

1 **Masting breakdown in European beech reduces fitness benefits of masting, partly explained by climate**
2 **change**

3 Cherine C. Jantzen^{1,2*}, Joseph B. Burant¹, Marlène Gamelon³, Elisabeth S. Bakker^{4,5}, Marcel E. Visser^{1,2}

4
5 *Corresponding author: c.jantzen@nioo.knaw.nl; ORCID: 0009-0006-0723-2682

6 ¹Department of Animal Ecology, Netherlands Institute of Ecology (NIOO-KNAW), Wageningen, The Netherlands

7 ²Groningen Institute for Evolutionary Life Sciences (GELIFES), University of Groningen, Groningen, The

8 Netherlands

9 ³Université Claude Bernard Lyon 1, CNRS, VetAgro Sup, UMR 5558 Biométrie et Biologie Evolutive, Bat. G

10 Mendel 43 bd du 11Novembre 1918, 69622, Villeurbanne Cedex, France.

11 ⁴Department of Terrestrial Ecology, Netherlands Institute of Ecology, Wageningen, The Netherlands

12 ⁵Wildlife Ecology and Conservation Group, Wageningen University & Research, Wageningen, The Netherlands

13
14 **Abstract**

- 15 1. Masting, which corresponds to highly synchronized but temporally variable seed production, is initiated
16 by weather cues and is thus highly sensitive to climate change. Changes in these cues can lead to a
17 masting breakdown, resulting in a reduction of the fitness benefits of masting through decreasing
18 pollination efficiency and increasing predation risk for seeds.
- 19 2. Here, for the first time, we use 50 years of individual tree data on annual seed production of European
20 beech (*Fagus sylvatica*) in the Netherlands to assess temporal changes in masting patterns and their
21 consequences for the selective benefits of masting. Additionally, we use a novel approach to identify
22 which weather cues initiate reproduction, assess their temporal changes, and test whether they
23 account for the observed changes in masting.
- 24 3. We show that synchrony and inter-annual variation in beechnut production have declined, resulting in
25 a masting breakdown in the mid-2000s, since which there has been constant, but low, seed production
26 each year. Consequently, predation risk increased more than three-fold, while pollination became less
27 efficient, together reducing the fitness benefits of masting. Seed production was driven by precipitation
28 and temperatures in the year of seed fall and the two preceding years, but the periods within the year
29 in which trees respond to each climate variable differ in both timing and duration. Interestingly, only

30 temperature, not precipitation, has changed over time, but this change only partly explained the
31 observed changes in masting patterns.

32 4. *Synthesis.* Masting breakdown is shown across the species range, but its fitness consequences remain
33 understudied, because detailed, individual-level, long-term data are required but still rare. By using
34 such a dataset, we here provide crucial evidence for the negative consequences of masting breakdown
35 for beeches through reduced pollination efficiency and increasing predation risk. Using a new
36 methodology, we further underline the strong effects of weather cues on reproduction, while showing
37 that changing climate alone cannot be driving the masting breakdown and must interact with currently
38 unidentified factors.

39

40 **Keywords**

41 Mast seeding, reproductive ecology, *Fagus sylvatica*, climate change, synchrony, weather cues, predator
42 satiation, pollen efficiency

43 **Introduction**

44 Many perennial plant species reproduce by mast seeding (or masting), i.e., individual seed production is highly
45 variable between years but strongly synchronized with conspecifics within a year, causing total annual seed
46 output to vary between high and low numbers of seeds produced (Herrera et al., 1998; Kelly, 1994). Driven by
47 weather cues, masting is sensitive to climate change (Bogdziewicz, Hacket-Pain, et al., 2021) and changes in the
48 magnitude and frequency of seed production can have strong effects on the fitness of the plants themselves
49 (Bogdziewicz, Kelly, Thomas, et al., 2020), as well as on a variety of species that depend on the seeds they
50 produce (Dri et al., 2025; Ostfeld & Keesing, 2000).

51

52 Seeds, such as acorns, chestnuts or beechnuts, provide a substantial food resource for a variety of consumer
53 species with strong bottom-up effects on their population dynamics (Dri et al., 2025; Perdeck et al., 2000; Touzot
54 et al., 2020; Zwolak et al., 2024). Through the high inter-annual variability in seed production, masting creates a
55 satiation-starvation cycle, in which seed consumers get satiated in years of high seed crop, increasing consumer
56 populations, and starved in the following year of low seed production, reducing consumer populations again
57 (Janzen, 1971). Thereby predation pressure is regulated by year-to-year variation in seed production, thus
58 enhancing the chance of surviving seeds in high seed years. If seed production changes in magnitude and
59 becomes less variable between years, the satiation-starvation cycle becomes less pronounced and predation
60 pressure increases. Such fitness benefits, where high reproductive investments yield larger returns, are defined
61 as economies of scale (EOS) and ultimately reduce the costs per surviving offspring (Kelly & Sork, 2002; Norton
62 & Kelly, 1988; Pearse et al., 2016).

63

64 Another well-supported EOS, besides predator satiation, is pollination efficiency: through the synchrony of tree
65 reproduction, and hence the simultaneous flowering, the amount (Kelly, 1994; Moreira et al., 2014) and genetic
66 diversity (Ascoli et al., 2021) of pollen in the air are maximised, which increases the chances of successful
67 pollination in the wind-pollinated species. Following the pollination efficiency hypothesis (Kelly, 1994), a
68 reduction in synchrony is expected to reduce successful pollination and thereby seed production. Pollination can
69 further be affected by climate change, as it is sensitive to weather events, such as heavy rainfall (Tradowsky et
70 al., 2023), that can reduce pollination success (Nussbaumer et al., 2020). These EOS make masting a beneficial
71 reproductive strategy, despite the selective disadvantages of missed reproductive opportunities in years where

72 no seeds are produced, and increased density-dependent competition in years of high seed production
73 (Bogdziewicz et al., 2024). The effectiveness of EOS is, therefore, crucial for masting to remain beneficial, leading
74 to selection for high inter-annual variability and high reproductive synchrony among individual trees
75 (Bogdziewicz, Kelly, Tanentzap, et al., 2020; Pesendorfer et al., 2024). With changes in climate, these masting
76 dynamics can change (Foest et al., 2024), but the consequences of these changes for the effectiveness of
77 economies of scale remain understudied.

78

79 Masting trees respond to environmental cues (i.e., environmental signals in certain periods of the year) to
80 initiate reproduction and individuals must respond similarly to the same cue to reproduce synchronously (Pearse
81 et al., 2016). Identifying such cues is crucial for understanding the proximate drivers of masting and forecasting
82 how masting may change with changing climate conditions (Journé et al., 2023). So far, many studies (e.g.,
83 Bogdziewicz, Kelly, Thomas, et al., 2020; Nussbaumer et al., 2020; Övergaard et al., 2007) have tested for climatic
84 effects on seed production by using climate variables from periods that are mostly arbitrarily predefined as
85 selected calendar months (e.g., summer being June and July). Only recently, studies have tested for the exact
86 periods in which trees are most sensitive to climatic cues (Bogdziewicz, Hacket-Pain, et al., 2021; Bogdziewicz,
87 Journé, et al., 2023; Journé et al., 2024), resulting in several different methodologies to identify cueing windows
88 (Journé, Simmonds, et al., 2025). While all of them give good approximations of these windows for a single
89 climate variable (Journé, Simmonds, et al., 2025), they mostly do not consider the complex interplay of different
90 climate variables on seed production (but see: Journé, Kelly, et al. 2025). Further development of cue
91 identification methodologies is therefore needed to incorporate the effects of different weather cues.

92

93 For European beech (*Fagus sylvatica*, hereafter beech), most commonly reported weather cues are
94 temperatures and precipitation in the two summers preceding seed fall (Bogdziewicz, Kelly, Tanentzap, et al.,
95 2020; Journé et al., 2024; Vacchiano et al., 2017), with the opening of the temperature window fixed on the
96 summer solstice (Journé et al., 2024; Journé, Simmonds, et al., 2025). Beech is an important forest forming
97 species in Central Europe (Leuschner & Ellenberg, 2017), which shows high spatiotemporal stability in its cue
98 sensitivity (Bogdziewicz, Journé, et al., 2023), inducing synchrony in reproduction over large distances, since
99 weather conditions are also synchronous over large spatial scales (Bogdziewicz, Hacket-Pain, et al., 2021).
100 However, trees still vary in their reproductive output despite experiencing the same climate, indicating that

101 weather alone cannot drive masting and suggesting an interplay with internal resource dynamics (Kelly et al.,
102 2025; Müller-Haubold et al., 2015; Pearse et al., 2016). Depending on resource availability, trees allocate a
103 fraction of the acquired resources to reproduction and reproduction can only be triggered by a cue if the
104 accumulated resources reach a minimum threshold (resource budget model; Crone & Rapp, 2014; Schermer et
105 al., 2019). If resource levels are low, reproduction can be largely suppressed even in presence of strong cues,
106 leading to reduced seed production (Kelly et al., 2025). When cue frequency increases, as through climate
107 change, the interval between cues is not sufficient to replenish depleted resources after reproduction and
108 reproductive output is reduced.

109

110 Such climate change effects on masting patterns are already visible in many beech populations where high seed
111 years became more frequent (Sweden: Övergaard et al., 2007, Germany: Gruber, 2003; Schmidt, 2006,
112 Switzerland, United Kingdom: Nussbaumer et al., 2016) and both inter-annual variability and synchrony of seed
113 production have declined (Bogdziewicz, Kelly, Thomas, et al., 2020; Foest et al., 2024, 2025). This so-called
114 *masting breakdown* can ultimately lead to strong declines in the fitness benefits of masting for the trees because
115 EOS are becoming less effective, leading to less successful pollination and higher predation (Bogdziewicz, Kelly,
116 Thomas, et al., 2020; Bogdziewicz, Kelly, et al., 2023). These negative consequences of changing masting patterns
117 have so far only been shown in a spatially limited set of beech populations of the UK (Bogdziewicz, Kelly, Thomas,
118 et al., 2020; Bogdziewicz, Hackett-Pain, et al., 2021), as long-term, individual-based data is needed that include
119 information on pollination and predation rates, which are still rare due to the logistical challenges and high-
120 effort data collection. It is therefore important to assess whether these climate-induced changes in masting
121 equally reduce the selective benefits of masting across the species range of European beech.

122

123 We here, for the first time, use an individual-level, 50-year dataset on beechnut production in the Netherlands
124 to investigate: 1) whether synchrony in, and inter-annual variability of, seed production changed over time, and
125 if so, how, 2) which consequences this had on the selective benefits of masting, specifically in relation to
126 pollination and predation, 3) which periods of temperature and precipitation within the year are most influential
127 for beechnut production, using a novel approach of identifying the windows of highest sensitivity, and 4)
128 whether temporal changes in the weather in these windows can explain the changes in masting patterns we
129 observe.

130 **Materials and Methods**

131 ***Study site and data collection***

132 Beechnuts have been collected annually from 1976 to 2025 in the National Park De Hoge Veluwe, Netherlands
133 (52.038°N, 5.857°E). The same set of around 30 to 35 trees (variation occurs throughout the years through
134 cutting or death of trees) is sampled yearly in mid-October by placing four metal squares (0.25 x 0.25 cm;
135 0.625m²) underneath each tree in a straight line of a fixed direction. The first square is located half a meter from
136 the trunk, the last square below the tip of the largest overhanging branch and the remaining two plots in equal
137 distances in between. In each plot, all full and partial nuts within a square are collected and categorised into
138 either full, empty or predated (split into nuts with caterpillars and predation by other animals), and the number
139 of nuts per category is counted. For a detailed description of the study site and data collection see Anonymous
140 (*under review*). This method cannot exclude a potential bias through predation or dispersal of nuts before
141 collection, leading to an underestimation of the number of nuts predated. However, generally the crop size
142 estimates of the ground plot method can be considered highly reliable (Touzot et al., 2018). For the analyses,
143 nuts in each category of all four plots per tree per year are summed and multiplied by four to get estimates of
144 the number of nuts per square metre. Permission for field work was granted by the National Park De Hoge
145 Veluwe.

146

147 ***Climate data***

148 Data on daily minimum and maximum temperatures and daily precipitation sums were retrieved from the Royal
149 Dutch Meteorological Institute (KNMI) for the weather station De Bilt as the closest weather station to the study
150 area (~ 47 km) for which homogenized data are available (i.e., corrected for changes in measurement methods
151 and relocation of the weather station; Note: the correlation of the mean daily temperatures between De Bilt
152 and Deelen, a weather station within one kilometre of the study area, is $r = 0.994$.) Mean daily temperature was
153 calculated as $(\text{minTemp} + \text{maxTemp}) / 2$.

154

155 ***Statistical analyses***

156 All statistical analyses were done in R (version 4.5.2) (R Core Team, 2025). Generalised linear mixed models
157 (GLMMs) were fitted with *glmmTMB* (version 1.1.13, Brooks et al., 2017) and generalised additive models
158 (GAMs) were fitted with *mgcv* (Wood, 2011). Model fits were assessed with the packages *DHARMA* for GLMMs

159 (Hartig, 2024) or *gratia* for GAMs (Simpson, 2022). Model predictions for GLMMs were calculated with *ggeffects*
160 (Lüdecke, 2018).

161

162 **a) Mastig metrics**

163 We tested whether the annual seed production of individual trees has changed over the study period by
164 modelling annual total beechnut counts (sum of full, empty and predated nuts) per tree as the response variable,
165 rather than using decadal averages (as suggested by Bogdziewicz, Kelly, et al., 2023). Although the large
166 between-tree and between-year variation in seed production included in annual values might mask part of the
167 temporal trend in seed production, we were also interested in temporal changes in the occurrence of years
168 without reproduction, which the decadal averages did not capture. We therefore use a zero-inflated negative
169 binomial GLMM to model both the probability of a year without reproduction (zero-inflation model; binomial)
170 and the trend in the number of nuts produced, if there is reproduction (conditional model; negative binomial).
171 When these two models are taken together, the temporal trend in annual seed production can be estimated.
172 Year as a continuous variable was fitted both as fixed effect and as a zero-inflation parameter. We fitted tree
173 identity (tree ID) as a random intercept to account for multiple measurements for a given tree over time and
174 accounted for the effect of previous year's seed production (Pesendorfer et al., 2020) by fitting the number of
175 nuts produced by the considered tree of the previous year (i.e., lag-1 nuts) as a fixed effect.

176

177 To assess temporal trends in inter-annual variation and synchrony in beechnut production, we used a sliding
178 window approach with a window size of five years and one year step size (similar to Foest et al., 2025). Within
179 each window, synchrony was calculated as the Pearson correlation coefficient in annual beechnut production
180 for each pair of trees. Tree-level synchrony was then calculated as the mean Pearson coefficient over all pairwise
181 correlations and population-level synchrony as the mean over all trees (for each window). Interannual variability
182 was calculated as the coefficient of variation (CV_i) of annual seed production for each tree (i) within each
183 window. Since the commonly used Pearson CV_i (${}^P\text{CV}_i = \text{SD}/\text{mean}$) has been criticised as being sensitive to outliers
184 (Lobry et al., 2023), we additionally calculated the newly proposed Kvålseth's CV_i (${}^K\text{CV}_i = \sqrt{{}^P\text{CV}_i^2 / (1 + {}^P\text{CV}_i^2)}$;
185 Kvålseth, 2017), which is a stabilized transformation of Pearson CV bounded between 0 and 1. For
186 comparability with other studies, we report ${}^P\text{CV}_i$ in Supporting Information A. Because we do not expect

187 synchrony and CV_i to change linearly over time, temporal trends were modelled using generalized additive
188 models (GAM) with restricted maximum likelihood (REML) and first year of the sliding window as a smoothing
189 term. For this part of the analysis, trees were excluded from a window if they had less than three out of five
190 years of observations in that window.

191

192 **b) Economies of scale**

193 To test for the selective benefits of masting for the trees through economies of scale, we used binomial GLMMs
194 with a logit link function, first-order temporal autocorrelation with a 1-year time lag to account for an effect of
195 the previous year's pollination and predation ratio, respectively, and tree ID as a random effect. We combined
196 the hypotheses of the *satiation effect*, which predicts that a smaller proportion of nuts is eaten when the number
197 of nuts is high, and the *starvation effect*, which predicts that a smaller proportion of nuts is eaten in a year when
198 the previous year had much fewer nuts, in one model. In that model, number of nuts predated against total
199 number of nuts was fitted as the response variable (using the *cbind()* specification in R) and explanatory variables
200 were the total number of nuts, year, their interaction, their quadratic terms, the ln-transformed ratio of total
201 number of nuts in year T to the total number of nuts in T1 (i.e., $\ln(\text{ratio} + 1)$ to avoid logarithms of zero), its
202 interaction with year and its quadratic term. Predated nuts included all nuts predated by caterpillars and other
203 animals.

204 To test the *pollination efficiency* hypothesis, which predicts higher pollination success under high flowering
205 synchrony between individuals and high flower production of conspecifics, full nuts (including full and predated
206 nuts) were used as a proxy for successful pollination, because beechnuts can only develop a kernel when
207 pollinated (Nilsson & Wastljung, 1987). Hence, the response variable (proportion of successfully pollinated nuts)
208 was defined as the total number of full nuts against the total number of nuts (using *cbind()* specification). Fixed
209 effects were the mean number of nuts produced by conspecifics (to account for the different numbers of
210 conspecifics measured across years), within-site within-year synchrony (defined as the coefficient of variation
211 within year between trees), their interaction, year, and the quadratic terms of number of nuts of conspecifics
212 and year.

213 For both models, the non-significant quadratic term for year was dropped, because this was only fitted to
214 explore whether there was a non-linear temporal trend, while other non-significant quadratic terms were kept

215 in the model, as they were fitted based on *a priori* assumptions. Year as an explanatory variable was used as an
216 ordinal variable to avoid convergence problems and data was restricted to observations with at least one nut
217 (as otherwise proportions cannot be calculated), resulting in $n = 1048$ observations (i.e., tree-year
218 measurements) for the pollination model, while for the predation model, infinite and non-defined values for the
219 $\log(\text{nut ratio})$ were additionally filtered out ($n = 691$ observations).

220

221 ***c) Sensitivity windows for climate variables***

222 To model the effects of climate variables on the annual beechnut production, we first identified the period of
223 the year in which seed production is most sensitive to the respective climate variables. First, to create a base
224 model, climate variables over three years (T_0 = year of seed fall, T_1 and T_2 = one and two years before seed fall,
225 respectively) were calculated for windows commonly described in the literature (e.g., Bogdziewicz, Kelly,
226 Thomas, et al., (2020)): maximum temperature and precipitation sum in summer (June & July) of T_1 , associated
227 with differentiation of flower primordia (Gruber, 2003), and in summer of T_2 , associated with resource
228 accumulation (Richardson et al., 2005), as well as mean temperatures in the growing season (1 May to 31 August)
229 of T_0 , and precipitation sum in spring of T_0 (1 March to 30 April), associated with flower abortion and pollination
230 success (Nussbaumer et al., 2020). These climate variables were fitted as fixed effects in a zero-inflated negative
231 binomial GLMM with annual total beechnut count per tree as the response variable, previous year's seed
232 production as a fixed effect, tree ID as a random effect, the `nbinom2()` function with a log-link was used to specify
233 the negative binomial error distribution, and the zero-inflation formula included all explanatory variables (= base
234 model). Note that this approach takes into account the error structure of the data (zero-inflated negative
235 binomial) whereas commonly used approaches do not (see also below). This base model was used as the start
236 model for an iterative sliding-window approach, in which we tested all possible windows between spring
237 equinox (21 March) and fall equinox (22 September) of a window length between one and 20 weeks (i.e., 7 to
238 140 days, 15343 tested windows). For each window, the mean of the minimum, maximum and mean daily
239 temperature, and the sum of the daily precipitation sum were calculated and z-transformed (i.e., $X - \text{mean}/\text{SD}$
240 within window over the years of the study) to bring the variables on the same scale and minimise convergence
241 problems.

242

243 To find the best window for the first climate variable (temperature in T1), we fitted a separate GLMM for every
244 window (= 15343 models) by modifying the base model so that temperature in T1 varied according to the
245 window tested. We tested all models separately for minimum, maximum and mean temperature in T1 to
246 determine which temperature measure describes seed production best. For that, we chose the best window for
247 each of them based on the lowest AIC value (Burnham & Anderson, 2002), respectively, compared the AIC values
248 of the three best models and chose the climate variable with the lowest AIC to be used in all further models (this
249 test was only done in the first iteration). For the next focal climate variable, we then replaced the values for
250 temperature in T1 in the base model with the ones of the selected best window. The same procedure was used
251 for temperatures in T2 and T0 (again also testing weather mean, minimum or maximum temperature is best
252 suited) and next, precipitation in the three years. After all climate variables of the base model have been
253 exchanged for the values in their respective best windows of the first iteration, we performed a second, third
254 and fourth iteration in the same way. In the starting models of each iteration, all non-focal variables were set to
255 the values of their respective best window of the previous iteration. This iterative approach ensured that the
256 initial order in which variables were tested did not influence the final windows selection. Since the best windows
257 of the fourth iteration clearly indicated a best window and only showed slight deviations from the ones of the
258 third iteration, if any, we did not run further iterations (details on the models are shown in Table S1).

259

260 Sliding windows have been shown to be a robust and reliable way to identifying sensitivity windows for masting
261 (Journé, Simmonds, et al., 2025). The models used within the sliding window approach differ however between
262 studies and mostly either calculate correlation coefficients (Journé et al., 2024) or linear regressions (Journé,
263 Simmonds, et al., 2025) between one climate variable and the number of beechnuts per window. Since these
264 approaches do not account for the interplay between different climatic drivers and the highly-zero inflated
265 nature of masting data, we used a more complex model to address these limitations. For comparability with
266 previous studies, we additionally applied the approach used in Journé et al. (2024) to our data (see Supporting
267 Information B for more details on comparing both approaches).

268

269 ***d) Climatic effects on seed production***

270 Since collinearity between all six climate variables was low (absolute $\rho < 0.4$; see Supporting Information B.4),
271 we fitted the six climate variables tested above, all set to their respective best window, back into the GLMM

272 used for the sensitivity window analysis to test how each affects annual seed production (= final model). We
273 corrected p-values of each climate variable for the number of windows (n = 15343) tested using a Bonferroni
274 correction. Model fit was assessed by comparing observed and predicted values using a Spearman rank
275 correlation and visually comparing seed production patterns. To better understand whether there is unexplained
276 temporal variation, we additionally fitted year back into the model. Temporal trends in climatic variables were
277 modelled with linear models of each climate variable against year.

278

279 **Results**

280 **a) Masting metrics**

281 Over the study period of 50 years, seed production of 81 trees has been recorded. Annual seed production
282 alternated between high seed years and years without any reproduction for at least part of the study period
283 (Figure 1a). The probability of a year in which a tree does not reproduce at all (i.e., a zero year) significantly
284 decreased over time ($\beta = -0.089 \pm 0.007$, $z = -13.49$, $p < 0.001$; Figure 1b), whereas there was no significant
285 change in the number of seeds produced per tree, for years in which a tree produced at least one nut ($\beta = 0.004$
286 ± 0.004 , $z = 1.12$, $p = 0.264$; Figure 1c, red line). When integrating these effects, i.e., the probability of seed
287 production and the average number of seeds produced per tree when there are seeds produced, seed
288 production of the population shows a slightly increasing trend (Figure 1c, black line). The previous year's seed
289 production (T1) was negatively affecting seed production in the current year (T0), when nuts were produced
290 ($\beta = -0.0009 \pm 0.0001$, $z = -7.268$, $p < 0.001$) and increased the probability of a zero-year in T0 ($\beta = 0.001 \pm 0.0002$,
291 $z = 5.654$, $p < 0.001$).

292

293 For the analyses of synchrony and inter-annual variability (CV_i), 46 windows of five years were used in the sliding
294 window approach, each containing at least 24 trees. The between-tree synchrony of beechnut production in the
295 population changed significantly throughout the study period (edf = 3.325, $F = 16.7$, $p < 0.001$, $n = 46$), decreasing
296 from between 0.77 and 0.98 in the early windows to around 0.39 in the last window (starting 2021; Figure 1d).
297 Already in the two windows starting in 2005 and 2006, respectively, synchrony showed a sudden strong decline
298 but increased back to its previous level afterwards until there is a second, even stronger decrease in synchrony
299 in the windows starting in 2019 and 2020, where trees have fallen out of synchrony almost completely with a
300 mean synchrony of 0.09. Beechnut production of the population additionally became less variable between

301 years ($\text{edf} = 4.517$, $F = 53.89$, $p < 0.001$, $n = 46$), with the mean ${}^k\text{CV}_i$ decreasing from around 0.87 at the start of
302 the study period to around 0.61 in the last window (Figure 1e). Similar to synchrony, the windows starting in
303 2019 and 2020 show a stronger drop in ${}^k\text{CV}_i$ to a value of 0.51. For ${}^p\text{CV}_i$ we found similar results, see Supporting
304 Information A.

305

306 **b) Economies of scale**

307 The proportion of predated nuts increased from 6% in the beginning of the study period to 20% in 2025 ($\beta = 0.04$
308 ± 0.006 , $z = 7.167$, $p < 0.001$; Figure 2a). In line with the predictions of the *satiation effect*, the proportion of
309 predated nuts declined with the total number of nuts produced and this relationship was better described by a
310 linear than a quadratic term (linear term: $\beta = -0.0027 \pm 0.0007$, $z = -3.755$, $p < 0.001$, quadratic term: $\beta = 4 \cdot 10^{-7}$
311 $\pm 3 \cdot 10^{-7}$, $z = 1.498$, $p = 0.134$) and changed over time (interaction: $\beta = 0.00004 \pm 0.00001$, $z = 2.928$, $p = 0.003$;
312 Figure 2c). Additionally, there is evidence for a *starvation effect*, since the proportion of predated nuts changed
313 with the ratio of seed production in two consecutive years (quadratic term: $\beta = -0.056 \pm 0.028$, $z = -1.990$, $p =$
314 0.047 , linear term: $\beta = 0.727 \pm 0.204$, $z = 3.566$, $p < 0.001$; Figure 2d). In the beginning of the study period, the
315 proportion of predated nuts increased when there were few seeds in the previous year and many seeds in the
316 current year, whereas this relationship was reversed in the second half of the study period (interaction: $\beta = -$
317 0.021 ± 0.004 , $z = -4.866$, $p < 0.001$).

318 *Pollination efficiency*, measured as percentage full nuts, decreased over time ($\beta = -0.010 \pm 0.005$, $z = -2.128$, $p =$
319 0.033) from around 47% in 1976 to 36% in 2025 (Figure 2b). The proportion of successfully pollinated seeds
320 changed quadratically with the number of nuts produced by conspecifics (quadratic term: $\beta = -9 \cdot 10^{-6} \pm 1 \cdot 10^{-6}$, z
321 $= -8.455$, $p < 0.001$, linear term: $\beta = 0.014 \pm 0.002$, $z = 8.539$, $p < 0.001$; Figure 2e) and the relationship between
322 pollination ratio and number of nuts of conspecifics changed with between-tree synchrony (interaction: $\beta = -$
323 0.005 ± 0.001 , $z = -3.990$, $p < 0.001$), here measured as the inverse of synchrony (i.e., CV_p). If trees are highly
324 synchronous (low CV_p), pollination success strongly increased with the number of nuts produced by conspecifics
325 (being a proxy for the number of pollen) until reaching a threshold after which pollination success declined again.
326 Under low synchrony (high CV_p), however, pollination success was overall lower and declined with increasing
327 seed production of conspecifics. Synchrony itself showed a negative effect on pollination success, i.e., the
328 proportion of successfully pollinated seeds was lower the more synchronous trees flower (i.e., the lower the

329 CV_p : $\beta = 0.31 \pm 0.15$, $z = 2.031$, $p = 0.042$). This seems to contradict the interaction effect of CV_p and the number
330 of nuts of conspecifics (showing a positive effect of synchrony on pollination success) and might be due to the
331 collinearity of CV_p and the number of nuts of conspecifics, since high numbers of nuts are only produced by
332 conspecifics under high synchrony and vice versa.

333

334 ***c) Sensitivity windows for climate variables***

335 While the best windows for individual climate variables still shifted strongly between the first and second
336 iteration, they started stabilizing in the third iteration and showed almost no change from the third to the fourth
337 iteration (maximum shift was three days; see table S2 for details on all iterations). Maximum daily temperature
338 was identified as the best temperature measure for T1 and T2 ($\Delta AIC = 40.73$ to mean temperature and $\Delta AIC =$
339 5.38 to minimum temperature for T1, $\Delta AIC = 33.95$ and $\Delta AIC = 15.06$, respectively, for T2) and mean daily
340 temperature for T0 ($\Delta AIC = 19.49$ to maximum temperature, $\Delta AIC = 51.26$ to minimum temperature). For all
341 climate variables in the fourth iteration, there was one distinct trough in the AIC values when plotted against
342 start day of each of the 15343 windows tested, making us confident that the selected windows (i.e., windows
343 with the minimum AIC) are not random (Figure S2).

344 Window size was specific to each climate variable, with a minimum of seven days (for precipitation in T0) and a
345 maximum of 60 days (for precipitation in T2). In the year of seeding, beeches showed the highest sensitivity to
346 mean temperatures in late June (22/06 to 01/07) and precipitation in the second half of August (14/08 to 20/08).
347 In the year prior to seeding (T1), maximum daily temperatures of late June to the end of July (20/06 to 29/07)
348 and precipitation in late May (25/05 to 04/06) were most relevant, while two years prior to seeding (T2) late
349 summer temperatures (26/08 to 02/09) and summer precipitation (05/06 to 03/08) were most related to seed
350 production (Figure 3a).

351

352 ***d) Climatic effects on seed production***

353 Throughout the study period, only temperatures in the best windows of T0 and T1 have significantly changed
354 (T0: $\beta = 0.028 \pm 0.008$, $t = 3.273$, $p = 0.002$; T1: $\beta = 0.025 \pm 0.009$, $t = 2.868$, $p = 0.006$; Figure 3b), while there
355 was no significant change in temperature in the window for T2 ($\beta = 0.011 \pm 0.009$, $t = 1.224$, $p = 0.227$) and

356 precipitation sums in any of the best windows (T0: $\beta = 0.008 \pm 0.009$, $t = 0.903$, $p = 0.371$; T1: $\beta = 0.013 \pm 0.009$,
357 $t = 1.414$, $p = 0.163$; T2: $\beta = 0.001 \pm 0.009$, $t = 0.139$, $p = 0.890$; Figure 3c).

358 Fitting all these climate variables set to their respective best windows of the fourth iteration into the final model
359 showed that a cold and wet climate in T2, a warm and dry climate in T1, and a warm and wet climate in T0
360 (despite the temperature effect not being significant) increase the number of nuts produced, if a tree reproduces
361 (conditional model, Table 1). The probability of a tree not reproducing in a year was also significantly affected
362 by all climate variables, with cold temperatures in the periods of T0 and T1 and warmer temperatures in T2
363 increasing the probability of a year without reproduction, as well as dry periods in the windows of T0 and T2,
364 and wet periods in T1 (zero-inflation model, table 1; note that positive estimates indicate that an increase in the
365 fixed effect leads to a higher probability of a zero-year, i.e., no reproduction, and vice versa for negative
366 estimates). Fitting year back into the model showed that there was no unexplained variation remaining for the
367 conditional model ($\beta = 0.0009 \pm 0.0031$, $z = 0.276$, $p = 0.783$), but year was significant in the zero-inflation model
368 ($\beta = -0.044 \pm 0.013$, $z = -3.437$, $p = < 0.001$), indicating that the climate variables considered here cannot fully
369 explain the variation in the probability of a year without reproduction.

370 Overall, the final model predicted annual seed production well ($\rho = 0.86$, Figure 4a) but generally underestimated
371 large seed crops. However, consistent with a transition in masting dynamics in the past two decades, the model
372 tended to overestimate the number of nuts produced in recent years and predictions deviated more strongly
373 from observed values. Comparing temporal patterns in observed and predicted seed production (Figure 4b)
374 showed that these were well aligned until the early 2000s, whereas deviations in the later 2000s were caused
375 by a more consistent pattern of peaks and valleys predicted by the model, while peaks declined more strongly
376 in the observed seed production. Although the model captured changes in masting, since peaks in annual seed
377 production became shorter and less frequent over time and predicted seed output rarely reaches zero in the
378 past decades, these predicted changes were less severe than observed changes and the model, therefore, could
379 not accurately capture the masting breakdown.

380

381 ***Discussion***

382 Based on 50 years of annual seed production data, we demonstrated that beech trees in the Netherlands show
383 a clear *masting breakdown* beginning around 2008. Since then, years without seed production become highly
384 unlikely, annual beechnut production stabilises at a low number, and between-tree synchrony and inter-annual
385 variability in seed production declined strongly. This ultimately led to a weak decline of pollination efficiency
386 and a strong increase in the proportion of predated seeds, since economies of scale became less efficient. In line
387 with the literature (Journé et al., 2024; Kelly et al., 2013; Vacchiano et al., 2017), seed production was driven by
388 temperatures and precipitation across three years (year of seed fall and both years before that), but we
389 identified partly different cueing periods for each of these weather variables than previously reported. From the
390 six weather cues, only temperatures in T0 (year of seed fall) and T1 (year prior to seed fall) showed significant
391 warming over time. However, these changes in temperature alone cannot explain the observed masting
392 breakdown, because the final model of annual beechnut production explained by climate cues predicted less
393 extreme changes in seed production.

394 Reproduction through masting can only be beneficial if costs per surviving offspring are reduced (Kelly & Sork,
395 2002), which is no longer the case when masting breaks down. With smaller seed crops in recent years, satiation
396 of seed consumers became less efficient and resulted in an increase in the proportion of predated beechnuts.
397 Consistent, low-level seed production should offer a more stable food supply for consumers, which can stabilise
398 or increase seed consumer populations and, thereby, further increase predation pressure. Due to the decline in
399 inter-annual variation in seed production we would have expected the predator starvation benefits (Janzen,
400 1971) to have weakened in recent years, since the ratio of seed production of two consecutive years became
401 much smaller. In contrast, the starvation effect became seemingly stronger, since the proportion of predated
402 beechnuts is reduced more strongly in recent years if there are many more nuts in one year compared to the
403 preceding year (i.e., $T/T-1$ is large). Comparing this relationship between early and late years of the study is,
404 however, difficult. While early years showed a more clustered distribution at the higher end of the range of
405 interannual differences in seed production, the distribution in recent years has become more uniform. Hence,
406 variation in seed production of two consecutive years was never low in early years, but is commonly low in
407 recent years, and overall seed predation is higher now than 50 years ago, making the decline in the predation
408 with increasing seed production ratio more visible now, but the starvation effect not necessarily stronger.

409 Consequently, the changing masting patterns can no longer maintain the mechanisms that reduce the
410 proportion of predated seeds.

411 Besides stabilising seed production, the breakdown of reproductive synchrony also reduces pollination
412 efficiency. While pollination success increased with higher reproductive effort by conspecifics under high
413 synchrony, it declined under low synchrony. This could result from only a small fraction of trees producing large
414 amounts of nuts (and pollen) under low synchrony, so that local pollen abundance is only high around these
415 trees and the overall pollen concentration in the air may be reduced for more distant trees, even though
416 dispersal distances of beech pollen can be very large (Belmonte et al., 2008). Hence, the chance of successful
417 pollination is reduced at the population level despite some individual trees producing many seeds. Even though
418 masting breakdown is reported throughout Europe (Foest et al., 2024), long-term individual-level data on
419 pollination and predation is rare, making our study valuable evidence for the severe consequences of masting
420 breakdown for the selective benefits of masting.

421 Masting breakdown is driven by changes in the weather cues and their effects on internal resource dynamics
422 (Kelly et al., 2025), which makes identifying these weather cues the first step towards understanding the
423 underlying mechanism of masting breakdown. Summer temperatures in the two years before seed fall have
424 repeatedly been shown to affect beechnut production (Bogdziewicz, Kelly, Thomas, et al., 2020; Kelly et al.,
425 2013; Nussbaumer et al., 2020; Vacchiano et al., 2017), but only recently have cueing windows in summer been
426 more precisely defined to start around the summer solstice (Journé et al., 2024). While we find the same for
427 temperatures in T1, the cueing window for temperature in T2 (two years prior to seed fall) is much later in
428 summer and, spanning only eight days, rather short in duration. For T0, we expected spring temperatures to
429 affect seed production (Nussbaumer et al., 2020), but found seed production to be most sensitive to a window
430 that also opened at the solstice. Temperatures in this period did not affect the number of nuts produced, but
431 only whether or not reproduction occurs. Given that seed fall starts in early autumn and seeds are therefore
432 almost fully formed in late summer, the mechanism behind this temperature effect on the occurrence of
433 reproduction remains unclear, as well as the effect of the short and late window for precipitation in T0.

434

435 Unlike temperature, precipitation cues were not anchored to the summer solstice, neither in our study nor in
436 previous ones (Journé et al., 2024) and while the window in T1 is unexpectedly short, both precipitation cues of
437 T1 and T2 fall into the expected summer period. All three precipitation cues significantly affected beechnut
438 production in our population, despite precipitation often being deemed lower priority as a driver of seed
439 production than temperature (Drobyshev et al., 2010; Vacchiano et al., 2017). Given that precipitation affects
440 trees through the soil water content, which also depends on local factors (e.g., soil porosity, drainage, water
441 holding capacity, evaporation and vegetation cover; Grayson et al., 1997), precipitation effects are expected to
442 show stronger local differences than temperature. Soil water content would therefore be a better measure of
443 water availability, but such information was not available for this study. Integrating this could strengthen the
444 model of environmental drivers of seed production further, especially as soil water content strongly affects
445 nutrient uptake rates in beeches (Geßler et al., 2005) and thereby also affects resource dynamics.

446 The differences in weather cues we find here compared to previous studies might reasonably be attributed to
447 methodological differences, rather than biological ones. Several methodologies exist to identify weather cues
448 and have all been shown to reliably detect a benchmark window (Journé, Simmonds, et al., 2025). However,
449 these methods mostly do not consider the interplay of different climatic drivers on seed production, nor do they
450 assess the differential effect of climate on the occurrence of reproduction and the number of nuts produced, if
451 reproduction occurs. Previous studies often log-transformed seed production data to normalise it (Journé et al.,
452 2024; Journé, Simmonds, et al., 2025), but given the highly zero-inflated nature of masting data, especially for
453 individual-level data, this leads to problems associated with log-transforming count data (O'Hara & Kotze, 2010)
454 or exclusion of zeros counts, which are an essential part of masting. Only recently, these limitations were partly
455 tackled by testing several climate variables in beta regression models (using absolute maximum standardisation
456 of seed counts to a 0-1 scale), to identify weather cues (Journé, Kelly, et al., 2025). While this standardisation
457 makes trends between populations more comparable, it does not allow comparison of the magnitudes of effects
458 on masting. Here, we applied an alternative method that accounted for these limitations (i.e., zero-inflation,
459 several climate variables, comparability of magnitudes) and additionally looked at the different effects of
460 environmental drivers on the number of seeds produced and the occurrence of reproduction. While we found
461 slightly different results using this approach, we were also able to replicate previous findings by combining our
462 data with one of the established approaches (as described in Journé et al. (2024)). As experimental tests on

463 masting are challenging (Bogdziewicz, Ascoli, Hackett-Pain, et al., 2020), we cannot assess which results are closer
464 to the real weather cues that affect masting, but we argue that the approach we use here is better suited to
465 accommodate the data structure and the complex biological processes driving masting. Testing our approach in
466 comparison to existing approaches across populations would therefore be interesting.

467 Despite the cueing windows being different, the overall effect of weather on beechnut production is the same
468 as commonly reported: a wet and cold climate in T2, under which resources can build up (Richardson et al.,
469 2005), followed by dry and warm climate in T1, which enhances formation of flower primordia (Gruber, 2003),
470 and warm and dry weather in T0, when flowers are pollinated, positively affect reproduction. While annual
471 variation in the number of produced seeds, if reproduction occurred, could be fully attributed to temperature
472 and precipitation, these climatic factors alone could not determine whether or not reproduction occurs.
473 Comparing model predictions with observed values of annual seed production showed that the severity of
474 temporal changes in the masting pattern is underestimated by the model, again indicating that there must be
475 other factors, in combination with temperature changes, driving the masting breakdown. One of them could be
476 internal resource dynamics, which can modulate the response to climate cues in beech (Kelly et al., 2025).

477 Since trees in our population respond to a warm temperature cue in the year before seeding, the increasing
478 temperatures in this cueing window increased cue frequency, which initiates reproduction more frequently
479 (Bogdziewicz et al., 2024). With a shorter period between reproductive events, resources may not sufficiently
480 be replenished (Kelly et al., 2025), weakening the tree's sensitivity to the temperature cue (Bogdziewicz, Hackett-
481 Pain, et al., 2021). Trees then produce seeds in proportion to the limited resources that they could build up in
482 this time, leading to smaller, less variable, and more frequent seed output. While trees could benefit from
483 increasing temperatures if they enhance resource availability, allowing them to frequently produce high crops
484 (Kelly et al., 2025), this is not the case in our population, as seed production seems to stabilise around 250
485 nuts/m². Increased cueing frequency can further magnify inter-individual differences in resource budgets
486 (Bogdziewicz, 2022), inducing different responses to the same cue, ultimately resulting in decreasing synchrony.

487

488 Although these interactive effects of weather cues and internal resource reserves in theory could explain the
489 observed changes in masting patterns of our population, we actually do not see a significant effect of resource

490 reserves both on the probability of reproduction and the number of nuts produced when fitting resource
491 reserves into the final model, including all six weather cues (Supporting Information C). Resources neither have
492 a significant effect in interaction with the temperature cue of the year prior to seeding (T1) nor as an additive
493 effect when all weather cues are included, while the interaction with the temperature cue in T1 is significant if
494 all other weather cues, besides temperature in T2, are excluded (following previous work; Kelly et al., 2025).
495 Resources may therefore only weakly modulate the tree's response to weather cues in this population,
496 increasing the tree's sensitivity to climate change effects (Bogdziewicz et al., 2024). The calculation of resource
497 reserves is here based on the cumulative reproductive effort (for more details: Kelly et al., 2025; Rees et al.,
498 2002), but as trees do not only rely on stored resources but also on resource uptake throughout the reproductive
499 cycle (Allen et al., 2017), this measure of resource availability may not sufficiently capture the overall effect of
500 available resources and thereby, mask the relationship with weather cues. Generally, nitrogen fertilisation
501 enhances seed production (Bogdziewicz et al., 2017) and nitrogen levels are highly correlated with the on-off
502 cycle of expression of flowering genes (Miyazaki et al., 2014). Under the predicted increasing anthropogenic
503 nitrogen deposition (Galloway et al., 2004), this suggests a further reduction of the probability of years without
504 reproduction and would thereby strengthen the masting breakdown. A more reliable method of measuring
505 resources (e.g., measuring the nutrient content in reproductive structures (Fernández-Martínez et al., 2017)),
506 and combining internal and external resource availability into future models describing annual beechnut
507 production is, therefore, crucial to more accurately capture masting breakdown and inter-individual variation
508 between trees. Consequently, we can currently not conclusively determine which other factors, besides changes
509 in temperature, are driving the observed masting breakdown.

510 Compared to UK populations, masting breakdown in our population occurred roughly at the same time around
511 the mid-2000s (Hackett-Pain et al., 2025), but, with 49%, the decrease in synchrony was stronger in our
512 population (compared to 30% in Bogdziewicz, Kelly, Thomas, et al., 2020), while the 30% decline in inter-annual
513 variability was less extreme (compared to 40% in UK). Despite populations across the species range showing
514 masting breakdowns (Foest et al., 2024), this highlights that the magnitude of these changes can differ between
515 populations, as they can differ in their response to local conditions (as shown in oak in Fleurot et al., 2023). Given
516 the negative fitness consequences caused by the masting breakdown, there is selection for individuals that are
517 highly sensitive to climate cues, because they will show higher synchrony and inter-annual variability in seed

518 production, and thereby selection for masting (Bogdziewicz, Kelly, Tanentzap, et al., 2020). While this could
519 allow beeches to cope with climate-induced changes, the long generation time of beech trees and other masting
520 tree species makes it challenging to keep up with the pace of climate change.

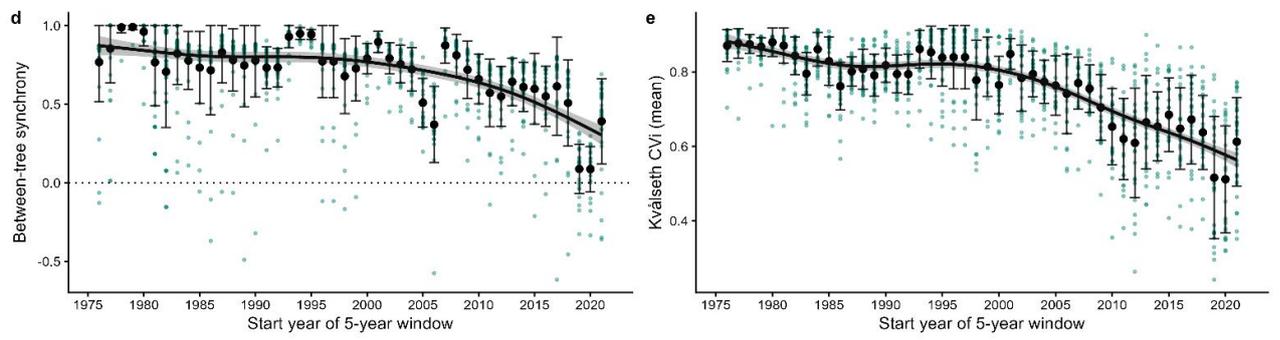
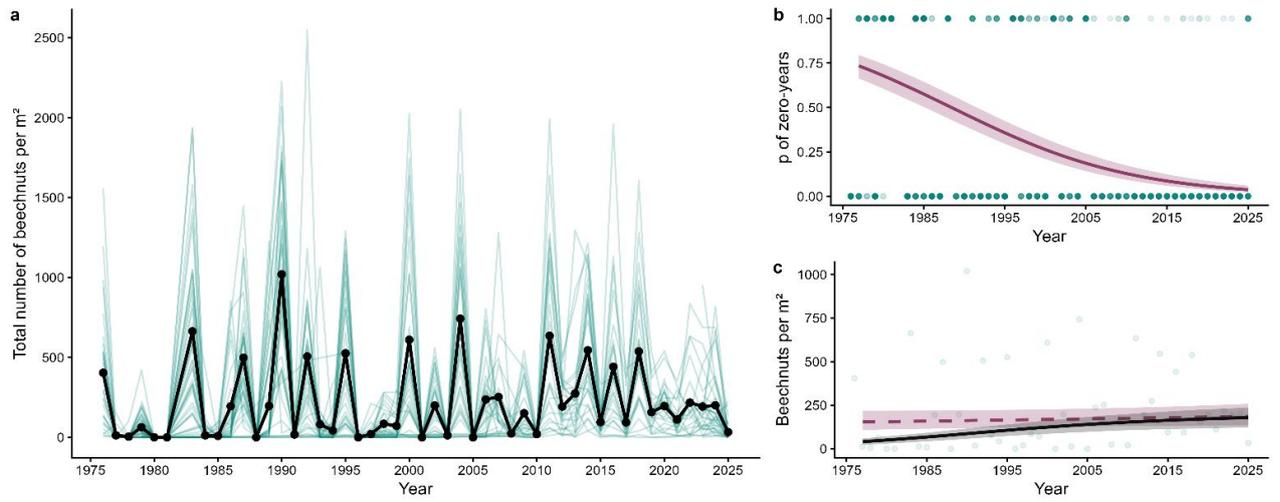
521 Despite looking at several aspects of masting, this first analysis of this valuable long-term dataset leaves room
522 for more elaborate questions, especially in connection with seed consumer populations. While masting
523 breakdown decreases the fitness of the trees, the more stable seed supply in the forest might positively affect
524 seed consumers, as beechnuts are a crucial food resource for many species, such as rodents (Zwolak et al., 2024),
525 wild boar (Touzot et al., 2020), or great tits (Perdeck et al., 2000). Changes in food supply can have yet
526 unforeseen effects on the forest, as changes in seed consumer population can further cascade through the wider
527 food web (Ostfeld & Keesing, 2000). However, given the decreasing pollination rate, many of the produced nuts
528 will be empty and therefore not of use for consumers. Investigating changes in viable seed supply linked to
529 population changes of forest species therefore provides an interesting avenue for further research.

530

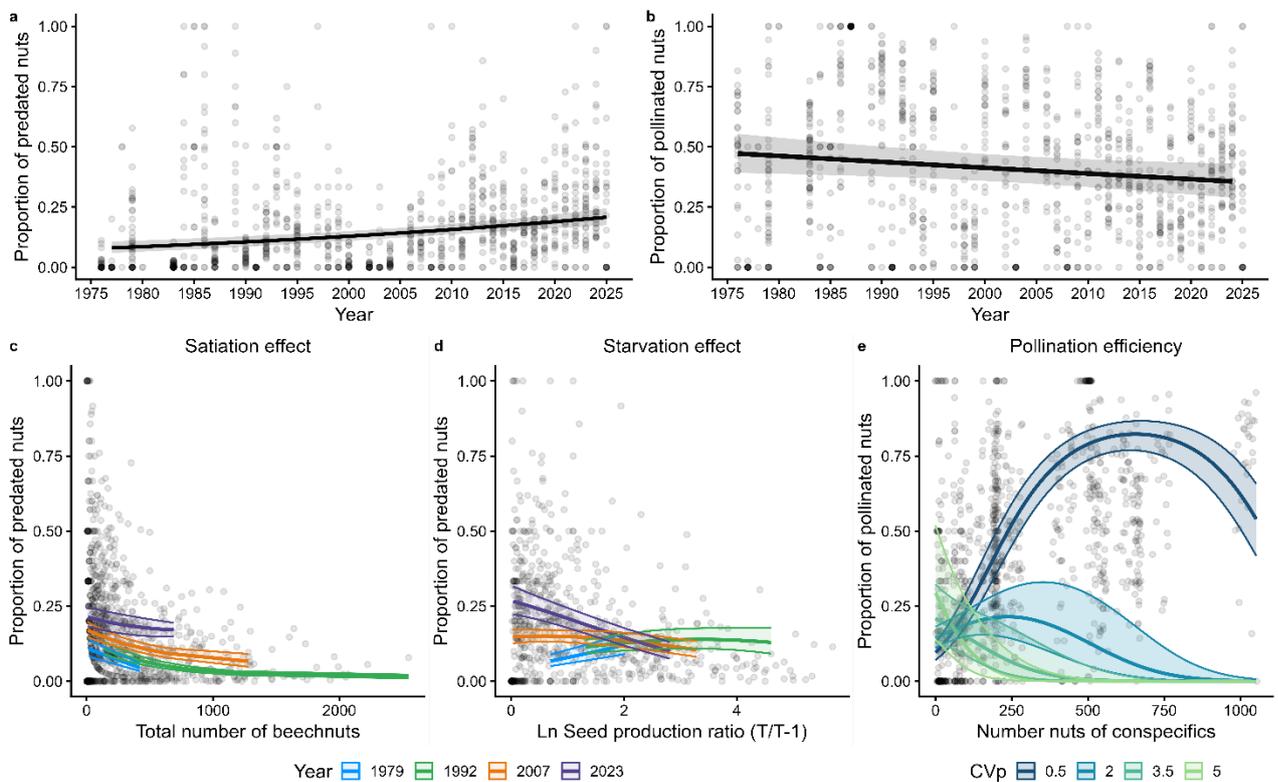
531

532 **Figures**

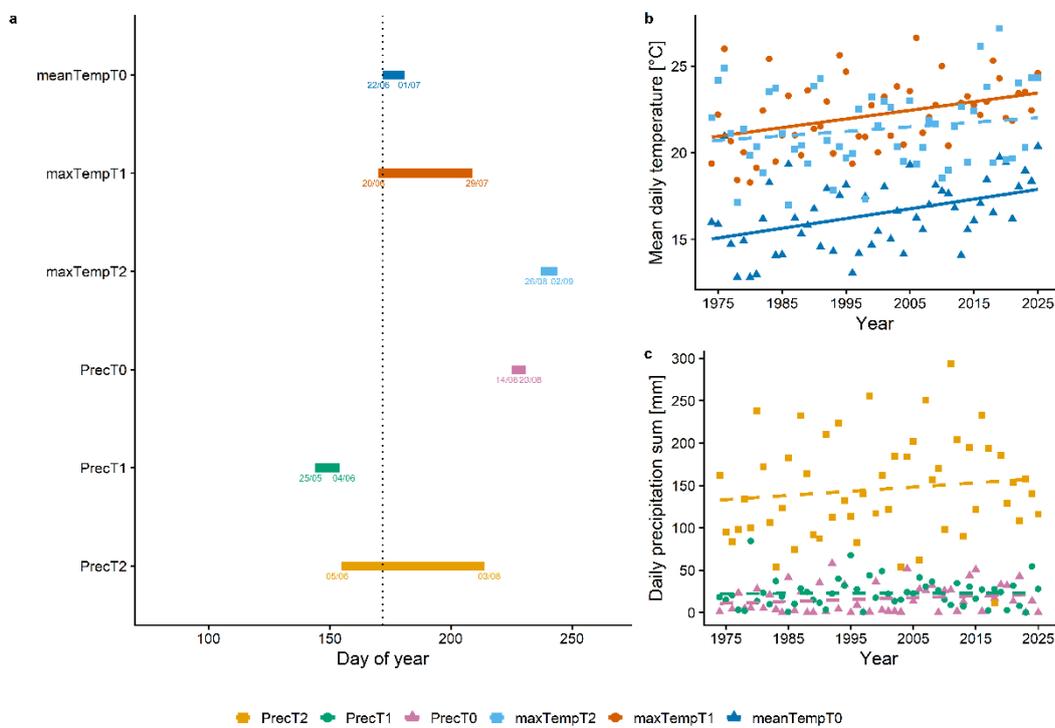
533 **Figure 1 Temporal changes of masting patterns.** (a) Total annual number of beechnuts per individual tree ($n =$
534 81; coloured lines) and the population mean (black). In 1982, no data was collected. (b & c) Model predictions
535 of the temporal trend of annual beechnut production of the population are based on a zero-inflated negative
536 binomial GLMM: (b) The predicted probability of the population to not reproduce at all in a year decreases
537 significantly over time, indicating that zero-years become less likely (based on zero-inflation part of model, red
538 line). Dots show raw data per tree, with 1 being a year without nuts (probability of 1 to be zero) and 0 a year
539 with at least one nut (probability of 0 to be zero). Dots are transparent and darker dots indicate many overlaying
540 data points. (c) The number of beechnuts produced by the population, if reproduction occurs, does not change
541 significantly over the study period (red, dashed line, based on conditional part of model). However, when
542 considering that years without beechnuts are getting rarer (see b), the annual seed output does show a positive
543 trend over time (black, solid line). (d & e) A sliding window approach with five-year window size and a step-size
544 of one year was used to assess the temporal trend in (d) between-tree synchrony (i.e., Spearman correlations
545 between conspecifics) and (e) inter-annual variation (i.e., Kvålseth CV_i) of total beechnut production. Raw data
546 per tree is shown as coloured dots (excluding trees from a window when they have less than three observations),
547 population mean as black dots together with their standard deviation. Fitted lines in d) and e) are based on
548 model predictions of GAMs with the population mean as a response.



551 **Figure 2 Economies of scale.** (a) The proportion of predated nuts significantly increased over time, while (b) the
 552 proportion of pollinated nuts shows a weakly significant decline. The proportion of predated seeds per tree
 553 changes with the (c) total number of nuts produced by that tree and (d) the ratio of seeds produced in one year
 554 to the seeds produced the previous year. For both relationships, these effects changed over time (coloured lines
 555 show model predictions for four different years and are truncated to the range of observed values of the
 556 respective year). The proportion of effectively pollinated nuts changes (e) both with the number of nuts
 557 produced by conspecifics and the within-site within-year synchrony in seed production (CV_p), where low CV_p
 558 indicates little variation and therefore high synchrony and vice versa (coloured lines show model predictions for
 559 different CV_p levels). All lines are fitted based on model predictions of binomial GLMMs.

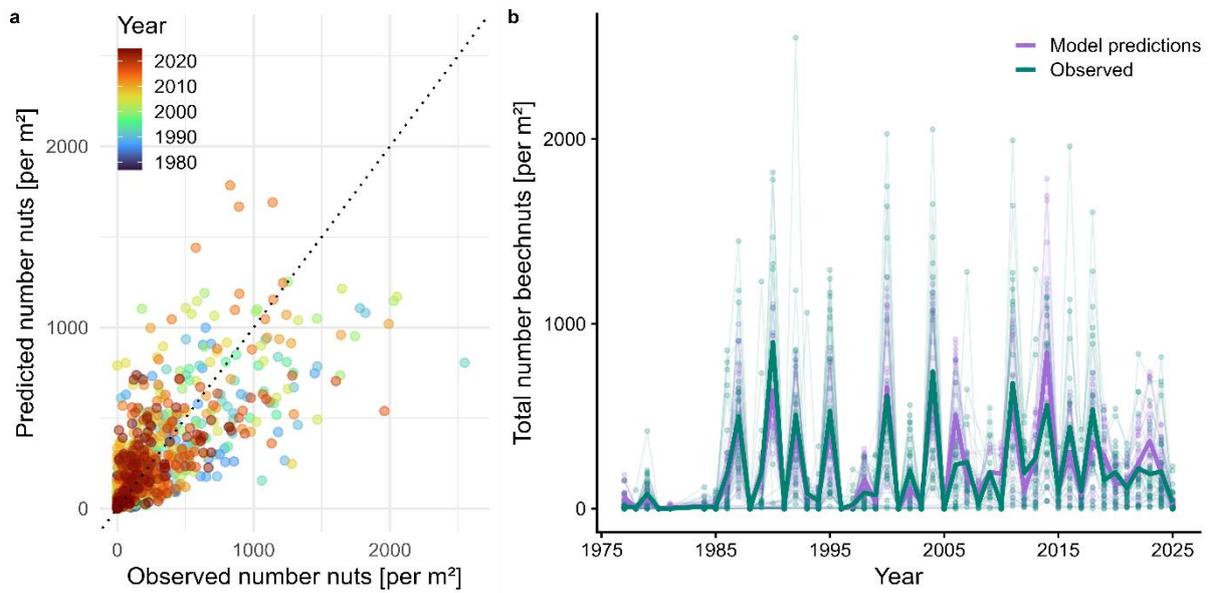


561 **Figure 3 Sensitivity windows and temporal change of climate variables in these windows.** (a) Beechnut
 562 production is sensitive to temperature and precipitation from different periods in the year of seeding (T0), one
 563 (T1) and two (T2) years prior. These sensitivity periods differ in length and are spread across the summer months.
 564 Point labels indicate the calendar date, assuming it is not a leap year (as dd/mm) and the dotted line indicates
 565 the summer solstice (DOY 172). (b) While mean daily mean (for T0) and maximum (for T1) temperatures in these
 566 windows significantly increased over the study period, there is no significant change in maximum temperatures
 567 of T2 and (c) the sum of daily precipitation sums in the sensitivity periods for the three years. Colours in panels
 568 a-c are according to the legend below panel a.



569

570 **Figure 4 Model predictions and observed number of beechnuts.** (a) Comparing the model predictions of the
571 final model with the observed number of beechnuts shows that the model underestimates large seed crops,
572 while it overestimates beechnut production more often in the most recent years. Predicted and observed values
573 are correlated with $\rho = 0.86$. Dots are individual trees; the dotted line represents a perfect correlation. (b) When
574 comparing temporal patterns in observed and predicted beechnut counts, it shows that inter-individual
575 differences (thin, transparent lines) are less well captured by the model and that temporal patterns, especially
576 in the population mean (thick lines), start diverging more strongly in the second half of the study.



578 **Table 1 Drivers of annual beechnut production.** Results from the final, negative binomial zero-inflated GLMM
579 in which all fixed effects are set to their respective best window. The model included annual beechnut
580 production per tree as the response variable, tree ID as random effect and previous year's number of nuts as a
581 fixed effect. T0 marks the year of seed fall, T1 and T2 one and two years prior, respectively. All fixed effects were
582 z-transformed within their respective window across the study period and estimates are on the log-link scale.
583 The zero-inflation part of the model estimates the probability of a year without reproduction (true zero), with
584 positive estimates indicating that an increase in the fixed effect leads to a higher probability of zero nuts, i.e.,
585 no reproduction. The conditional model explains the effects on the number of nuts produced, if reproduction
586 occurs, with positive estimates indicating that an increase in the fixed effect increases the number of nuts
587 produced. P-values for climate variables are corrected by multiplying them with the number of windows tested
588 ($p * 15343$ windows, corrected p-values are indicated with *).

Fixed effects	Conditional model			Zero-inflation model		
	Estimate (\pm se)	z	p	Estimate (\pm se)	z	p
Mean temperature T0	0.198 (0.043)	4.57	0.076*	-0.800 (0.102)	-7.812	< 0.001*
Max temperature T1	0.306 (0.041)	7.41	< 0.001*	-0.847 (0.126)	-6.713	< 0.001*
Max temperature T2	-0.411 (0.037)	-11.17	< 0.001*	1.494 (0.245)	6.104	< 0.001*
Precipitation T0	0.449 (0.035)	12.95	< 0.001*	-1.993 (0.256)	-7.774	< 0.001*
Precipitation T1	-0.765 (0.044)	-17.49	< 0.001*	2.295 (0.245)	9.380	< 0.001*
Precipitation T2	0.528 (0.039)	13.46	< 0.001*	-2.881 (0.369)	-7.799	< 0.001*
Number nuts T1	-0.0005 (0.0001)	-4.42	<0.001	0.0008 (0.0003)	2.717	0.007

589

590 ***Acknowledgements***

591 We want to thank everyone that contributed to the collection and processing of the beechnuts over time,
592 especially J.H. van Balen, L.J. Holleman and B. van Lith and J. M. Loy, and J. Risse for managing and curating the
593 database. We especially thank the board and director of the National Park De Hoge Veluwe for letting us collect
594 the data over the past 50 years.

595

596 ***Author Contributions***

597 **Cherine C. Jantzen:** Conceptualization, Data Curation, Formal Analysis, Methodology, Visualization, Writing -
598 Original Draft; **Joseph B. Burant:** Conceptualization, Methodology, Supervision, Writing - Review & Editing;
599 **Marlène Gamelon:** Methodology, Writing - Review & Editing; **Elisabeth S. Bakker:** Conceptualization,
600 Methodology, Supervision, Writing - Review & Editing; **Marcel E. Visser:** Conceptualization, Data Curation,
601 Methodology, Supervision, Writing - Review & Editing

602

603 ***Data availability***

604 The data and code for all statistical analyses will be stored on suitable repositories upon acceptance of this
605 manuscript and can for now be found on GitHub: [https://github.com/CherineJ/Masting-breakdown_Jantzen-](https://github.com/CherineJ/Masting-breakdown_Jantzen-et-al)
606 [et-al.](https://github.com/CherineJ/Masting-breakdown_Jantzen-et-al)

607

608 ***Funding***

609 This publication is part of the project *The heartbeat of the forest* with file number OCENW.M.22.426 of the
610 research programme NWO Open Competition Domain Science which is financed by the Dutch Research
611 Council (NWO). M.G. was funded by the French National Research Agency ANR PURE project (ANR-23-CE02-
612 0028).

613

614 ***Conflict of Interest***

615 The authors declare no conflict of interest.

616 **References**

- 617 Allen, R. B., Millard, P., & Richardson, S. J. (2017). A Resource Centric View of Climate and Mast
618 Seeding in Trees. In F. M. Cánovas, U. Lüttge, & R. Matyssek (Eds), *Progress in Botany Vol. 79*
619 (Vol. 79, pp. 233–268). Springer International Publishing.
620 https://doi.org/10.1007/124_2017_8
- 621 Anonymous. (n.d.). Long-term annual seed production data of individual European beech (*Fagus*
622 *sylvatica*) trees in the Netherlands. [*Manuscript under review*]
- 623 Ascoli, D., Hacket-Pain, A., Pearse, I. S., Vacchiano, G., Corti, S., & Davini, P. (2021). Modes of climate
624 variability bridge proximate and evolutionary mechanisms of masting. *Philosophical*
625 *Transactions of the Royal Society B: Biological Sciences*, 376(1839), 20200380.
626 <https://doi.org/10.1098/rstb.2020.0380>
- 627 Belmonte, J., Alarcón, M., Avila, A., Scialabba, E., & Pino, D. (2008). Long-range transport of beech
628 (*Fagus sylvatica* L.) pollen to Catalonia (north-eastern Spain). *International Journal of*
629 *Biometeorology*, 52(7), 675–687. <https://doi.org/10.1007/s00484-008-0160-9>
- 630 Bogdziewicz, M. (2022). How will global change affect plant reproduction? A framework for mast
631 seeding trends. *New Phytologist*, 234(1), 14–20. <https://doi.org/10.1111/nph.17682>
- 632 Bogdziewicz, M., Ascoli, D., Hacket-Pain, A., Koenig, W. D., Pearse, I., Pesendorfer, M., Satake, A.,
633 Thomas, P., Vacchiano, G., Wohlgemuth, T., & Tanentzap, A. (2020). From theory to
634 experiments for testing the proximate mechanisms of mast seeding: An agenda for an
635 experimental ecology. *Ecology Letters*, 23(2), 210–220. <https://doi.org/10.1111/ele.13442>
- 636 Bogdziewicz, M., Crone, E. E., Steele, M. A., & Zwolak, R. (2017). Effects of nitrogen deposition on
637 reproduction in a masting tree: Benefits of higher seed production are trumped by negative
638 biotic interactions. *Journal of Ecology*, 105(2), 310–320. [https://doi.org/10.1111/1365-](https://doi.org/10.1111/1365-2745.12673)
639 [2745.12673](https://doi.org/10.1111/1365-2745.12673)

640 Bogdziewicz, M., Hacket-Pain, A., Ascoli, D., & Szymkowiak, J. (2021). Environmental variation drives
641 continental-scale synchrony of European beech reproduction. *Ecology*, *102*(7), e03384.
642 <https://doi.org/10.1002/ecy.3384>

643 Bogdziewicz, M., Hacket-Pain, A., Kelly, D., Thomas, P. A., Lageard, J., & Tanentzap, A. J. (2021).
644 Climate warming causes mast seeding to break down by reducing sensitivity to weather
645 cues. *Global Change Biology*, *27*(9), 1952–1961. <https://doi.org/10.1111/gcb.15560>

646 Bogdziewicz, M., Journé, V., Hacket-Pain, A., & Szymkowiak, J. (2023). Mechanisms driving
647 interspecific variation in regional synchrony of trees reproduction. *Ecology Letters*, *26*(5),
648 754–764. <https://doi.org/10.1111/ele.14187>

649 Bogdziewicz, M., Kelly, D., Ascoli, D., Caignard, T., Chianucci, F., Crone, E. E., Fleurot, E., Foest, J. J.,
650 Gratzer, G., Hagiwara, T., Han, Q., Journé, V., Keurinck, L., Kondrat, K., McClory, R.,
651 LaMontagne, J. M., Mundo, I. A., Nussbaumer, A., Oberklammer, I., ... Hacket-Pain, A. J.
652 (2024). Evolutionary ecology of masting: Mechanisms, models, and climate change. *Trends in*
653 *Ecology & Evolution*. <https://doi.org/10.1016/j.tree.2024.05.006>

654 Bogdziewicz, M., Kelly, D., Tanentzap, A. J., Thomas, P. A., Lageard, J. G. A., & Hacket-Pain, A. (2020).
655 Climate Change Strengthens Selection for Mast Seeding in European Beech. *Current Biology*,
656 *30*(17), 3477–3483. e2. <https://doi.org/10.1016/j.cub.2020.06.056>

657 Bogdziewicz, M., Kelly, D., Tanentzap, A. J., Thomas, P., Foest, J., Lageard, J., & Hacket-Pain, A.
658 (2023). Reproductive collapse in European beech results from declining pollination efficiency
659 in large trees. *Global Change Biology*, *29*(16), 4595–4604.
660 <https://doi.org/10.1111/gcb.16730>

661 Bogdziewicz, M., Kelly, D., Thomas, P. A., Lageard, J. G. A., & Hacket-Pain, A. (2020). Climate warming
662 disrupts mast seeding and its fitness benefits in European beech. *Nature Plants*, *6*(2), 88–94.
663 <https://doi.org/10.1038/s41477-020-0592-8>

664 Brooks, M. E., Kristensen, K., Benthem, K. J., van, Magnusson, A., Berg, C. W., Nielsen, A., Skaug, H. J.,
665 Mächler, M., & Bolker, B. M. (2017). glmmTMB Balances Speed and Flexibility Among

666 Packages for Zero-inflated Generalized Linear Mixed Modeling. *The R Journal*, 9(2), 378.
667 <https://doi.org/10.32614/RJ-2017-066>

668 Burnham, K. P., & Anderson, D. R. (Eds). (2002). *Model Selection and Multimodel Inference*. Springer
669 New York. <https://doi.org/10.1007/b97636>

670 Crone, E. E., & Rapp, J. M. (2014). Resource depletion, pollen coupling, and the ecology of mast
671 seeding. *Annals of the New York Academy of Sciences*, 1322(1), 21–34.
672 <https://doi.org/10.1111/nyas.12465>

673 Dri, G. F., Bogdziewicz, M., Hunter, M., Witham, J., & Mortelliti, A. (2025). Coupled effects of forest
674 growth and climate change on small mammal abundance and body weight: Results of a 39-
675 year field study. *Journal of Animal Ecology*, 94(10), 2118–2129.
676 <https://doi.org/10.1111/1365-2656.70114>

677 Drobyshev, I., Övergaard, R., Saygin, I., Niklasson, M., Hickler, T., Karlsson, M., & Sykes, M. T. (2010).
678 Masting behaviour and dendrochronology of European beech (*Fagus sylvatica* L.) in southern
679 Sweden. *Forest Ecology and Management*, 259(11), 2160–2171.
680 <https://doi.org/10.1016/j.foreco.2010.01.037>

681 Fernández-Martínez, M., Vicca, S., Janssens, I. A., Espelta, J. M., & Peñuelas, J. (2017). The role of
682 nutrients, productivity and climate in determining tree fruit production in European forests.
683 *New Phytologist*, 213(2), 669–679. <https://doi.org/10.1111/nph.14193>

684 Fleurot, E., Lobry, J. R., Boulanger, V., Debias, F., Mermet-Bouvier, C., Caignard, T., Delzon, S., Bel-
685 Venner, M.-C., & Venner, S. (2023). Oak masting drivers vary between populations
686 depending on their climatic environments. *Current Biology*, 33(6), 1117-1124.e4.
687 <https://doi.org/10.1016/j.cub.2023.01.034>

688 Foest, J. J., Bogdziewicz, M., Pesendorfer, M. B., Ascoli, D., Cutini, A., Nussbaumer, A., Verstraeten,
689 A., Beudert, B., Chianucci, F., Mezzavilla, F., Gratzner, G., Kunstler, G., Meesenburg, H.,
690 Wagner, M., Mund, M., Cools, N., Vacek, S., Schmidt, W., Vacek, Z., & Hacket-Pain, A. (2024).

691 Widespread breakdown in masting in European beech due to rising summer temperatures.
692 *Global Change Biology*, 30(5), e17307. <https://doi.org/10.1111/gcb.17307>

693 Foest, J. J., Szymkowiak, J., Dyderski, M. K., Jastrzębowski, S., Fuchs, H., Ratajczak, E., Hacket-Pain, A.,
694 & Bogdziewicz, M. (2025). No Refuge at the Edge for European Beech as Climate Warming
695 Disproportionately Reduces Masting at Colder Margins. *Ecology Letters*, 28(12), e70284.
696 <https://doi.org/10.1111/ele.70284>

697 Galloway, J. N., Dentener, F. J., Capone, D. G., Boyer, E. W., Howarth, R. W., Seitzinger, S. P., Asner,
698 G. P., Cleveland, C. C., Green, P. A., Holland, E. A., Karl, D. M., Michaels, A. F., Porter, J. H.,
699 Townsend, A. R., & Vörösmarty, C. J. (2004). Nitrogen Cycles: Past, Present, and Future.
700 *Biogeochemistry*, 70(2), 153–226. <https://doi.org/10.1007/s10533-004-0370-0>

701 Geßler, A., Jung, K., Gasche, R., Papen, H., Heidenfelder, A., Börner, E., Metzler, B., Augustin, S.,
702 Hildebrand, E., & Rennenberg, H. (2005). Climate and forest management influence nitrogen
703 balance of European beech forests: Microbial N transformations and inorganic N net uptake
704 capacity of mycorrhizal roots. *European Journal of Forest Research*, 124(2), 95–111.
705 <https://doi.org/10.1007/s10342-005-0055-9>

706 Grayson, R. B., Western, A. W., Chiew, F. H. S., & Blöschl, G. (1997). Preferred states in spatial soil
707 moisture patterns: Local and nonlocal controls. *Water Resources Research*, 33(12), 2897–
708 2908. <https://doi.org/10.1029/97WR02174>

709 Gruber, F. (2003). Steuerung und Vorhersage der Fruktifikation bei der Rotbuche (*Fagus sylvatica* L.)
710 für den Standort Zierenberg 38A und den Level I Flächen von Hessen durch die Witterung.
711 *Allgemeine Forst- und Jagdzeitung*, 174(4), 67–79.

712 Hacket-Pain, A., Szymkowiak, J., Journé, V., Barczyk, M. K., Thomas, P. A., Lageard, J. G. A., Kelly, D.,
713 & Bogdziewicz, M. (2025). Growth decline in European beech associated with temperature-
714 driven increase in reproductive allocation. *Proceedings of the National Academy of Sciences*,
715 122(5), e2423181122. <https://doi.org/10.1073/pnas.2423181122>

716 Hartig, F. (2024). *DHARMA: Residual Diagnostics for Hierarchical (Multi-Level / Mixed) Regression*
717 *Models*. <https://github.com/florianhartig/dharma>

718 Herrera, C. M., Jordano, P., Guitián, J., & Traveset, A. (1998). Annual Variability in Seed Production by
719 Woody Plants and the Masting Concept: Reassessment of Principles and Relationship to
720 Pollination and Seed Dispersal. *The American Naturalist*, *152*(4), 576–594.
721 <https://doi.org/10.1086/286191>

722 Janzen, D. H. (1971). Seed Predation by Animals. *Annual Review of Ecology and Systematics*, *2*(1),
723 465–492. <https://doi.org/10.1146/annurev.es.02.110171.002341>

724 Journé, V., Hacket-Pain, A., Oberklammer, I., Pesendorfer, M. B., & Bogdziewicz, M. (2023).
725 Forecasting seed production in perennial plants: Identifying challenges and charting a path
726 forward. *New Phytologist*, *239*(2), 466–476. <https://doi.org/10.1111/nph.18957>

727 Journé, V., Kelly, D., Hacket-Pain, A., Pearse, I. S., Szymkowiak, J., Foest, J. J., Kondrat, K.,
728 Oberklammer, I., Pesendorfer, M. B., Satake, A., & Bogdziewicz, M. (2025). Weather drivers
729 of reproductive variability in perennial plants and their implications for climate change risks.
730 *Nature Communications*, *16*(1), 9226. <https://doi.org/10.1038/s41467-025-64300-6>

731 Journé, V., Simmonds, E. G., Barczyk, M. K., & Bogdziewicz, M. (2025). Comparing statistical methods
732 for detecting weather cues of mast seeding in European beech (*Fagus sylvatica*) across
733 Europe. *Agricultural and Forest Meteorology*, *375*, 110857.
734 <https://doi.org/10.1016/j.agrformet.2025.110857>

735 Journé, V., Szymkowiak, J., Foest, J., Hacket-Pain, A., Kelly, D., & Bogdziewicz, M. (2024). Summer
736 solstice orchestrates the subcontinental-scale synchrony of mast seeding. *Nature Plants*,
737 *10*(3), 367–373. <https://doi.org/10.1038/s41477-024-01651-w>

738 Kelly, D. (1994). The evolutionary ecology of masting. *Trends in Ecology & Evolution*, *9*(12).
739 [https://doi.org/10.1016/0169-5347\(94\)90310-7](https://doi.org/10.1016/0169-5347(94)90310-7)

740 Kelly, D., Geldenhuis, A., James, A., Penelope Holland, E., Plank, M. J., Brockie, R. E., Cowan, P. E.,
741 Harper, G. A., Lee, W. G., Maitland, M. J., Mark, A. F., Mills, J. A., Wilson, P. R., & Byrom, A. E.

742 (2013). Of mast and mean: Differential-temperature cue makes mast seeding insensitive to
743 climate change. *Ecology Letters*, *16*(1), 90–98. <https://doi.org/10.1111/ele.12020>

744 Kelly, D., & Sork, V. L. (2002). Mast Seeding in Perennial Plants: Why, How, Where? *Annual Review of*
745 *Ecology and Systematics*, *33*(1), 427–447.
746 <https://doi.org/10.1146/annurev.ecolsys.33.020602.095433>

747 Kelly, D., Szymkowiak, J., Hacket-Pain, A., & Bogdziewicz, M. (2025). Fine-tuning mast seeding: As
748 resources accumulate, plants become more sensitive to weather cues. *New Phytologist*,
749 *246*(5), 1975–1985. <https://doi.org/10.1111/nph.70092>

750 Kvålseth, T. O. (2017). Coefficient of variation: The second-order alternative. *Journal of Applied*
751 *Statistics*, *44*(3), 402–415. <https://doi.org/10.1080/02664763.2016.1174195>

752 Leuschner, C., & Ellenberg, H. (2017). Beech and Mixed Beech Forests. In *Ecology of Central*
753 *European Forests: Vegetation Ecology of Central Europe, Volume I* (pp. 351–441). Springer
754 International Publishing. https://doi.org/10.1007/978-3-319-43042-3_5

755 Lobry, J. R., Bel-Venner, M., Bogdziewicz, M., Hacket-Pain, A., & Venner, S. (2023). The CV is dead,
756 long live the CV! *Methods in Ecology and Evolution*, *14*(11), 2780–2786.
757 <https://doi.org/10.1111/2041-210x.14197>

758 Lüdecke, D. (2018). ggeffects: Tidy Data Frames of Marginal Effects from Regression Models. *Journal*
759 *of Open Source Software*, *3*(26), 772. <https://doi.org/10.21105/joss.00772>

760 Miyazaki, Y., Maruyama, Y., Chiba, Y., Kobayashi, M. J., Joseph, B., Shimizu, K. K., Mochida, K., Hiura,
761 T., Kon, H., & Satake, A. (2014). Nitrogen as a key regulator of flowering in *Fagus crenata*:
762 Understanding the physiological mechanism of masting by gene expression analysis. *Ecology*
763 *Letters*, *17*(10), 1299–1309. <https://doi.org/10.1111/ele.12338>

764 Moreira, X., Abdala-Roberts, L., Linhart, Y. B., & Mooney, K. A. (2014). Masting promotes individual-
765 and population-level reproduction by increasing pollination efficiency. *Ecology*, *95*(4), 801–
766 807. <https://doi.org/10.1890/13-1720.1>

767 Müller-Haubold, H., Hertel, D., & Leuschner, C. (2015). Climatic Drivers of Mast Fruiting in European
768 Beech and Resulting C and N Allocation Shifts. *Ecosystems*, *18*(6), 1083–1100.
769 <https://doi.org/10.1007/s10021-015-9885-6>

770 Nilsson, S. G., & Wastljung, U. (1987). Seed Predation and Cross-Pollination in Mast-Seeding Beech
771 (*Fagus Sylvatica*) Patches. *Ecology*, *68*(2), 260–265. <https://doi.org/10.2307/1939256>

772 Norton, D. A., & Kelly, D. (1988). Mast Seeding Over 33 Years by *Dacrydium cupressinum* Lamb.
773 (*rimu*) (Podocarpaceae) in New Zealand: The Importance of Economies of Scale. *Functional*
774 *Ecology*, *2*(3), 399. <https://doi.org/10.2307/2389413>

775 Nussbaumer, A., Meusburger, K., Schmitt, M., Waldner, P., Gehrig, R., Haeni, M., Rigling, A., Brunner,
776 I., & Thimonier, A. (2020). Extreme summer heat and drought lead to early fruit abortion in
777 European beech. *Scientific Reports*, *10*(1), 5334. [https://doi.org/10.1038/s41598-020-62073-](https://doi.org/10.1038/s41598-020-62073-0)
778 [0](https://doi.org/10.1038/s41598-020-62073-0)

779 Nussbaumer, A., Waldner, P., Etzold, S., Gessler, A., Benham, S., Thomsen, I. M., Jørgensen, B. B.,
780 Timmermann, V., Verstraeten, A., Sioen, G., Rautio, P., Ukonmaanaho, L., Skudnik, M.,
781 Apuhtin, V., Braun, S., & Wauer, A. (2016). Patterns of mast fruiting of common beech,
782 sessile and common oak, Norway spruce and Scots pine in Central and Northern Europe.
783 *Forest Ecology and Management*, *363*, 237–251.
784 <https://doi.org/10.1016/j.foreco.2015.12.033>

785 O’Hara, R. B., & Kotze, D. J. (2010). Do not log-transform count data. *Methods in Ecology and*
786 *Evolution*, *1*(2), 118–122. <https://doi.org/10.1111/j.2041-210X.2010.00021.x>

787 Ostfeld, R. S., & Keesing, F. (2000). Pulsed resources and community dynamics of consumers in
788 terrestrial ecosystems. *Trends in Ecology & Evolution*, *15*(6), 232–237.
789 [https://doi.org/10.1016/S0169-5347\(00\)01862-0](https://doi.org/10.1016/S0169-5347(00)01862-0)

790 Övergaard, R., Gemmel, P., & Karlsson, M. (2007). Effects of weather conditions on mast year
791 frequency in beech (*Fagus sylvatica* L.) in Sweden. *Forestry*, *80*(5), 555–565.
792 <https://doi.org/10.1093/forestry/cpm020>

793 Pearse, I. S., Koenig, W. D., & Kelly, D. (2016). Mechanisms of mast seeding: Resources, weather,
794 cues, and selection. *New Phytologist*, 212(3), 546–562. <https://doi.org/10.1111/nph.14114>

795 Perdeck, A. C., Visser, M. E., & Balen, J. H. V. (2000). Great Tit *Parus major* survival, and the beech-
796 crop cycle. *Ardea*, 88(1), 99–108.

797 Pesendorfer, M. B., Bogdziewicz, M., Oberklammer, I., Nopp-Mayr, U., Schwagrzyk, J., & Gratzner, G.
798 (2024). Positive spatial and temporal density-dependence drive early reproductive economy-
799 of-scale effects of masting in a European old-growth forest community. *Journal of Ecology*,
800 112(8), 1872–1884. <https://doi.org/10.1111/1365-2745.14368>

801 Pesendorfer, M. B., Bogdziewicz, M., Szymkowiak, J., Borowski, Z., Kantorowicz, W., Espelta, J. M., &
802 Fernández-Martínez, M. (2020). Investigating the relationship between climate, stand age,
803 and temporal trends in masting behavior of European forest trees. *Global Change Biology*,
804 26(3), 1654–1667. <https://doi.org/10.1111/gcb.14945>

805 R Core Team. (2025). *R: A Language and Environment for Statistical Computing*. R Foundation for
806 Statistical Computing. <https://www.R-project.org/>

807 Rees, M., Kelly, D., & Bjørnstad, O. N. (2002). Snow Tussocks, Chaos, and the Evolution of Mast
808 Seeding. *The American Naturalist*, 160(1), 44–59. <https://doi.org/10.1086/340603>

809 Richardson, S. J., Allen, R. B., Whitehead, D., Carswell, F. E., Ruscoe, W. A., & Platt, K. H. (2005).
810 Climate And Net Carbon Availability Determine Temporal Patterns Of Seed Production By
811 *Nothofagus*. *Ecology*, 86(4), 972–981. <https://doi.org/10.1890/04-0863>

812 Schermer, É., Bel-Venner, M., Fouchet, D., Siberchicot, A., Boulanger, V., Caignard, T., Thibaudon, M.,
813 Oliver, G., Nicolas, M., Gaillard, J., Delzon, S., & Venner, S. (2019). Pollen limitation as a main
814 driver of fruiting dynamics in oak populations. *Ecology Letters*, 22(1), 98–107.
815 <https://doi.org/10.1111/ele.13171>

816 Schmidt, W. (2006). Zeitliche Veränderung der Fruktifikation bei der Rotbuche (*Fagus sylvatica* L.) in
817 einem Kalkbuchenwald (1981–2004). *Allgemeine Forst- und Jagdzeitung*, 177(1), 9–19.

818 Simpson, G. (2022). *gratia: Graceful ggplot-Based Graphics and Other Functions for GAMs Fitted*
819 *using mgcv*.

820 Touzot, L., Bel-Venner, M.-C., Gamelon, M., Focardi, S., Boulanger, V., Débias, F., Delzon, S., Saïd, S.,
821 Schermer, E., Baubet, E., Gaillard, J.-M., & Venner, S. (2018). The ground plot counting
822 method: A valid and reliable assessment tool for quantifying seed production in temperate
823 oak forests? *Forest Ecology and Management*, 430, 143–149.
824 <https://doi.org/10.1016/j.foreco.2018.07.061>

825 Touzot, L., Schermer, É., Venner, S., Delzon, S., Rousset, C., Baubet, É., Gaillard, J., & Gamelon, M.
826 (2020). How does increasing mast seeding frequency affect population dynamics of seed
827 consumers? Wild boar as a case study. *Ecological Applications*, 30(6), e02134.
828 <https://doi.org/10.1002/eap.2134>

829 Tradowsky, J. S., Philip, S. Y., Kreienkamp, F., Kew, S. F., Lorenz, P., Arrighi, J., Bettmann, T.,
830 Caluwaerts, S., Chan, S. C., De Cruz, L., De Vries, H., Demuth, N., Ferrone, A., Fischer, E. M.,
831 Fowler, H. J., Goergen, K., Heinrich, D., Henrichs, Y., Kaspar, F., ... Wanders, N. (2023).
832 Attribution of the heavy rainfall events leading to severe flooding in Western Europe during
833 July 2021. *Climatic Change*, 176(7), 90. <https://doi.org/10.1007/s10584-023-03502-7>

834 Vacchiano, G., Hacket-Pain, A., Turco, M., Motta, R., Maringer, J., Conedera, M., Drobyshev, I., &
835 Ascoli, D. (2017). Spatial patterns and broad-scale weather cues of beech mast seeding in
836 Europe. *New Phytologist*, 215(2), 595–608. <https://doi.org/10.1111/nph.14600>

837 Wood, S. N. (2011). Fast Stable Restricted Maximum Likelihood and Marginal Likelihood Estimation
838 of Semiparametric Generalized Linear Models. *Journal of the Royal Statistical Society Series*
839 *B: Statistical Methodology*, 73(1), 3–36. <https://doi.org/10.1111/j.1467-9868.2010.00749.x>

840 Zwolak, R., Celebias, P., Zduniak, M., Bogdziewicz, M., & Wróbel, A. (2024). Scatterhoarder
841 abundance and advantages of seed burial drive dynamics of a tree–rodent interaction.
842 *Journal of Ecology*, 1365-2745.14356. <https://doi.org/10.1111/1365-2745.14356>
843

844
845
846
847
848
849

Supporting information to

**Masting breakdown in European beech reduces fitness benefits of masting, partly explained by
climate change**

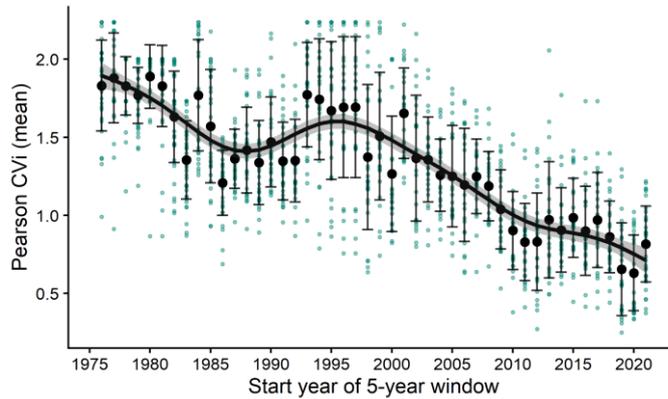
850 **Contents**

- 851 A. Inter-annual variability: Pearson $^P CV_i$ and Kvålseth $^K CV_i$
852 B. Sensitivity windows
853 B.1. Limitations of existing methodologies to identify weather cues of masting
854 B.2. Our approach
855 B.3. Comparison with approach used in Journé et al., (2024)
856 B.4. Collinearity test of explanatory climate variables
857 C. Effect of internal resource reserves on seed production

858
859

860 **A. Inter-annual variability: Pearson $^P\text{CV}_i$ and Kvålseth $^K\text{CV}_i$**

861 The Pearson CV_i ($^P\text{CV}_i$) shows a significant decrease over the study period (edf = 6.343, $F = 41.99$, $p <$
862 0.001 , $n = 46$). While it was between 1.83 and 1.88 at the beginning of the study period, it declines to
863 0.82 in the last window starting 2021. In the two windows starting in 2019 and 2020, there is an even
864 stronger decline to 0.65 and 0.63, respectively. Compared to Kvålseth $^K\text{CV}_i$ reported in the main text,
865 the predicted trend shows more variation, with an initial decline until 1988, followed by a small
866 increase until 1996 and another overall decrease until the end of the study period.



867

868 **Figure S2 Pearson CV_i ($^P\text{CV}_i$) of annual seed production.** $^P\text{CV}_i$ is calculated using a sliding window approach with
869 a window size of five years and one year step-size (see main text). Blue dots indicate the CV_i of individual trees,
870 the black dots the population mean per year. There was a significant effect of year based on a GAM fitting $^P\text{CV}_i$
871 against the year in which the window opened ($\sim s(\text{year})$). The fitted line is based on model predictions; the ribbon
872 indicates their standard error.

873

874

875 **B. Sensitivity windows**

876 **B.1. Limitations of existing methodologies to identify weather cues of masting**

877 As explained in more detail in the main text, several methodologies exist to identify those periods in
878 the year in which a biological response variable is most sensitive to climate variables (methodologies
879 for finding sensitivity windows of annual seed production are compared in Journé et al., (2025)).
880 Commonly, these methodologies analyse the window of highest sensitivity for each climate variable
881 separately, which neglects the complex interplay of other climatic factors that influence the response
882 variable. Additionally, many methods require normally distributed data, often resulting in
883 normalization of the data by log-transformation or other standardisation methods. For highly zero-
884 inflated data, such as annual seed counts (especially on an individual tree level), log-transformation
885 leads to an exclusion of zero counts (as the logarithm of zero is not defined), which will bias the results.
886 Additionally, most of these methods fit linear regressions or Spearman rank correlations between seed
887 counts and the climate variable, assuming linearity in this relationship, which can only be assumed if
888 zero-counts are excluded. Lastly, climate variables might act differently on the number of annually
889 produced seeds than on whether or not reproduction occurs. This can however be accounted for by
890 using more complex, zero-inflated models, which model these two components separately and
891 integrate their effects.

892 Most recent approaches start to account for some of these limitations (Journé, Kelly, et al., 2025) by
893 standardising data through absolute maximum standardisation (i.e., dividing each value by the
894 maximum value of the time series), making seed counts bound between 0 and 1, and fitting beta
895 regressions testing multiple climate variables in the same model. This accounts better for the data
896 structure and does not exclude zero-counts, but magnitudes of effects between time series (i.e.,
897 populations) become less comparable through this standardisation.

898

899 **B.2. Our approach**

900 To account for the limitations described in B.1., we developed a new approach, as describes in detail
901 in the main text. It is based on a sliding window approach, which has been shown to be robust and
902 reliable in identifying weather cues for masting (Journé, Simmonds, et al., 2025) and is commonly used
903 through the R package *climwin* (Bailey & van de Pol, 2016) for many biological response variables (e.g.,
904 egg laying dates in birds). As *climwin* currently cannot fit zero-inflated negative binomial models,
905 which are needed to accurately capture the structure of our dataset, we developed our own approach
906 (for a general description see methods 3 main text).

907 Our sliding window approach tested a total of 15343 windows for each climate variable, and this was
908 repeated four times to make sure that climate windows were stable and not selected randomly. For
909 each window a zero-inflated negative binomial GLMM was fitted using *glmmTMB* (Brooks et al., 2017),
910 with total annual seed count per tree as the response variable, temperatures in T1, T2 and T0 and
911 precipitation in T1, T2 and T0, as well as total number of nuts of the previous year as explanatory
912 variables and tree ID as random effect. The zero-inflated model was dependent on all fixed effects.
913 For each set of 15343 tested windows, the focal climate variable could vary according to the window
914 tested, while all other variables were kept constant to a certain window (i.e., first to the value of the
915 window defined by the literature in the base model, then to the value of their best window of the
916 previous iteration). Once a best window was defined for the focal variable by choosing the window
917 with the lowest AIC, this variable was set to the value of this best window, and the next climate
918 variable was tested. Hence, after the first iteration, all climate variables were set to their respective

919 best windows defined by the first iteration, and the same procedure was followed for the second,
 920 third, and fourth iteration. To determine which measure for temperature is best suited to explain seed
 921 production, we additionally tested all windows for maximum, minimum and mean temperatures in
 922 the first iteration for all three years (T1, T2, T0) and chose the temperature measure that had the best
 923 window with the lowest AIC to be further used in the remaining models. For better understandability
 924 of this process, see Table S1 showing the exact specifications of climate variables that have been used
 925 in the four iterations.

926

927 **Table S2 Model specifications of iterative process to find sensitivity windows for each climate variable.** Each
 928 model includes a temperature and precipitation variable for year T0 (year of seed fall), T1 (year prior to seed fall)
 929 and T2 (two years prior to seed fall). Each model is run for every window of a length of 7 to 140 days between
 930 March 21 and September 22, resulting in 15343 windows. The focal variable varies according to the respective
 931 window, while the remaining variables are kept constant to a specified window. The window for which each
 932 variable is kept constant is specified in brackets: “base” = window commonly defined in the literature (June &
 933 July for T1 and T2 and May to August for temperature T0, March to April for precipitation T0), “best win In” =
 934 best window found in nth iteration (n can be first, second, third or fourth), and “test” indicating the focal variable
 935 that can vary (additionally bold). After model 3, 6 and 9 it was decided based on the lowest AIC value, which of
 936 the temperature variables is used (min, max or mean temperature) in further models (i.e., max temperature for
 937 T1 and T2, mean temperature for T0). For the second iteration, all variables are set to their respective best
 938 window found in the first iteration and each variable is tested again and the same process is used for the third
 939 and fourth iteration (i.e., setting all variables to the best window of the previous iteration). All models are zero-
 940 inflated negative binomial GLMMs with tree ID as a random effect, number of nuts of the previous year as fixed
 941 effect, and individual-level total annual number of beechnuts as response variable (see methods c) in main text).

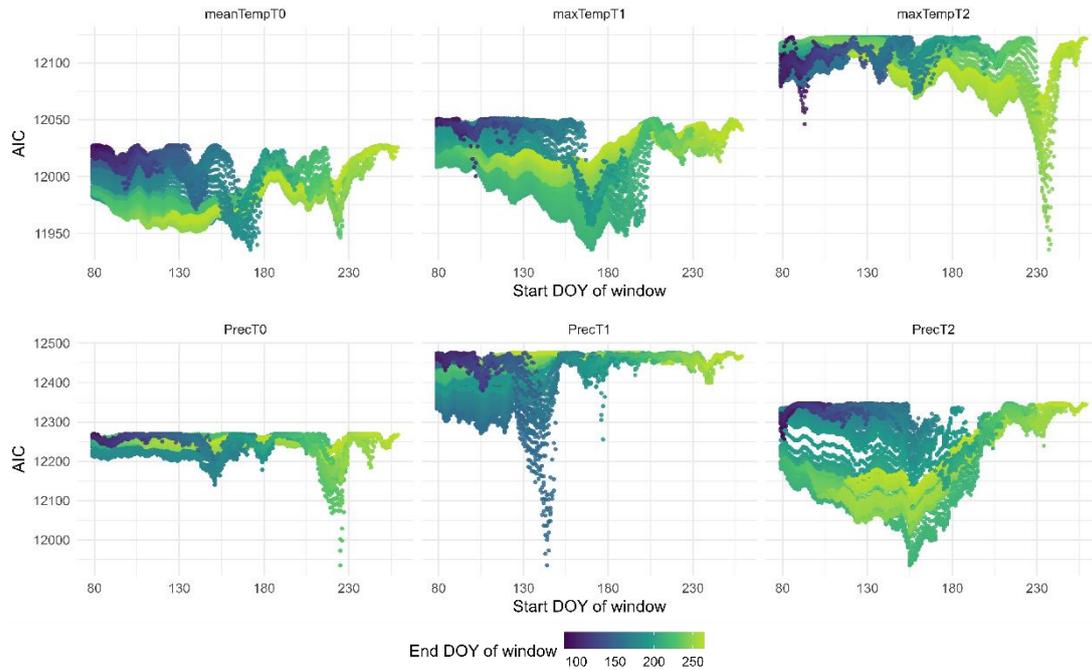
First iteration (I1)	
1.	meanTempT0 (base) + maxTempT1 (test) + maxTempT2 (base) + precT0 (base) + precT1 (base) + precT2 (base)
2.	meanTempT0 (base) + minTempT1 (test) + maxTempT2 (base) + precT0 (base) + precT1 (base) + precT2 (base)
3.	meanTempT0 (base) + meanTempT1 (test) + maxTempT2 (base) + precT0 (base) + precT1 (base) + precT2 (base)
4.	meanTempT0 (base) + maxTempT1 (best win I1) + maxTempT2 (test) + precT0 (base) + precT1 (base) + precT2 (base)
5.	meanTempT0 (base) + maxTempT1 (best win I1) + minTempT2 (test) + precT0 (base) + precT1 (base) + precT2 (base)
6.	meanTempT0 (base) + maxTempT1 (best win I1) + meanTempT2 (test) + precT0 (base) + precT1 (base) + precT2 (base)
7.	meanTempT0 (test) + maxTempT1 (best win I1) + maxTempT2 (best win I1) + precT0 (base) + precT1 (base) + precT2 (base)
8.	maxTempT0 (test) + maxTempT1 (best win I1) + maxTempT2 (best win I1) + precT0 (base) + precT1 (base) + precT2 (base)
9.	minTempT0 (test) + maxTempT1 (best win I1) + maxTempT2 (best win I1) + precT0 (base) + precT1 (base) + precT2 (base)
10.	meanTempT0 (best win I1) + maxTempT1 (best win I1) + maxTempT2 (best win I1) + precT0 (base) + precT1 (test) + precT2 (base)
11.	meanTempT0 (best win I1) + maxTempT1 (best win I1) + maxTempT2 (best win I1) + precT0 (base) + precT1 (best win I1) + precT2 (test)
12.	meanTempT0 (best win I1) + maxTempT1 (best win I1) + maxTempT2 (best win I1) + precT0 (test) + precT1 (best win I1) + precT2 (best win I1)
Second iteration (I2)	
13.	meanTempT0 (best win I1) + maxTempT1 (test) + maxTempT2 (best win I1) + precT0 (best win I1) + precT1 (best win I1) + precT2 (best win I1)
14.	meanTempT0 (best win I1) + maxTempT1 (best win I2) + maxTempT2 (test) + precT0 (best win I1) + precT1 (best win I1) + precT2 (best win I1)
15.	meanTempT0 (test) + maxTempT1 (best win I2) + maxTempT2 (best win I2) + precT0 (best win I1) + precT1 (best win I1) + precT2 (best win I1)
16.	meanTempT0 (best win I2) + maxTempT1 (best win I2) + maxTempT2 (best win I2) + precT0 (best win I1) + precT1 (test) + precT2 (best win I1)
17.	meanTempT0 (best win I2) + maxTempT1 (best win I2) + maxTempT2 (best win I2) + precT0 (best win I1) + precT1 (best win I2) + precT2 (test)
18.	meanTempT0 (best win I2) + maxTempT1 (best win I2) + maxTempT2 (best win I2) + precT0 (test) + precT1 (best win I2) + precT2 (best win I2)
Third iteration (I3)	
19.	meanTempT0 (best win I2) + maxTempT1 (test) + maxTempT2 (best win I2) + precT0 (best win I2) + precT1 (best win I2) + precT2 (best win I2)
20.	meanTempT0 (best win I2) + maxTempT1 (best win I3) + maxTempT2 (test) + precT0 (best win I2) + precT1 (best win I2) + precT2 (best win I2)
21.	meanTempT0 (test) + maxTempT1 (best win I3) + maxTempT2 (best win I3) + precT0 (best win I2) + precT1 (best win I2) + precT2 (best win I2)
22.	meanTempT0 (best win I3) + maxTempT1 (best win I3) + maxTempT2 (best win I3) + precT0 (best win I2) + precT1 (test) + precT2 (best win I2)
23.	meanTempT0 (best win I3) + maxTempT1 (best win I3) + maxTempT2 (best win I3) + precT0 (best win I2) + precT1 (best win I3) + precT2 (test)
24.	meanTempT0 (best win I3) + maxTempT1 (best win I3) + maxTempT2 (best win I3) + precT0 (test) + precT1 (best win I3) + precT2 (best win I3)
Fourth iteration	
25.	meanTempT0 (best win I3) + maxTempT1 (test) + maxTempT2 (best win I3) + precT0 (best win I3) + precT1 (best win I3) + precT2 (best win I3)
26.	meanTempT0 (best win I3) + maxTempT1 (best win I4) + maxTempT2 (test) + precT0 (best win I3) + precT1 (best win I3) + precT2 (best win I3)
27.	meanTempT0 (test) + maxTempT1 (best win I4) + maxTempT2 (best win I4) + precT0 (best win I3) + precT1 (best win I3) + precT2 (best win I3)
28.	meanTempT0 (best win I4) + maxTempT1 (best win I4) + maxTempT2 (best win I4) + precT0 (best win I3) + precT1 (test) + precT2 (best win I3)
29.	meanTempT0 (best win I4) + maxTempT1 (best win I4) + maxTempT2 (best win I4) + precT0 (best win I3) + precT1 (best win I4) + precT2 (test)
30.	meanTempT0 (best win I4) + maxTempT1 (best win I4) + maxTempT2 (best win I4) + precT0 (test) + precT1 (best win I4) + precT2 (best win I4)

942

943 As there was only little variation in the selected best windows between the third and fourth iteration,
 944 if any, we did not do further iterations. The best window for each climate variable in each iteration is
 945 shown in Table S2. Given the strong signal and clear trough in AIC values for all climate variables in
 946 the fourth iteration (Figure S2), we are certain that selected windows are not random.

947 **Table S3 Best windows from all four iterations.** For each climate variable, the start and end days of the best
 948 window are given as DOY of the year and calendar date in parentheses (assuming the year was not a leap
 949 year), as well as the window length and the AIC of the model for all four iterations.

Climate variable	Windows start	Window End	Windows length	AIC
First iteration				
Mean temperature T0	131 (12/05)	245 (03/09)	115	12473.12
Max temperature T1	165 (15/06)	204 (24/07)	40	12588.18
Max temperature T2	208 (28/07)	215 (04/08)	8	12511.33
Precipitation T0	225 (14/08)	231 (20/08)	7	12163.27
Precipitation T1	144 (25/05)	154 (04/06)	11	12297.49
Precipitation T2	157 (07/06)	253 (11/09)	97	12248.7
Second iteration				
Mean temperature T0	176 (26/06)	182 (02/07)	7	12042.9
Max temperature T1	171 (21/06)	209 (29/07)	39	12093.34
Max temperature T2	237 (26/08)	243 (01/09)	7	12060.68
Precipitation T0	225 (14/08)	231 (20/08)	7	11948.17
Precipitation T1	144 (25/05)	154 (04/06)	11	12042.9
Precipitation T2	155 (05/06)	214 (03/08)	60	11948.17
Third iteration				
Mean temperature T0	172 (22/06)	181 (01/07)	10	11936.9
Max temperature T1	170 (20/06)	212 (01/08)	43	11947.28
Max temperature T2	237 (26/08)	244 (02/09)	8	11937.62
Precipitation T0	225 (14/08)	231 (20/08)	7	11936.47
Precipitation T1	144 (25/05)	154 (04/06)	11	11936.9
Precipitation T2	155 (05/06)	215 (04/08)	61	11936.47
Fourth iteration				
Mean temperature T0	172 (22/06)	181 (01/07)	10	11935.7
Max temperature T1	170 (20/06)	209 (29/07)	40	11935.7
Max temperature T2	237 (26/08)	244 (02/09)	8	11935.7
Precipitation T0	225 (14/08)	231 (20/08)	7	11935.62
Precipitation T1	144 (25/05)	154 (04/06)	11	11935.7
Precipitation T2	155 (05/06)	214 (03/08)	60	11935.62

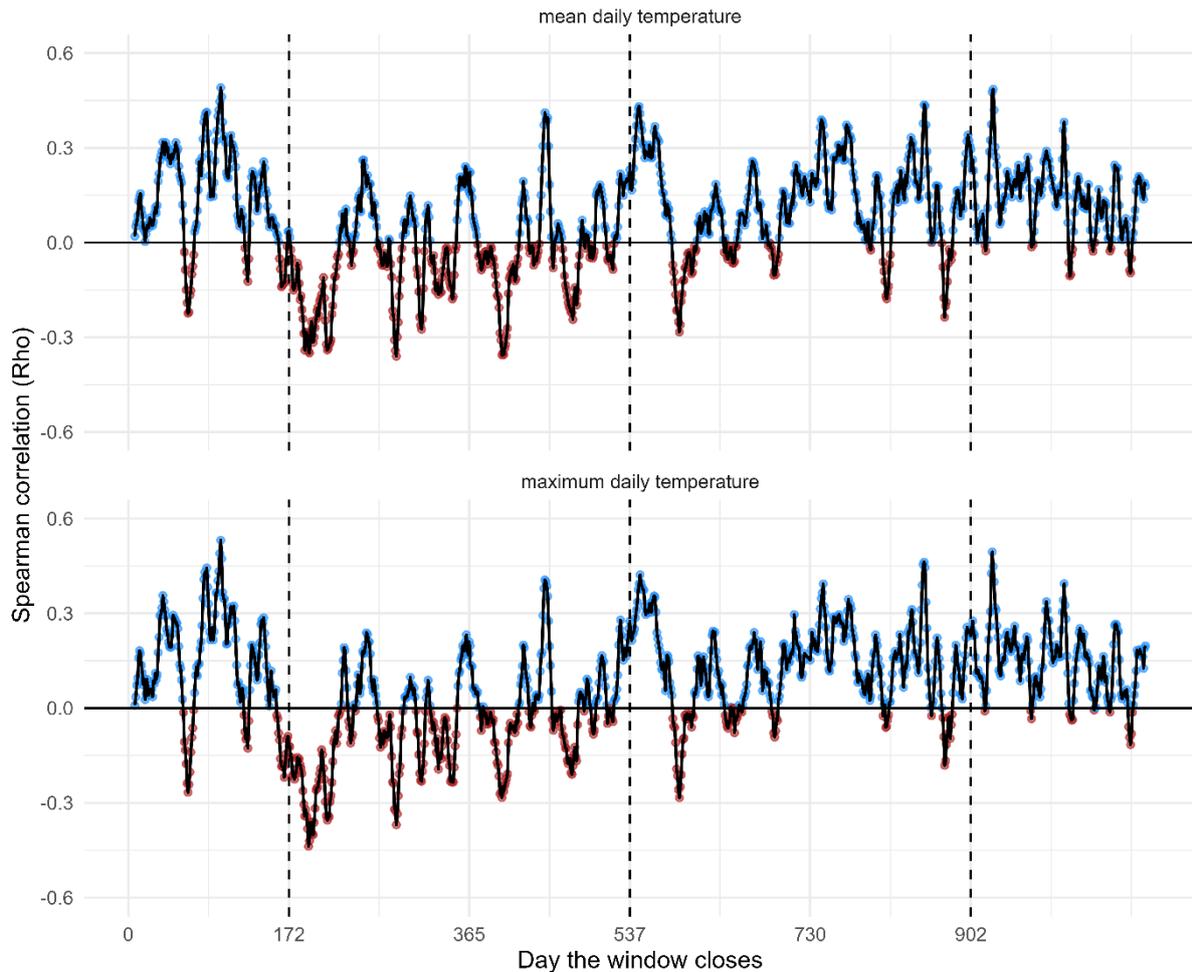


951 **Figure S2 AIC values of all tested windows per climate variable in the fourth iteration.** For each of the 15 343
 952 windows tested per climate variable, the AIC value of the respective model is plotted against the start day of the
 953 window. For all variables, there is a clear pattern in AIC values with a through around a certain date visible. This
 954 indicates that the window with the lowest AIC, which is selected as the best window, is not selected by chance,
 955 as all windows around the same date with a similar length have low AIC values as well.

956

957 **B.3. Comparison with approach used in Journé et al., (2024)**

958 Our new approach gives slightly different results than commonly reported in the literature when using
 959 established methodologies. To test whether the differences we find are based on biological or
 960 methodological differences, we additionally used the approach of Journé et al. (2024) on our data to
 961 assess whether we can replicate the results of Journé et al. for our data when using their approach,
 962 indicating a methodological difference, or still find different results, indicating a biological difference.
 963 We chose this study as a comparison, as their main finding, that temperature windows are anchored
 964 to the summer solstice, has broadly been picked up in following studies and because they use a large
 965 dataset covering several populations, rather than simulated data (Journé, Simmonds, et al., 2025),
 966 making it a good baseline to compare our findings to.



967

968 **Figure S3 Spearman correlation of log-transformed annual beechnut count and temperatures in a seven-day**
 969 **sliding window with one-day step-size.** The x-axis shows the day of the year on which the seven-day window
 970 closes, starting at January 1 of year T-2 (DOY 1) until December 31 of the year of seed fall (DOY 1095). All years
 971 are treated as non-leap years. The dashed lines indicate the day of the summer solstice in all three years. Blue
 972 dots show positive correlations, red negative correlations. The upper panel shows the correlation with mean daily
 973 temperatures (as used in Journé et al. (2024), the lower panel the correlation with maximum daily temperatures
 974 as used in the main analysis.

975 Journé et al., (2024) also used a sliding window approach with a fixed window size of seven days and
 976 a one-day step size. For each window, they calculated Spearman rank correlations between the log-
 977 transformed mean total annual number of nuts per tree and the mean daily temperature over three
 978 years (year of seed fall and the two years before that). They found a clear peak in correlations for both
 979 years (negative for T2, positive for T1) opening at the summer solstice. As shown in Figure S3, we find
 980 a similar pattern, i.e., a negative peak in correlation coefficients starting at the summer solstice in T2
 981 and a positive peak starting around the solstice in T1. In our population, these peaks are however not
 982 stronger than other peaks throughout the years, just slightly wider. For comparability with the results
 983 of our main approach, we additionally did the same analysis using maximum daily temperatures,
 984 resulting in neglectable differences compared to mean daily temperature (Figure S3, lower panel).

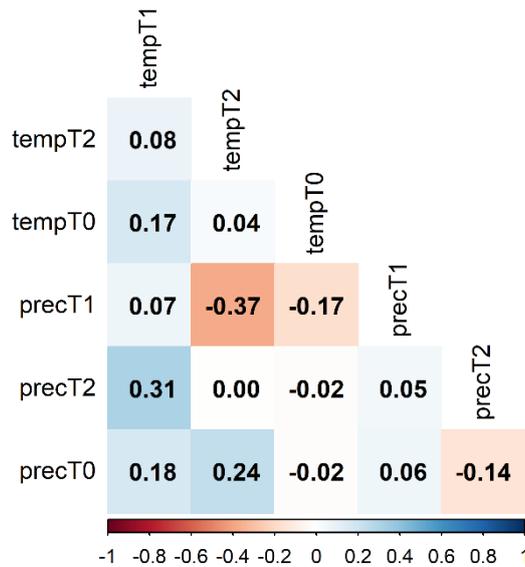
985 Given that Journé et al. report the means over many populations, the slight deviations in the results
 986 that we see for our populations are within an acceptable range, making our results (when using their
 987 methodology) comparable with theirs. This indicates that our study population, which has previously
 988 not been included in comparative analyses, is not an outlier, further suggesting that the differences in

989 the identified cueing windows we find by using our new approach are due to methodological rather
 990 than biological differences. For further discussion see discussion of the main text.

991

992 **B.4. Collinearity test of explanatory climate variables**

993 To exclude collinearity of our explanatory climatic variables, we calculated the Pearson correlation
 994 coefficient between each pair of climate variables, all set to their respective best window as
 995 determined in the fourth iteration (Figure S4). Correlation coefficients are low (i.e., absolute $\rho < 0.4$)
 996 for all variables, allowing us to fit all six variables into the final model.



997

998 **Figure S4 Correlation matrix of the six explanatory climate variables.** Numbers show the Pearson correlation
 999 coefficient (ρ) between each pair of climate variables. Colours facilitate the interpretation of these values, with
 1000 darker red squares indicating more negative correlations, and darker blue squares more positive correlations.
 1001 The correlation between climate variables is overall low, allowing to fit all variables into the same model.

1002

1003 **C. Effect of internal resource reserves on seed production**

1004 As the predictions of the final model could not accurately capture the masting breakdown and there
 1005 was unexplained temporal variance remaining when fitting year back into the final model, climate
 1006 variables likely interact with another factor, currently missing in the model. Since the internal resource
 1007 budget of the tree determines its sensitivity to the climatic cues and ultimately determines how many
 1008 resources can be allocated to reproduction, we additionally tested for the effect of individual resource
 1009 availability on annual seed production. Following the method described in Kelly et al. (2025) and Rees
 1010 et al. (2002), we calculated the resource reserves per individual tree based on its cumulative annual
 1011 seed production throughout the study period. We then fitted resources of the previous year as an
 1012 explanatory variable into the final model in an interaction with temperature in T1 for the conditional
 1013 part of the model and as an additive effect for the zero-inflated part of the model¹, as we expected

¹ Model specification in R using *glmmTMB*: `glmm(no. nuts ~ tempT0 + tempT1 * resources + tempT2 + precT0 + precT1 + precT2 + no. nutsT1 + (1| TreeID), ziformula = tempT0 + tempT1 + resources + tempT2 + precT0 + precT1 + precT2 + no. nutsT1 + (1| TreeID), family = nbinom2(), data = data)`

1014 resources to be mainly related to the temperature cue in the year before seed fall, which is when the
1015 resources are used to form the flower primordia.

1016 Resources did not show a significant effect on the number of nuts produced (interaction effect in
1017 conditional part: $\beta = 6.041 * 10^{-5} \pm 4.870 * 10^{-5}$, $z = 1.24$, $p = 0.215$) nor on the probability of
1018 reproducing (additive effect in zero-inflated part: $\beta = 8.978 * 10^{-5} \pm 1.764 * 10^{-4}$, $z = 0.509$, $p = 0.611$).
1019 When removing the interaction effect in the conditional model, resources did still not affect the
1020 number of nuts produced per tree per year (conditional part: $\beta = 6.741 * 10^{-5} \pm 4.730 * 10^{-5}$, $z = 1.43$,
1021 $p = 0.154$; zero-inflated part: $\beta = 9.097 * 10^{-5} \pm 1.767 * 10^{-4}$, $z = 0.515$, $p = 0.607$).

1022 As previous studies (Kelly et al., (2025)) only included temperatures in T1 and T2 (one and two years
1023 prior to seed fall, respectively) in the model when testing for an effect of resources, we additionally
1024 also excluded temperatures in the year of seed fall and all three precipitation cues from the model.
1025 We again fitted an interaction of resources with temperatures in T1 for the conditional part of the
1026 model, and an additive effect for the zero-inflated part. In this model, resources still do not have an
1027 effect on the probability of a zero-year ($\beta = 9.784 * 10^{-5} \pm 1.203 * 10^{-4}$, $z = 0.813$, $p = 0.416$), while the
1028 interaction with temperature of T1 shows a significantly positive effect on the number of nuts
1029 produced, if there is reproduction ($\beta = 1.671 * 10^{-4} \pm 5.966 * 10^{-5}$, $z = 2.80$, $p = 0.005$). The temperature
1030 cue in T2 does then however no longer affect the number of nuts ($\beta = -0.04428 \pm 0.03670$, $z = -1.16$, p
1031 $= 0.247$).

1032

1033 **References**

1034 Bailey, L. D., & van de Pol, M. (2016). climwin: An R Toolbox for Climate Window Analysis. *PLOS ONE*,
1035 11(12). <https://doi.org/10.1371/journal.pone.0167980>

1036 Brooks, M. E., Kristensen, K., Benthem, K. J., van, Magnusson, A., Berg, C. W., Nielsen, A., Skaug, H. J.,
1037 Mächler, M., & Bolker, B. M. (2017). glmmTMB Balances Speed and Flexibility Among
1038 Packages for Zero-inflated Generalized Linear Mixed Modeling. *The R Journal*, 9(2), 378.
1039 <https://doi.org/10.32614/RJ-2017-066>

1040 Journé, V., Kelly, D., Hacket-Pain, A., Pearse, I. S., Szymkowiak, J., Foest, J. J., Kondrat, K.,
1041 Oberklammer, I., Pesendorfer, M. B., Satake, A., & Bogdziewicz, M. (2025). Weather drivers
1042 of reproductive variability in perennial plants and their implications for climate change risks.
1043 *Nature Communications*, 16(1), 9226. <https://doi.org/10.1038/s41467-025-64300-6>

1044 Journé, V., Simmonds, E. G., Barczyk, M. K., & Bogdziewicz, M. (2025). Comparing statistical methods
1045 for detecting weather cues of mast seeding in European beech (*Fagus sylvatica*) across
1046 Europe. *Agricultural and Forest Meteorology*, 375, 110857.
1047 <https://doi.org/10.1016/j.agrformet.2025.110857>

1048 Journé, V., Szymkowiak, J., Foest, J., Hacket-Pain, A., Kelly, D., & Bogdziewicz, M. (2024). Summer
1049 solstice orchestrates the subcontinental-scale synchrony of mast seeding. *Nature Plants*,
1050 *10*(3), 367–373. <https://doi.org/10.1038/s41477-024-01651-w>

1051 Kelly, D., Szymkowiak, J., Hacket-Pain, A., & Bogdziewicz, M. (2025). Fine-tuning mast seeding: As
1052 resources accumulate, plants become more sensitive to weather cues. *New Phytologist*,
1053 *246*(5), 1975–1985. <https://doi.org/10.1111/nph.70092>

1054 Rees, M., Kelly, D., & Bjørnstad, O. N. (2002). Snow Tussocks, Chaos, and the Evolution of Mast
1055 Seeding. *The American Naturalist*, *160*(1), 44–59. <https://doi.org/10.1086/340603>

1056