruct sampling distributions are rarely
536 reported. Of the 91 source publications we initially targeted, only a small subset provided
537 sufficient information to model uncertainty in survival or reproductive transitions. Improved
538 reporting of transition-specific sample sizes or standard errors, particularly on the scale of the
539 linear predictor, would greatly enhance future researchers' ability to assess robustness and
540 propagate uncertainty where necessary (see Supplementary Methods section S1.7 for detailed
541 recommendations). At the same time, the absence of a single, universally "correct" approach to
542 modelling transition rates should not deter progress. Even partial information can be sufficient
543 to identify cases where inference is likely to be robust or fragile, and recognising the limits of
544 inference is often as important as refining point estimates. Viewed in this light, accounting for
545 sampling uncertainty is not simply a technical exercise but a means of understanding when
546 demographic inference from compiled databases can be trusted and when it should be treated
547 with caution.

548 *Conclusions*

549 Our analyses show that the consequences of transition-rate sampling uncertainty in published
550 matrix population models are highly trait-dependent. For many commonly used integrated
551 demographic metrics, such as population growth rate and generation time, sampling uncertainty
552 accounted for only a small fraction of total variation, suggesting that inference from published
553 MPMs is often robust for these traits. By contrast, parameters related to mortality trajectories,
554 including life expectancy and, especially, the shape of the mortality distribution, were
555 substantially more sensitive to sampling uncertainty. In these cases, a large proportion of

556 apparent variation among published matrices is indistinguishable from statistical noise, and point
557 estimates can be systematically biased by boundary estimates of stage-specific survival.

558 These results indicate that accounting for sampling uncertainty will not universally alter
559 conclusions drawn from analyses of published MPMs, but that its importance depends critically
560 on the demographic traits being analysed and on features of model structure and data availability.
561 Although our study examined only a subset of MPM-derived parameters and a limited fraction
562 of the matrices available in COMPADRE, it highlights specific conditions under which inference
563 is likely to be fragile and provides practical guidance for assessing robustness in comparative and
564 time-series analyses.

565 More broadly, the growing availability of large, compiled demographic databases presents both
566 new opportunities and new challenges. The data underlying these databases were collected for
567 diverse purposes, using heterogeneous sampling designs and statistical approaches, resulting in
568 variation not only in sampling uncertainty but also in underlying assumptions and context.
569 Recognising and accounting for these limitations through improved reporting in primary studies
570 and careful, trait-specific interpretation in secondary analyses will be essential to ensure that
571 comparative demography based on published projection matrices yields reliable and biologically
572 meaningful inferences.

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580 **Author contributions**

581 ORJ conceived and designed the study, curated data from the COMPADRE database and data
582 provided by authors of papers selected as case studies, conducted the analysis, and wrote the first
583 draft of the manuscript. EGS contributed to conceptualization and interpretation of results, and
584 reviewed and edited the manuscript. Both authors approved the final version for submission.

585 **Conflict of interest**

586 The authors declare no conflict of interest.

587 **Data availability**

588 Data and code required to reproduce our analyses are archived on Zenodo:
589 <https://doi.org/10.5281/zenodo.19474729>.

590 **Supplementary Materials**

591 The supplementary materials include the following items:

- 592 • **Supplementary Methods S1.1:** Sampling-distribution workflow used for each MPM.
- 593 • **Supplementary Methods S1.2:** Age-from-stage derivation, truncation, and shape
594 metric.
- 595 • **Supplementary Methods S1.3:** Pooled and mean MPM handling.
- 596 • **Supplementary Methods S1.4:** Bayesian model priors and fitting settings.
- 597 • **Supplementary Methods S1.5:** Variance-component summary (sampling-to-other
598 ratio).
- 599 • **Supplementary Methods S1.6:** Simplified spring-temperature model (analysis 2).
- 600 • **Supplementary Methods S1.7:** Reporting recommendations for future MPM studies.

- 601 • **Supplementary Methods S1.8:** Species identities in analysis 1.
- 602 • **Supplementary Methods S1.9:** Multi-site extension for analysis 3 (*Astragalus scaphoides*).
- 603 • **Supplementary Results S2.1:** Sensitivity to point-estimate weighting choice (no weights
- 604 vs SSD weights).
- 605 • **Table S1:** Single-MPM derived-parameter summary for the focal *Agrimonia eupatoria*
- 606 example.
- 607 • **Table S2:** Mean-MPM derived-parameter summary for the focal *Agrimonia eupatoria*
- 608 example (population B).
- 609 • **Table S3:** Analysis 1 posterior summary for the slope parameter (β) in the life
- 610 expectancy-shape model.
- 611 • **Table S4:** Analysis 2 Gaussian-process regression coefficients (monthly lags) for
- 612 recruitment responses to temperature anomalies.
- 613 • **Table S5:** Analysis 2 posterior summary for the spring-temperature coefficient.
- 614 • **Table S6:** Primary studies represented in the analysis subset, as identified in
- 615 COMPADRE.
- 616 • **Table S7:** Analysis 3 site coverage and spring-temperature effect summaries for
- 617 *Astragalus scaphoides*.
- 618 • **Table S8:** Analysis 3 posterior summary for the spring-temperature coefficient by site
- 619 and model formulation.
- 620 • **Table S9:** Analysis 2 stage-specific survival-temperature coefficients under no weights
- 621 and SSD weights.
- 622 • **Table S10:** Analysis 3 stage-specific survival-temperature coefficients under no weights
- 623 and SSD weights.
- 624 • **Figure S1:** Boundary-estimate diagnostic.

625

- **Figure S2:** Illustration of survivorship trajectories under boundary survival transitions.

626

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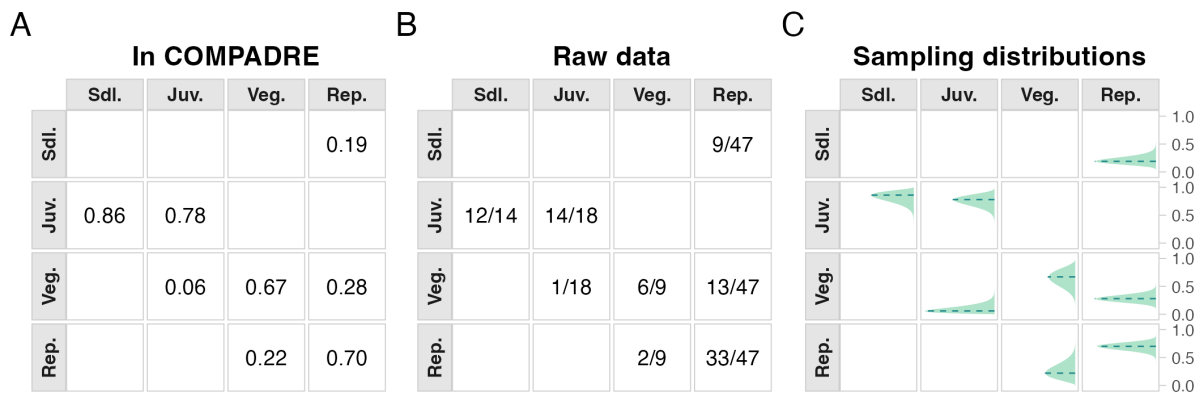
751

752 **Tables**

753 *Table 1: Sampling-to-other variance ratios for key demographic parameters in analysis 1. Ratios greater than 1*
754 *indicate that variance attributable to transition-rate sampling uncertainty exceeds residual biological variance*
755 *across population-level entries (including both among-population and among-species variation). The table shows*
756 *that this ratio is highest for shape, intermediate for life expectancy, and lowest for log population growth rate.*

Parameter	Sampling:other variance ratio
Shape	3.31
Life expectancy	1.93
Log population growth rate	1.01
Damping ratio	1.12
Generation time	1.63
Probability of reaching maturity	1.62

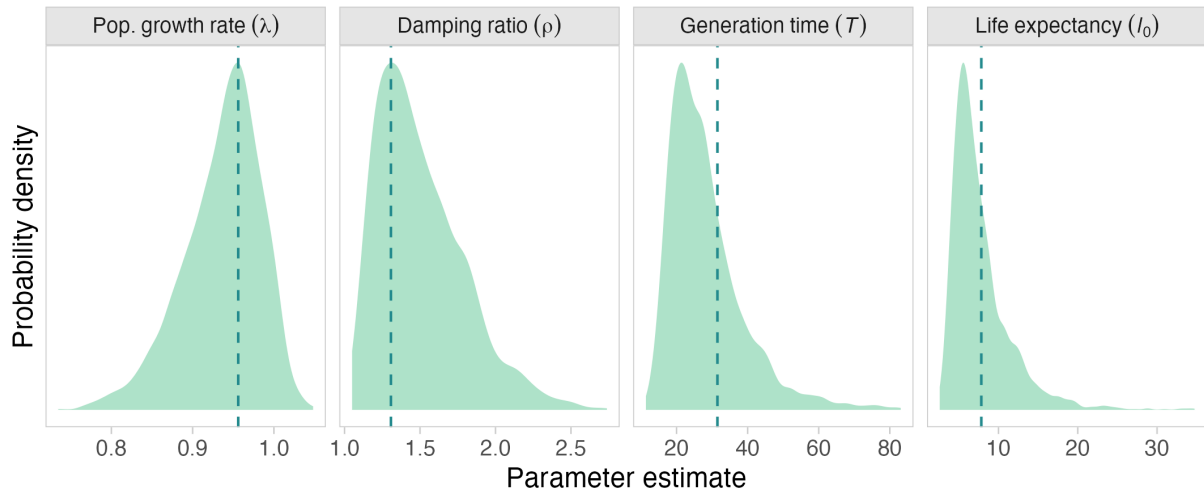
757



759

760 *Figure 1: Sampling distributions of transition rates for the focal single-MPM example (Agrimonia eupatoria,*
 761 *population B), showing transition rates as reported in COMPADRE (A), the underlying transition counts used*
 762 *for estimation (B), and bootstrap-based sampling distributions (C). In panel C, green density ribbons show*
 763 *transition-rate sampling distributions, red dashed vertical lines show point estimates, and grey dotted vertical lines*
 764 *show posterior medians. Transition rates are probabilities (bounded 0-1). Together, these panels show that several*
 765 *matrix elements are weakly estimated from finite samples.*

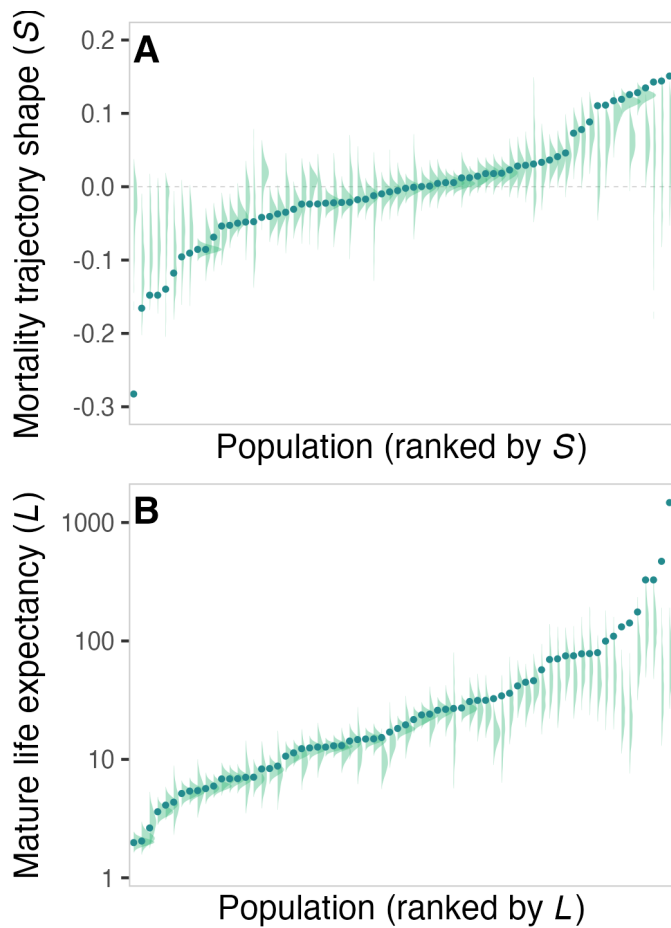
766



767

768 *Figure 2: Sampling distributions of demographic quantities derived from the focal single MPM, including*
 769 *population growth rate (λ), damping ratio (ρ), generation time (T), and life expectancy (l_0). Orange density*
 770 *ribbons show derived-quantity sampling distributions; red dashed vertical lines show point estimates; grey dotted*
 771 *vertical lines show posterior medians. Quantities are shown on their natural scales (T and l_0 are in years).*
 772 *Uncertainty in transition rates propagates to substantial uncertainty in these derived parameters, with the largest*
 773 *spread in quantities that depend on long-term survivorship and reproduction.*

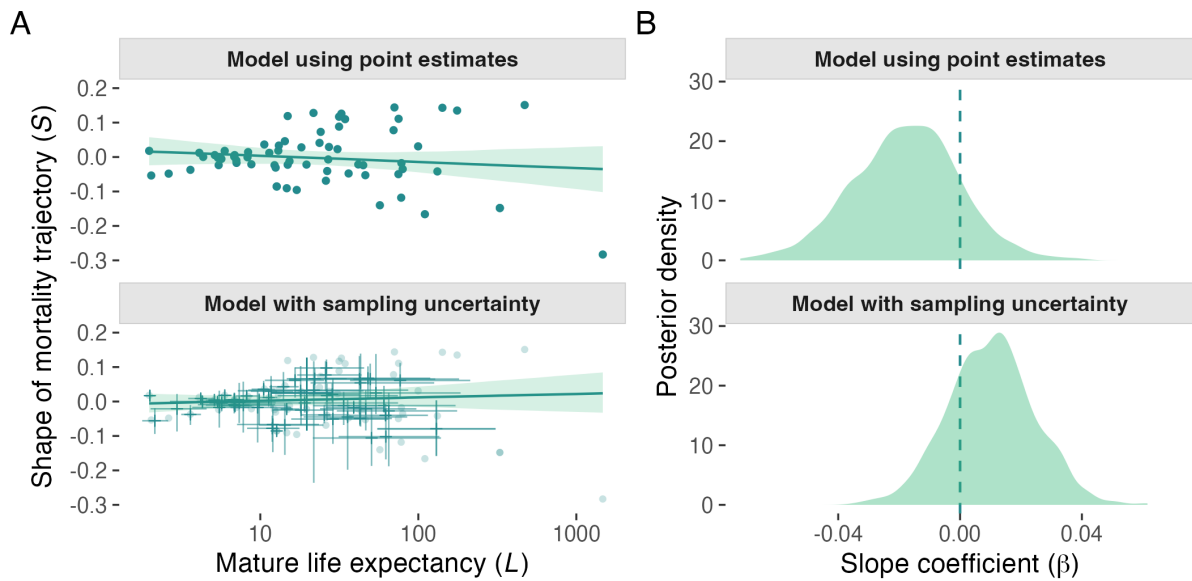
774



775

776 *Figure 3: Analysis 1 comparison of point estimates versus sampling-distribution summaries for population-level*
 777 *life expectancy and shape. Panel A shows mortality-trajectory shape, and panel B shows mature life expectancy*
 778 *(years). Purple distributions/points correspond to estimates from matrix point estimates, and green*
 779 *distributions/points correspond to uncertainty-aware summaries from transition-rate sampling distributions. The*
 780 *figure shows systematic shifts and shrinkage in extreme values when sampling uncertainty is propagated.*

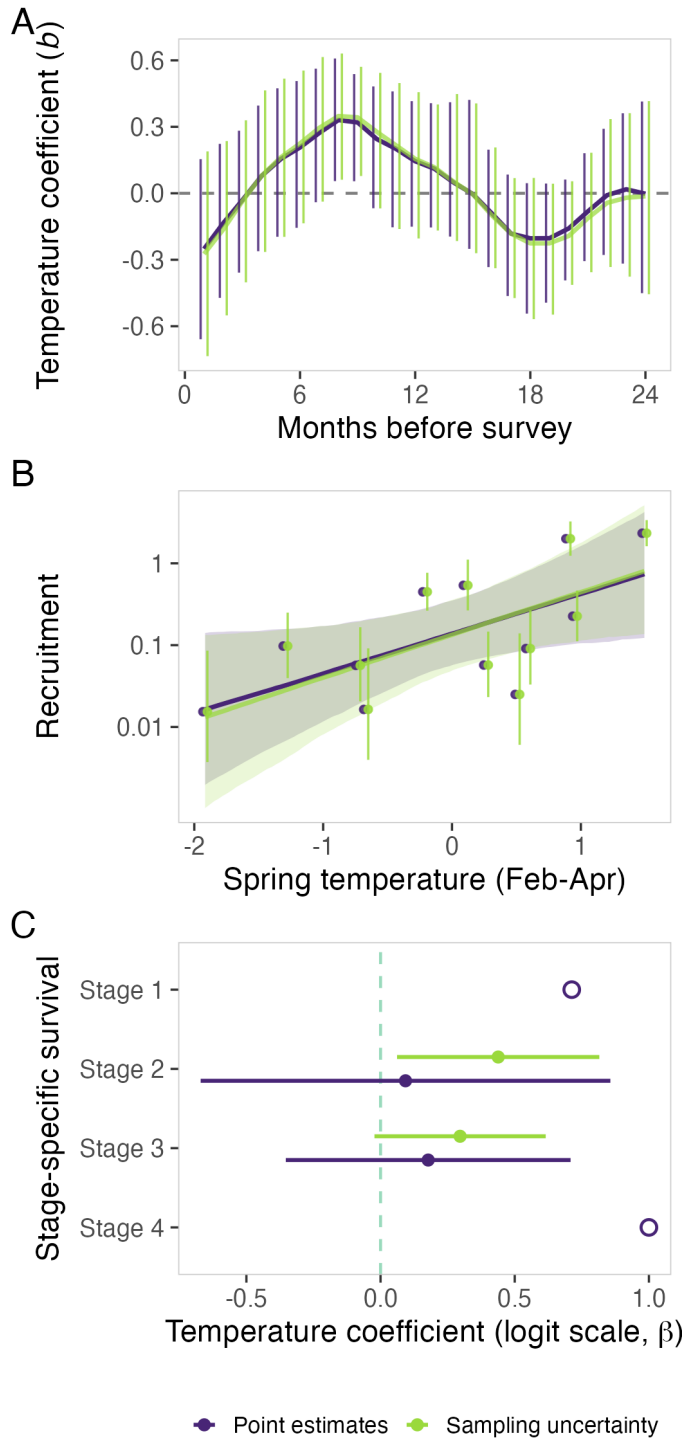
781



782

783 *Figure 4: Analysis 1 relationship between life expectancy and shape under models with and without sampling*
 784 *uncertainty propagation. In panel A, purple points/line/ribbon show the model using point estimates. Green*
 785 *crosses/line/ribbon show the uncertainty-aware model; crosses are centred on sampling-distribution medians with*
 786 *horizontal and vertical bars showing 95% intervals. The displacement between faint reference points and crosses*
 787 *indicates that propagated sampling distributions can shift central tendency away from point estimates, consistent*
 788 *with the offsets between dashed lines and density peaks in Figure 3. Axes are mature life expectancy (years, log*
 789 *scale) and mortality-trajectory shape. Panel B shows posterior densities for the slope parameter β (unit change in*
 790 *shape per unit change in log life expectancy). This figure shows that accounting for sampling uncertainty shifts*
 791 *posterior support towards a positive association.*

792



793

794 *Figure 5: Analysis 2 climate effects on recruitment and stage-specific survival in *Silene spaldingii*. Panels A-B*
 795 *show monthly-lag and spring-temperature recruitment models for point-estimate (purple) and uncertainty-aware*
 796 *(green) formulations, with similar posterior coefficient estimates and interval widths. In panel A, y-values are*
 797 *temperature coefficients b for monthly temperature deviations (months before survey on the x-axis). In panel B,*

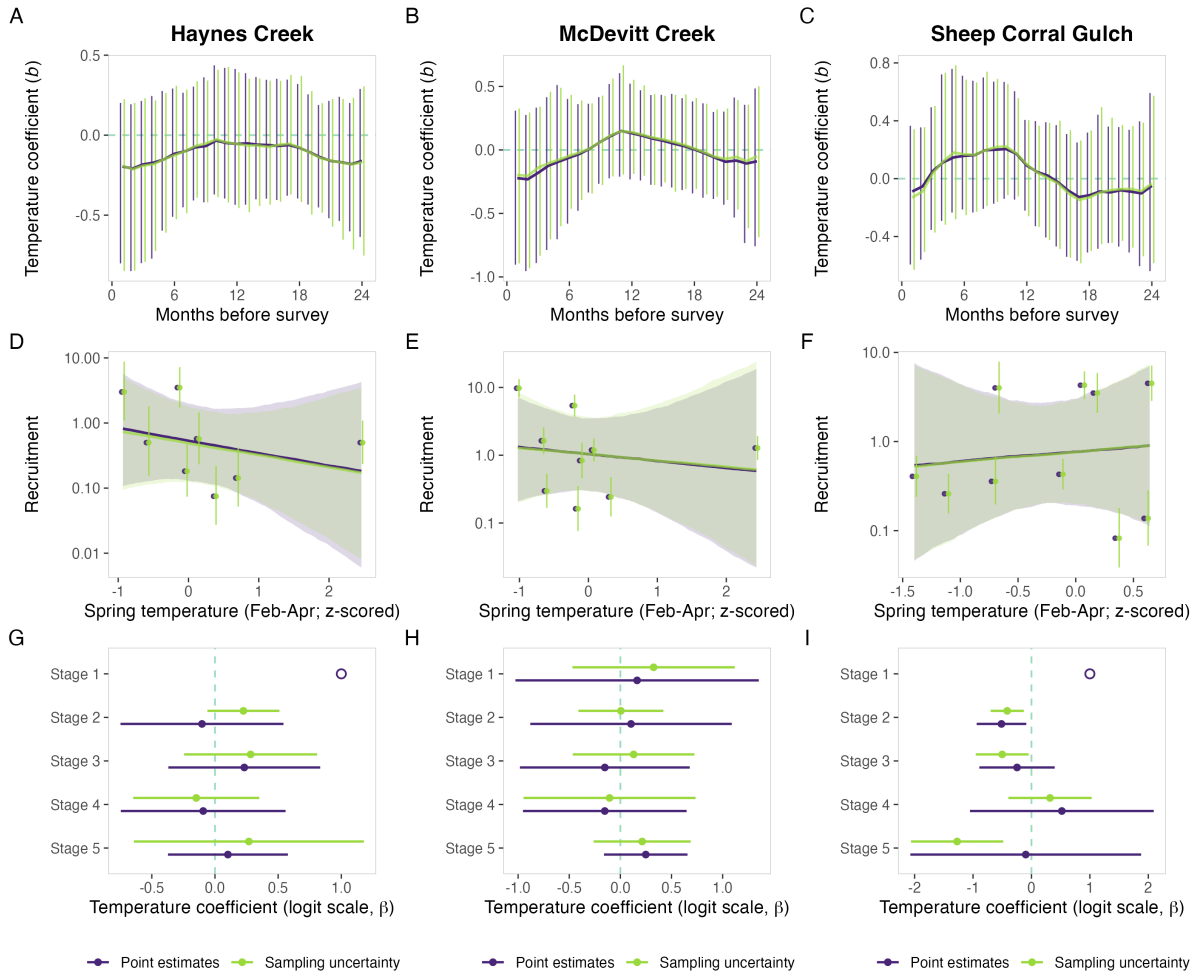
798 *recruitment is seedlings per reproductive plant and spring temperature is February-April temperature deviation.*

799 *Panel C shows stage-specific survival-temperature coefficients on the logit scale (beta) where estimable (stages 2-3).*

800 *Open circles in panel C are descriptive markers without interval bars for non-estimable stages: stage 1 shows mean*

801 *survival across years, and stage 4 shows survival = 1 (100% survival in all usable years).*

802



803

804 *Figure 6: Analysis 3 multisite climate-demography comparison for Astragalus scaphoides. Panel A (Haynes*
 805 *Creek), panel B (McDevitt Creek), and panel C (Sheep Corral Gulch) each show monthly-lag temperature*
 806 *coefficients, recruitment-temperature fits, and stage-specific survival-temperature coefficients for point-estimate*
 807 *(purple) and uncertainty-aware (green) formulations. Recruitment is seedlings per reproductive plant, and stage-*
 808 *specific survival coefficients are on the logit scale (β). Across sites, the two formulations remain closely aligned,*
 809 *showing that among-site ecological differences are larger than differences induced by uncertainty propagation.*

810

811 **Supplementary Methods**

812 *S1.1 Sampling-distribution workflow used for each MPM*

813 For each focal matrix population model (MPM), we reconstructed transition-level sampling
814 distributions from reported transition counts (or equivalent sample-size information). We then
815 generated 2,000 parametric bootstrap draws per MPM by sampling each transition from its
816 likelihood (multinomial for survival/growth/retrogression transitions and Poisson for
817 recruitment transitions), rebuilding the full MPM for each draw, and recalculating each derived
818 demographic quantity from that draw. We treated structural zeros (biologically impossible
819 transitions) as fixed zeros in all draws, and treated sampling zeros (possible but unobserved
820 transitions) as uncertain parameters under the same likelihood models used for non-zero
821 transitions. When a recovered stage-specific sample size was zero, we retained the published
822 point-estimate transition column unchanged in all bootstrap draws rather than treating it as
823 structurally fixed at zero; this preserves the published matrix while acknowledging that no
824 additional sampling variance can be propagated for that column.

825 This produced a sampling distribution for each derived quantity used downstream (e.g. life
826 expectancy, shape, log population growth rate). Point estimates in the manuscript correspond to
827 values derived directly from the reported matrix entries; uncertainty-aware summaries
828 correspond to moments of these bootstrap distributions.

829 *S1.2 Age-from-stage derivation, truncation, and shape metric*

830 Life expectancy and shape were derived from survivorship trajectories obtained using age-from-
831 stage methods, starting from the first reproductive stage. Shape was calculated using the pace-
832 shape framework metric (Baudisch and Stott 2019). Because stage-based survivorship can
833 converge to an artefactual mortality plateau, shape was calculated on truncated trajectories up to
834 quasi-stable convergence (distance threshold 1%).

835 *S1.3 Pooled and mean MPM handling*

836 Analysis 1 used one population-level MPM per population. Where a pooled population MPM
837 was available, it was used; otherwise, a mean population MPM was used; if neither was available,
838 a population-level matrix was constructed as the element-wise mean of the available annual
839 individual MPMs. When COMPADRE included both site-specific population rows and
840 aggregate rows that combined multiple populations, only the site-specific population rows were
841 retained, so that each population contributed a single analysis unit. For uncertainty propagation
842 in pooled or mean matrices, stage-specific sample sizes were pooled across years before
843 bootstrap simulation. This captures within-year sampling variation but does not fully represent
844 between-year process variation, so the resulting uncertainty should be interpreted conservatively.

845 *S1.4 Bayesian model priors and fitting settings*

846 Stan models and priors are defined in `models/*.stan`. For the two main analyses:

- 847 • Analysis 1 models use weakly informative normal priors for location/slope terms and
848 normal priors for scale terms (see `models/regress_hier.stan` and
849 `models/regress_hier_error.stan`).
- 850 • Analysis 2 uses two models with Gaussian-process regression forms and normal priors
851 on GP hyperparameters and coefficients (see `models/movbeta_gprc.stan` and
852 `models/movbeta_gprc_err.stan`).
- 853 • Analysis 3 uses the same Gaussian-process recruitment model pair as analysis 2
854 (`models/movbeta_gprc.stan` and `models/movbeta_gprc_err.stan`) together with
855 a simpler spring-temperature regression pair using weakly informative normal priors on
856 regression coefficients and scale terms (`models/regress.stan` and
857 `models/regress_log_err.stan`).

858 Fitting settings used in scripts:

- 859 • Analysis 1 (`scripts/S07_case_study_1_analysis.R`): `chains = 2, iter = 4000,`
860 `warmup = 2000-3000, seed = 5654.`
- 861 • Analysis 2 (`scripts/S12_case_study_2.R`): `iter = 5000` for key models, `seed =`
862 `5654`, with additional control settings for GP fits.
- 863 • Analysis 3 (`scripts/S14_case_study_3_astragalus.R`): `iter = 5000, seed =`
864 `5654`, using the same GP control settings as analysis 2 for recruitment models and
865 matched regression settings for the simplified spring-temperature models.

866 *S1.5 Variance-component summary (sampling-to-other ratio)*

867 For each parameter, the total observed variance was partitioned into:

- 868 • sampling component: mean within-MPM sampling variance from bootstrap draws;
- 869 • residual component: remaining variance among populations/species.

870 The reported summary is Eq. (15):

$$871 \quad \text{ratio} = \frac{v_{\text{sampling}}}{v_{\text{residual}}} \quad (15)$$

872 Ratios above 1 indicate that sampling uncertainty exceeds residual biological variation for that
873 parameter.

874 *S1.6 Simplified spring-temperature model (analysis 2)*

875 In addition to the monthly-lagged GP model, we fit a simplified spring summary model with
876 mean spring temperature deviation as the predictor, under both point-estimate and uncertainty-
877 aware formulations. The contrasts reported in the manuscript use posterior medians and 95%
878 interval widths from these paired models.

879 For the stage-specific survival panel in analysis 2 (Fig. 5C), we derived annual stage-specific
880 survival as the column sums of the U matrix for each year and stage, then compared two logit-

881 scale models per stage: (i) a stable-stage distribution (SSD)-weighted quasibinomial point-
882 estimate model, $\text{surv} \sim \text{tmp}$, where weights are the stable stage distribution from each annual U
883 matrix; and (ii) a binomial count model, $\text{cbind}(n_{\text{surv}}, n_{\text{fail}}) \sim \text{tmp}$, where n_{surv} and n_{fail} are
884 survived and non-survived counts implied by stage sample size and U -column survival
885 probability.

886 *S1.7 Reporting recommendations for future MPM studies*

887 To support uncertainty propagation in secondary analyses, primary studies should report (at a
888 minimum):

- 889 • transition-level sample sizes for each matrix element (or reconstructable transition count
890 tables);
- 891 • clear identification of pooled/mean matrices and pooling rules;
- 892 • uncertainty estimates on model-appropriate scales (e.g. linear predictor where relevant);
- 893 • explicit projection interval, stage definitions, and treatment filters used to derive
894 published matrices.

895 These items are usually sufficient for parametric uncertainty reconstruction in compiled
896 databases.

897 *S1.8 Study and species identities in analysis 1*

898 Analysis 1 used a subset of the full compiled dataset: 68 population-level entries representing 25
899 species. Study-level and species-level identities (including COMPADRE `SpeciesAuthor` labels)
900 are provided in Table S6.

901 *S1.9 Multi-site extension for analysis 3 (*Astragalus scaphoides*)*

902 To evaluate whether inference robustness in analysis 2 generalised across populations, we
903 repeated the same paired modelling framework for all available *Astragalus scaphoides* sites with

904 sufficient annual coverage. For each site, we fit (i) a point-estimate model and (ii) a sampling-
905 uncertainty model for recruitment-temperature relationships, using both the monthly-lag
906 Gaussian-process specification and the simplified spring-temperature specification. We then
907 compared posterior medians and 95% intervals between model formulations within site. This
908 design isolates the contribution of uncertainty propagation while preserving site-specific
909 ecological context.

910 For stage-specific survival in analysis 3, we used the same matched pair as in analysis 2: an SSD-
911 weighted quasibinomial point-estimate model ($\text{surv} \sim \text{tmp}$, weights from the stable stage
912 distribution of each annual U matrix) and a binomial count model ($\text{cbind}(n_{\text{surv}}, n_{\text{fail}}) \sim \text{tmp}$).
913 This keeps the linear predictor and response scale aligned across formulations, differing only in
914 whether survival is treated as a plug-in proportion or as counts with explicit sampling
915 uncertainty.

916 For sensitivity analyses intended to mimic database-only use (no access to true stage sample
917 sizes), we re-fit stage-specific point-estimate models under two alternatives: (i) no weighting and
918 (ii) SSD weighting derived from each annual U matrix.

919 **Supplementary Results**

920 *S2.1 Sensitivity to point-estimate weighting choice*

921 Across estimable stages in analysis 2, switching from no weighting to SSD weighting changed
922 temperature-coefficient medians by at most 0.142 on the logit scale, with interval widths
923 changing by -0.045 to 0.614 (Table S9).

924 Across estimable site-stage combinations in analysis 3, the corresponding maximum absolute
925 median shift was 0.912, and interval-width differences ranged from -0.512 to 1.526 (Table S10).

926 These sensitivity checks show how weighting assumptions alter point-estimate uncertainty when
927 true stage sample sizes are unavailable.

928 **Supplementary Tables**

929 Table S1. Single-MPM derived-parameter summary for the focal *Agrimonia eupatoria* example.
930 Reported values are point estimates and uncertainty intervals from parametric bootstrap
931 sampling distributions, showing substantial uncertainty for several derived demographic
932 quantities.

Parameter	Median	Lower 95%	Upper 95%	Point estimate
Pop. growth rate (λ)	0.943	0.826	1.010	0.956
Damping ratio (ρ)	1.440	1.110	2.200	1.310
Reproductive value (vegetative stage)	2.800	0.803	15.000	
Elasticity	0.345	0.145	0.558	
Generation time (T)	25.900	15.300	58.100	31.500
Life expectancy (l_0)	6.650	3.600	17.900	7.840

933

934 Table S2. Mean-MPM derived-parameter summary for the focal *Agrimonia eupatoria* example
 935 (population B). Point estimates are compared with sampling-distribution summaries to show
 936 how pooling/averaging changes central tendency and uncertainty relative to the single-MPM
 937 case.

Parameter	Median	Lower 95%	Upper 95%	Point estimate
Life expectancy (l_0)	1.920	1.69	2.16	1.930
Pop. growth rate (λ)	0.929	0.82	1.03	0.958
Damping ratio (ρ)	1.430	1.25	1.68	1.370

938

939 Table S3. Analysis 1 posterior summary for the slope parameter (β) in the life expectancy-shape
940 model. The uncertainty-aware model shows greater posterior support for $\beta > 0$ than the point-
941 estimate model.

Model	Median beta	Lower 95%	Upper 95%	P(beta > 0)
Model using point estimates	-0.017	-0.053	0.017	0.146
Model with sampling uncertainty	0.011	-0.017	0.038	0.770

942

943 Table S4. Analysis 2 Gaussian-process regression coefficients (monthly lags) for recruitment
 944 responses to temperature anomalies. Lag-specific posterior summaries are similar between point-
 945 estimate and uncertainty-aware model formulations.

Lag	Model	Median beta	Lower 95%	Upper 95%	lag_err
1	Point estimates	-0.251	-0.659	0.154	0.82
2	Point estimates	-0.135	-0.472	0.223	1.82
3	Point estimates	-0.030	-0.358	0.282	2.82
4	Point estimates	0.078	-0.261	0.396	3.82
5	Point estimates	0.155	-0.196	0.474	4.82
6	Point estimates	0.207	-0.156	0.507	5.82
7	Point estimates	0.272	-0.040	0.563	6.82
8	Point estimates	0.331	0.055	0.608	7.82
9	Point estimates	0.320	0.054	0.538	8.82
10	Point estimates	0.245	-0.068	0.482	9.82
11	Point estimates	0.200	-0.154	0.460	10.80
12	Point estimates	0.145	-0.151	0.416	11.80
13	Point estimates	0.111	-0.154	0.407	12.80
14	Point estimates	0.049	-0.195	0.411	13.80
15	Point estimates	-0.003	-0.251	0.421	14.80
16	Point estimates	-0.091	-0.333	0.197	15.80
17	Point estimates	-0.182	-0.464	0.084	16.80
18	Point estimates	-0.203	-0.543	0.046	17.80
19	Point estimates	-0.203	-0.494	0.044	18.80
20	Point estimates	-0.159	-0.393	0.062	19.80
21	Point estimates	-0.082	-0.306	0.181	20.80
22	Point estimates	-0.009	-0.278	0.291	21.80
23	Point estimates	0.018	-0.317	0.361	22.80
24	Point estimates	-0.002	-0.451	0.414	23.80
1	Sampling uncertainty	-0.274	-0.735	0.189	1.18

Lag	Model	Median beta	Lower 95%	Upper 95%	lag_err
2	Sampling uncertainty	-0.163	-0.551	0.236	2.18
3	Sampling uncertainty	-0.035	-0.402	0.329	3.18
4	Sampling uncertainty	0.079	-0.265	0.464	4.18
5	Sampling uncertainty	0.162	-0.197	0.517	5.18
6	Sampling uncertainty	0.225	-0.136	0.554	6.18
7	Sampling uncertainty	0.295	-0.038	0.615	7.18
8	Sampling uncertainty	0.348	0.061	0.631	8.18
9	Sampling uncertainty	0.342	0.078	0.571	9.18
10	Sampling uncertainty	0.277	-0.068	0.543	10.20
11	Sampling uncertainty	0.216	-0.161	0.497	11.20
12	Sampling uncertainty	0.155	-0.204	0.456	12.20
13	Sampling uncertainty	0.117	-0.168	0.401	13.20
14	Sampling uncertainty	0.053	-0.215	0.448	14.20
15	Sampling uncertainty	-0.004	-0.270	0.406	15.20
16	Sampling uncertainty	-0.100	-0.338	0.207	16.20
17	Sampling uncertainty	-0.182	-0.473	0.068	17.20
18	Sampling uncertainty	-0.226	-0.568	0.069	18.20
19	Sampling uncertainty	-0.226	-0.548	0.043	19.20
20	Sampling uncertainty	-0.193	-0.413	0.054	20.20
21	Sampling uncertainty	-0.112	-0.356	0.183	21.20
22	Sampling uncertainty	-0.043	-0.333	0.335	22.20
23	Sampling uncertainty	-0.020	-0.376	0.362	23.20
24	Sampling uncertainty	-0.013	-0.456	0.416	24.20

947 Table S5. Analysis 2 posterior summary for the spring-temperature coefficient. Point-estimate
948 and uncertainty-aware models yield similar medians and interval widths, indicating limited effect
949 of uncertainty propagation on this coefficient.

Model	Median	Lower 95%	Upper 95%
Point estimates	1.12	0.135	2.10
Sampling uncertainty	1.21	0.151	2.33

950

951 Table S6. Primary studies represented in the analysis subset, as identified in COMPADRE. For
 952 matrix-level metadata and source-link details beyond this summary, refer to the COMPADRE
 953 database record for each study.

Authors	Year	Journal	DOI	Species (COMPADRE label)
Andrello	2010			Dracocephalum_austriacum
Dangremond; Knight	2010	Ecology	10.1890/09-0418.1	Lupinus_tidestromii
Davison; Jacquelyn; Adriaens; Honnay; de Kroon; Tuljapurkar	2010	J Ecol	10.1111/j.1365-2745.2009.01611.x	Anthyllis_vulneraria
Fortini; Zarin	2010	Forest Ecol Manag	10.1016/j.foreco.2010.11.007	Callycophyllum_spruceanum; Carapa_guianensis; Licania_heteromorpha; Licaria_mahuba; Mora_paraensis; Platymiscium_filipes; Virola_surinamensis
Gaoue; Ticktin	2010	Conserv Biol	10.1111/j.1523-1739.2009.01345.x	Khaya_senegalensis
Jacquelyns; Brys; Jongejans	2010	Ecology	10.1890/08-2321.1	Orchis_purpurea
Picard; Mortier; Chagneau	2010	Ecol Model	10.1016/j.ecolmodel.2010.06.010	Dicorynia_guianensis
Sampaio; Scariot	2010	J Trop Ecol	10.1017/S0266467409990599	Geonoma_schottiana
Sletvold; Øien; Moen	2010	Biol Conserv	10.1016/j.biocon.2009.12.017	Dactylorhiza_lapponica
Bullock; White; Prudhomme; Tansey; Perea; Hooftman	2011	J Ecol	10.1111/j.1365-2745.2011.01910.x	Cirsium_acaule_2
Chen; Visser; Jongejans; van Breugel; Zuidema; Kassim; de Kroon	2011	J Ecol	10.1111/j.1365-2745.2011.01825.x	Shorea_leprosula
Csergő; Molnár; García	2011	Popul Ecol	10.1007/s10144-010-0249-y	Saponaria_bellidifolia
Dominguez; Lozano; Moreno Saiz; Schwartz	2011	J Nat Conserv	10.1016/j.jnc.2011.05.005	Vella_pseudocytisus_subsp._pau_2

Authors	Year	Journal	DOI	Species (COMPADRE label)
Navarro; Galeano; Bernal	2011	Trop Conserv Sci	10.1177/1940082 91100400104	Lepidocaryum_tenuis
Norghauer; Newbery	2011	Ecol Monogr	10.1890/10- 2268.1	Microberlinia_bisulcata; Tetraberlinia_bifoliolata
Thorpe; Stanley; Kayne; Latham	2011			Cypripedium_fasciculatum
Andrello; Bizoux; Barbet- Massin; Gaudeul; Nicolè; Till- Bottraud	2012	Biol Conserv	10.1016/j.biocon.2 011.12.012	Eryngium_alpinum
Ellis	2012	Ecology	10.1890/11- 1052.1	Liatris_scariosa
Ellis; Williams; Lesica; Bell; Bierzych udek; Bowles; Crone; Doak; Ehrlén; Ellis- Adam; McEache rn; Ganesan ; Latham; Luijten; Kaye; Knight; Menges; Morris; den Nijs; Oosterm eijer; Quintana - Ascencio ; Shelly; Stanley; Thorpe; Ticktin; Valverde; Weekley	2012	Ecology	10.1890/11- 1052.1	Cirsium_pitcheri_4; Phyllanthus_emblica_3
Scanga; Leopold	2012	Biol Conserv	10.1016/j.biocon.2 012.01.061	Trollius_laxus
Ticktin; Ganesan ; Parame sha; Setty	2012	J Appl Ecol	10.1111/j.1365- 2664.2012.02156. x	Phyllanthus_emblica_2; Phyllanthus_indofischeri
Gornish	2013			Pityopsis_aspera
Matlaga; Davis	2013	J Appl Ecol	10.1111/1365- 2664.12057	Miscanthus_giganteus
Osunkoy a; Perrett; Fernand o; Clark; Raghu	2013	Popul Ecol	10.1007/s10144- 013-0364-7	Lantana_camara

Authors	Year	Journal	DOI	Species (COMPADRE label)
Villellas; Ehrlén; Olesen; Braza; García	2013	Ecography	10.1111/j.1600-0587.2012.07425.x	Plantago_coronopus_2
Scanga	2014	Plant Ecol	10.1007/s11258-014-0344-9	Trollius_laxus_2
Baldauf; Correa; Ferreira; Santos	2015	Forest Ecol Manag	10.1016/j.foreco.2015.06.022	Himatanthus_drasticus
Gibert; Magda; Hazard	2015	PLOS ONE	10.1371/journal.pone.0139919	Festuca_eskia
Berry; Natalie L Cleavitt	2021	Popul Ecol	10.1002/1438-390X.12092	Platanthera_macrophylla_ Platanthera_orbiculata_

954

955 Table S7. Analysis 3 site coverage and spring-temperature effect summaries for *Astragalus*
 956 *scaphoides*. Point-estimate and sampling-uncertainty models give similar spring-temperature
 957 coefficients within each site, showing that between-site heterogeneity is larger than within-site
 958 method differences.

Site	Years (total)	Years (modelled)	Years excluded	Spring beta (point)	Spring beta (sampling)
Haynes Creek	11	8	3	-0.441	-0.424
Sheep Corral Gulch	11	10	1	0.244	0.258
McDevitt Creek	9	9	0	-0.238	-0.212

959

960 Table S8. Analysis 3 posterior summary for the spring-temperature coefficient by site and model
 961 formulation. Across all sites, 95% intervals overlap zero under both formulations, showing
 962 similar uncertainty conclusions with and without explicit sampling-uncertainty propagation.

Site	Model	Median	Lower 95%	Upper 95%
Haynes Creek	Point-estimate model	-0.441	-1.76	0.899
Haynes Creek	Sampling-uncertainty model	-0.424	-1.74	0.872
McDevitt Creek	Point-estimate model	-0.238	-1.45	1.110
McDevitt Creek	Sampling-uncertainty model	-0.212	-1.52	1.160
Sheep Corral Gulch	Point-estimate model	0.244	-1.69	2.100
Sheep Corral Gulch	Sampling-uncertainty model	0.258	-1.51	2.080

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964 Table S9. Analysis 2 stage-specific survival-temperature coefficients under two point-estimate
 965 weighting scenarios. “No weights” treats annual stage-specific survival proportions equally,
 966 whereas “SSD weights” applies stable-stage-distribution weights from the annual U matrix.

Stage	Weighting	Years used	Median	Lower 95%	Upper 95%
Stage 1	No weights	4			
Stage 1	SSD weights	4			
Stage 2	No weights	12	0.235	-0.221	0.691
Stage 2	SSD weights	12	0.093	-0.670	0.856
Stage 3	No weights	12	0.238	-0.315	0.791
Stage 3	SSD weights	12	0.178	-0.353	0.708
Stage 4	No weights	12			
Stage 4	SSD weights	12			

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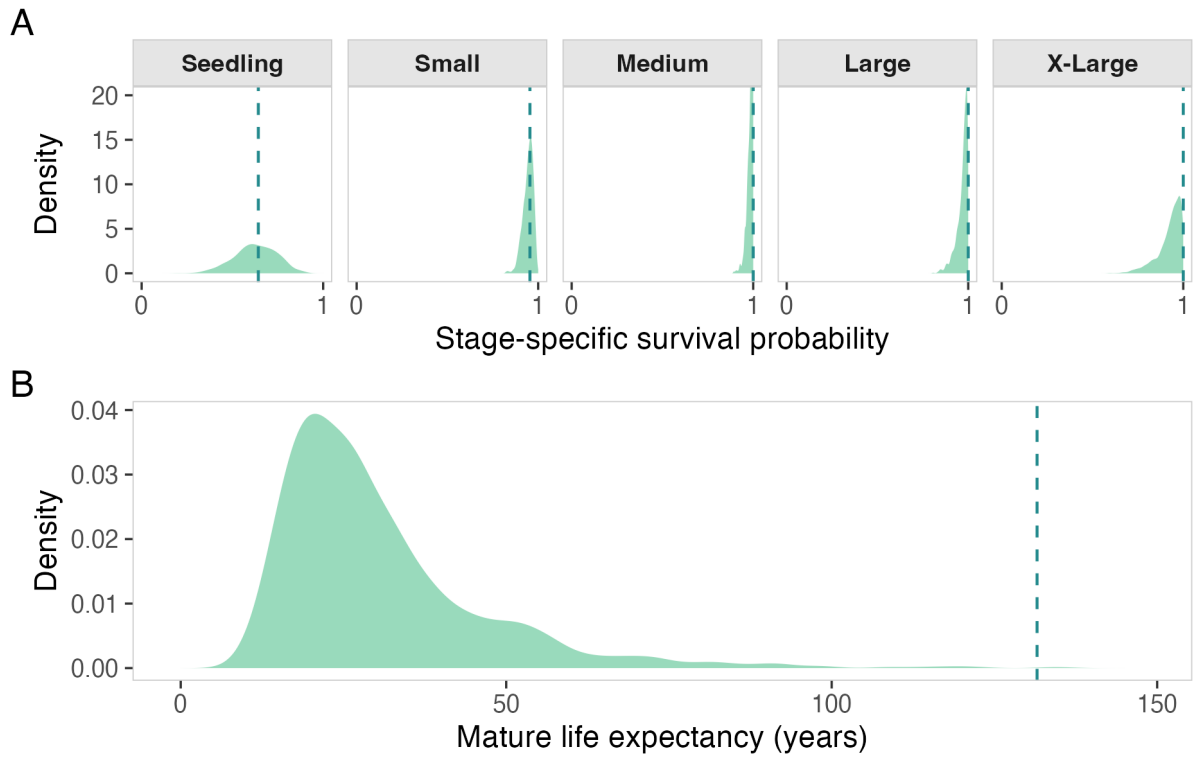
968 Table S10. Analysis 3 stage-specific survival-temperature coefficients under two point-estimate
 969 weighting scenarios across sites. The comparison shows how coefficient estimates and interval
 970 widths change when no stage sample-size information is used and SSD weights are substituted.

Site	Stage	Weighting	Years used	Median	Lower 95%	Upper 95%
Haynes Creek	Stage 1	No weights	8			
Haynes Creek	Stage 1	SSD weights	8			
Haynes Creek	Stage 2	No weights	8	0.180	-0.401	0.761
Haynes Creek	Stage 2	SSD weights	8	-0.103	-0.748	0.542
Haynes Creek	Stage 3	No weights	8	0.232	-0.207	0.670
Haynes Creek	Stage 3	SSD weights	8	0.232	-0.369	0.833
Haynes Creek	Stage 4	No weights	7	-0.210	-0.625	0.206
Haynes Creek	Stage 4	SSD weights	7	-0.093	-0.745	0.559
Haynes Creek	Stage 5	No weights	8	0.028	-0.702	0.759
Haynes Creek	Stage 5	SSD weights	8	0.102	-0.372	0.577
McDevitt Creek	Stage 1	No weights	9	0.120	-0.485	0.725
McDevitt Creek	Stage 1	SSD weights	9	0.164	-1.030	1.350
McDevitt Creek	Stage 2	No weights	9	0.020	-0.601	0.642
McDevitt Creek	Stage 2	SSD weights	9	0.104	-0.881	1.090
McDevitt Creek	Stage 3	No weights	9	0.086	-0.605	0.777
McDevitt Creek	Stage 3	SSD weights	9	-0.152	-0.982	0.678
McDevitt Creek	Stage 4	No weights	9	-0.211	-0.608	0.186
McDevitt Creek	Stage 4	SSD weights	9	-0.152	-0.953	0.648
McDevitt Creek	Stage 5	No weights	9	0.283	-0.371	0.937
McDevitt Creek	Stage 5	SSD weights	9	0.248	-0.160	0.657
Sheep Corral Gulch	Stage 1	No weights	10			
Sheep Corral Gulch	Stage 1	SSD weights	10			
Sheep Corral Gulch	Stage 2	No weights	10	-0.317	-0.889	0.255
Sheep Corral Gulch	Stage 2	SSD weights	10	-0.513	-0.938	-0.089
Sheep Corral Gulch	Stage 3	No weights	10	-0.460	-1.010	0.092

Site	Stage	Weighting	Years used	Median	Lower 95%	Upper 95%
Sheep Corral Gulch	Stage 3	SSD weights	10	-0.247	-0.892	0.398
Sheep Corral Gulch	Stage 4	No weights	10	0.329	-0.877	1.530
Sheep Corral Gulch	Stage 4	SSD weights	10	0.520	-1.050	2.090
Sheep Corral Gulch	Stage 5	No weights	10	-1.010	-2.220	0.204
Sheep Corral Gulch	Stage 5	SSD weights	10	-0.098	-2.070	1.880

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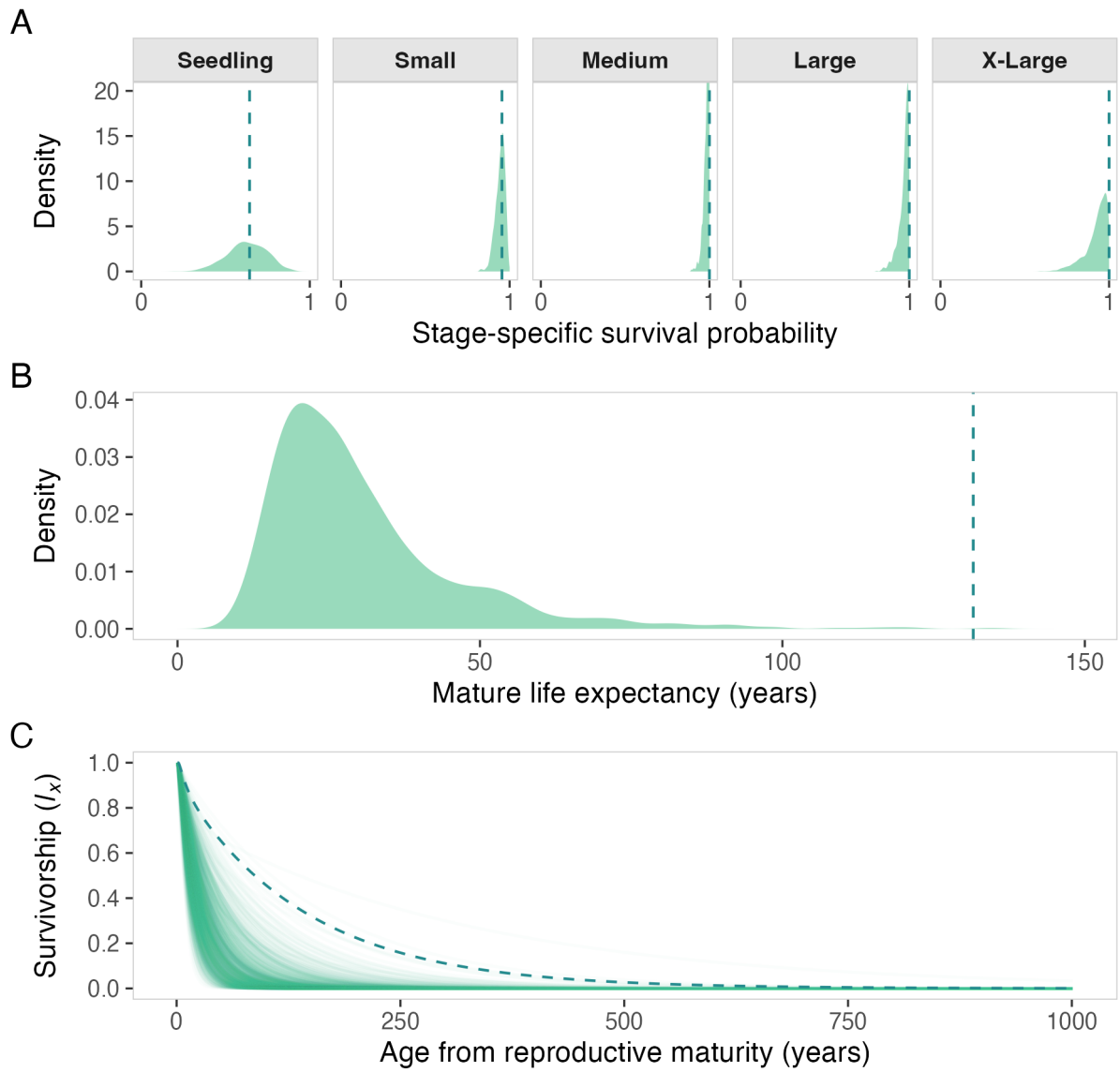
972 **Supplementary Figures**



973

974 *Figure 7: Boundary-estimate diagnostic. Purple lines indicate point-estimate-derived quantities and green densities*
975 *indicate uncertainty-aware summaries from sampling distributions. When stage-specific survival estimates are near*
976 *0 or 1, derived-parameter sampling distributions can be displaced relative to point estimates, indicating potential*
977 *inferential fragility.*

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979

980 *Figure 8: Illustration of survivorship trajectories under boundary survival transitions. Age/time is shown in years.*

981 *These cases can induce artefactual trajectory shapes and help explain mismatches between point estimates and*

982 *uncertainty-aware summaries.*

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