

High-resolution modelling of biodiversity under the shared socio-economic scenarios

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Abstract

Impacts on biodiversity from global climate and land use change manifest through changes in habitat availability and suitability for species. We currently lack fine grain predictions about how species and their habitats respond at the local level to global drivers of change, particularly under future scenarios of coupled climate and land use change. Policies for biodiversity management often place an emphasis on management and protection of threatened and socially important species and their habitat. This requires modelling many species individually to inform national policy design and decisions. However, most multi-species modelling systems focus on the role of bioclimatic drivers of species distribution and abundance to the exclusion of the land use variables needed to understand and predict outcomes of socio-economic drivers of biodiversity change. We provide high-quality predictive species distribution models that capture the impacts of change in land use and other environmental drivers, and a comprehensive biodiversity dataset for Australia for thousands of species at high spatial resolution. We model the response of 1,488 species, including 592 birds, 254 mammals and 642 reptiles under coupled land-use (Shared Socioeconomic Pathways) and climate (Representative Concentration Pathways) scenarios. Our analyses present new possibilities for developing bespoke species-specific models when analysing many species at large spatial scales, that combine disciplinary ideas from data-science, high performance computing and species distribution modelling. Our assessment provides a new understanding of how projected socio-economic and climate scenarios play out for biodiversity.

40 Introduction

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42 Scenarios and model predictions are useful for helping map pathways and shape policy.
43 Making robust predictions of biodiversity response to climate and land use, drivers that
44 operate at the global scale yet their impacts manifest at the local/regional scale, is a step
45 towards developing realistic strategies for achieving national targets such as the 30 by 30
46 target of the Global Biodiversity Framework (GBF)¹. Though many studies map global
47 pathways to achieving GBF targets²⁻⁴, they fail to inform decisions at the national scale. This
48 is because these analyses do not comprehensively capture the biodiversity of any specific
49 country, nor do they provide a nuanced understanding of the local level dynamics and impacts
50 of drivers of change on biodiversity. We need nationally relevant analyses that consider
51 scenarios and constraints suited to the national context to inform policy decisions and
52 articulate the trade-offs of competing environmental or biodiversity targets⁵.

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54 The Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services
55 (IPBES) recognises the need to understand biodiversity impacts under future environmental
56 and anthropogenic conditions⁶. Governments are developing national plans to align with
57 internationally agreed sustainability goals and biodiversity targets and the private sector's
58 appetite to integrate biodiversity impacts into project design is growing^{7,8}. However,
59 biodiversity detail in current approaches for mapping or evaluating policy or planning
60 pathways are either taxonomically incomplete or overlook local-level dynamics in biodiversity
61 impacts. State and regional policies for biodiversity management often emphasise
62 management and protection of individual, threatened or locally significant, species and their
63 habitats. As such, species are a critically important attribute for conservation planning but
64 very few continental or global-scale analyses model individual species responses to future
65 scenarios^{9,10}. We need better approaches to (i) accurately assess individual species' outcomes
66 under future change scenarios and (ii) aggregate predicted outcomes for many species into
67 indices of biodiversity change to enable more reliable guidance for national and continental
68 scale environmental policy design and evaluation.

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70 Most national-scale biodiversity assessments for Australia model aggregated biodiversity, e.g.
71 dissimilarity, richness, or mean species abundance^{11,12}. Top-down approaches that rely on
72 modelling macro-ecological relationships between the whole of biodiversity and land
73 use/climate cannot reveal individual species responses, their vulnerability to changing
74 conditions or the predicted future spatial distribution of their habitats. Few studies have
75 evaluated spatial biodiversity patterns under coupled global climate and land use change
76 scenarios for large numbers of species using species distribution models that explicitly models
77 individual species responses to individual climate and land use change variables¹³⁻¹⁵. The
78 absence of species-level (bottom up *sensu*¹⁶) model inference in national or continental
79 analyses of biodiversity change under land use futures limits our ability to understand species-

80 level response to new and emerging drivers of change, to infer which elements of biodiversity
81 are declining or improving, or to identify the winners and losers and future land use and
82 climate scenarios. Ignoring the intimate link between land use change and species response
83 downplays the influence of a direct driver of biodiversity change, both for habitat and species
84 persistence. It also hampers the kind of understanding and communication of change that
85 most people can connect with, which tends to individual species or ecosystem-level
86 connections¹⁷.

87

88 Scenario analyses that describe future outcomes for biodiversity based on shared socio-
89 economic pathways (SSP)¹⁸ often model biodiversity at the global scale using proxies¹⁹, coarse
90 range maps for species²⁰ or macro-ecological models such as PREDICTS that rely on
91 correlations between land use and biodiversity intactness^{14,21}. In Australia, where significant
92 effort has been made to develop biodiversity data repositories for use in conservation
93 planning and climate change impact analysis at the state and federal levels, a high spatial
94 resolution and taxonomically rich analysis of biodiversity under SSP scenarios is currently
95 lacking. Building a picture of biodiversity outcomes under coupled climate and land use
96 change scenarios from the bottom up, starting with species and aggregating to biodiversity,
97 should arguably provide a more accurate and interrogable picture of future biodiversity
98 trends because individual species responses to environmental change is accommodated,
99 rather than assuming all elements of biodiversity will respond in approximately similar ways
100 to environmental change.

101

102 Individual species responses to environmental change can be modelled using species
103 distribution models (SDMs) to assess likely habitats and for use within conservation decision
104 making^{22,23}. Using SDMs within scenario modelling requires high resolution spatial
105 information on the environment and biodiversity. With the aid of cutting-edge, high-
106 performance computing and data storage solutions, we conduct a comprehensive national
107 scale biodiversity assessment for Australia under global climate and land use change
108 projections, for a suite of coupled Shared Socioeconomic Pathways (SSPs) and Representative
109 Concentration Pathways (RCPs), focussing on outcomes for species. We model current and
110 future predicted distributions of 1,488 species at a high spatial resolution (250m) to help
111 tease apart individual species response and whole-of-biodiversity outcomes under the
112 scenarios. Our assessment encompasses all known, recorded and valid species of birds,
113 mammals and reptiles for Australia. We highlight differential response across scenarios and
114 taxonomic groups as well as emphasize the likely impacts for threatened species.
115 Furthermore, the SDMs developed are operationalised for reproducibility and computational
116 efficiency with respect to data and compute requirements. We provide our high-quality,
117 open-source biodiversity modelling workflow and comprehensive data to allow users to
118 model many species at once and at high spatial resolution to represent whole-of-biodiversity
119 outcomes of scenarios and policies.

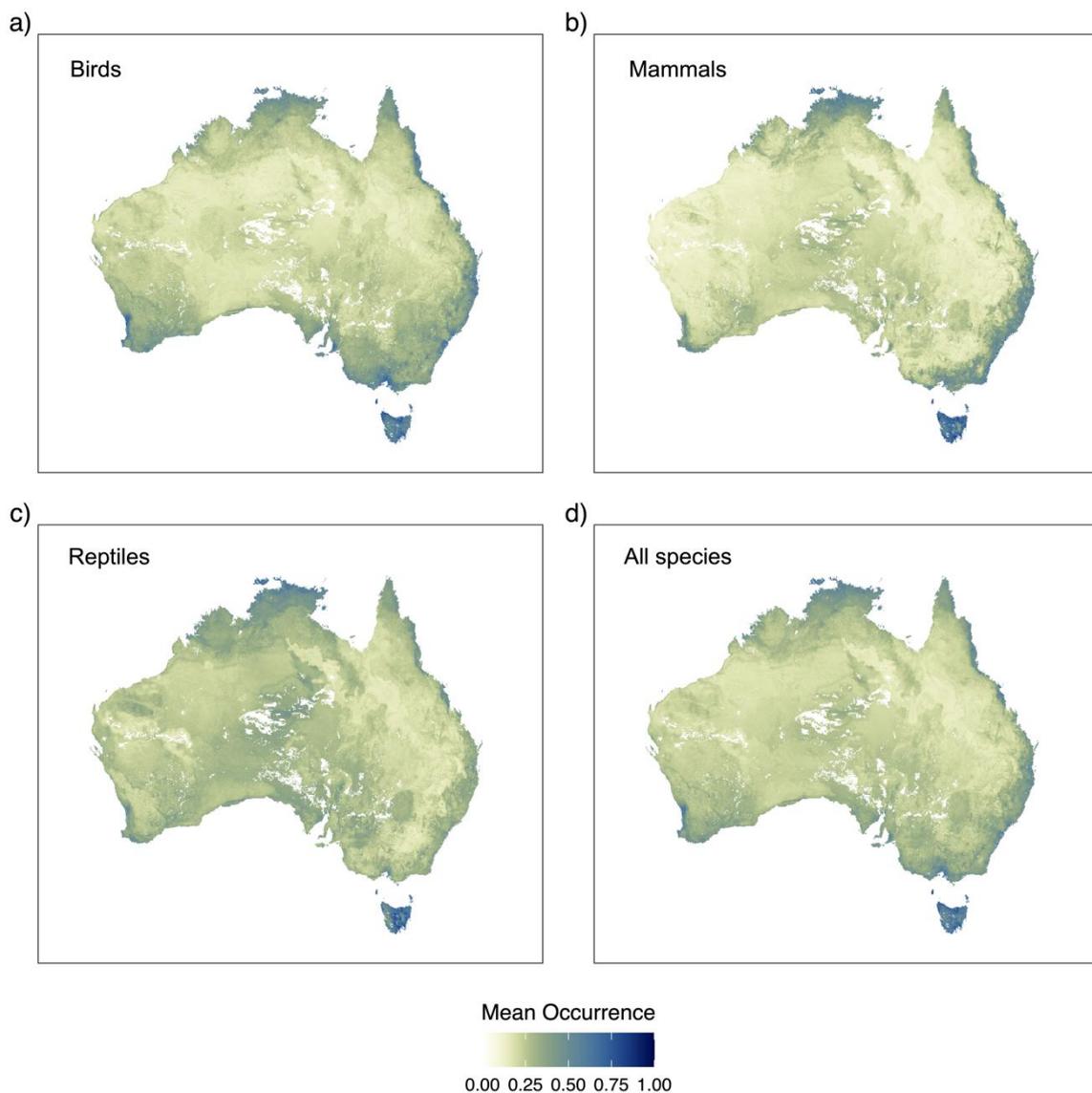
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121 **Results**

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123 Mean relative likelihood of occurrence for the current time step (2020) for each taxonomic
124 group and across all species for the entire study region highlights clear geographic variation
125 in biodiversity patterns across Australia (Figure 1). Areas of high mean relative likelihood of
126 occurrence are concentrated primarily along the east coast with additional pockets around
127 Perth and Darwin for birds and mammals, with a stronger coastal preference for mammals.
128 Reptile species have approximately equivalent preferences for arid and semi-arid areas as
129 they do for coastal areas. All three taxa groups have a very high mean likelihood of occurrence
130 values for Tasmania.

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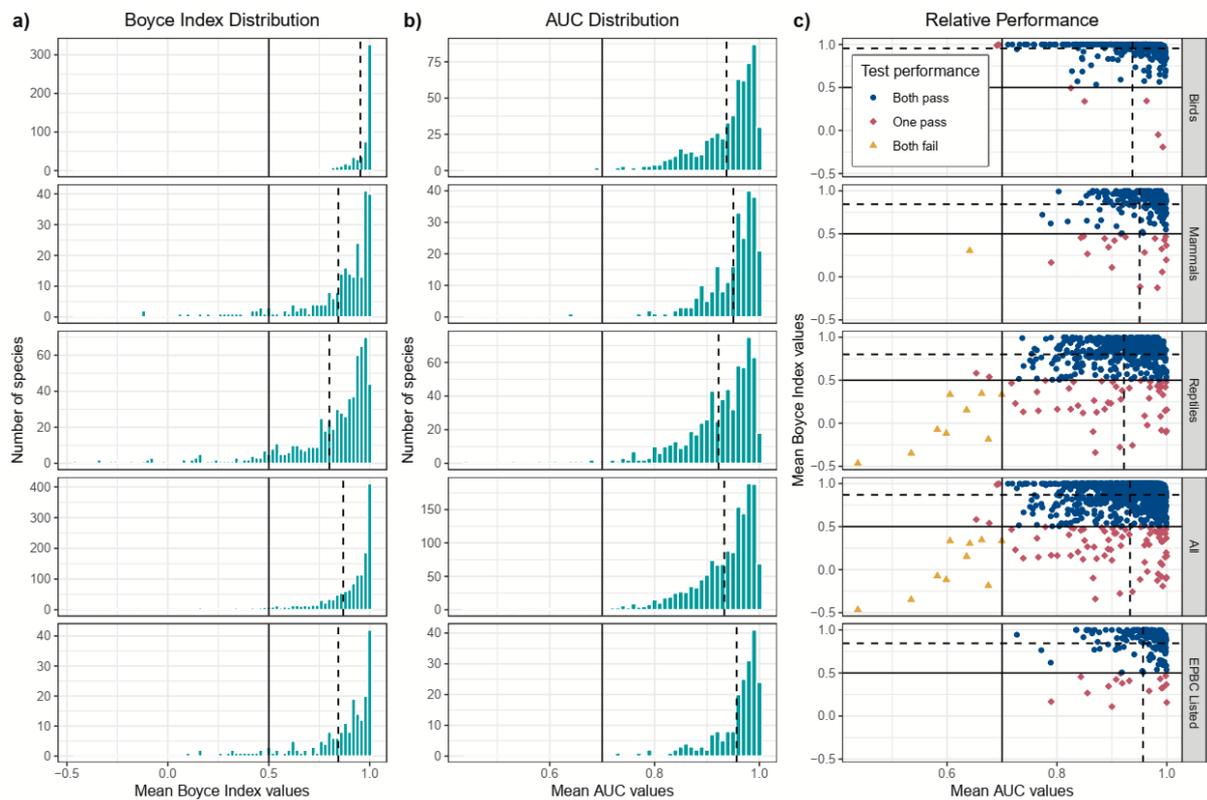


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133 *Figure 1: Mean relative likelihood of occurrences for each taxonomic group, including a) birds, b) mammals, c)*
134 *reptiles and d) all species for the current time period. d) Performance metric summaries (additional metrics*
135 *provided in the supplementary information).*

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Model performance is typically high across most species. Mean Boyce Index (BI) values across all species, taxa groups and EPBC listed species is greater than 0.8 indicating good model performance overall (Figure 2, extended data Table E2.1). Only 5.8% of the models (n = 86 including 5 birds, 21 mammals and 60 reptiles) performed poorly with a BI value equal to or less than 0.5 (Table E2.2). Notably, birds performed consistently well: 99% of species with BI values greater than 0.5. 99% of the models (n = 1472) had an AUC value greater than 0.7. 16 out of the 201 EPBC listed species models had a BI value of less than 0.5. Model outputs from the three GCMs exhibit similar predicted patterns (Figures E2.1 a-c), though the projected values vary (Figure E2.1 d-f). A similar trend was observed when comparing individual GCM projections to the averaged projections from all three models (Figures E2.1 g-h). This variation reflects model uncertainty in future climate projections.

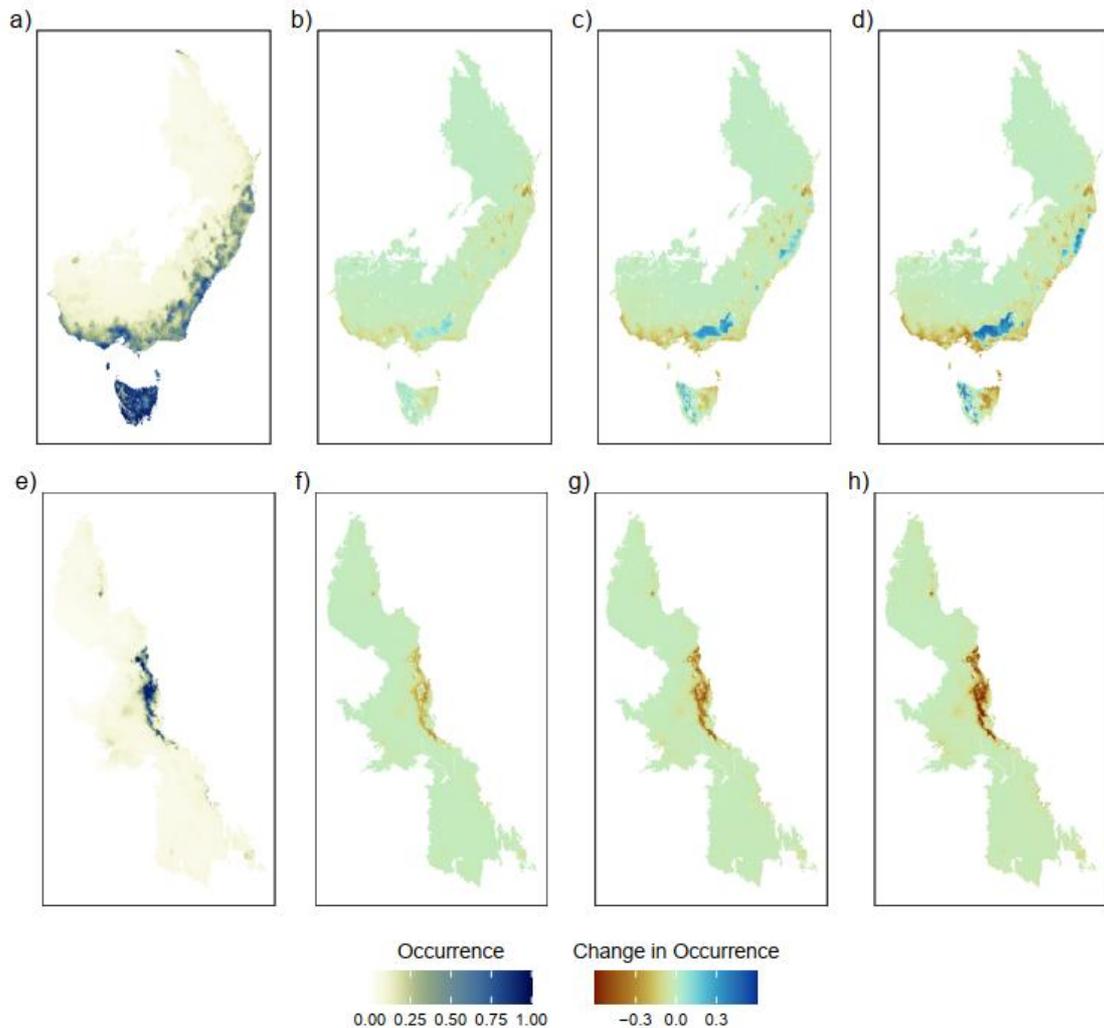


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Figure 2: Distribution plots of model performance metrics (a) Boyce Index (BI) and (b) the Area Under the Curve (AUC) across species for birds, mammals, reptiles, all species, and for EPBC listed species (indicated in rows). Dotted lines are mean performance values and solid lines indicate the 0.5 and 0.7 minimum thresholds for good model performance for BI and AUC, respectively. The scatter plots (c) indicate overall performance showing that very few models performed poorly with respect to both metrics examined here.

Predictions under each coupled land use and climate scenarios (SSP-RCPs) for 2030, 2050 and 2070 show spatial patterns of increase or decrease in habitat suitability across the landscape and we can explore these variations for individual species (Figure 3) and across taxon groups (Figures 4, 5, extended data Figure E2.2). For instance, some species showed divergent

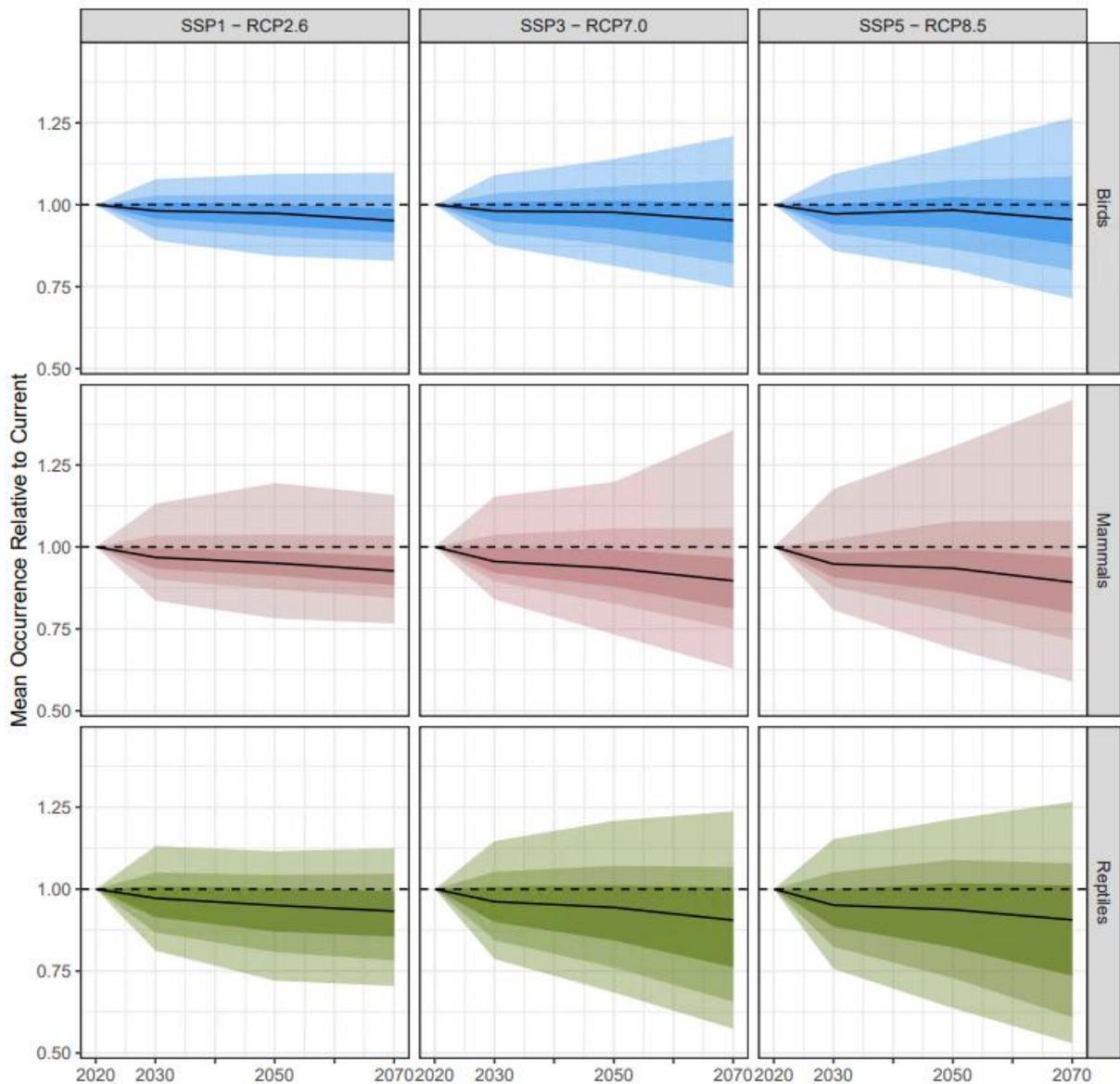
161 responses under alternate climate and land-use scenarios. The mean relative likelihood of
 162 occurrence of the green ringtail possum (*Pseudochirops archeri*) declined under all future
 163 scenarios, but declines were most severe for SSP5 (Figure 3a). By contrast, the spotted-tailed
 164 quoll (*Dasyurus maculatus*) showed both increases and decreases in the relative likelihood of
 165 occurrence across its range under the future scenarios modelled (Figure 3b). Most species in
 166 our analysis, on average, are expected to decline in both mean relative likelihood of
 167 occurrence and range extent under all SSPs (Figure 6).
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 170 *Figure 3: The current estimated relative likelihood of occurrence (a) and expected change in relative likelihood of*
 171 *occurrence under the SSP 5 - RCP 8.5 climate scenario in the years b) 2030, c) 2050 and d) 2070 for the spot-*
 172 *tailed quoll (*Dasyurus maculatus*), and the current estimated relative likelihood of occurrence (e) and expected*
 173 *change in relative likelihood of occurrence under the SSP5 (RCP 8.5) climate scenario in the years f) 2030, g) 2050*
 174 *and h) 2070 for the green ringtail possum (*Pseudochirops archeri*).*
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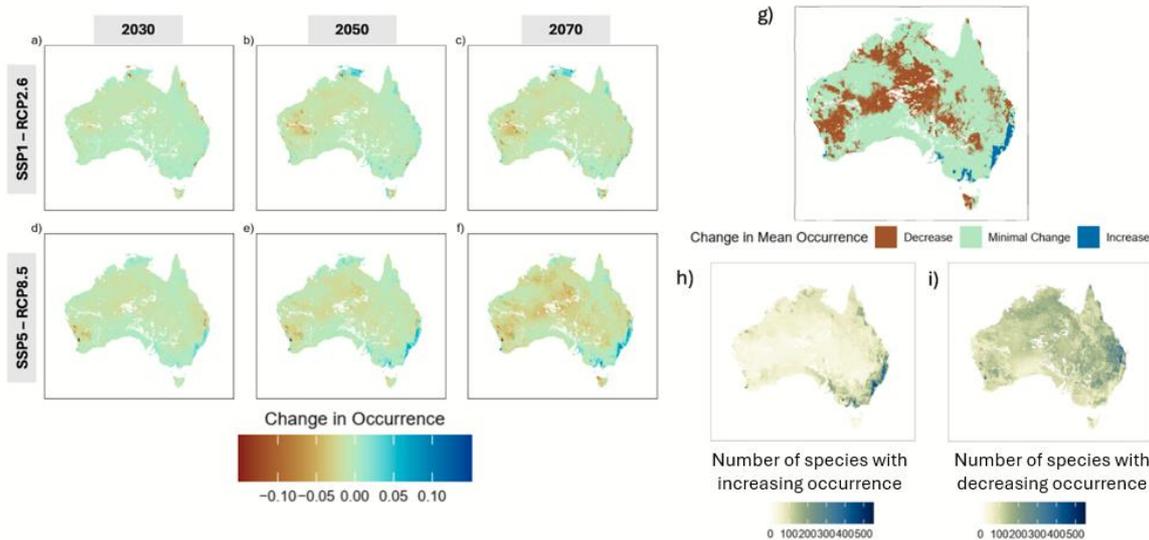
176 When looking across taxonomic groups in their entirety, we found that on average the relative
 177 likelihood of occurrence of all species declined under the future scenarios, and these declines
 178 were most severe for SSP 5 - RCP 8.5(Figure 4). Across the species in our study, at least 75%
 179 are predicted to decline in their average likelihood of occurrence by 2070 under the SSP 5 -

180 RCP 8.5 scenario (Figure 4). On average, the relative likelihood of occurrence for all taxon
 181 groups in our study is expected to decline by 5-10% by 2070 under SSP 5 - RCP 8.5 (Figure 4).
 182 However, these declines were not ubiquitous across Australia. Rather, declines in the relative
 183 likelihood of occurrence for all species were centred in central Australia in the arid and semi-
 184 arid regions, as well as localised declines in western Tasmania, mountain regions of far-north
 185 Queensland and parts of coastal Western Australia (Figure 5g,i). By contrast, our models
 186 predict increases in the mean relative likelihood of occurrence of all taxa in the eastern
 187 foothills and coast of south-eastern Australia, especially under the most severe future
 188 scenarios (SSP 5 - RCP 8.5) (Figure 5g,h). More specifically, locations with the highest richness
 189 of species expected to decline under SSP 5 - RCP 8.5 were in southern Queensland and the
 190 western Australian coast (Figure 5i). Most of the high elevation parts of Australia's eastern
 191 coast also had >100 species expected to decline (Figure 5i). Locations with the highest
 192 richness of species expected to increase under SSP 5 - RCP 8.5 were on the eastern coast of
 193 Australia, as well as central Victoria (Figure 5h).
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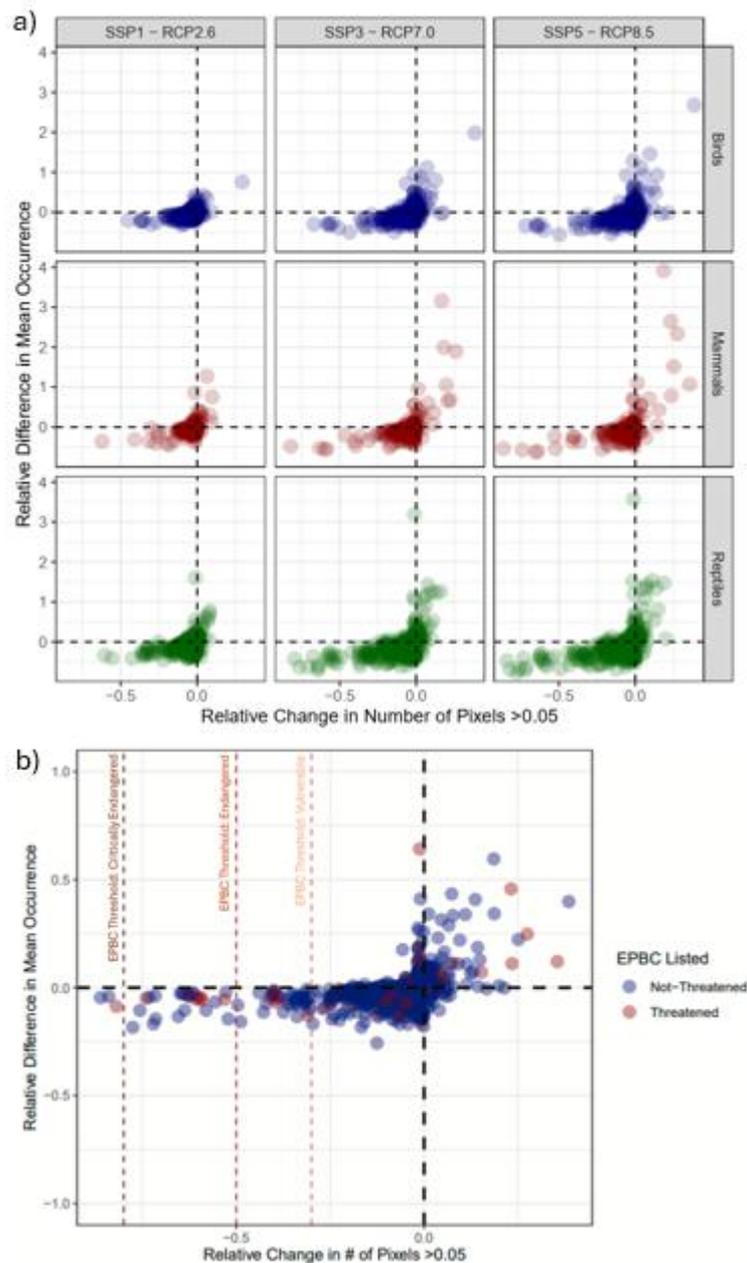
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196 Figure 4: Change in the relative likelihood of occurrence for each taxonomic group in 2030, 2050 and 2070
 197 compared to current time step (2020) under each SSP scenario. Dotted line indicates no change and the solid
 198 black line indicates the median change in the likelihood of occurrence for each taxonomic group under each
 199 scenario and the ribbons indicate the 50%, 75% and 90% quantiles of individual species trajectories under
 200 averaged projections from the three GCMs.
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 204 Figure 5: Projections under the RCP-SSP scenarios for all species showing (a) change in mean relative likelihood
 205 of occurrence in 2030, 2050 and 2070) relative to mean relative likelihood of occurrence in 2020 for SSP1 - RCP2.5
 206 (a-c) and for SSP5-RCP8.5 (d-f). Predicted increase versus decrease and areas of minimal change in the mean
 207 relative likelihood of occurrence under the RCP8.5-SSP5 scenario (g). Number of species with a predicted increase
 208 in relative likelihood of occurrence under the RCP8.5-SSP5 scenario (h). Number of species with a predicted
 209 decrease in relative likelihood of occurrence under the RCP8.5-SSP5 scenario (i). Projections for SSP3 can be seen
 210 in supplementary Figure E2.2.
 211

212 We also related the change in habitat suitability and extent of suitable habitat across
 213 taxonomic groups to identify types of biodiversity response under future scenarios. As
 214 expected, most species show decreased habitat suitability and range contraction while some
 215 show increase in habitat suitability and range (Figure 6b). Examples of species showing range
 216 contraction despite increased habitat suitability and vice versa were few (e.g. *Amaurornis*
 217 *cinerea* and *Motacilla tschutschensis*). Under SSP 5 - RCP 8.5, 39 species are predicted to
 218 experience range declines by between 30-50% (consistent with a vulnerable EPBC listing), 26
 219 species between 50-80% (consistent with an endangered EPBC listing) and three species,
 220 including the green ringtail possum (*Pseudochirops archeri*) and chameleon gecko
 221 (*Carphodactylus laevis*) by >80% (consistent with a critically endangered EPBC listing, Figure
 222 6b). When looking at EPBC listed species, we find that the impacts on certain species under
 223 SSP 5 - RCP 8.5 in 2070 are likely to be severe. For example, the Adelaide pygmy blue-tongue
 224 skink (*Tiliqua adelaidensis*), northern greater glider (*Petaurus minor*) and lesser sooty owl
 225 (*Tyto multipunctata*) are all predicted to have a reduction in their range (number of pixels
 226 >0.05) of at least 70% under SSP 5 - RCP 8.5 (Figure 6b).



228
 229 *Figure 6: Scatter plots showing the spread of change in species geographic range (on the x axis – proportional*
 230 *change in range size, positive values indicating expansion and negative values indicating contraction) as*
 231 *impacted by change in the species’ habitat suitability (on the y axis – proportional change in probability of*
 232 *occurrence, positive values indicating increase in suitability and vice versa) for across taxa groups. In a) Change*
 233 *is estimated from the current time step to 2070 under each climate scenario. Relative change indicates the*
 234 *change as a proportion of overall probability of occurrence and range size of a species within its species mask.*
 235 *Each point indicates a species, colour coded according to the corresponding taxon group. In b) Change is*
 236 *estimated from the current time step to 2070 under RCP8.5-SSP5. Relative change indicates the change as a*
 237 *proportion of overall probability of occurrence and range size of a species within its species mask. Each point*
 238 *indicates a species, colour coded according to the current EPBC status of the species. Vertical coloured dashed*
 239 *lines represent EPBC Act threat status range reduction thresholds at 30% (light red), 50% (red) and 80% (dark*
 240 *red) for vulnerable, endangered and critically endangered, respectively.*
 241

242 **Discussion**

243 Our modelling results highlight spatial and taxonomic variation in biodiversity responses
244 under future climate and land-use scenarios. When assessing cumulative effects on
245 biodiversity – whether across all species or within taxonomic groups – predicted patterns
246 remain consistent across scenarios, though impact severity generally increases with
247 increasing climate change intensity, with SSP 5 - RCP 8.5 showing the most adverse outcomes
248 for species. Future projections indicate overall declines in biodiversity across taxonomic
249 groups, with expected declines of 5-10% on average in the worst-case scenario and reduced
250 habitat suitability and range contractions for most (approximately 75%) of the species.
251 Impacts on EPBC-listed species were particularly concerning, with several species projected
252 to experience severe declines. These results give further warning of likely extinctions for
253 Australia given current economic growth trajectories if adequate environmental and
254 sustainability targets are not recognised and legislated.

255
256 At the species level, responses varied, with some, like the green ringtail possum, experiencing
257 widespread declines, while others, such as the spotted-tailed quoll, exhibited mixed
258 responses depending on the scenario and section of their range (Figure 3). Under the most
259 severe climate scenario (SSP 5 - RCP 8.5), we predict that 69 mammal, bird and reptile species
260 will experience range declines of at least 30%. Many of these species are not currently listed
261 under the EPBC Act, highlighting the conservation implications of future climate and land-use
262 change. Further, many of these species co-occur in hotspots of decline. For example, our
263 models suggest that far north Queensland and central Australia are likely to experience
264 significant declines in species occurrence (e.g. green ringtail possum, spot-tailed quoll and
265 northern greater glider) under future climate. Indeed, many species in these locations are
266 already experiencing declines. For example, arboreal mammals in far North Queensland have
267 declined in abundance over the past two decades mostly due to increasing
268 temperatures^{24,25}. These insights strengthen the rationale for developing individual species
269 SDMs in scenario modelling as it reveals underlying dynamics in species responses that
270 contribute to the emergent trends in aggregated biodiversity metrics (e.g. overall suitability,
271 richness). Additionally, the ability to identify likely risks and the severity of land use and
272 climate change pressures for vulnerable species or those of national interest (Figure 6b) is
273 crucial for targeted management and threatened species conservation.

274
275 Understanding the role of land use and land use change in shaping species response is vital
276 for shaping biodiversity policy and targets under future scenarios of change. Our models
277 provide the first step towards achieving this in national biodiversity assessments.

278 In response to coupled climate and land-use change, we identify locations across Australia
279 that are likely to face the biggest biodiversity declines and therefore are of immediate
280 conservation concern (e.g., central Australia, parts of Tasmania and coastal Western
281 Australia) as well as biodiversity hotspots in the future (e.g., mountainous regions of
282 southeastern Australia). Our results suggest that the coastal regions of eastern Australia and

283 parts of southern Australia may experience a net increase in the relative likelihood of
284 occurrence of species, indicating the potential of these locations as biodiversity refuges under
285 a changing climate and landscape and warrant recognition in priority planning for habitat
286 protection measures. The locations identified in our results generally mirror previous studies
287 that aim to identify Australian climate refugia²⁶. However, previous studies focus only on
288 climate change, rather than the joint drivers of both climate change and land-use change²⁶.
289 By incorporating land use change variables explicitly in the SDMs coupled with other climate
290 and environmental drivers, we allow for more accurate prediction of likely risks to species
291 habitats as countries move through future stages of economic, social and demographic
292 change.

293

294 The national-scale SDM workflows developed in our study provide three *key* improvements
295 to existing approaches because we: (1) incorporate land use variables from global SSP
296 projections²⁷ explicitly in the SDMs, (2) model the response of individual species under
297 coupled climate-land use scenarios, and (3) provide a reproducible integrated data and
298 analysis pipeline. We also provide technical innovations in national scale biodiversity
299 modelling by developing automated workflows for data processing, allowing expert input into
300 model variable selection²⁸, creating dynamic vegetation layers that vary in response to
301 predicted land use change to use in the SDMs for future predictions, and make explicit
302 modelling decisions to achieve computation efficiency (e.g. use of species masks, hierarchical
303 variable multicollinearity checks) without losing model robustness. Our framework is modular
304 to enable complete transparency and reproducibility for users to review and improve
305 iteratively²⁸. Each step in the workflow is sequential, can be set up for parallel processing if
306 resources allow, and generates standalone outputs that can be examined enroute to final
307 predictions. Moreover, our outputs can be summarised in various ways to address at-scale
308 policy objectives²⁹.

309

310 The challenges of conducting large scale assessments at high spatial resolution, such as the
311 one presented in this study, cannot be understated. Such analyses can often be discouraging
312 due to their considerable data and computation requirements, need for detailed deliberation
313 on modelling assumptions to capture the appropriate response of many species, and the lack
314 of expertise in high performance computing within ecological research. While we overcome
315 many of these challenges, our study presents some limitations that give way to suggestions
316 to help improve biodiversity assessments for national environmental policy. Achieving the
317 best possible model for individual species requires species-specific tailored approaches that
318 are infeasible when modelling large quantities of species at scale. Instead, our priority was to
319 develop a robust modelling workflow to generate predictions for many species at once.
320 Therefore, trade-offs needed to be made between ecological accuracy and computational
321 efficiency while retaining model robustness. This includes the automation of species mask
322 generation, outlier detection, and data cleaning following a set of rules that apply consistently
323 across all species and may miss species-specific nuance. Our modular workflow design does

324 allow fine-tuning of individual species models, if required, without having to re-run the full
325 analyses, albeit with some advice from the authors.

326

327 The exclusion of marine species potentially underestimates overall impacts on Australian
328 biodiversity. While the majority of mammalian and reptilian marine species are entirely
329 marine bound there are several species that are partially terrestrial. These are limited to the
330 coast or around inland rivers and are likely driven by the marine conditions not captured by
331 our existing environmental variables. Extending the analysis to marine species presents
332 significant computational challenges due to the larger spatial extents required for the
333 analysis, however, collaboration with existing marine modelling research provides some
334 opportunities for such assessments³⁰. Marine avian species, on the other hand, especially
335 migratory species, can be found further inland as they are potentially selecting breeding
336 grounds based on terrestrial conditions and thus can be more easily included using our
337 workflow. It is important to note that such areas are often highly productive for agriculture
338 (e.g. the southern Australia riverina) and therefore likely to be in potential conflict between
339 biodiversity and land use change; as such it is important they be included. Our approach can
340 also easily be adapted to include other taxa such as plants, invertebrates or freshwater
341 species where data is available^{31,32}, and with the inclusion of appropriate environmental
342 variables and modelling decisions^{33,34}.

343

344 While we use SSP land use predictions from the global LUH2 project²⁷ to align with global
345 scenario modelling efforts worldwide, using more accurate land use predictions for
346 Australia³⁵ can greatly improve our predictions. For example, some spatial artefacts do appear
347 in our maps of net biodiversity change (e.g. straight-line borders around locations with a net
348 increase in the relative likelihood of occurrence) that could be refined with more precise land
349 use predictions. Alternatively, coupling a land use change model with our SDMs can help
350 better capture land use dynamics and its effect on species distributions¹⁵. This will overcome
351 the shortcomings of current integrated approaches that consider biodiversity simplistically
352 from static range maps (e.g. IUCN range maps¹⁶), aggregated metrics (e.g. mean species
353 abundance²⁰), or proxies (e.g. biodiversity services via environmental plantings⁵), thereby
354 obscuring differential species' response to threats. Used within integrated assessment
355 approaches, our SDMs can represent biodiversity in a more appropriate and meaningful way
356 for policy and impact evaluation, as well as provide the necessary biodiversity detail sought
357 after in footprint analyses⁸.

358

359 Planning towards environmentally sustainable targets and economic activity requires at scale
360 biodiversity assessments that translate the impacts of global socio-economic drivers to local
361 level land use. At the same time, we need to model many species at sufficient spatial and
362 thematic resolution to capture patterns in biodiversity response across the landscape. Our
363 study presents insights on likely future impacts on targeted species and locations across
364 Australia under coupled climate and land use scenarios. By providing individual species

365 models and aggregated whole-of-biodiversity patterns through metrics such as average
366 habitat suitability and species richness, we illustrate the versatility of our approach in
367 informing species-focussed versus habitat-focussed national policy design. The key strength
368 of our approach is the provision of reproducible and adaptable workflows, while drawing on
369 advances in data science and high-performance computing to enable modelling thousands of
370 species at the national scale.

371
372

373 **Methods**

374 ***Species data***

375 We collated all records of terrestrial birds, mammals and reptiles from the Atlas of Living
376 Australia for the years 1970 – 2023 (Atlas of Living Australia website
377 at <http://www.ala.org.au>. Accessed 5 December 2023) using the *galah* package in R (version
378 2.0.2³⁶). Data was screened for geographic and taxonomic quality, resulting in a clean dataset
379 with 15,173,393 records from 1,488 species (Supplementary Table S1.1).

380

381 ***Species masks***

382 Cleaned species occurrence records were overlaid on the IBRA 7 regions and a mask was
383 created for each species by including the IBRA regions with known presences for that species
384 and their adjoining regions. Models were trained using species-specific raster masks and likely
385 distributions were predicted within the mask for a species.

386

387 ***Environmental variables***

388 We considered three coupled socio-economic (Shared Socioeconomic Pathways, SSP) and
389 climate (Representative Concentration Pathways, RCP) scenarios at 2030, 2050 and 2070 for
390 modelling: SSP 1 – RCP 2.6, SSP 3 – RCP 7.0 and SSP 5 – RCP 8.5. To use a consistent set of
391 global climate models (GCMs) across our scenarios, we considered three out of the 14 GCMs
392 from WorldClim that had data for all scenarios (i.e., INM-CM5-0, MIROC6, MRI-ESM2-0) for
393 2021-2040, 2041-2060 and 2061-2080 (referred to as 2030, 2050 and 2070 in our analysis,
394 respectively). We ran all nine GCM-RCP combinations instead of averaging across GCMs. This
395 is because species respond to a combination of rainfall and temperature, and the other
396 variables in SDMs, and pre-averaging GCMs smooths out this important interaction.
397 Vegetation classes in the National Vegetation Information System (NVIS) dataset³⁷ were
398 reclassified to six vegetation classes: rainforests, eucalyptus forests, woodlands, shrublands,
399 grasslands and wetlands, based on expert advice. Future projections of the vegetation
400 variables considered in this study (i.e., tree cover, NDVI, NVIS variables, native neighbourhood
401 and canopy height) were not available. To account for future change in vegetation and non-
402 vegetation classes (i.e., cropland and urban) under future scenarios, we estimated the change
403 based on predicted land use (see supplementary information for details).

404

405 All spatial datasets for environmental variables were obtained from publicly accessible
406 repositories (Table S1.2). Layers were resampled to 250m resolution, reprojected in
407 GDA2020/Australian Albers (EPSG 9473) and clipped using a land/water mask of the
408 Australian coastline.

409

410 ***Bias layers***

411 A unique sampling bias layer was created for each taxon group by combining a kernel density
412 estimate layer and the inverse-distance to features layers (roads, protected areas and water
413 bodies). Each source of bias was normalised to be 0-1 before summing all bias sources
414 together to give each source equal weighting, before a final normalisation 0-1 to serve as the
415 bias layer for background point generation. For each species, the appropriate taxon group
416 bias layer was clipped using the species mask such that the values within the mask were
417 indicative of the sampling bias in the data for that species.

418

419 ***Background points***

420 We generated 50,000 weighted-random background points³⁸ for each species using the
421 ppmData package in R³⁹ (version 1.0.0) using the species-specific bias layer. Range-restricted
422 species, with less than 50,000 non-NA pixels within their species mask, generated background
423 points up to the number of available pixels.

424

425 ***Variable multicollinearity checks***

426 We set up a hierarchical multicollinearity testing method for each species individually to
427 account for expert advice. Variables were split into land use variables, variables considered
428 informative in species distribution modelling as per experts and published literature, and all
429 other remaining variables (Table S1.3). Our hierarchical testing method involved running
430 three multicollinearity tests per species. First, we evaluated multicollinearity within just the
431 land use variables set, then those passing the first stage combined with the expert set
432 (retaining all variables that passed the preceding test), and then a final evaluation using the
433 variables that passed the preceding test with all remaining variables (again retaining all
434 variables that passed the preceding test). All multicollinearity evaluations consisted of 10,000
435 random samples of the environmental variables within the species' mask using Pearson
436 correlation ($<|0.7|$) and Variance Inflation Factor (<10) tests using the usdm R package⁴⁰
437 (version 2.1-7) and removing all variables that failed one or both tests.

438

439 ***Species distribution models***

440 We used a Poisson point process-based approach to fit species distribution models to species
441 with ≥ 20 presence records after data cleaning (592 birds, 254 mammals and 642 reptiles,
442 1488 total) using the *glmnet* R package⁴¹ (version 4.1-8). We fit Infinitely Weighted Logistic
443 Regression models with a logit link function and penalised maximum likelihood, which is
444 equivalent to an inhomogeneous Poisson point process⁴². In this approach the data points are
445 weighted such that a presence point carries a weight of 1 and the weight of a background

446 point is equal to the ratio of presences to background points. We also utilise lasso
447 regularisation to penalise complex model structures and reduce the risk of overfitting the
448 data. Model formulae were constrained by the number of data points available for a species:
449 20-89 presence records used linear terms only and could select up to one variable per five
450 presence records and ≥ 90 presence records used linear and quadratic terms and could select
451 one variable per ten presence records (as each variable has two model terms). Variables were
452 randomly selected up to this limit from the variables that passed the multicollinearity tests
453 for each species in order from: 1) random selection from land use set variables up to a
454 maximum of 50% of the total allowed variables; 2) random selection from the expert set
455 variables; and 3) random selection from the remaining variables. Distributions were modelled
456 and predicted at 250m resolution and predictions were made for the current time period as
457 well as the 27 permutations of the three time periods, three SSP-RCP scenarios, and three
458 GCMs.

459

460 ***Model cross validation***

461 We used “p-thinning” to implement cross-validation for model evaluation of point-process
462 based SDMs⁴³. We implement “five-fold cross-validation” using five simulations of “p-
463 thinning” to randomly generate 75:25 splits of training and testing data. Model performance
464 of the underlying models is evaluated using the Boyce Index (BI) representing deviation of
465 predictions from randomness such that positive values away from zero indicate a good model.
466 Corresponding AUC values are also reported, with values closer to 1 indicating a good model,
467 however these are less informative than the Boyce Index for presence-only based
468 predictions⁴⁴.

469

470 ***Estimating biodiversity response***

471 Predictions under each coupled land use and climate scenarios generated 27 outputs per
472 species at the three time points (2030, 2050, 2070) under the three SSPs and for the three
473 GCM per SSP. For each species-scenario-time period combination, we estimated the mean,
474 minimum and maximum predicted occurrence per pixel across the three GCMs. Then, to
475 summarise across species, we calculated and mapped the relative likelihood of occurrence of
476 a species within each 250m pixel. Mean relative likelihood of occurrence was calculated for
477 all species, each taxonomic group (birds, mammals, reptiles) and for EPBC-listed species.

478

479 ***Model uncertainty***

480 We examine the influence of uncertainty in the climate scenarios on our SDM outputs. We
481 compare the mean predicted occurrence (relative likelihood of occurrence) for all species
482 averaged across the three GCMs to the minimum and maximum predicted values across the
483 three GCM), and to the predicted values obtained under each individual GCM. This provides
484 us with a range of biodiversity response maps under alternate climate scenarios.

485

486 Further details on methods can be found in the supplementary information.

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Data availability

Accompanying scripts in R^{45,46} (versions 4.4.0 and 4.4.1) and python (version 3.10.12) and the sequence in which they are implemented can be accessed at <pending-zenodo-link>. All data used in the study is available at <pending-figshare-link>.

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Author contributions

Conceptualisation: PB, BW. Methodology: PB, DW. Formal analysis: PB, DW, DAR, WLG
Computation: PB, DW, DAR, RV, UN. Writing – original draft: PB, DW. Writing – review and editing: all.

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