

1 **Toward a participatory and adaptive ecology of biodiversity conservation**

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12 Spinoza

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14 GitHub at <https://github.com/istaude/participatory-adaptive-conservation>. Data from Eichenberg et al. (2020)
15 used in Figure 2 are publicly available via the iDiv Data Repository at
16 <https://idata.idiv.de/ddm/Data/ShowData/1875?version=9>.

17 **Abstract**

18 Conservation biology emphasizes, with good reason, the harmful impacts of human activity
19 but often extends the same antagonism to novel, potentially beneficial biodiversity that also
20 arises through human involvement. This asymmetry is rooted in a pervasive nature/culture
21 dualism that affords ecological value primarily to processes considered “natural.” Such a
22 framing constrains conservation’s ability to engage productively with adaptive cultural and
23 societal processes that could support global efforts to prevent extinction under rapid
24 environmental change. Using gardens as an example, we illustrate how cultural processes can
25 accelerate opportunities for species persistence, facilitate assisted movement under climate
26 change, and foster adaptive dynamics, whilst taking into account potential ecological risks.
27 Integrating human-mediated and human-dependent processes expands conservation’s
28 opportunity space for buffering the many ecological uncertainties ahead and highlights how
29 societal participation can help reduce net extinction rates.

30 **In a nutshell**

- 31
- 32 • Although a bird and a human may both disperse a seed, current conservation paradigms
33 value human processes very differently from “natural” ones. We illustrate how society
34 and culture could support biodiversity if such dualistic thinking is overcome.
 - 35 • This is highly contentious, as precautionary principles dominate conservation practice
36 and broader societal participation is often seen as risky. Yet under rapid global change
37 and ecological uncertainty, precaution that resists change may itself be risky.
 - 38 • Wrestling with how to value culturally mediated (novel) biodiversity, both empirically
39 and normatively, will be essential to prepare conservation for the challenges of a
radically new future.

40 Main text

41 Diversifying conservation strategies in a changing world

42 Conservation biology, as a scientific discipline, developed in the late 20th century in response
43 to accelerating losses of species and ecosystems (Soulé 1985). It was built on a “nature for
44 itself” paradigm (Mace 2014), which held that nature should be protected from human
45 influence, reflecting early conservation ideals that envisioned wilderness as landscapes
46 “untrammelled by man” (Cronon 1996; Marris 2013). This idea rests on an ontological
47 nature/human dualism that defines the natural in opposition to the cultural, thereby situating
48 humans outside the natural realm, and is often coupled with the assumption that nature exists
49 in an undisturbed state of equilibrium in the absence of human intervention (Dussault 2016).
50 This dualism continues to shape ecological research and conservation’s goal of protecting
51 natural systems from humans, preserving historical baselines, and privileging biological
52 nativism (Lundgren *et al.* 2026). While this mindset brought major successes (Langhammer *et*
53 *al.* 2024), it also narrows how we imagine opportunities to support biodiversity in a world that
54 has long been—and is increasingly—shaped and overlaid by human culture.

55 From Late Pleistocene extinctions (Svenning *et al.* 2024) to anthropogenic “gardens” in the
56 Amazon (Rostain *et al.* 2024), the Agave gardens of the Hohokam in Arizona (Hodgson *et al.*
57 2023), and many forest and grassland types in Europe, humans have long co-created
58 ecosystems now considered “natural”. Recognizing that Holocene (“wild”) baselines have long
59 been shaped by people has opened new ways of thinking about conservation, producing several
60 coexisting framings (Kareiva and Marvier 2012; Marris 2013; Mace 2014). “Nature despite
61 people” acknowledges pervasive human impacts and aims to conserve in spite of them, “nature
62 for people” foregrounds ecosystem services and human benefits, and more recent “people and
63 nature” approaches emphasize sustainable, resilient interactions between human societies and
64 the environment (Mace 2014). Despite broadened perspectives, much of conservation remains
65 focused on preserving or reinstating (pre-)historic baselines, functionality and nativeness (even
66 in dynamic approaches such as rewilding), and continues to value human processes differently
67 than those considered “natural” (henceforth, referred to as mainstream conservation).

68 The implications of this prevailing paradigm are evident across many areas of conservation but
69 are particularly visible in plant conservation, which provides a useful illustration. Human-
70 mediated plant dispersal, for instance, is typically cast as an external interference that
71 “disrupts” native floras or causes biological invasions (e.g., Cai *et al.* 2025), even though it can
72 also be seen as an ecological process in its own right, one that acts in both directions and can
73 equally facilitate species persistence (Segar *et al.* 2022; Staude *et al.* 2025; Lundgren *et al.*
74 2026). For example, the common snowdrop (*Galanthus nivalis*) is listed as Near Threatened in
75 Germany even though it flowers in millions of gardens and parks each spring (Metzing *et al.*
76 2018). Such cases exemplify that mainstream conservation typically recognizes only “wild”,
77 populations as valid, while human-mediated dispersal and habitats, even when they support
78 species, fall outside what conservation recognizes as legitimate nature. This framing limits how
79 conservation biology can learn from and work with these human-established, human-
80 dependent forms of supporting species and the novel biodiversity they produce.

81 A plethora of more holistic and dynamic views of nature exists today (Dussault 2016), but their
82 origins in Western philosophy can already be found in Spinoza’s 17th-century philosophy
83 (Spinoza 2016; Notes S1). In modern ecological language (Stephano 2017; Notes S1), Spinoza
84 can be understood as conceiving nature, including humans, as a single system in which each
85 part strives to maintain and increase its ability to persist and adapt. Value does not derive from
86 imposed ideals such as the preservation of historical baselines chosen by humans but arises
87 internally through the effects that species have on one another’s ability to persist (Stephano
88 2017; Notes S1). What is “good” is whatever enables adaptive dynamics that maintain or
89 increase *average* fitness (*sensu* Carroll *et al.* 2023). Human activities are therefore not judged
90 as artificial but simply as one of nature’s many processes that can accelerate disruption yet also
91 accelerate opportunities for species to persist or evolve. This reframes human-mediated and
92 human-dependent biodiversity as legitimate expressions of nature and opens space for
93 conservation to engage with cultural processes in ways that support biodiversity.

94 Here, we build on these Spinozist principles and turn to their practical application in
95 conservation amid the dynamic and inevitable changes in biodiversity under rapid global
96 change. These changes require a rethinking of conservation concepts grounded in historical
97 baselines and equilibrium assumptions, as such concepts can no longer be expected to function
98 invariably under the uncertainties of change (Holling and Meffe 1996; Grainger *et al.* 2025).
99 This creates a need to spread the risk of failure in conservation by diversifying our strategies.
100 We use gardens to illustrate how non-dualistic thinking allows cultural processes to contribute
101 to conservation by supporting the persistence, movement, and adaptive potential of species that
102 might otherwise decline. We also address the tensions this raises within mainstream
103 conservation and show how an approach informed by Spinozist principles recognizes human-
104 created ecological novelty as part of how diversity is expressed in nature, treating ecological
105 risks arising from human participation not as reasons for exclusion but as information for
106 understanding and guiding it.

107 **Climate change and human participation in species redistribution**

108 Climate is shifting at rates of several kilometers per year in many regions (Kosanic *et al.* 2019),
109 yet most plant species disperse only tens of meters per generation (Lososová *et al.* 2023).
110 Animal dispersers which play key roles in long-distance seed movement, such as birds and
111 mammals, are in decline, and this loss is estimated to have reduced plants’ capacity to track
112 climate change by up to 60 % (Fricke *et al.* 2022). Additional barriers, both human-made (e.g.,
113 fragmentation) and natural (e.g., mountains and water bodies) (Sanczuk *et al.* 2025), further
114 constrain movement, and even when seeds reach new areas, unsuitable soils often prevent
115 establishment (Ni and Vellend 2024). Together, this severely limits species’ capacity to disperse
116 and expand their ranges, with effects on ecosystem functioning and human well-being (Gardner
117 and Bullock 2025). Restoring the conditions that enable dispersal and mitigating the human
118 processes that disrupt them is important. Yet in doing so, we must not overlook the human
119 processes that also move species and that could be adaptive.

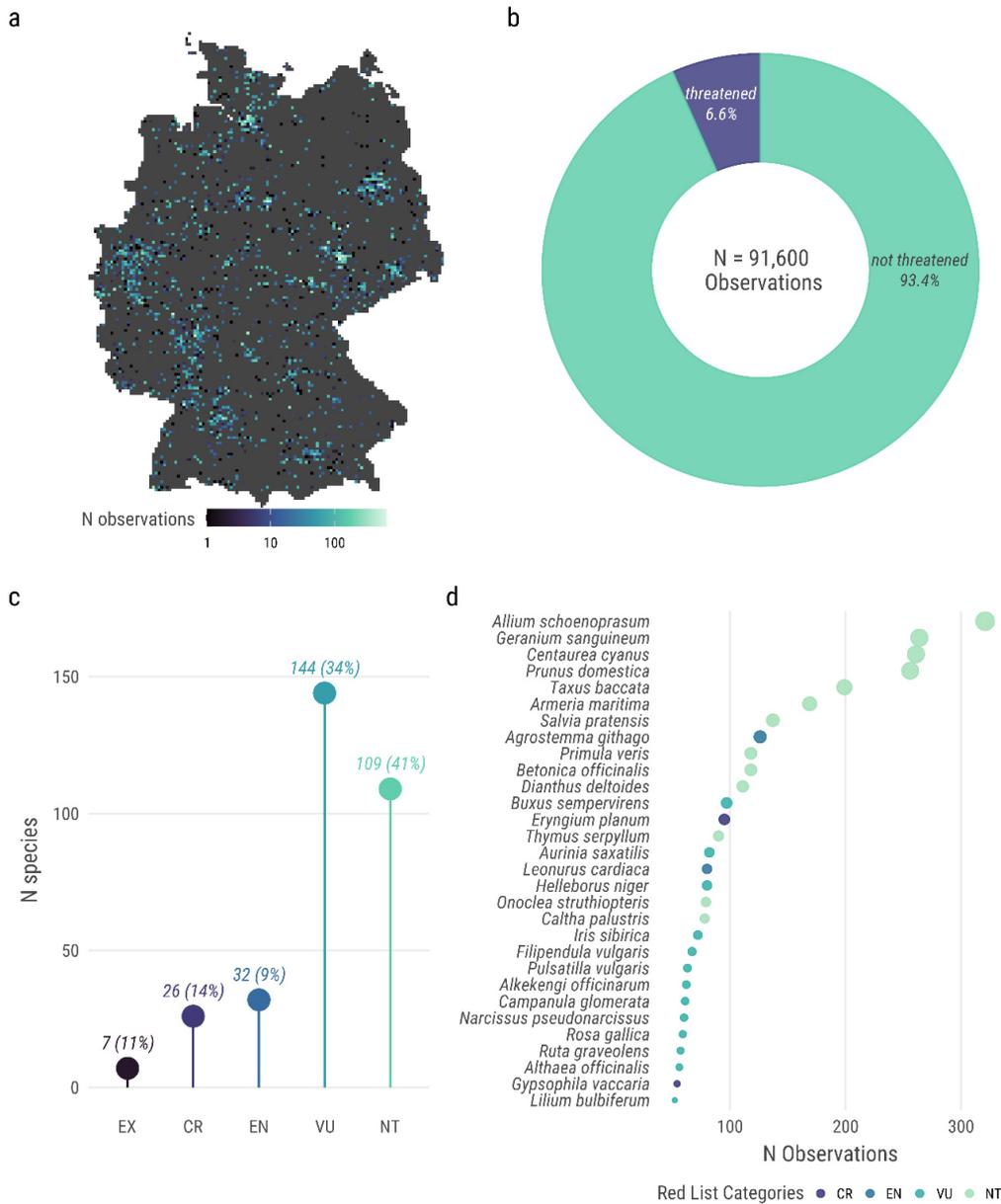
120 Narratives of constrained “natural” dispersal and ongoing scientific debates over assisted
121 colonization (Twardek *et al.* 2023) tend to bracket out the fact that humans move species all

122 the time. The issue is not a lack of awareness of potential ecological harm, but a neglect of how
123 this human process could be beneficial. In this context, gardens (and the many millions of
124 people who garden) form an extensive yet underappreciated cultural infrastructure for species
125 movement that could help prevent extinctions and foster adaptive dynamics. While invasion
126 biology has long emphasized the potential negative ecological consequences of horticultural
127 plant movements (Dehnen-Schmutz *et al.* 2007; Van Kleunen *et al.* 2018), recent work shows
128 that gardens can also support the conservation of species (Segar *et al.* 2022; Staude 2024;
129 Bucher *et al.* 2025). Given a world in which human activity has reshaped the very conditions
130 under which species now persist, focusing only on “natural” processes and removing human
131 processes from the equation becomes inconsistent. Gardens offer a tangible example and
132 pathway for integrating cultural processes to support species.

133 **Gardens as distributed infrastructure for plant movement and persistence**

134 First, gardens can act as refugia for native species declining in their native range, thereby
135 expanding ex-situ conservation efforts (Ismail *et al.* 2021; Segar *et al.* 2022). A notable
136 example is the Wollemi pine (*Wollemia nobilis*), endemic to Australia, which is expected to go
137 extinct in the wild soon but persists in home gardens (Offord and Zimmer 2023). In Germany,
138 the community-science project GartenDiv¹ has begun documenting garden plant diversity
139 using the Flora Incognita plant identification app (Mäder *et al.* 2021). In its pilot phase, ~7%
140 of 91,600 observations from 1,502 gardens were nationally red-listed species (Fig. 1). An
141 extrapolation (ignoring, for now, the many biases of community-science data) to Germany’s
142 ~17 million gardens suggests millions of individuals across hundreds of declining native
143 species may already occur in these gardens. Commercial availability of native plants is
144 increasingly well developed in Central Europe (Munschek *et al.* 2023), Australia (Esperon-
145 Rodriguez *et al.* 2025) and North America (Fertakos *et al.* 2026), and is beginning to emerge
146 in megadiverse countries such as Brazil (Flavia *et al.* under review).

¹ <https://floraincognita.de/gartendiv/>



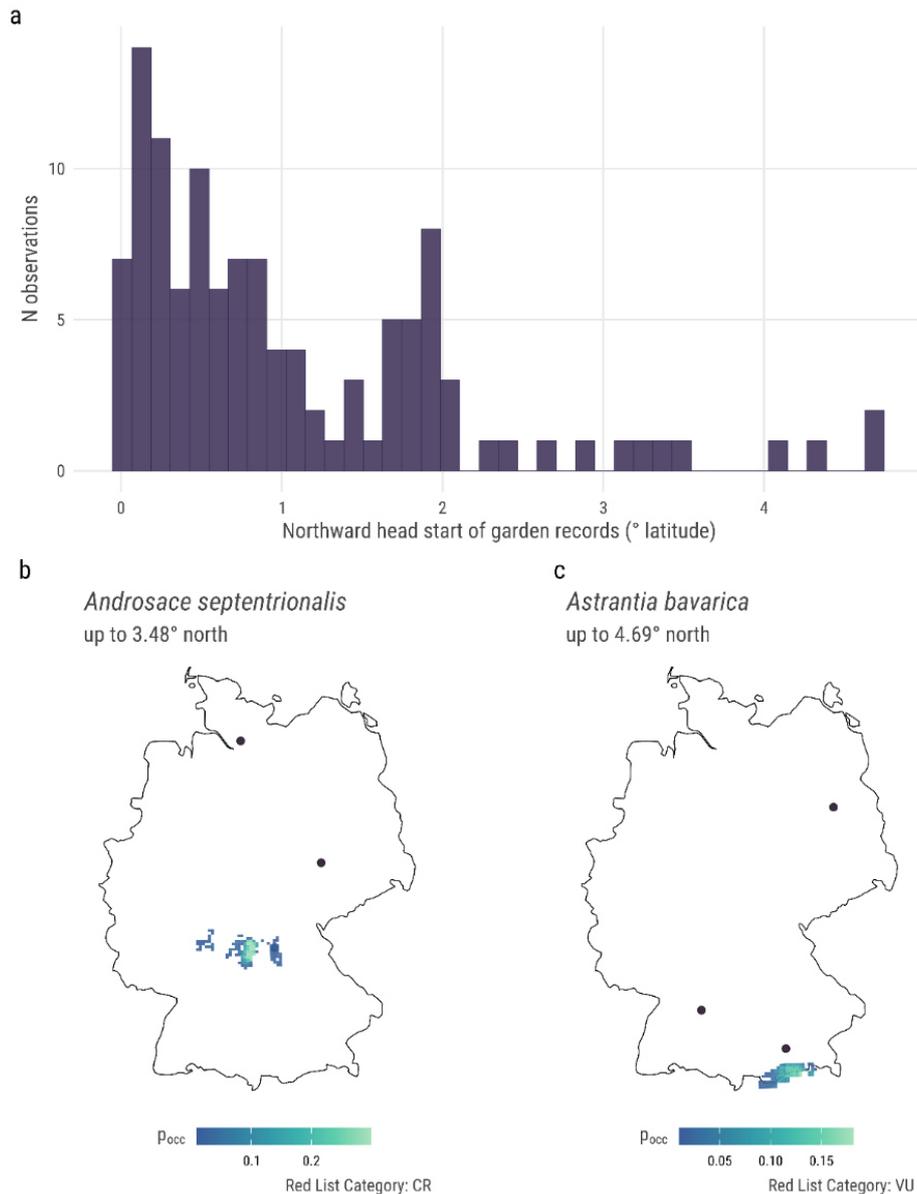
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148 **Fig. 1: Data from a Germany-wide community science project monitoring plant diversity in gardens.** a)
 149 Spatial distribution of observations recorded during the pilot phase of GartenDiv (May–October 2025). b) Of the
 150 91,600 garden observations, approximately 7% belong to threatened or declining plant species. c) Number and
 151 proportion of species across red list categories. d) Number of observations of the 30 most frequently observed
 152 declining native plant species in German gardens. Red list categories refer to the German national assessment: EX
 153 = Extinct or lost, CR = Critically Endangered, EN = Endangered, VU = Vulnerable, NT = Near Threatened.

154 Second, gardens (and nurseries; Veken *et al.* 2008) can give species that cannot move
 155 northward fast enough a head start in their range shifts. A key contentious issue here, however,
 156 is who has the power to decide what counts as assisted migration (Lundgren *et al.* 2025). There
 157 is a tendency for conservationists in positions of authority to control the narrative, whereas
 158 democratic dispersion of assisted migration is often considered rash. A productive example
 159 illustrating this tension is the threatened Florida Torreya (*Torreya taxifolia*), an endemic conifer
 160 moved northward by citizen volunteers (Barlow and Martin 2004). Species most likely to

161 benefit from assisted migration through gardens are those with contracting or very narrow
162 ranges that have difficulty shifting their ranges, e.g., due to dispersal barriers or slow growth
163 (*T. taxifolia* is a case in point; Thomas 2011), as well as near-native species with low risks of
164 unintended ecological consequences that help maintain ecosystem functioning, especially in
165 regions where barriers limit species range expansions from further south (e.g., surrounding seas
166 in the case of Great Britain; Gardner and Bullock 2025).

167 Third, gardens, unlike wild areas, provide largely controlled environments where potential
168 risks of assisted migration (e.g., rapid spread) can be quickly detected and addressed.
169 GartenDiv already shows that gardens extend the northward distributions of several native
170 plant species with contracting ranges (Fig. 2). If the right digital platforms exist, contracting
171 and/or climate-threatened native and non-native species could be monitored through
172 community science (enabled by species identification apps) to assess their persistence and
173 performance under new conditions, detect early signs of invasiveness, track ecological
174 interactivity, and identify species that could provide functional insurance when climate-
175 sensitive native species decline. Such platforms and emerging social media networks for
176 gardeners, could further support the creation of participatory digital ex-situ collection databases
177 and mobilize hands-on cultivation knowledge for many species. Creating digital feedback
178 loops, e.g., alerts when a planted species is invasive elsewhere, could connect gardeners'
179 decisions with scientific inputs to assist in conservation efforts.



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181 Fig. 2: **Gardens as sites for participatory assisted migration.** a) Histogram showing the number of observations
 182 of 39 threatened native plant species recorded in the pilot phase of GartenDiv that occurred north of their native
 183 range limits in Germany (x-axis shows the latitudinal difference (°) between each garden record north of the limit
 184 and the species' northern native range boundary). As an example, (b) *Androsace septentrionalis* (CR) was recorded
 185 in gardens up to 3.5° north of its range limit, and (c) *Astrantia bavarica* (VU), was recorded up to 4.7° north.
 186 Native range in A–C is based on Eichenberg (2022) data for the period 1997–2017, which provide distribution
 187 estimates for >2000 species on a 5×5 km grid across Germany with an occurrence probability (P_{occ}) for each grid
 188 cell. For all calculations, cells with $P_{occ}<0.01$ were removed. Purple dots represent northern GartenDiv
 189 occurrences.

190 Together, these functions show how millions of gardens could embed adaptive cultural and
 191 societal processes within conservation under a Spinozist framework, once we stop seeing these
 192 processes as hopelessly alien to nature. A good example that elements of such cultural
 193 conservation infrastructures already exist is provided by the National Plant Collections² in

² <https://www.plantheritage.org.uk/>

194 Britain. This network currently holds c. 730 collections that together safeguard c. 100,000 taxa,
195 ranging from cultivars of garden ornamentals to rare wild species. Some are maintained by
196 botanic gardens or commercial nurseries, but large numbers are maintained by enthusiastic
197 amateurs. Although these collections are organized mainly by plant group and horticultural
198 interest³ rather than by conservation status, they provide a useful foundation and model that
199 could be leveraged for conservation purposes. For example, with scientific input on which
200 species are climate-threatened and with low invasion risk (Dehnen-Schmutz 2011), such
201 collections could serve as valuable sources for assisted migration. Yet despite this potential,
202 gardens are largely overlooked in mainstream conservation strategies.

203 **Tensions with mainstream conservation**

204 Plant conservation has long viewed gardens as artificial environments and “biological deserts”
205 (Elton 1966), as sources of problematic non-native species, and, even when native species are
206 planted, as sites where the use of non-local provenance and hybridization between cultivated
207 and wild taxa may erode the genetic integrity of native populations (cf. Skowronek *et al.* 2023).
208 Maintaining historical community composition and local adaptation is important for
209 conserving (genetic) diversity. Yet what may hold under equilibrium assumptions becomes
210 increasingly uncertain under conditions of rapid change (Grainger *et al.* 2025). These concerns
211 also rest, at least in part, on normative ideas about authenticity in nature, i.e., that “unchanged”
212 populations better conserve (genetic) diversity. A Spinozist view would not categorically
213 exclude these processes because of their risks but would recognize them as processes that may
214 contribute to the ongoing expression of natural diversity (Fig. 3). Importantly, acknowledging
215 that such processes can generate adaptive novelty does not mean dismissing their risks but
216 rather using them as information to refine human participation.

217 A first implication of this view concerns non-native populations. Roughly a quarter of all plant
218 species established outside their native range are endangered in parts of their native range
219 (Staudé *et al.* 2025; Lundgren *et al.* 2026). For example, Moon’s agave (*Agave vera-cruz*,
220 Mexico), and Angel’s trumpet (*Brugmansia suaveolens*, Brazil), are both considered globally
221 extinct in the wild yet have established wild introduced populations in several countries via
222 gardening. The possibility that such introduced organisms may have conservation value is
223 highly contentious under current conservation paradigms. Non-dualist thinking opens space to
224 engage with these cases as part of global efforts to prevent extinction. Endangered, non-native
225 plants (e.g., climate-threatened endemics facing dispersal barriers) could increasingly be part
226 of nursery offerings in biogeographically near-native regions, while measures are taken to limit
227 risks (e.g., avoiding garden plantings in regions rich in native endemics). In this way,
228 conservation-minded gardeners could, whilst still prioritizing native species, plant non-native
229 species, taking into account both risks and benefits, and help reduce their risk of extinction.

230 The same reasoning extends to provenance and genetic mixing. Although conservation
231 prioritizes local provenance to preserve genetic integrity (Skowronek *et al.* 2023), evidence
232 shows that mixing with non-local provenances enhances fitness in changing environments

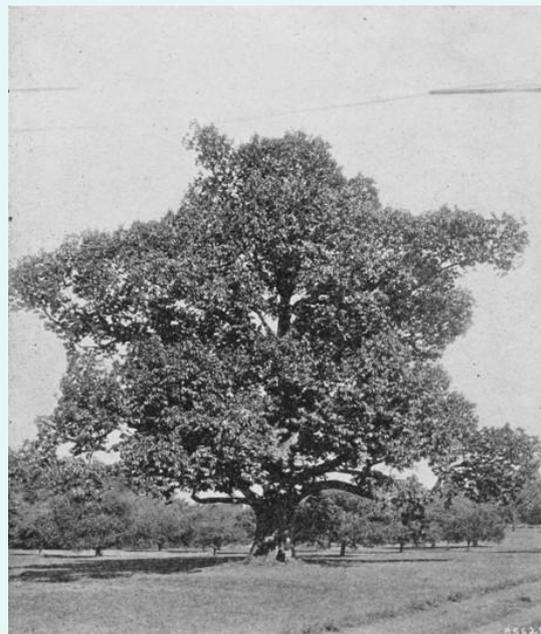
³ <https://www.planteritage.org.uk/national-plant-collections/search-the-national-plant-collections/>

233 (Kottler *et al.* 2021). Similarly, gardens with their mixed and uncertain provenance plant
234 material are seen as having no conservation value and as posing risks of introgression into wild
235 populations. Yet the genetic diversity arising from countless individual choices and multiple
236 sources of plant material (e.g., seed exchanges) can also provide non-trivial amounts of
237 potentially adaptive genetic variation that remains largely unstudied. Rather than excluding
238 gardens on the premise of purity and risk, a Spinozist conservation refines participation. This
239 could mean enabling nurseries (so far sidelined from conservation) to produce local-
240 provenance seeds at scale (so that these are widely available to gardeners), without rigidly
241 restricting genetic mixing. This would increase genetic diversity via widespread production of
242 local-provenance seeds while allowing the gene flow needed for adaptation under change.

243 Horticultural practices, including intentional hybridization, offer similar ambiguities. While
244 mainstream conservation treats hybridization as a threat to species identity, evolutionary
245 ecology recognizes it as a driver of adaptation (Zhang *et al.* 2025) and diversification (Thomas
246 2015; Marques *et al.* 2019). Introductions of non-native plants are known to produce new taxa
247 via hybridization with native congeners (Vallejo-Marín and Hiscock 2016; Faurby *et al.* 2022;
248 Staude and Ebersbach 2023), and the number of garden hybrid varieties that represent new
249 human associated/dependent species is unknown. Such processes are usually dismissed as
250 artificial, yet they expand genetic and phenotypic diversity within species. Some hybrid
251 varieties and cultivars reduce ecological value, while others add value, e.g., by supporting more
252 pollinators than their wild relatives (Ricker *et al.* 2019). While much of this still needs to be
253 better understood, the point here is to highlight that a Spinozist view would not dismiss such
254 processes based on being artificial. Instead, it would recognize them as part of how diversity is
255 generated in a world where human and non-human processes are intertwined.



The **wild apple** (*Malus sylvestris*) is a valuable component of cultural landscapes, yet its existence as a distinct species has become uncertain. Numerous cultivated apple forms, as well as wild relatives from Asia, are closely related to *M. sylvestris* and frequently hybridize with it, producing individuals with traits intermediate between wild and cultivated forms. We are thus losing the wild apple through hybridization with its domesticated relatives, while also gaining many new genotypes that may be better adapted to human-shaped landscapes than the original *M. sylvestris*, which relied on seed dispersal via the digestive tracts of large herbivores, now largely absent. From a mainstream conservation viewpoint, the erosion of the species' genetic integrity is concerning; from a Spinozist viewpoint, it presents diversification and adaptation due to human processes.



The **American chestnut** (*Castanea dentata*) was once a cornerstone of eastern North American forests but was driven to near extinction in the wild by chestnut blight (*Cryphonectria parasitica*), a fungal pathogen introduced from Asia. Breeding programs have sought to restore the species by crossing surviving American individuals with Asian relatives (*C. mollissima* and *C. crenata*), which carry natural resistance to the blight. Successive backcrosses aim to retain the American chestnut's characteristics while incorporating disease resistance. Today, many of these hybrids have been planted in forests, restoration sites, and gardens, where they serve as both experimental populations for evaluating blight resistance and as reservoirs of genetic diversity. These efforts illustrate how horticultural processes, such as intentional hybridization, can be part of conservation.

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Fig.3: **Human participation and adaptive processes.** A Spinozist perspective shifts the focus from judging human-mediated ecological novelty as artificial, to recognizing it as part of how diversity is formed and re-formed in a world where human and non-human processes are inseparable. In the two examples, it is not about uncritical acceptance of processes such as hybridization, but to consider in what way human participation can, and at times already does, support adaptive processes. This could include, for instance, broader collaboration between conservation and horticulture to support endangered species of both ecological and cultural importance. Photo credits: *Malus sylvestris* and *Castanea dentata*, CC0/Public Domain.

264 **Toward a participatory and adaptive ecology of conservation**

265 A participatory and adaptive ecology of conservation is not without risks: ill-considered species
266 choices, inadvertent spread or mixing with harmful ecological consequences, ethical
267 complacency, or economic participation that is performative rather than ecologically
268 meaningful. Yet these risks do not stand in opposition to the potential benefits of involving
269 cultural processes. Instead, they form part of the same ecological process and provide feedback
270 through which participation can be guided and improved. From this perspective, risks are
271 neither ignored nor grounds for excluding participation but are incorporated into an adaptive

272 approach that refines how participation unfolds. Moreover, traditional strategies that are biased
273 toward resisting change also carry risks under rapid changes in global climate (e.g., losing
274 populations altogether, risking loss of ecosystem function). These risks can be mitigated by
275 relying on a diversity of approaches. It is clear that gardens, as an example of a participatory
276 approach, are not a panacea for biodiversity conservation. But this is not an “either–or” choice
277 but a “both/and,” a synergy of approaches that buffer uncertainties under change.

278 Many human activities undeniably drive the biodiversity crisis. We argue, however, that this
279 crisis narrative is also shaped by an asymmetry in attention. That is, conservation frameworks
280 readily recognize the harmful ecological outcomes of human activity while treating novel,
281 potentially beneficial ones with antagonism or ambivalence (Thomas 2017). Such frameworks
282 can restrict our ability to develop creative and productive directions of thinking. Embracing
283 ecological novelty that is mediated or dependent on humans in conservation challenges us to
284 examine how assumptions and advocacy shape scientific interpretation (Cardou and Vellend
285 2023), encouraging a shift from judging processes as “natural” or “artificial” to assessing their
286 actual, empirical contributions to species persistence. Here, we have outlined how humans,
287 while often accelerating disruption beyond many species’ limits, can also participate in
288 accelerating opportunities that help prevent species extinctions. Creating an atmosphere that
289 enables societal participation in a societal problem, i.e., the biodiversity crisis, will be key to
290 broader transformative change.

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12 Notes S1: Conceptual Background

13 Although several Indigenous communities, notably the Indigenous North American creation
14 stories of Sky Woman and Turtle Island (Kimmerer 2013), have long lived by an ethos of co-
15 creation, this perspective was rather revolutionary for Western thought. About two centuries
16 before Darwin's *Origin of Species* established humans as part of evolutionary processes,
17 Baruch Spinoza (1632–1677) articulated a monist vision of nature in his *Ethics* (1677). In his
18 philosophy, which was radical at the time, everything that exists is a modification of a single
19 reality. Spinoza argued that this reality consists of a single substance, which he called *Deus*
20 *sive Natura* (God or Nature). In *Ethics*, he describes it as “a substance consisting of infinite
21 attributes, each expressing eternal and infinite essence” (*Ethics* I p11). Everything we
22 encounter in the world is a *mode* of this substance, i.e., a finite expression of Nature that exists
23 only through its relations with other modes (*Ethics* I d5). These relations unfold through *affects*,
24 understood as the capacities of modes to affect and be affected. Through these affects, modes
25 persist, change, and interact. Every mode strives “as far as it can by its own power, to persevere
26 in its being” (*conatus*; *Ethics* III, p6). Yet its actual power (*potentia*) always depends on the
27 situation and the relations it enters into. Relations that enhance a mode's ability to persist and
28 act increase *potentia*; relations that constrain this ability diminish it. Spinoza rejects the idea
29 that humans stand outside nature. As he writes in the Preface to Part III, many authors “seem
30 to conceive man in nature as a dominion within a dominion,” as if human actions were external
31 to nature.

32 Spinoza therefore provides an ontology in which humans are internal participants in nature
33 rather than external agents acting upon it. His concepts of *conatus* and *potentia* offer a dynamic
34 account of persistence. Beings strive to continue in existence, and their success depends on the
35 relations that expand or diminish their power to act. Stephano (2017) builds on this ontology
36 to formulate an immanent ecological ethic that evaluates relations not by externally imposed
37 standards (for example, historical baselines or ideals of “pristine” nature) but by how they
38 modulate the capacities of beings to persist, adapt, and thrive. As Stephano puts it, “what is
39 *good* is whatever increases a body's capacity to act, and what is *bad* is whatever diminishes
40 this capacity” (2017, p. 157). This shifts ethical attention from protecting nature from humans,
41 preserving historical baselines and privileging biological nativism to fostering relations that
42 enhance collective capacities to persist and evolve. Carroll *et al.* (2023) complement these
43 insights with an ecological framework (adaptive community dynamics) that describes how
44 changes in community composition can increase the match between species' traits and shifting
45 environments and thereby maintain or increase community-level fitness. In that vein, we
46 propose that conservation should not default to treating human-mediated novelty with default
47 antagonism or ambivalence. Instead, conservation can evaluate human-mediated and human-
48 dependent forms of biodiversity (gardening, plant propagation, assisted movement,
49 hybridization) through their effects on the capacities of species and communities to persist and
50 adapt. Such an approach would situate humans and the novel biodiversity they create within
51 the same adaptive dynamics through which ecosystems respond to change.

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