

1 **Title:** Contribution and applications of demographic concepts to conservation
2 Le Coeur C., Perret J., Besnard A., Cubaynes S., Monchy C., Naciri M., Nakamura M.,
3 Gimenez O.

4 **Abstract**
5 Studying the demographic processes that shape how populations respond to environmental
6 changes has long provided insights for conservation biology. Recent theoretical advances
7 have deepened our understanding of these processes, yet their application in conservation
8 remains unclear. We conducted a literature search to examine how six key demographic
9 concepts — life-history trade-offs, the fast–slow continuum, temporal covariation among
10 demographic parameters, demographic buffering and lability, individual heterogeneity and
11 transient dynamics — have been used in conservation, and discussed their potential benefits
12 and limitations.
13 Their applications fall into three main categories: improving estimates of demographic
14 parameters, population dynamics, and extinction risk; predicting the magnitude and duration
15 of population responses to disturbances or conservation actions; and identifying the
16 demographic processes most relevant for guiding conservation decisions. Individual
17 heterogeneity and the fast–slow continuum were widely used, likely due to their low data and
18 analytical requirements, allowing broad predictions of species' vulnerability and informing
19 conservation decisions. Trade-offs explained how populations adapt to anthropogenic
20 disturbances, invasions or conservation actions. Conversely, temporal covariation and
21 buffering–lability were rarely applied, despite their value for improving projections and
22 assessing populations' capacity to cope with environmental variability. Limited use reflects
23 data and modelling needs, and, for temporal covariation, lack of direct conservation guidance.
24 Transient dynamics, highlighting short-term responses and demographic resilience, are
25 relevant because they match the timescale of many conservation projects.

26 We argue that even modest monitoring efforts can capture essential demographic processes,
27 and that their systematic integration, directly or via inference from related systems, could
28 strengthen long-term conservation outcomes.

29 **Keywords:** conservation action, environmental perturbations, extinction risk, life history,
30 population dynamics, vital rates

31

32 **1. Introduction**

33 Conservation biology is often described as a crisis discipline focused on preventing
34 biodiversity loss and supporting urgent, evidence-based management decisions (Kareiva &
35 Marvier 2012; Primack 2008; Soulé 1985). Its core objectives include avoiding species
36 extinctions, maintaining population viability, and preserving ecological functions. Throughout
37 this paper, we use the term “conservation” primarily in the sense of population-based
38 conservation, i.e., approaches that rely on demographic data and models to understand
39 population dynamics and inform management decisions. While conservation decisions often
40 occur under uncertainty and limited knowledge, a deeper understanding of the mechanisms
41 driving population dynamics is essential to mitigating risks and achieving successful
42 outcomes. In this context, demography provides essential insights into the structure and
43 functioning of populations (Caswell 2001; Lande 1988) and offers a rich set of quantitative
44 tools to inform conservation decisions (Speakman *et al.* 2025).

45 Demography investigates how individuals contribute to population growth, decline, or
46 stability by analysing the parameters and life-cycle transitions that drive population dynamics.
47 In a conservation context, it provides a framework for understanding which processes—such
48 as reduced survival, lower reproductive output, or limited recruitment—are most responsible
49 for population change and, therefore, where management actions can be most effective. Over
50 the past two decades, tools such as matrix population models (Caswell 2001), integral
51 projection models (Ellner *et al.* 2016), individual-based models (Grimm & Railsback 2005),
52 and integrated population models (Schaub & Kéry 2021) have gained increasing use in applied

53 conservation (e.g., Heinrichs *et al.* 2023; Messerman *et al.* 2023). These tools are particularly
54 useful for projecting population trajectories and evaluating extinction risks under different
55 management scenarios, providing a quantitative basis for decision-making.

56 In parallel, theoretical advances in demography have deepened our understanding of
57 population trajectories and their underlying mechanisms. Among these, demographic
58 concepts such as demographic buffering and lability (e.g., Gascoigne *et al.* 2025), temporal
59 correlations among demographic parameters (e.g., Fay *et al.* 2022b), individual heterogeneity
60 (e.g., Hamel *et al.* 2018a), life-history trade-offs (e.g., Bliard *et al.* 2025), the fast-slow
61 continuum (e.g., Stott *et al.* 2024), and transient dynamics (e.g., Hinrichsen 2025) have
62 received growing attention in theoretical studies. Despite their potential to improve population
63 projections and risk assessments and to inform adaptive management (Buhnerkempe *et al.*
64 2011; Gerber & Kendall 2016), these concepts appear underused in applied conservation, as
65 reflected by persistent demographic data and knowledge gaps across many taxa of
66 conservation concern (Conde *et al.* 2019; Paniw *et al.* 2021). Calls have also been made to
67 better integrate demography with other disciplines such as evolution (Metcalf & Pavard 2007),
68 population genetics and genomics (Lowe *et al.* 2017), climate change ecology (Paniw *et al.*
69 2021), and functional ecology (Salguero-Gómez *et al.* 2018) to foster theoretical developments
70 and practical applications. We argue that closer integration between conservation biology and
71 demography could provide a promising avenue to translate recent theoretical advances in
72 demography into practical applications for biodiversity conservation.

73 Broadly, demography can inform conservation through two complementary
74 approaches: (1) a comparative approach that positions species and populations along general
75 axes of life-history variation (e.g., speed of life), and (2) a mechanistic, system-specific
76 perspective that models the processes driving population dynamics under real-world
77 constraints (e.g., environmental forcing, small population size, isolation, or ongoing decline).
78 In this framework, the fast-slow continuum and classic life-history trade-offs primarily underpin
79 the comparative approach, whereas concepts such as demographic buffering and lability,

80 temporal correlations, individual heterogeneity, and transient dynamics are useful in the
81 mechanistic approach. Importantly, life-history trade-offs are central to both perspectives: they
82 define the evolutionary constraints that generate broad axes of variation across species (Healy
83 *et al.* 2019), and they also determine how individuals allocate limited resources when facing
84 ecological stressors, thereby shaping demographic responses of populations (Kentie *et al.*
85 2020).

86 Here, we examine how six key demographic concepts, listed above, can contribute to
87 conservation biology. Using examples from the literature and from our own work, we (i) review
88 how these concepts have been applied in conservation; and (ii) assess their potential benefits
89 for conservation practice while discussing limitations that may constrain their broader
90 application, including data requirements, modelling complexity, and disciplinary boundaries.
91 With this approach, our study aims to foster closer integration between demographic theory
92 and conservation practice, and highlight promising directions for future research and
93 application.

94 **2. Literature search**

95 We examined six demographic concepts: (i) life-history trade-offs, ii) the fast–slow continuum,
96 iii) temporal covariation among demographic parameters, (iv) demographic buffering and
97 lability, (v) individual heterogeneity, and (iv) transient dynamics (Fig. 1). We selected them for
98 their central role in recent demography and their potential to shape population responses to
99 environmental changes and management interventions. To identify the primary conservation
100 applications associated with each concept, we conducted a literature search using the Web of
101 Science Core Collection. We restricted this search to a predefined set of conservation journals,
102 including those ranked in 2022 according to Bradshaw & Brook's journal-ranking method
103 (2016), and two additional journals publishing applied conservation studies (full list in Appendix
104 S1). For each concept, a set of keywords was selected based on their definition (main
105 keywords in Table 1; full list in Appendices S2-S7, S9), meaning that our search mainly
106 identified studies explicitly applying these concepts. Articles selected according to the

107 screening criteria (Appendix S9) were assigned to application categories based on the
108 purpose for which the concept was used within the study context. Studies focusing on captive
109 populations and laboratory experiments were excluded.

110 **3. Applications of demographic concepts in conservation**

111 **3.1 Life-history trade-offs**

112 Individuals at the expanding edges of their range — whether in the context of biological
113 invasion or climate-driven range shifts — often exhibit higher reproductive and/or dispersal
114 abilities than conspecifics in core populations (Chuang & Peterson 2016). However, greater
115 energy investment in reproduction and/or dispersal often comes at the expense of other
116 functions, such as survival. Such negative correlations between life-history traits are known
117 as life-history trade-offs (LHTOs). Most often, they arise because organisms must allocate the
118 limited amount of energy they acquire to different functions (Stearns 1992; see Table 1 for
119 other mechanisms). Investing more in one trait inevitably comes at the expense of another,
120 and natural selection should favour the allocation strategy that maximizes fitness. LHTOs have
121 been widely studied to explore variation in life-history strategies at multiple levels, ranging
122 from individual-level processes to interspecific patterns. Spatial or temporal variation in
123 environmental conditions, such as those experienced during range expansion and biological
124 invasions, can induce shifts in optimal allocation strategies, altering the observed correlation
125 among life-history traits. Investigating these shifts within and across populations helps clarify
126 the ecological and evolutionary mechanisms by which environmental changes affect
127 populations. This knowledge can inform effective conservation and management measures,
128 such as preventing the expansion of introduced species.

129 We identified three primary applications of LHTO in conservation (Table 2). Firstly, it
130 has been used to understand and predict plastic or micro-evolutionary responses of
131 populations to anthropogenic pressures, especially climate change and harvest (Application
132 1, n=14 articles). For instance, LHTOs between growth, survival, and reproduction were
133 accounted for by Holt & Jørgensen (2014) to better predict life-history adaptations of Atlantic

134 cod (*Gadus morhua*) in response to warming temperatures. Secondly, this concept provides
135 a framework to understand the ability of introduced species to invade their new environment
136 (Application 2, n=4). For instance, plants introduced into new habitats and released from their
137 co-evolved herbivores tend to reallocate energy from defense to reproduction (Rotter &
138 Holeski 2018). Lastly, the concept has been used to assess the demographic consequences
139 of management, and to inform future actions, particularly restoration efforts (Application 3,
140 n=11). For example, the effectiveness of coral transplantation is influenced by how species
141 resolve the survival–growth trade-off (Montero-Serra *et al.* 2018). Most applications focused
142 on comparisons between populations (n=12) or on interspecific patterns (n=8). Fewer studies
143 examined individual variation (n=5), or temporal (n=2) and environmental shifts (n=3) in LHTO
144 within populations.

145 Because LHTOs arise from constraints — most notably energetic constraints — that
146 can be modulated by anthropogenic pressures and interventions, and because LHTOs
147 themselves condition population and species responses to these changes, considering them
148 can be key to understanding and predicting demographic responses to environmental changes
149 or to conservation and management actions. Given their central role in the evolution of life-
150 history strategies, LHTOs also help clarify how species adapt to rapidly changing
151 environments (e.g., Wang *et al.* 2017). Their strong theoretical basis and valuable insights
152 make them particularly relevant for broader application in conservation.

153 **3.2 Fast-slow continuum of life histories**

154 Can we predict a species' vulnerability to environmental disturbances based on its life-history
155 traits, and thereby anticipate both its extinction risk and that of other species with similar traits?
156 Answering this question is certainly central to the application of the fast-slow continuum in
157 conservation. This concept quantifies how species vary along a gradient of co-varying life-
158 history traits, shaped by ecological and evolutionary pressures, including life-history trade-offs
159 (Table 1). At the fast end of this continuum, species tend to mature early and have high
160 reproductive rates and short lifespans, while those at the slow end display the opposite traits.

161 Understanding the diversity of life-history strategies, along with the ecological drivers and
162 adaptive mechanisms that shape them, has been an important focus in evolution, ecology and
163 conservation (Ducatez & Shine 2019; Stott *et al.* 2024).

164 Across the 55 reviewed conservation studies, six main applications emerged (Table
165 2). These range from studies describing the life-history strategy of single species and
166 predicting their extinction risk or vulnerability to disturbances (Application 1, n=15; Waldron *et*
167 *al.* 2013); to methods-oriented applications, such as guiding data imputation for species with
168 incomplete life-history information; or accounting for variation in life histories when quantifying
169 population- or species-level trends (Applications 2, n=3; Horswill *et al.* 2019). The most
170 common application involves comparing populations' and species' responses, often quantified
171 as vulnerability or resilience to ongoing threats (e.g., land use and climate changes,
172 overfishing/exploitation), thereby inferring outcomes for other species along the fast-slow
173 continuum and informing conservation measures (Application 3, n=18; S3; Schindler *et al.*
174 2002). The continuum has also been used to assess whether a species' or population's
175 position along it can explain the effectiveness of conservation actions, such as translocation
176 success (Application 4, n=9; Ducatez & Shine 2019), as well as their vulnerability to extinction
177 and susceptibility to threats (Application 5, n=5; Koleček *et al.* 2014), and infer responses in
178 other species or populations.

179 While the concept was initially defined at the species level, it has also been applied at
180 both the community and within-species levels, and may be useful not only among populations
181 but also among individuals (Del Giudice 2020; but see Van De Walle *et al.* 2023). Comparison
182 of life-history strategies between populations was studied mainly to identify local adaptive
183 responses to disturbances (Appendix S3). At the community level, the composition and
184 diversity of life-history strategies serve as indicators of community health and functioning, and
185 can also reveal community shifts or successions in responses to environmental stressors (e.g.,
186 ocean warming, agricultural intensification; Application 6, n=5; Guerrero *et al.* 2024).

187 Overall, the fast-slow continuum has been widely used in conservation to help predict long-
188 term viability and guide conservation efforts. However, it only partially captures the full

189 spectrum of life-history variation, highlighting the need to consider additional axes, such as
190 developmental or reproductive patterns (Stott *et al.* 2024; Fig. 1A), and local ecological
191 processes to improve its predictive value in conservation. For some organisms, other
192 continuums may be more relevant (n=13), including the Equilibrium-Periodic-Opportunistic
193 continuum of life-history strategies for fish and bivalves (n=8; Table S3.4; Winemiller & Rose
194 1992).

195 **3.3 Temporal covariation among demographic parameters**

196 At the population level, demographic parameters such as growth and survival rarely vary
197 independently over time. Instead, they covary, and the strength and direction of these
198 covariances can lead to substantial changes in population dynamics and long-term growth
199 (Tuljapurkar 1982). Covariation is positive (negative) when two or more demographic
200 parameters in a population increase or decrease simultaneously (in opposite direction) over
201 time. Such covariation is shaped by environmental stochasticity, along with other processes
202 such as life-history trade-offs and density-dependence (Fay *et al.* 2022b). When positive, it
203 amplifies the benefits of favourable years (e.g., years with high food availability) when multiple
204 demographic parameters exceed their long-term mean, while also exacerbating the negative
205 effects experienced during unfavourable years. Strong positive covariance among
206 demographic rates amplifies population fluctuations and, in some cases, reduces long-term
207 growth, thereby increasing extinction risk. By contrast, negative covariance buffers population
208 responses against environmental and demographic variability. Ignoring temporal (co)variation
209 among demographic parameters in population models can hinder the identification of
210 parameters that most influence long-term population growth, and may result in less reliable
211 estimates of extinction risk (Earl 2019).

212 Four conservation studies investigated temporal covariation to quantify population
213 dynamics more accurately in response to environmental variability and, in turn, improve
214 extinction risk estimates (Application 1, n=4, Table 2; Doak *et al.* 1994). This allows for more
215 effective conservation planning and reliable assessment of the effectiveness of conservation

216 and management actions (Application 2, n=2; Johnson *et al.* 2010). Temporal covariation has
217 also been explored to understand how populations may buffer the effects of climate change
218 (Kissel *et al.* 2019) or incidental take (McGowan *et al.* 2011) through density-dependent
219 processes. This phenomenon, known as demographic compensation, arises when negative
220 effects on certain demographic parameters (e.g., reduced survival of large fishes due to
221 harvesting) are offset by density-dependent increases in other parameters (e.g., increased
222 survival of smaller individuals; Application 3, n=3). Identifying demographic compensation in
223 a population can help refine management interventions (e.g., fishing quotas, management
224 strategies for invasive species).

225 Despite its importance for accurately predicting population trajectories, relatively few
226 conservation studies have explicitly investigated temporal covariation (9 studies, 2 with only
227 brief mention; Table S4.1). Some studies, however, may have empirically accounted for it
228 when modelling population dynamics over several years (e.g., Nakaoka 1996). This highlights
229 the importance of multi-year population monitoring to accurately capture temporal
230 environmental variation and demographic covariation, which contribute to more effective
231 conservation decisions.

232 **3.4 Demographic buffering-lability**

233 In populations of slow-living species, adult survival rates are generally high and show little
234 annual variations. Because population growth in these species is particularly sensitive to even
235 small changes in adult survival, this rate is expected to be buffered against environmental
236 variation by natural selection (Hilde *et al.* 2020; Table 1). This process, known as demographic
237 buffering, corresponds to the low temporal variation of some demographic parameters (at the
238 population level) in response to environmental variability, while others fluctuate more widely,
239 a process known as demographic lability (Fig. 1C). Lability and buffering are adaptive when
240 they confer positive effects on long-term population growth rate (Le Coeur *et al.* 2022).
241 According to the ‘demographic buffering hypothesis’, selection should favour reduced variation
242 in the demographic parameters with the strongest influence on population growth, such as

243 adult survival in slow-living species. Understanding the demographic buffering-lability
244 strategies in a population provides insights into its overall capacity to cope with environmental
245 variability and short-term perturbations. This is particularly relevant for conservation, as it
246 helps anticipate population responses to perturbations, and assess whether a population's
247 buffering-lability capacity is maintained or challenged when environmental conditions deviate
248 from their typical range of variability (e.g., increased environmental variability or more frequent
249 extreme events associated with climate change). Insights into these dynamics are critical for
250 assessing and mitigating extinction risk in species of conservation concern.

251 In this context, four conservation studies investigated whether the demographic buffering
252 capacity of a population was maintained or challenged under increases in climate variability
253 and local extreme events in mammals and seabirds (e.g., Forcada *et al.* 2008), or following a
254 major human-induced shift in food availability in a population of Eurasian vultures (Almaraz *et*
255 *al.* 2022; Application 1). Another type of application involves assessing species' responses to
256 conservation or management actions and their effectiveness in maintaining or restoring
257 populations' demographic buffering capacity, for example in restoration projects or harvest
258 management plans (Application 2, n=2). Size-selective harvesting, for instance, can modify
259 populations' buffering-lability strategies over the long term. It can shift populations toward size
260 classes that are more sensitive to environmental variability, thereby increasing the overall
261 population vulnerability (Gamelon *et al.* 2019). Such effects can be mitigated through
262 appropriate management actions (Goto 2023). To date, conservation studies have not
263 specifically addressed adaptive lability. Investigating both buffering and lability provides a
264 framework to identify species likely to be vulnerable, as well as those potentially resilient to
265 increased environmental variability under climate change — a perspective that remains largely
266 unexplored empirically.

267 **3.5 Individual heterogeneity**

268 Individuals often respond differently to anthropogenic disturbances or relocation to unfamiliar
269 environments. These differences are frequently linked to intrinsic characteristics – such as

270 age, sex, physiological condition, or personality – that influence individual fitness. For
271 instance, bolder individuals may fare worse in human-dominated landscapes due to greater
272 exposure to disturbances, provided that shyer individuals can access suitable refuges
273 (Assandri *et al.* 2017). Such among-individual differences, whether associated with fitness-
274 related traits (Fay *et al.* 2022a) or not (Hamel *et al.* 2018b; Table 1), are referred to as
275 individual heterogeneity. Individual heterogeneity can scale up to shape population-level
276 patterns in life-history traits and fitness outcomes, with implications that extend to higher
277 ecological levels. Recognizing and studying this variation can reveal traits associated with
278 success under specific conditions, thereby informing more effective conservation and
279 management strategies at broader ecological scales. Although typically implemented at the
280 population, species, or community level, incorporating among-individual variation could
281 significantly enhance their effectiveness (Jolles *et al.* 2020).

282 Conservation applications relying on this concept include: assessing behavioural traits
283 with high individual variation that may influence conservation and management outcomes, in
284 order to inform and refine strategies (Application 1, n=30; Moseby *et al.* 2023); reducing bias
285 in demographic parameter estimates, thereby improving inference on population dynamics
286 (Application 2, n=19; Halstead *et al.* 2012); evaluating stressor factors on specific individuals
287 that can impact population structure and persistence, and potentially affect ecosystem-level
288 processes (Application 3, n=13; Milenkaya *et al.* 2013); identifying key individuals to support
289 effective population management or invasive species control (Application 4, n=20; Lopez *et*
290 *al.* 2012); and assessing the individual traits that affect species' or population's reproductive
291 success to improve conservation and management outcomes (Application 5, n=5; Hamel *et*
292 *al.* 2012). Most studies accounted for non-fitness (n=54) and fitness (n=30) traits, and a few
293 for both traits (n=3; Table S6.2). While early studies focused on observed (or measured) traits
294 like age or sex (Milenkaya *et al.* 2013), there has been increasing emphasis on quantifying
295 unobserved (latent) factors that may be attributed to individual behaviour, morphology, life-
296 history or physiology (Exposito-Granados *et al.* 2020; Table S6.4).

297 Selecting individuals can improve ecosystem management and conservation
298 outcomes. For instance, the success of translocation and reintroduction programs depends
299 on pre- or post-release differences in individual behaviour and personality, which influence
300 survival (West *et al.* 2019). Similarly, invasive species control can benefit from targeting
301 individuals with specific traits (e.g., more prone to disperse), including recalcitrant individuals
302 under adaptive control strategies (Johnstone *et al.* 2024). However, it is crucial to evaluate
303 how practitioner actions (e.g., agricultural, management) shape population composition and
304 intraspecific diversity to avoid counterproductive phenotypic selection (Mensinger *et al.* 2021).
305 Individual heterogeneity is likewise central in human–wildlife conflict mitigation, where
306 identifying conflict-prone individuals—such as damage-making brown bears (Berezowska-
307 Cnota *et al.* 2023)—enables more targeted and effective interventions.

308 Future studies should account for individual heterogeneity to reduce bias in population
309 dynamics estimates (Cubaynes *et al.* 2010), although data limitations remain a major
310 constraint in conservation and demographic studies (Conde *et al.* 2019). Evidence from the
311 literature suggests that considering among-individual traits in conservation planning has been
312 useful to directly guide actions. Further research into neurobiological, genetic, and disease-
313 related factors that shape these traits will deepen our understanding and guide more evidence-
314 based conservation (Firth *et al.* 2018; Gamble *et al.* 2020).

315 **3.6 Transient dynamics**

316 At equilibrium, a population reaches a stable age or stage structure, meaning the proportion
317 of individuals in each (st)age and the population's (asymptotic) growth rate are constant over
318 time. In protected areas like no-take marine reserves, fish populations are expected to tend
319 towards this equilibrium, and the factors favouring long-term growth and stability are well
320 documented (White *et al.* 2013). However, over the short term, i.e. the first years or decades
321 following reserve establishment, fish populations can undergo surprising dynamics: population
322 abundance can remain stable, decline or oscillate periodically, regardless of the long-term
323 outcome. Such patterns can all stem from the same underlying process: transient population

324 dynamics (Table 1, Fig. 1E). It occurs before a population reaches its stable distribution, a
325 condition that may never be achieved in a disturbed environment, or when it is pushed away
326 from it (Capdevila *et al.* 2020b; Stott *et al.* 2011). Deviations from the stable distribution are
327 caused by disturbances or human interventions that differentially affect some life stages, for
328 example when the establishment of a marine reserve enhances survival of mature individuals,
329 or conversely, when fishing pressure preferentially targets them over immature stages
330 (Anderson *et al.* 2008). Depending on which life stages are affected (e.g., mature individuals)
331 and how their demographic parameters changed (e.g., increased or decreased survival and/or
332 fecundity), the perturbation will differently influence transient dynamics, causing oscillations in
333 population size and structure of varying duration and intensity until the population reaches its
334 asymptotic state. Understanding the consequences of these deviations and populations'
335 demographic resilience to perturbations lies at the core of conservation biology.

336 Transient dynamics is of particular importance in conservation because: 1/ a species'
337 potential for transient dynamics can guide more efficient conservation or management actions
338 by preferentially targeting specific classes of individuals ; 2/ small populations undergoing a
339 transient dynamics phase may face high risk of extinction, as oscillations in abundance can
340 periodically bring population size near the quasi-extinction threshold (Table 2; Ezard *et al.*
341 2010). The concept of transient dynamics has been applied in conservation to better estimate
342 the extinction risk and key demographic parameters of populations of threatened species,
343 sometimes in the context of population reintroduction or reinforcement programmes
344 (Application 1, n=17; Gaoue 2016; Wong & Ticktin 2015). It has been also used to identify the
345 best sustainable exploitation strategy (for practical examples, see Buhnerkempe *et al.* 2011;
346 Goto 2023) or the best management strategy for invasive species (Application 2, n=18; Miller
347 & Tenhumberg 2010). In these contexts, populations are likely to deviate from their expected
348 stable distribution, and studying the resulting transient dynamics helps quantify their
349 demographic resilience to disturbance or management actions (Capdevila *et al.* 2020b). In a
350 broader sense, demographic imbalances (e.g., biased sex ratios or skewed age/stage
351 structure) are already commonly addressed in the literature on small populations and

352 conservation translocations, although they are not always explicitly identified as transient
353 dynamics. We believe transient dynamics and demographic resilience warrant greater
354 attention in conservation, as transient effects occur on timescales more realistic and relevant
355 to many conservation projects than asymptotic dynamics (Ezard *et al.* 2010). The main
356 challenge is that it requires (st)age-specific demographic and abundance data, and therefore
357 involves monitoring many individuals over several years. This can be difficult for rare or elusive
358 species (Couturier *et al.* 2013), or may require more time than allowed by management
359 decisions.

360 **Discussion**

361 Conservation biology aims to protect and maintain biodiversity by supporting evidence-based
362 practice and decision-making (Primack 2008; Soulé 1985). Our literature search highlights
363 how demographic concepts can contribute to this fundamental goal, primarily by addressing
364 three objectives: 1) providing accurate estimates of demographic parameters and population
365 dynamics, thereby enabling reliable assessment of population trends and extinction risks; 2)
366 predicting the strength, nature, and duration of population and species responses to
367 disturbances, whether negative or positive; and 3) understanding the demographic and
368 ecological processes most relevant for guiding or refining conservation decisions. Several
369 concepts have been applied beyond their original ecological scale (e.g., fast–slow continuum
370 at intraspecific and community levels; individual heterogeneity up to ecosystem implications).
371 This scaling-up of demographic concepts aligns with broader trends in conservation biology
372 over the past decades, which increasingly embrace integrative and cross-scale perspectives
373 (Kareiva & Marvier 2012; Mace 2014).

374 The six demographic concepts examined range from long-established to more recently
375 developed, and from more theoretical concepts aimed at understanding the mechanisms
376 driving changes (e.g., LHTO), to those with *a priori* a more practical applicability for
377 conservation (e.g., transient dynamics). They also differ in maturity, typical scale of

378 application, and practical cost (in terms of data requirements, modelling effort, and
379 interpretability), which makes direct comparisons across concepts inherently imperfect.
380 Among the six concepts, individual heterogeneity and the fast-slow continuum, two long-
381 established concepts, were by far the most studied. The strong emphasis on individual
382 heterogeneity (87 studies) likely reflects that it directly informs conservation decisions,
383 encompasses diverse traits, and can be analysed with only minimal theoretical requirements,
384 using methods that range from relatively simple multivariate models to more complex models
385 with latent states (Hamel *et al.* 2018a). For the fast-slow continuum (50 studies), its
386 widespread use likely reflects modest data needs and modelling complexity, enabling rapid
387 comparisons among species and generating broad expectations of vulnerability and
388 responses. Its simplicity allows quick assessment and integration into conservation planning;
389 though its limited resolution yields only general expectations (Fig. 2). Many applications of the
390 LHTO have also been reported (31 studies). As a fundamental concept in ecology and
391 evolution, the LHTO informs conservation by providing predictive insight regarding the
392 adaptive responses of populations to anthropogenic disturbances, invasions or management
393 actions. Nevertheless, several studies refer to demographic concepts only as a way to
394 interpret potential mechanisms, rather than formally incorporating them in analyses. Given
395 their relevance for management, we encourage their more systematic use.

396 By contrast, concepts of temporal covariation among demographic parameters and
397 buffering-lability have received limited attention (< seven articles). While temporal covariation
398 is crucial for understanding mechanisms and obtaining reliable estimates that can inform
399 conservation decisions, it does not provide direct conservation guidance *per se*, which may
400 explain its lower prevalence in conservation studies. Buffering-lability, on the other hand, is
401 still benefiting from ongoing developments and can be technically difficult to quantify,
402 especially for lability. Both concepts rely on detailed, long-term demographic data, which may
403 also constrain their current application in conservation (Fig. 2).

404 Two concepts that are increasingly discussed in the literature, buffering-lability and
405 transient dynamics have been used in conservation articles to characterise a population's
406 capacity to be buffered against, or recover from perturbations (past or future), respectively, as
407 well as the demographic processes underlying these responses. Transient processes are
408 critical because populations and communities targeted by conservation are often small and
409 subject to stochastic or persistent disturbances (e.g., climate change), making stable stage
410 equilibrium unlikely (Ezard *et al.* 2010). While these two concepts provide complementary
411 insights, their application requirements and the nature of their outputs differ, which may explain
412 differences in use (35 studies for transient dynamics, six for buffering-lability). Medium-term
413 monitoring is necessary to detect a disturbance and to quantify its transient effects on
414 population dynamics. This provides disturbance-specific insights that can guide or refine
415 conservation actions. In contrast, quantifying buffering-lability requires long-term monitoring
416 and provides general predictions about how a population may respond to future disturbances.

417 Conservation practice faces multiple constraints, including limited resources, urgent
418 timelines, and the need to act under high uncertainty and within complex socio-political
419 contexts (Sabo *et al.* 2024). Biological and methodological challenges add further obstacles,
420 such as working with small, vulnerable, or cryptic populations, or the difficulty of monitoring
421 large numbers of individuals across extended periods and spatial scales (White 2019).
422 Conservation must navigate these constraints by balancing data collection with information
423 gain to support effective planning and long-term outcomes, from anticipation and decision-
424 making to evaluating success (Watts *et al.* 2020). From a demographic perspective,
425 understanding the mechanisms underlying a population's or species' responses to
426 disturbances or conservation actions depends on repeated measurements of a sufficiently
427 large number of individuals, which may not always be compatible with conservation
428 constraints. As a result, some concepts remain underused despite their clear predictive value
429 and potential to inform management (Fig. 2). Importantly, however, "underuse" does not

430 necessarily imply suboptimal practice: in some applied contexts, simpler demographic
431 summaries or models may already provide sufficient information for robust decisions.

432 Nevertheless, we believe that with a moderate monitoring effort of 4-6 years, many
433 metrics associated with demographic concepts can be quantified with reasonable reliability
434 across a wide range of life histories (e.g., individual heterogeneity, transient dynamics, LHTO).
435 While others require longer time series to be estimated accurately (e.g., temporal covariance,
436 demographic buffering-lability), they can be at least implicitly accounted for even if they are
437 not the primary focus. For instance, population models built from multi-year monitoring data
438 (such as matrix population models) inherently capture covariation among demographic
439 parameters within each year and the observed variation between years, although the
440 interannual correlations are not explicitly estimated. Producing more robust information on
441 demographic and ecological processes strengthens evidence-based planning and decisions,
442 and reduces the risk of obtaining misleading ecological predictions. In that respect, delaying
443 immediate action when possible may improve conservation strategies and their outcomes
444 (Iacona *et al.* 2017). When replication across individuals or time is limited, we encourage
445 practitioners to familiarize themselves with these concepts, as this knowledge can guide the
446 design and implementation of effective conservation actions. For example, knowing that some
447 behavioural traits can influence reintroduction success allows managers to optimize selection
448 strategies for release (e.g., West *et al.*, 2019). Other important concepts not included in this
449 search, such as demographic senescence, the Allee effect or carry-over effects (e.g., Robert
450 *et al.* 2015; Stephens & Sutherland 1999; Sutton *et al.* 2021), could further support
451 conservation goals. We therefore recommend reviewing demographic studies, including those
452 on the target species and on species with similar life histories, to better anticipate the range
453 of possible demographic processes relevant to each conservation project. General
454 demographic knowledge and concepts can be integrated into population projection models
455 even when data for the target population are incomplete, helping with parameter imputation
456 and the inclusion of known mechanisms into predictions.

457 Overall, conservation biology and demography have been closely interconnected for
458 decades, and our review shows that demographic knowledge and tools still contribute
459 substantially to conservation goals. Both disciplines focus on long-term population stability
460 and viability, which are central to biodiversity conservation across ecological levels. Even
461 under constraints of urgency and uncertainty, decisions should be guided, wherever possible,
462 by ecological, demographic, and evolutionary processes. Doing so will improve the robustness
463 and effectiveness of conservation strategies and help preserve biodiversity into the future.

464 **References**

465 Almaraz, P., Martínez, F., Morales-Reyes, Z., Sánchez-Zapata, J.A. & Blanco, G. (2022).
466 Long-term demographic dynamics of a keystone scavenger disrupted by human-
467 induced shifts in food availability. *Ecological Applications*, 32, e2579.

468 Anderson, C.N., Hsieh, C., Sandin, S.A., Hewitt, R., Hollowed, A., Beddington, J., et al. (2008).
469 Why fishing magnifies fluctuations in fish abundance. *Nature*, 452, 835–839.

470 Assandri, G., Giacomazzo, M., Brambilla, M., Griggio, M. & Pedrini, P. (2017). Nest density,
471 nest-site selection, and breeding success of birds in vineyards: management
472 implications for conservation in a highly intensive farming system. *Biological
473 Conservation*, 205, 23–33.

474 Beissinger, S.R. & McCullough, D.R. (2002). *Population Viability Analysis*. University of
475 Chicago Press.

476 Berezowska-Cnota, T., Konopiński, M.K., Bartoń, K., Bautista, C., Revilla, E., Naves, J., et al.
477 (2023). Individuality matters in human–wildlife conflicts: Patterns and fraction of
478 damage-making brown bears in the north-eastern Carpathians. *Journal of Applied
479 Ecology*, 60, 1127–1138.

480 Blaustein, L., Martin, J.S., Paniw, M., Blumstein, D.T., Martin, J.G.A., Pemberton, J.M., et al.
481 (2025). Detecting context dependence in the expression of life history trade-offs.
482 *Journal of Animal Ecology*, 94, 379–393.

483 Bradshaw, C.J. & Brook, B.W. (2016). How to rank journals. *PLoS one*, 11, e0149852.

484 Buhnerkempe, M.G., Burch, N., Hamilton, S., Byrne, K.M., Childers, E., Holfelder, K.A., et al.
485 (2011). The utility of transient sensitivity for wildlife management and conservation:
486 bison as a case study. *Biological Conservation*, 144, 1808–1815.

487 Cam, E., Gimenez, O., Alpizar-Jara, R., Aubry, L.M., Authier, M., Cooch, E.G., et al. (2013).
488 Looking for a needle in a haystack: inference about individual fitness components in a
489 heterogeneous population. *Oikos*, 122, 739–753.

490 Capdevila, P., Beger, M., Blomberg, S.P., Hereu, B., Linares, C. & Salguero-Gómez, R.
491 (2020a). Longevity, body dimension and reproductive mode drive differences in
492 aquatic versus terrestrial life-history strategies. *Functional Ecology*, 34, 1613–1625.

493 Capdevila, P., Stott, I., Beger, M. & Salguero-Gómez, R. (2020b). Towards a comparative
494 framework of demographic resilience. *Trends in Ecology & Evolution*, 35, 776–786.

495 Caswell, H. (2001). Matrix population models: construction, analysis, and interpretation. In:
496 *Matrix population models: construction, analysis, and interpretation*. pp. 722–722.

497 Chuang, A. & Peterson, C.R. (2016). Expanding population edges: theories, traits, and trade-
498 offs. *Global Change Biology*, 22, 494–512.

499 Conde, D.A., Staerk, J., Colchero, F., Da Silva, R., Schöley, J., Baden, H.M., et al. (2019).
500 Data gaps and opportunities for comparative and conservation biology. *Proc. Natl.
501 Acad. Sci. U.S.A.*, 116, 9658–9664.

502 Couturier, T., Cheylan, M., Bertolero, A., Astruc, G. & Besnard, A. (2013). Estimating
503 abundance and population trends when detection is low and highly variable: A
504 comparison of three methods for the Hermann's tortoise. *J Wildl Manag*, 77, 454–462.

505 Cubaynes, S., Pradel, R., Choquet, R., Duchamp, C., Gaillard, J.-M., Lebreton, J.-D., et al.
506 (2010). Importance of accounting for detection heterogeneity when estimating
507 abundance: the case of French wolves. *Conservation Biology*, 24, 621–626.

508 Del Giudice, M. (2020). Rethinking the fast-slow continuum of individual differences. *Evolution
509 and Human Behavior*, 41, 536–549.

510 Doak, D., Kareiva, P. & Klepetka, B. (1994). Modeling Population Viability for the Desert
511 Tortoise in the Western Mojave Desert. *Ecological Applications*, 4, 446–460.

512 Duceatz, S. & Shine, R. (2019). Life-history traits and the fate of translocated populations.
513 *Conservation Biology*, 33, 853–860.

514 Earl, J.E. (2019). Evaluating the assumptions of population projection models used for
515 conservation. *Biological conservation*, 237, 145–154.

516 Ellner, S.P., Childs, D.Z. & Rees, M. (2016). *Data-driven Modelling of Structured Populations: 517 A Practical Guide to the Integral Projection Model*. Lecture Notes on Mathematical 518 Modelling in the Life Sciences. Springer International Publishing, Cham.

519 Exposito-Granados, M., Parejo, D., Chastel, O. & Aviles, J.M. (2020). Physiological stress and 520 behavioural responses of European Rollers and Eurasian Scops Owls to human 521 disturbance differ in farming habitats in the south of Spain. *Bird Conservation 522 International*, 30, 220–235.

523 Ezard, T.H.G., Bullock, J.M., Dagleish, H.J., Millon, A., Pelletier, F., Ozgul, A., et al. (2010). 524 Matrix models for a changeable world: the importance of transient dynamics in 525 population management: Transient dynamics and population management. *Journal of 526 Applied Ecology*, 47, 515–523.

527 Fay, R., Authier, M., Hamel, S., Jenouvrier, S., Van De Pol, M., Cam, E., et al. (2022a). 528 Quantifying fixed individual heterogeneity in demographic parameters: Performance of 529 correlated random effects for Bernoulli variables. *Methods Ecol Evol*, 13, 91–104.

530 Fay, R., Hamel, S., van de Pol, M., Gaillard, J.M., Yoccoz, N.G., Acker, P., et al. (2022b). 531 Temporal correlations among demographic parameters are ubiquitous but highly 532 variable across species. *Ecology Letters*, 25, 1640–1654.

533 Firth, J.A., Cole, E.F., Ioannou, C.C., Quinn, J.L., Aplin, L.M., Culina, A., et al. (2018). 534 Personality shapes pair bonding in a wild bird social system. *Nature ecology & 535 evolution*, 2, 1696–1699.

536 Forcada, J., Trathan, P.N. & Murphy, E.J. (2008). Life history buffering in Antarctic mammals 537 and birds against changing patterns of climate and environmental variation. *Global 538 Change Biology*, 14, 2473–2488.

539 Gamble, A., Bazire, R., Delord, K., Barbraud, C., Jaeger, A., Gantelet, H., et al. (2020). 540 Predator and scavenger movements among and within endangered seabird colonies: 541 Opportunities for pathogen spread. *Journal of Applied Ecology*, 57, 367–378.

542 Gamelon, M., Sandercock, B.K. & Sæther, B. (2019). Does harvesting amplify environmentally 543 induced population fluctuations over time in marine and terrestrial species? *Journal of 544 Applied Ecology*, 56, 2186–2194.

545 Gaoue, O.G. (2016). Transient dynamics reveal the importance of early life survival to the 546 response of a tropical tree to harvest. *Journal of Applied Ecology*, 53, 112–119.

547 Garland, T., Downs, C.J. & Ives, A.R. (2022). Trade-Offs (and Constraints) in Organismal 548 Biology. *Physiological and Biochemical Zoology*, 95, 82–112.

549 Gascoigne, S.J.L., Kajin, M., Tuljapurkar, S., Santos, G.S., Compagnoni, A., Steiner, U.K., et 550 al. (2025). Structured Demographic Buffering: A Framework to Explore the 551 Environmental Components and Demographic Mechanisms Underlying Demographic 552 Buffering. *Ecology Letters*, 28, e70066.

553 Gerber, B.D. & Kendall, W.L. (2016). Considering transient population dynamics in the 554 conservation of slow life-history species: an application to the sandhill crane. *Biological 555 Conservation*, 200, 228–239.

556 Goto, D. (2023). Transient demographic dynamics of recovering fish populations shaped by 557 past climate variability, harvest, and management. *Global Change Biology*, 29, 6018– 558 6039.

559 Grimm, V. & Railsback, S.F. (2005). *Individual-based Modeling and Ecology*: Princeton 560 University Press.

561 Guerrero, I., Duque, D., Onate, J.J., Pärt, T., Bengtsson, J., Tscharntke, T., et al. (2024). 562 Agricultural intensification affects birds' trait diversity across Europe. *Basic and Applied 563 Ecology*, 74, 40–48.

564 Halstead, B.J., Wylie, G.D., Coates, P.S., Valcarcel, P. & Casazza, M.L. (2012). "Exciting 565 statistics": the rapid development and promising future of hierarchical models for 566 population ecology. *Animal Conservation*, 15.

567 Hamel, S., Craine, J.M. & Towne, E.G. (2012). Maternal allocation in bison: co-occurrence of 568 senescence, cost of reproduction, and individual quality. *Ecological Applications*, 22, 569 1628–1639.

570 Hamel, S., Gaillard, J., Douhard, M., Festa-Bianchet, M., Pelletier, F. & Yoccoz, N.G. (2018a).
571 Quantifying individual heterogeneity and its influence on life-history trajectories:
572 different methods for different questions and contexts. *Oikos*, 127, 687–704.

573 Hamel, S., Gaillard, J.-M. & Yoccoz, N.G. (2018b). Introduction to: Individual heterogeneity—
574 the causes and consequences of a fundamental biological process. *Oikos*, 127.

575 Hastings, A. (2004). Transients: the key to long-term ecological understanding? *Trends in*
576 *ecology & evolution*, 19, 39–45.

577 Healy, K., Ezard, T.H.G., Jones, O.R., Salguero-Gómez, R. & Buckley, Y.M. (2019). Animal
578 life history is shaped by the pace of life and the distribution of age-specific mortality
579 and reproduction. *Nat Ecol Evol*, 3, 1217–1224.

580 Heinrichs, J.A., Marcot, B.G., Linnell, M.A. & Lesmeister, D.B. (2023). Characterizing long-
581 term population conditions of the elusive red tree vole with dynamic individual-based
582 modeling. *Conservat Sci and Prac*, 5, e12938.

583 Hilde, C.H., Gamelon, M., Sæther, B.-E., Gaillard, J., Yoccoz, N.G. & Pélabon, C. (2020). The
584 Demographic Buffering Hypothesis: Evidence and Challenges. *Trends in Ecology &*
585 *Evolution*, 35, 523–538.

586 Hinrichsen, R.A. (2025). Solving three core challenges in transient dynamics analysis of matrix
587 population models. *Methods Ecol Evol*, 16, 2517–2530.

588 Holt, R.E. & Jørgensen, C. (2014). Climate warming causes life-history evolution in a model
589 for Atlantic cod (*Gadus morhua*). *Conservation Physiology*, 2, cou050.

590 Horswill, C., Kindvater, H.K., Juan-Jordá, M.J., Dulvy, N.K., Mangel, M. & Matthiopoulos, J.
591 (2019). Global reconstruction of life-history strategies: A case study using tunas.
592 *Journal of Applied Ecology*, 56, 855–865.

593 Iacona, G.D., Possingham, H.P. & Bode, M. (2017). Waiting can be an optimal conservation
594 strategy, even in a crisis discipline. *Proc. Natl. Acad. Sci. U.S.A.*, 114, 10497–10502.

595 Johnson, H.E., Mills, L.S., Stephenson, T.R. & Wehausen, J.D. (2010). Population-specific
596 vital rate contributions influence management of an endangered ungulate. *Ecological*
597 *Applications*, 20, 1753–1765.

598 Johnstone, K.C., Garvey, P. & Hickling, G.J. (2024). Invasive mammal control selects for trap-
599 recalcitrant behaviour and personality. *Biol Invasions*, 26, 549–564.

600 Jolles, J.W., King, A.J. & Killen, S.S. (2020). The role of individual heterogeneity in collective
601 animal behaviour. *Trends in ecology & evolution*, 35, 278–291.

602 Kareiva, P. & Marvier, M. (2012). What is conservation science? *BioScience*, 62, 962–969.

603 Kentie, R., Clegg, S.M., Tuljapurkar, S., Gaillard, J.M. & Coulson, T. (2020). Life-history
604 strategy varies with the strength of competition in a food-limited ungulate population.
605 *Ecology Letters*, 23, 811–820.

606 Kissel, A.M., Palen, W.J., Ryan, M.E. & Adams, M.J. (2019). Compounding effects of climate
607 change reduce population viability of a montane amphibian. *Ecological Applications*,
608 29, e01832.

609 Koleček, J., Albrecht, T. & Reif, J. (2014). Predictors of extinction risk of passerine birds in a
610 Central European country. *Animal Conservation*, 17, 498–506.

611 Koons, D.N., Pavard, S., Baudisch, A. & Jessica, C. (2009). Is life-history buffering or lability
612 adaptive in stochastic environments? *Oikos*, 118, 972–980.

613 Lande, R. (1988). Genetics and Demography in Biological Conservation. *Science*, 241, 1455–
614 1460.

615 Le Coeur, C., Yoccoz, N.G., Salguero-Gómez, R. & Vindenes, Y. (2022). Life history
616 adaptations to fluctuating environments: Combined effects of demographic buffering
617 and lability. *Ecology Letters*, 25, 2107–2119.

618 Lopez, D.P., Jungman, A.A. & Rehage, J.S. (2012). Nonnative African jewelfish are more fit
619 but not bolder at the invasion front: a trait comparison across an Everglades range
620 expansion. *Biological Invasions*, 14, 2159–2174.

621 Lowe, W.H., Kovach, R.P. & Allendorf, F.W. (2017). Population Genetics and Demography
622 Unite Ecology and Evolution. *Trends in Ecology & Evolution*, 32, 141–152.

623 Mace, G.M. (2014). Whose conservation? *Science*, 345, 1558–1560.

624 McGowan, C.P., Ryan, M.R., Runge, M.C., Millspaugh, J.J. & Cochrane, J.F. (2011). The role
625 of demographic compensation theory in incidental take assessments for endangered
626 species. *Biological Conservation*, 144, 730–737.

627 Mensinger, M.A., Brehm, A.M., Mortelliti, A., Blomberg, E.J. & Zydlewski, J.D. (2021).
628 American eel personality and body length influence passage success in an
629 experimental fishway. *Journal of Applied Ecology*, 58, 2760–2769.

630 Messerman, A.F., Clause, A.G., Gray, L.N., Krkošek, M., Rollins, H.B., Trenham, P.C., *et al.*
631 (2023). Applying stochastic and Bayesian integral projection modeling to amphibian
632 population viability analysis. *Ecological Applications*, 33, e2783.

633 Metcalf, C.J.E. & Pavard, S. (2007). Why evolutionary biologists should be demographers.
634 *Trends in ecology & evolution*, 22, 205–12.

635 Milenkaya, O., Weinstein, N., Legge, S. & Walters, J.R. (2013). Variation in body condition
636 indices of crimson finches by sex, breeding stage, age, time of day, and year.
637 *Conservation physiology*, 1, cot020.

638 Miller, T.E.X. & Tenhumberg, B. (2010). Contributions of demography and dispersal
639 parameters to the spatial spread of a stage-structured insect invasion. *Ecological
640 Applications*, 20, 620–633.

641 Montero-Serra, I., Garrabou, J., Doak, D.F., Figuerola, L., Hereu, B., Ledoux, J., *et al.* (2018).
642 Accounting for Life-History Strategies and Timescales in Marine Restoration.
643 *CONSERVATION LETTERS*, 11, e12341.

644 Moseby, K., Van Der Weyde, L., Letnic, M., Blumstein, D.T., West, R. & Bannister, H. (2023).
645 Addressing prey naivety in native mammals by accelerating selection for antipredator
646 traits. *Ecological Applications*, 33, e2780.

647 Nakaoka, M. (1996). Dynamics of age- and size-structured populations in fluctuating
648 environments: Applications of stochastic matrix models to natural populations. *Res
649 Popul Ecol*, 38, 141–152.

650 Paniw, M., James, T.D., Ruth Archer, C., Römer, G., Levin, S., Compagnoni, A., *et al.* (2021).
651 The myriad of complex demographic responses of terrestrial mammals to climate
652 change and gaps of knowledge: A global analysis. *Journal of Animal Ecology*, 90,
653 1398–1407.

654 Primack, R.B. (2008). *A primer of conservation biology*. Sinauer Associates Sunderland.

655 Robert, A., Chantepie, S., Pavard, S., Sarrazin, F. & Téplitsky, C. (2015). Actuarial
656 senescence can increase the risk of extinction of mammal populations. *Ecological
657 Applications*, 25, 116–124.

658 Rotter, M.C. & Holeski, L.M. (2018). A meta-analysis of the evolution of increased competitive
659 ability hypothesis: genetic-based trait variation and herbivory resistance trade-offs. *Biol
660 Invasions*, 20, 2647–2660.

661 Sabo, A.N., Berger-Tal, O., Blumstein, D.T., Gregg, A.L. & Swaddle, J.P. (2024).
662 Conservation practitioners' and researchers' needs for bridging the knowledge–action
663 gap. *Frontiers in Conservation Science*, 5, 1415127.

664 Salguero-Gómez, R., Viole, C., Gimenez, O. & Childs, D. (2018). Delivering the promises of
665 trait-based approaches to the needs of demographic approaches, and *vice versa*.
666 *Functional Ecology*, 32, 1424–1435.

667 Schaub, M. & Kéry, M. (2021). *Integrated population models: Theory and ecological
668 applications with R and JAGS*. Academic Press.

669 Schindler, D.E., Essington, T.E., Kitchell, J.F., Boggs, C. & Hilborn, R. (2002). Sharks and
670 tunas: fisheries impacts on predators with contrasting life histories. *Ecological
671 Applications*, 12, 735–748.

672 Soulé, M.E. (1985). What is conservation biology? *BioScience*, 35, 727–734.

673 Speakman, C.N., Bull, S., Cubaynes, S., Davis, K.J., Devillard, S., Fryxell, J.M., *et al.* (2025).
674 Understanding and Predicting Population Response to Anthropogenic Disturbance:
675 Current Approaches and Novel Opportunities. *Ecology Letters*, 28, e70198.

676 Stearns, S.C. (1976). Life-history tactics: a review of the ideas. *The Quarterly Review of
677 Biology*, 51, 3–47.

678 Stearns, S.C. (1992). *The Evolution Of Life Histories*. Oxford University Press.

679 Stephens, P.A. & Sutherland, W.J. (1999). Consequences of the Allee effect for behaviour,
680 ecology and conservation. *Trends in Ecology & Evolution*, 14, 401–405.

681 Stott, I., Salguero-Gómez, R., Jones, O.R., Ezard, T.H., Gamelon, M., Lachish, S., *et al.*
682 (2024). Life histories are not just fast or slow. *Trends in Ecology & Evolution*, 39, 830–
683 840.

684 Stott, I., Townley, S. & Hodgson, D.J. (2011). A framework for studying transient dynamics of
685 population projection matrix models. *Ecology Letters*, 14, 959–970.

686 Sutton, A.O., Strickland, D., Freeman, N.E. & Norris, D.R. (2021). Climate-driven carry-over
687 effects negatively influence population growth rate in a food-caching boreal passerine.
688 *Global Change Biology*, 27, 983–992.

689 Tuljapurkar, S.D. (1982). Population dynamics in variable environments. II. Correlated
690 environments, sensitivity analysis and dynamics. *Theoretical Population Biology*, 21,
691 114–140.

692 Van De Walle, J., Fay, R., Gaillard, J.-M., Pelletier, F., Hamel, S., Gamelon, M., *et al.* (2023).
693 Individual life histories: neither slow nor fast, just diverse. *Proc. R. Soc. B.*, 290,
694 20230511.

695 Waldron, J.L., Welch, S.M., Bennett, S.H., Kalinowsky, W.G. & Mousseau, T.A. (2013). Life
696 history constraints contribute to the vulnerability of a declining North American
697 rattlesnake. *Biological Conservation*, 159, 530–538.

698 Wang, H., Chen, Y., Hsu, C. & Shen, S. (2017). Fishing-induced changes in adult length are
699 mediated by skipped-spawning. *Ecological Applications*, 27, 274–284.

700 Watts, K., Whytock, R.C., Park, K.J., Fuentes-Montemayor, E., Macgregor, N.A., Duffield, S.,
701 *et al.* (2020). Ecological time lags and the journey towards conservation success.
702 *Nature Ecology & Evolution*, 4, 304–311.

703 West, R.S., Blumstein, D.T., Letnic, M. & Moseby, K.E. (2019). Searching for an effective pre-
704 release screening tool for translocations: can trap temperament predict behaviour and
705 survival in the wild? *Biodivers Conserv*, 28, 229–243.

706 White, E.R. (2019). Minimum time required to detect population trends: the need for long-term
707 monitoring programs. *BioScience*, 69, 40–46.

708 White, J.W., Botsford, L.W., Hastings, A., Baskett, M.L., Kaplan, D.M. & Barnett, L.A.K. (2013).
709 Transient responses of fished populations to marine reserve establishment.
710 *Conservation Letters*, 6, 180–191.

711 Winemiller, K.O. & Rose, K.A. (1992). Patterns of Life-History Diversification in North
712 American Fishes: implications for Population Regulation. *Can. J. Fish. Aquat. Sci.*, 49,
713 2196–2218.

714 Wong, T.M. & Ticktin, T. (2015). Using population dynamics modelling to evaluate potential
715 success of restoration: a case study of a Hawaiian vine in a changing climate.
716 *Environmental Conservation*, 42, 20–30.

717

718 **Figure captions**

719 **Figure 1.** Schematic view of the relationships between demographic concepts and the
720 ecological levels at which they are defined and commonly applied. The demographic
721 parameters (mean, variance) and migratory tactics shown in the figure are hypothetical and
722 were chosen for illustration purposes.

723

724 Species differ in their life-history strategies, often aligning with the **fast–slow continuum of life**
725 **histories** (A; adapted from Capdevila *et al.* 2020a). For example, sandhill cranes (*Grus canadensis*)
726 represent a relatively slow strategy, characterized by delayed maturity, a long lifespan, and low
727 fecundity. These trait combinations are mainly driven by **life-history trade-offs** that emerge at the
728 individual level and scale up to shape population- and species-level patterns (B). In slow-living species,
729 natural selection tends to favour strategies that allocate relatively more resources to survival,
730 particularly among mature stages (e.g., stage 5) over immediate reproductive output. Trade-offs and
731 temporal environmental variability cause these demographic parameters (stage-specific survival,
732 growth, fecundity) to vary and co-vary through time (positive or negative **temporal covariance**; C).
733 According to the **demographic buffering** hypothesis, the demographic parameters that most strongly
734 influence long-term population growth are expected to be more buffered against environmental variation
735 (e.g., low variation in survival of stage 5) to help maintain long-term population growth under fluctuating
736 conditions. Variation (**lability**) in some demographic parameters can also be beneficial when natural
737 selection favours responsiveness to environmental conditions that disproportionately enhance those
738 parameters when conditions improve, outweighing the fitness costs of reduced rates in poor years.
739 **Individual heterogeneity**, driven by physiological, morphological and/or behavioural differences, such
740 as migratory tactics (D), further contributes to this (co)variation. For example, some individuals may
741 consistently survive and reproduce more successfully than others in a population, or survival rates
742 across different age classes may vary in synchrony with environmental conditions. Together, these
743 processes shape population dynamics and influence long-term growth. When a disturbance occurs,
744 whether positive (e.g., following a conservation action) or negative (e.g., sport harvesting), it can
745 temporarily disrupt the population's structure, leading to **transient dynamics** before reaching again the
746 long-term, asymptotic growth (E; effects of initial crane stage structure on transient population growth,
747 adapted from Gerber and Kendall, 2016. Lines represent different scenarios of initial stage structure

748 and for each scenario, colours indicate the initial stage with the majority of individuals, $\geq 50\%$). The
749 magnitude and duration of the transient phase reflect the population's demographic resilience to this
750 perturbation.

751

752 **Figure 2.** Limitations and contributions of the six demographic concepts to conservation
753 practice, evaluated through key criteria: (i) Data requirements (considering monitoring
754 duration, sample size, and number of demographic traits measured; Table S8.1), (ii) Modelling
755 complexity (type of models and number of analytical steps required for models of minimal
756 complexity, see detail in Table S8.2), (iii) concept maturity (long-established concept in
757 demography VS under active development), and (iv) Operability for conservation applications.
758 Levels are ranked as low, medium and high. The operability of each concept was assessed
759 according to three criteria: (i) Ability to set general ecological and conservation expectations
760 based on limited demographic outcomes, (ii) Capacity to provide robust and reliable estimates
761 of key demographic metrics; and (iii) Usefulness for guiding conservation decisions by
762 providing detailed demographic outcomes. This qualitative synthesis is intended to be
763 evaluative rather than prescriptive, and should be interpreted in light of the study goals and
764 practical constraints.

765

766

767

768

769

770

771

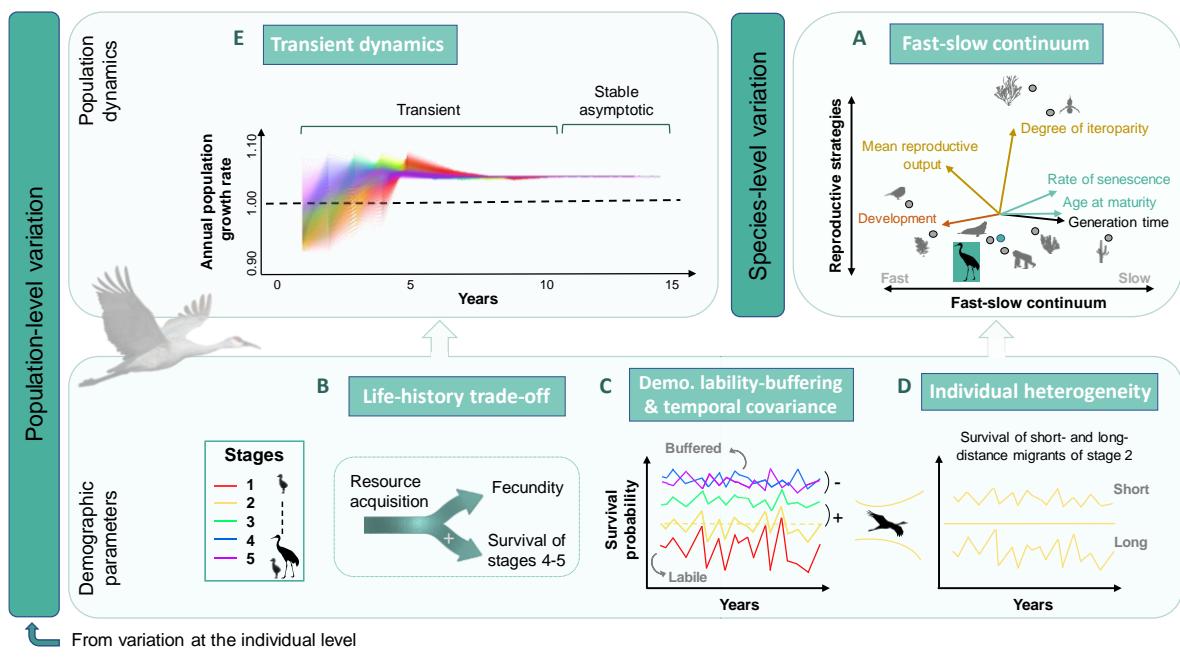
772

773

774

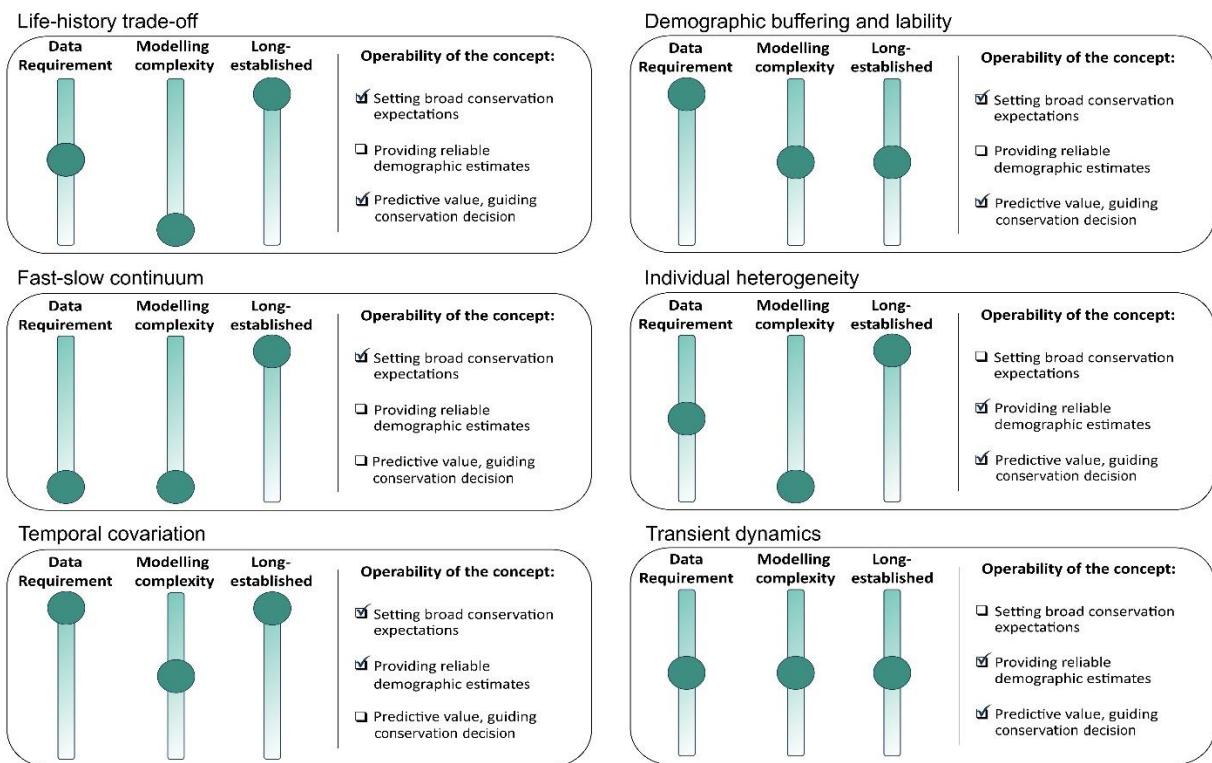
775

776 **Figure 1**



777

778 **Figure 2**



779

Table 1. Definitions, main keywords, and key studies recommended as a first introduction to each of the six demographic concepts.

Concept [main keywords]	Definition	Getting started with the concept
Life-history trade-off	Direct or lagged negative correlation(s) between two or more fitness-related traits (such as age at maturity, growth, reproduction, survival, and lifespan) due to the limited amount of energy that organisms can acquire and must allocate among traits (Stearns 1992). Life-history trade-offs occur at the individual level, but are often measured at the population level. They arise not only from limited resources, but from a combination of genetic, developmental, physiological, structural, and ecological constraints, all of which restrict the simultaneous optimization of multiple fitness components (Garland <i>et al.</i> 2022).	Stearns (1992)
Fast-slow continuum [fast-slow continuum; r-/K-selected species; pace of life; variation in life-history strategies]	The fast-slow continuum refers to one major axis of life-history variation across species , reflecting different strategies in speed of life (Stearns 1976). This continuum ranges from short-lived, fast-growing species with high fecundity at one end, to long-lived species with low fecundity and late age at first reproduction at the other, reflecting a diversity of life-history strategies throughout.	Stott <i>et al.</i> (2024)
Temporal covariation [temporal/demographic correlation/covariation; demographic compensation]	Population-level covariation between demographic parameters within a population over time. Temporal covariation is positive (negative) when two or more demographic parameters — either different (e.g., reproduction and adult survival) or the same across life stages (e.g., survival of juveniles and yearlings) — increase or decrease synchronously (in opposite direction). Environmental stochasticity, together with other processes such as density-dependence and life-history trade-offs, influence population-level covariation in demographic parameters (Tuljapurkar 1982).	Fay <i>et al.</i> (2022b)
Demographic buffering and lability [Environmental canalization; demographic buffering; demographic lability]	In a population, buffered and labile demographic parameters (e.g., age-specific survival or fecundity) are characterized by low and high fluctuations, respectively, in response to temporal variation in the environment. Ability and buffering are adaptive when they lead to an overall increase (for lability) or stable long-term population growth in varying environments (Hilde <i>et al.</i> 2020; Koons <i>et al.</i> 2009; Le Coeur <i>et al.</i> 2022). According to the demographic buffering hypothesis , natural selection is expected to favour low variance in demographic parameters that have the strongest influence on population growth in stable environmental conditions.	Hilde <i>et al.</i> (2020) Koons <i>et al.</i> (2009)
Individual heterogeneity [individual heterogeneity; individual quality; frailty; among individual variation; personality; individual behaviour; temperament]	Individual heterogeneity refers to any observed or unobserved (i.e. measured or latent) source of variation in a given trait among individuals, irrespective of its influence on fitness (i.e. fitness and non-fitness-related traits; Hamel <i>et al.</i> 2018b). The variation in traits within and among individuals has also been referred to as among-individual variation, and individual quality, frailty, personality and temperament (e.g., Firth <i>et al.</i> 2018; Halstead <i>et al.</i> 2012). These terms have been used focusing, for example, on the among-individual variance in demographic parameters (Fay <i>et al.</i> 2022a), or on differences among individuals only associated	Cam <i>et al.</i> (2013); Hamel <i>et al.</i> (2018b)

with traits that underlie fitness components (Milenkaya *et al.* 2013). Individual heterogeneity can be **fixed** or **dynamic** whether individual differences are shaped early in life conditions and persist or change throughout life, respectively (see Cam *et al.* 2013 for a discussion of this terminology).

Transient dynamics [transient demography; transient dynamics; demographic resilience]	<p>Transient dynamics capture the short-term, non-stable dynamics of a population that arise from temporary shifts in its age or stage structure (Hastings 2004). These changes occur when the population is not in a stable state, for instance, following a perturbation that affects certain stages or ages more than others. Transient dynamics can be used to quantify demographic resilience and anticipate population's response to disturbances.</p> <p>Demographic resilience refers to the ability of populations to resist and recover from alterations in their demographic structure, usually with concomitant change in population size (Capdevila <i>et al.</i> 2020b). Different metrics can be used to quantify the demographic resilience, including the damping ratio (Caswell 2001).</p>	Capdevila <i>et al.</i> (2020b); Stott <i>et al.</i> (2011)
--	--	---

Table 2. Main conservation applications associated with each demographic concept, and the number of conservation articles referring to each application category. Studies that rely on a modelling or theoretical approach are indicated in brackets.

Concepts	Conservation applications	Number of studies
Life-history trade-offs	<ol style="list-style-type: none"> Understand and predict plastic or micro-evolutionary responses of populations affected by anthropogenic disturbances Understand how species establish and spread in novel or altered environments and use this knowledge to guide risk assessments and early detection efforts of invasion Understand the demographic consequences of management actions and inform future management actions, most notably restoration efforts 	11 (3) 4 11
Fast-slow continuum	<ol style="list-style-type: none"> Describe a single species, justify the study's relevance based on its life history, predict its vulnerability and extinction risk to disturbances, and identify conservation needs Serve methods-oriented applications Compare species' responses to environmental and/or anthropogenic disturbances to infer outcomes for other species along the fast-slow continuum and guide conservation efforts Explain and predict species' responses to conservation actions, assess their effectiveness, and inform future conservation measures Explain variation in vulnerability to extinction Serve as an index of community health and functioning, and measure changes in community composition or succession in response to environmental changes 	14 (1) 2 (1) 14 (4) 6 (3) 5 4 (1)
Temporal covariation	<ol style="list-style-type: none"> Accurately assess the contribution of demographic parameters to population growth and provide realistic, unbiased estimates of population dynamics, resilience and extinction risk Better assess and improve the efficiency of management plans or conservation actions Identify compensatory mechanisms (“demographic compensation”) 	3 (1) 2 3
Demographic lability – buffering	<ol style="list-style-type: none"> Assess whether a population's buffering capacity is maintained following environmental or anthropogenic perturbations, and to forecast extinction risks Assess population's recovery capacity and responsiveness to conservation/management actions 	4 2
Individual heterogeneity	<ol style="list-style-type: none"> Identify behavioural traits for conservation/management Reduce bias in demographic parameter estimates, thereby improving inference on population dynamics Evaluate stressor factors on specific individuals that can impact population structure and persistence, and potentially affect ecosystem-level processes Identify key individuals to support effective population management or invasive species control 	30 19 13 20

	5. Assess individual traits that affect reproductive success of population/species to improve conservation and management outcomes	5
Transient dynamics	1. Better estimate the extinction risk and key demographic parameters of populations and their demographic resilience to perturbations	16 (1)
	2. Identify and design the best conservation or management strategy and/or assess responsiveness to conservation/management actions	17 (1)

Appendix

S1 - Set of journals included in the six Web of Science searches

The search was restricted to a predefined set of journals in conservation ecology (listed below), including the 31 conservation journals ranked in 2022 according to Bradshaw & Brook's journal-ranking method (2016) and two additional journals, The Journal of Wildlife Management and Global Change Biology. It covered publications with cut-off dates ranging from February 2024 to January 2025, depending on the concept (see S2-S7).

- Ambio
- Animal conservation
- Aquatic conservation: Marine and Freshwater Ecosystems
- Basic and applied ecology
- Biodiversity and Conservation
- Biological conservation
- Biological invasions
- Bird conservation international
- Conservation biology
- Conservation genetics
- Conservation letters
- Conservation physiology
- Conservation science and practice
- Conservation & society
- Ecological applications
- Ecological Management & Restoration
- Ecology and society
- Emu – Austral Ornithology
- Endangered species research
- Environmental conservation
- Frontiers in Ecology and the Environment
- Global Change Biology
- Insect Conservation and Diversity
- Journal for Nature Conservation
- Journal of applied ecology
- Journal of insect conservation
- Nature Ecology & Evolution
- One earth
- Oryx
- People and nature
- Restoration ecology
- The Journal of Wildlife Management
- Tropical Conservation Science

S2 - Life-history trade-off

- **Full list of keywords used in Web of Science query (up to Apr. 2024):**

(ALL=((life-history OR life history) AND (trade-off* OR tradeoff*)) OR (survival AND (trade-off* OR tradeoff*)) OR (reproduction AND (trade-off* OR tradeoff*)) OR (growth AND (trade-off* OR tradeoff*)) OR (fitness AND (trade-off* OR tradeoff*)))

- **Number of conservation studies relying on the ‘Life-history trade-off’ concept, categorized by conservation applications and studied organismal groups (Table S2.1)**

Table S2.1 Number of articles categorized by studied organismal groups and type of conservation applications. Only articles in which the concept was quantified are included, excluding theoretical studies. Studies predominantly based on modelling are indicated in parentheses.

<i>Organismal groups</i>	<i>Conservation applications</i>			<i>Total</i>
	<i>1</i>	<i>2</i>	<i>3</i>	
Arthropods (mixed)	(1)			(1)
Actinopterygii	2(1)	1	1	4(1)
Amphibia	1			1
Anthozoa (corals)	1(1)		2	3(1)
Aves	2		1	3
Mammalia	1			1
Bivalvia	1			1
Reptilia	1			1
Magnoliopsida (trees)		1	5	6
Magnoliopsida (shrubs)		1	1	2
Magnoliopsida (herbaceous)	1		1	2
Plants mixed	1	1		2
Total	11(3)	4	11	26(3)

- Number of conservation studies categorized by conservation applications and trade-offs defined in terms of life-history traits (Table S2.2)

Table S2.2 Life-history traits involved in trade-offs across the selected articles, categorized by type of conservation applications. Articles referring to "maintenance" and "longevity" are grouped together under "survival". Two articles dealing with fitness-related traits (resistance to bleaching, competitive ability) that could not be linked to a single life-history trait are included in the final selection but excluded from this table. Numbers in square brackets indicate the number of articles that examined the three types of trade-offs.

<i>Life-history trade-offs</i>	<i>Conservation applications</i>			<i>Total</i>
	<i>1</i>	<i>2</i>	<i>3</i>	
Growth / Survival	5	1	6	12
Reproduction / Survival	3		3	6
Growth / Reproduction	2	2		4
Total	10 [3]	3	9 [2]	22 [5]

Note: Applications 1-3 correspond to: 1) Understand and predict plastic or micro-evolutionary responses of populations affected by anthropogenic disturbances; 2) Understand how species establish and spread in novel or altered environments and use this knowledge to guide risk assessments and early detection efforts of invasion; and 3) Understand the demographic consequences of management actions and inform future management actions, most notably restoration efforts

S3 - Fast-slow continuum of life histories

- **Full list of keywords used in Web of Science query (up to Dec. 2024):**

ALL="("slow-fast continuum") OR ALL="("fast-slow continuum") OR ALL="("fast-slow life-history continuum") OR ALL="("variation in life history strategies") OR ALL="("Variation in life history traits") OR ALL="("life-history strategies") OR ALL="("life history strategies") OR ALL="("life-history strategy") OR ALL="("life history strategy") OR ALL="("axis of life history strategies") OR ALL="("r-/K-Selected") OR ALL="("r-K Strategies") OR ALL="("K-selected species") OR ALL="("r-selected species") OR ALL="("r-Strategy") OR ALL="("K-Strategy") OR ALL="("fast-living species") OR ALL="("slow-living species") OR ALL="("fast life histories") OR ALL="("slow life histories") OR ALL="("fast life history") OR ALL="("slow life history") OR ALL="("fast-slow life history") OR ALL="("slow-fast life history") OR ALL="("pace of life continuum") OR ALL="("pace of life") OR ALL="(" fast-slow life history spectrum")

- **Number of conservation studies relying on the ‘fast-slow continuum’ concept, categorized by conservation applications, main keywords and organismal groups (Tables S3.1 and S3.2, respectively)**

Table S3.1 Number of articles categorized by keywords and type of conservation applications. Only articles in which the concept was quantified (rather than discussed) are included, excluding theoretical studies.

<i>Keywords used in articles</i>	<i>Conservation applications</i>						<i>Total</i>
	<i>1</i>	<i>2</i>	<i>3</i>	<i>4</i>	<i>5</i>	<i>6</i>	
Fast-slow continuum	12	1	9	5	3	2	32
r-/K-selected species	2		1				3
r/K & fast-slow continuum		1			1	1	3
Pace of life			1				1
Pace of life & Fast-slow continuum			1		1	1	3
Long-lived/short-lived species			1				1
Fast-slow growing species		1	1				2
Total	14	2	14	6	5	4	45

Table S3.2 Number of articles categorized by studied organismal groups and type of conservation applications. Only articles in which the concept was quantified are included, excluding theoretical studies. Numbers indicate the number of articles that focused on a specific organismal group, with the number “(1)” representing one study involving two different organismal groups (below, Mammalia and Insecta).

<i>Organismal groups</i>	<i>Conservation applications</i>						<i>Total</i>
	<i>1</i>	<i>2</i>	<i>3</i>	<i>4</i>	<i>5</i>	<i>6</i>	
Mammalia	2		2 (1)	1	2		7 (1)
Aves	3	1	2	1	1	1	9
Actinopterygii	2	1	1			1	5
Chondrichthyes			1				1
Reptilia	4		2				6
Amphibia	2		2				4
Malacostraca	1						1
Benthic invertebrates					1		1
Woody species (>1900 species)					1		1
Woody and herbaceous species			2	2			4
Mixed: mostly forbs and woody species; cacti, ferns, graminoids				1			1
Insecta			(1) Diptera, Coleoptera, Lepidoptera		1	Lepidoptera	1 (1)
Mixed: Anthozoa (corals), Demospongiae, Liliopsida (seagrass)				1			1
Mixed: other			1 Mammalia, Reptilia, Aves		1 Vertebrates, plants, invertebrates		2

Note: Conservation applications 1-6 correspond to: 1) Species description and predictability on a species' vulnerability to extinction or ability to adapt; 2) To serve methods-oriented applications; 3) to compare and assess species' responses to anthropogenic disturbances and infer outcomes for other species along the fast-slow continuum and guide conservation efforts; 4) To predict species' responses to conservation actions and assess effectiveness; 5) To explain variation in vulnerability to extinction; 6) To serve as an index of community health and functioning, and measure changes in community composition or succession in response to environmental changes .

Additional information:

Application 3 - Studies comparing the impact of anthropogenic disturbances on species (quantified as species' vulnerability, resilience, demographic responses or energetic investment/cost) along the fast-slow continuum of life histories often encompass a broad range of disturbance sources (n= 14, Table 2). These disturbances mostly include climate and land use changes (e.g., Albaladejo-Robles et al. 2023); habitat degradation and agricultural intensification (e.g., Harper et al. 2008); fishing/harvesting/overexploitation/individual collection (e.g., Schindler et al. 2002 with consequences at the species and food web levels - applications 3 and 6); and hydraulic disturbance (e.g., McManamay et al. 2015, using the Equilibrium-Periodic-Opportunistic continuum, Table S3.4).

Application 4 - Most studies using the fast-slow continuum to test and assess the success of conservation actions between species refer to translocation or reintroduction programs (n= 6; e.g., Ducatez & Shine, 2019).

- **Ten theoretical/modelling articles with conservation goal and applications:** In these articles, the authors simulated a range of life history strategies or two contrasting strategies (slow vs fast-living species) to explore mainly whether life history traits can help to determine a species' resistance and resilience to disturbance (n= 4), or to identify which conservation measure can be efficient regarding the considered life histories (n=3).

Table S3.3 Number of theoretical studies grouped by category of conservation application.

<i>Conservation applications</i>	<i>N articles</i>
1. Species description and predictability on a species' vulnerability to extinction, ability to adapt	1
2. To serve methods-oriented applications	1
3. To compare and assess species' responses to anthropogenic disturbances	4
4. To predict species' responses to conservation actions and assess effectiveness	3
5. To explain variation in vulnerability to extinction	0
6. To assess community functioning in regards of environmental changes	1

- **From the literature search, eleven articles (and two articles that only discussed the concept) referred to the fast-slow continuum (or r/K concept) at the intraspecific scale.**

Three main applications were identified by comparing life histories and pace of life between populations. It was commonly applied to explain, predict, or highlight disturbance-induced changes in life history traits between populations, as an adaptive response to new environment(s). These disturbances include climate change, disease, urbanization or invasion that induce changes in life history strategies of i) native species (e.g., comparison between populations with or without the presence of an invasive species e.g., Sharma et al. 2021) or ii) newly established and potentially invasive species (comparison between locations, e.g., Dean et al. 2023). The two articles that discussed the concept of fast-slow continuum also used it in the context of invasion-driven and establishment-driven shifts in life histories between populations.

The concept was also useful to evaluate and then improve efficiency of conservation action (e.g., conservation stocking of eels as a recovery tool, where matching the life-history characteristics of donor and recipient sub-populations was found to be important; Stacey et al. 2015). The third and last application category focuses on quantifying intraspecific variation in life histories to reliably assess extinction risk of an endangered species in the face of multivariate environmental stressors, and guide conservation actions (Monnet et al. 2022).

- Other continuums of life history variation

Thirteen studies relied on different continuum of life history variation among fish and bivalves, including the Equilibrium-Periodic-Opportunistic continuum (n= 8; Table S3.4).

Table S3.4 Other continuums reported in studies grouped by conservation application. Brackets indicate if a study (and the use of a specific continuum) belongs to two categories of conservation application.

<i>Organismal groups</i>	<i>Conservation applications</i>				<i>Total</i>
	2	3	4	6	
Equilibrium-Periodic-Opportunistic continuum	1	4	3		8
Continuum based on functional groups (traits related to reproduction, dispersal, development time and synchronisation)			1		1
Continuum based on species' foraging behavior/guilds		1			1
Continuum based on species' reproductive strategy, diet specialization and foraging behaviour		1			1
Continuum based on competitiveness and stress tolerance		(1)		(1)	1
Continuum based on competitiveness, reproductive effort and lifespan			(1)	(1)	1

References:

Albaladejo-Robles, G., Böhm, M., & Newbold, T. (2023). Species life-history strategies affect population responses to temperature and land-cover changes. *Global Change Biology*, 29(1), 97-109. <https://doi.org/10.1111/gcb.16454>

Dean, E.K., Drake, D.A.R. & Mandrak, N.E. (2023) Non-linear effects on the population performance of Bighead Carp under different maturation schedules. *Biological Invasions* 25, 3567–3581. <https://doi.org/10.1007/s10530-023-03126-z>

Ducatez, S., & Shine, R. (2019). Life-history traits and the fate of translocated populations. *Conservation Biology*, 33(4), 853-860. <https://doi.org/10.1111/cobi.13281>

Harper, E. B., Rittenhouse, T. A., & Semlitsch, R. D. (2008). Demographic consequences of terrestrial habitat loss for pool-breeding amphibians: predicting extinction risks associated with inadequate size of buffer zones. *Conservation Biology*, 22(5), 1205-1215. <https://doi.org/10.1111/j.1523-1739.2008.01015.x>

McManamay, R. A., & Frimpong, E. A. (2015). Hydrologic filtering of fish life history strategies across the United States: implications for stream flow alteration. *Ecological Applications*, 25(1), 243-263. <https://doi.org/10.1890/14-0247.1>

Monnet, G., Corse, E., Archambaud-Suard, G., Grenier, R., Chappaz, R., & Dubut, V. (2022). Growth variation in the endangered fish Zingel asper: Contribution of substrate quality, hydraulics, prey abundance, and water temperature. *Aquatic Conservation: Marine and Freshwater Ecosystems*, 32(7), 1156-1170. <https://doi.org/10.1002/aqc.3818>

Sharma, A., Dubey, V.K., Johnson, J.A. et al. (2021) Introduced, invaded and forgotten: allopatric and sympatric native snow trout life-histories indicate brown trout invasion effects in the Himalayan hinterlands. *Biological Invasions* 23, 1497–1515. <https://doi.org/10.1007/s10530-020-02454-8>

Schindler, D. E., Essington, T. E., Kitchell, J. F., Boggs, C., & Hilborn, R. (2002). Sharks and tunas: fisheries impacts on predators with contrasting life histories. *Ecological Applications*, 12(3), 735-748. [https://doi.org/10.1890/1051-0761\(2002\)012\[0735:SATFIO\]2.0.CO;2](https://doi.org/10.1890/1051-0761(2002)012[0735:SATFIO]2.0.CO;2)

Stacey, J. A., Pratt, T. C., Verreault, G., & Fox, M. G. (2015). A caution for conservation stocking as an approach for recovering Atlantic eels. *Aquatic Conservation: Marine and Freshwater Ecosystems*, 25(4), 569-580. <https://doi.org/10.1002/aqc.2498>

S4 - Temporal covariation among demographic parameters

- **Full list of keywords used in Web of Science query (up to Jan. 2025):**

(ALL=("temporal correlation\$" AND "demographic parameter\$") OR ALL=("temporal correlation\$" AND "demographic rate\$") OR ALL=("temporal correlation\$" AND "vital rate\$") OR ALL=("correlation\$" AND "demographic parameter\$") OR ALL=("correlation\$" AND "parameter") OR ALL=("correlation\$" AND "vital rate\$") OR ALL=("demographic correlation\$") OR ALL=("vital rate correlation\$") OR ALL=("temporal covariation\$" AND "demographic parameter\$") OR ALL=("temporal covariation\$" AND "demographic rate\$") OR ALL=("temporal covariation\$" AND "vital rate\$") OR ALL=("temporal covariation\$" AND "parameter\$") OR ALL=("covariation\$" AND "demographic parameter\$") OR ALL=("covariation\$" AND "vital rate\$") OR ALL=("demographic covariation\$") OR ALL=("vital rate covariation\$") OR ALL=("temporal covariance\$" AND "demographic parameter\$") OR ALL=("temporal covariance\$" AND "demographic rate\$") OR ALL=("temporal covariance\$" AND "vital rate\$") OR ALL=("temporal covariance\$" AND "parameter\$") OR ALL=("covariance\$" AND "demographic parameter\$") OR ALL=("covariance\$" AND "vital rate\$") OR ALL=("demographic covariance\$") OR ALL=("vital rate covariance\$") OR ALL=("demographic compensation"))

We did not include articles on temporal covariation of demographic parameters between populations, compared along geographical and environmental gradients (i.e., ‘demographic compensation’ defined as changes in opposite directions in mean demographic parameters across populations - see Villegas et al. 2015). Only articles examining compensatory mechanisms within a population were considered.

- **Number of conservation studies using the ‘temporal covariation’ concept, categorized by conservation applications and organismal groups (Tables S4.1)**

Table S4.1 Number of articles categorized by studied organismal groups and type of conservation applications. The number of articles refers to those in which the concept was quantified rather than discussed (theoretical studies in square brackets).

<i>Organismal groups</i>	<i>Conservation applications</i>			<i>Total</i>
	<i>1</i>	<i>2</i>	<i>3</i>	
Magnoliopsida	1	1		2
Actinopterygii	1 [1]			1 [1]
Reptilia	1			1
Mammalia		1	1	2
Amphibia			1	1
Aves			1	1
Total	3 [1]	2	3	8 [1]

Note: Applications 1-3 correspond to: 1) Accurately assess the contribution of demographic parameters to population growth and provide realistic, unbiased estimates of population dynamics, resilience and extinction risk; 2) Better assess and improve the efficiency of management plans or conservation actions; and 3) Identify compensatory mechanisms (“demographic compensation”).

Reference:

Villegas, J., Doak, D. F., García, M. B., & Morris, W. F. (2015). Demographic compensation among populations: what is it, how does it arise and what are its implications? *Ecology letters*, 18(11), 1139-1152. <https://doi.org/10.1111/ele.12505>

S5 - Demographic buffering and lability

- **Full list of keywords used in Web of Science query (up to Dec. 2024):**

(ALL=("Environmental canalization" OR "Canalized fitness component" OR "Canalization of survival" OR "Canalization of fertility" OR "Life-history buffering" OR "Life history buffering" OR "Demographic buffering" OR "Buffering hypothesis" OR "Buffering of demographic rates" OR "Buffering of demographic parameters" OR "Buffering in the vital rates" OR "Buffered against environmental variability" OR "temporal variation in vital rates" OR "temporal variation in demographic parameters" OR "Demographic lability" OR "life history lability" OR "life-history lability" OR ("lability" AND "vital rates") OR ("lability" AND "demographic parameters") OR "Labile fertility" OR "Labile demography"))

- **Keyword occurrence (Tables S5.1) and list of organismal groups studied (Table S5.2) across the six selected articles on the ‘demographic buffering and lability’ concept, categorized by type of conservation application**

Table S5.1 Keyword occurrence across the selected articles, categorized by type of conservation application.

<i>Keywords used in articles</i>	<i>Conservation applications</i>	
	<i>1</i>	<i>2</i>
Environmental canalization	2	
Demographic buffering	2	2
Demographic lability		
Buffering hypothesis	2	

Table S5.2 Number of articles categorized by studied organismal groups across the six selected articles, with (1) indicating one study involving two different organismal groups.

<i>Organismal groups</i>	<i>Conservation applications</i>		<i>Total</i>
	<i>1</i>	<i>2</i>	
Actinopterygii	0	1	1
Aves	2 (1)	0	2 (1)
Mammalia	1 (1)	0	2 (1)
Forb (Equisetopsida)	0	1	1
Total			6 studies

Note: Application 1: Assessing whether a population’s buffering capacity is maintained following environmental or anthropogenic perturbations, and to forecast extinction risks; **Application 2:** Assessing population’s recovery capacity and responsiveness to conservation/management actions

S6 - Individual heterogeneity

- **Full list of keywords used in Web of Science query (up to Feb. 2024):**

ALL=(“individual heterogeneity” OR “individual quality” OR “frailty” OR “among individual variation” OR “personality” OR “individual behavior” OR “individual behaviour” OR “temperament”)

- **Article classification on the ‘individual heterogeneity’ concept by keyword occurrence (Tables S6.1), definition use (Table S6.2), organismal groups studied (Table S6.3), and type of heterogeneity (Table S6.4) categorized by type of conservation application**

Note: Applications 1-5 correspond to: 1) Assessing behavioural traits with high individual variation that may influence conservation and management outcomes, in order to inform and refine strategies; 2) Reducing bias in demographic parameter estimates, thereby improving inference on population dynamics; 3) Evaluating stressor factors on specific individuals that can impact population structure and persistence, and potentially affect ecosystem-level processes; 4) Identifying key individuals to support effective population management or invasive species control; 5) Assessing individual traits that affect reproductive success of population/species to improve conservation and management outcomes.

Table S6.1 Number of articles categorized by keywords and type of conservation applications. Only articles in which the concept was quantified (rather than discussed) are included, excluding theoretical studies.

<i>Keyword used in articles</i>	<i>Conservation applications</i>					<i>Total</i>
	<i>1</i>	<i>2</i>	<i>3</i>	<i>4</i>	<i>5</i>	
Indiv. heterogeneity	1	13	1	1		16
Indiv. quality	1		3	1	3	8
Frailty		2				2
Among indiv. variation			1		1	2
Indiv. behaviour	4	2	1			7
Personality	4		1	7		12
Temperament	1		2			3
Indiv. heterogeneity & Indiv. quality					1	1
Indiv. heterogeneity & Indiv. quality & Personality	1					1
Indiv. heterogeneity & Frailty		1				1
Indiv. heterogeneity & Among indiv. variation	3			1		4
Indiv. heterogeneity & Among indiv. variation & Personality				1		1
Indiv. heterogeneity & Personality		1				1
Indiv. quality & Indiv. behaviour			1			1
Indiv. quality & Personality	1			1		2
Among indiv. variation & Indiv. behaviour	1			1		2
Among indiv. variation & Indiv. behaviour & Personality	3			1		4
Among indiv. variation & Indiv. behaviour & Personality & Temperament	2		1			3
Among indiv. variation & Personality	1			3		4
Among indiv. variation & Personality & Temperament	1				1	1
Indiv. behaviour & Personality	3		2	1		6
Personality & Indiv. behaviour						0
Personality & Temperament	3			2		5
Total	30	19	13	20	5	87 studies

Table S6.2 Number of articles categorized by definition use (Fitness or Non-Fitness related trait) and type of conservation applications. Only articles in which the concept was quantified (rather than discussed) are included, excluding theoretical studies.

<i>Definition used</i>	<i>Conservation applications</i>					<i>Total</i>
	<i>1</i>	<i>2</i>	<i>3</i>	<i>4</i>	<i>5</i>	
Fitness	8	7	5	5	5	30
Non-fitness	22	12	7	13		54
Fitness & Non-fitness			1	2		3
Total	30	19	13	20	5	87 studies

Table S6.3 Number of articles categorized by organismal groups studied and type of conservation applications. Only articles in which the concept was quantified (rather than discussed) are included, excluding theoretical studies.

<i>Organismal group</i>	<i>Conservation applications</i>					<i>Total</i>
	<i>1</i>	<i>2</i>	<i>3</i>	<i>4</i>	<i>5</i>	
Mammalia - terrestrial	13	13	2	5	1	34
Mammalia - marine		2				2
Aves	12		6	4	3	25
Actinopterygii	2	1	3	5		11
Reptilia	2	2	1			5
Amphibia	1					1
Insecta		1		1		2
Arachnida			1	1		2
Malacostraca				1		1
Liliopsida				1		1
Mixed: Actinopterygii & Malacostraca				1		1
Mixed: Magnoliopsida & Insecta					1	1
Mixed: Actinopterygii & Amphibia				1		1
Total	30	19	13	20	5	87 studies

Table S6.4 Number of articles categorized by type of heterogeneity (Measured* or Latent**; with the number of variables considered, in brackets) and type of conservation applications. Only articles in which the concept was quantified (rather than discussed) are included, excluding theoretical studies.

<i>Type of heterogeneity</i>	<i>Conservation applications</i>					<i>Total</i>
	<i>1</i>	<i>2</i>	<i>3</i>	<i>4</i>	<i>5</i>	
Measured [1]	1	1	4	1	1	8
Measured [2]		1		1		2
Measured [3]	2	1	1	1	1	6
Measured [4]				1		1
Latent [1]	17	15	3	9	1	45
Latent [2]				1		1
Latent [3]				1		1
Measured&Latent [2]	6		3	4		13
Measured & Latent [3 to 4]	3	1				4
Measured & Latent [5 to 7]	1		2	2	1	6
Total	30	19	13	20	5	87 studies

* Measured variables include: i) Sex; ii) Age; iii) Morphological measurements (e.g., size, length, gut fullness), iv) Mass; v) Body condition; vi) Reproduction output; vii) Reproduction state or life stage; viii) Physiology, epidemiology or stress; ix) Other (e.g., enclosure type, generation in captivity).

** Latent variables include: i) Behaviour; ii) Personality, boldness, temperament or tolerance; iii) Movement, spatial use or foraging; iv) Reproduction investment, parental behaviour or maternal allocation; v) Latent heterogeneity; vi) detection.

Additional information: supplementary studies/further reading (individual heterogeneity):

In our study, we excluded review or fully experimental studies. However, we recommend further reading with recommendation of new approaches for future studies that reveal the importance of the concept of individual heterogeneity not only at the population but also at the ecosystem level:

- Include functional multi-trait covariation (e.g., affected by sex, long-term selection history and short-term environmental conditions) that can have a cascading ecological response to anthropogenic global changes (Pauli et al., 2020).
- Individual phenotypic diversity is a complex phenomenon that needs to be considered in ecosystem-based studies. The ultimate realization is that maintaining or increasing individual trait diversity may enhance the resilience of not only species but also entire ecosystems to environmental perturbations. Individuals are of central importance for ecosystem-based approaches (Ward et al., 2016).

References:

Pauli, B.D., Edeline, E., Evangelista, C., 2020. Ecosystem consequences of multi-trait response to environmental changes in Japanese medaka, *Oryzias latipes*. *Conserv. Physiol.* 8. <https://doi.org/10.1093/conphys/coaa011>

Ward, T.D., Algera, D.A., Gallagher, A.J., Hawkins, E., Horodysky, A., Jørgensen, C., Killen, S.S., McKenzie, D.J., Metcalfe, J.D., Peck, M.A., Vu, M., Cooke, S.J., 2016. Understanding the individual to implement the ecosystem approach to fisheries management. *Conserv. Physiol.* 4, cow005. <https://doi.org/10.1093/conphys/cow005>

S7 Transient dynamics

- **Full list of keywords used in Web of Science query (up to Dec 2024):**

(ALL=("transient demography" OR "transient dynamics" OR "demographic resilience" OR "short-term population dynamics"))

- **Number of conservation studies relying on the 'transient dynamics' concept, categorized by main keywords and type of conservation applications, and organismal groups (Tables S7.1 and S7.2, respectively)**

Table S7.1 Number of articles categorized by keywords and type of conservation applications. Theoretical studies that didn't include the analysis of field data are indicated in square brackets.

<i>Keywords used in articles</i>	<i>Conservation applications</i>		<i>Total</i>
	<i>I</i>	<i>2</i>	
Transient dynamics	15 [1]	16 [1]	31 [2]
Transient demography	0	0	0
Demographic resilience	0	0	0
Short-term population dynamics	1	0	1
Transient dynamics AND Demographic resilience	0	1	1
<i>Total</i>		33[2] studies	

Table S7.2 Number of articles categorized by studied organismal groups and type of conservation applications. Two theoretical studies that didn't include the analysis of field data were excluded. Numbers indicate the number of articles that focused on each organismal group. One article focused on both Mammalia and Magnoliopsida, another article focused on Algae, Echinodermata and Crustacea, and a third article focused on Insecta, Reptilia, Aves and Mammalia.

<i>Organismal groups</i>	<i>Conservation applications</i>		<i>Total</i>
	<i>I</i>	<i>2</i>	
Actinopterygii	0	8	8
Algae	0	1	1
Aves	3	2	5
Crustacea	0	1	1
Echinodermata	0	1	1
Insecta	2	2	4
Magnoliopsida	7	3	10
Mammalia	3	3	6
Reptilia	1	1	2

Note: Applications 1 and 2 correspond to 1) Better estimate the extinction risk and key demographic parameters of populations of threatened species and their demographic resilience to perturbations; 2) Identify the best conservation or management strategy and/or assess responsiveness to conservation/management actions.

S8 - Criteria for data requirements and modelling complexity across demographic concepts

Table S8.1 Data requirements for studying each concept and quantifying associated metrics were evaluated based on three criteria: 1) sample size; 2) monitoring duration (i.e., no temporal replicate needed, medium, or long time series); 3) number of demographic traits measured. TS= time series.

<i>Demographic concepts</i>	<i>Sample size</i>	<i>Monitoring duration</i>	<i>Number of distinct demographic traits</i>	<i>Score used in figure 2 (low, medium, high)</i>
Life-history trade-off	Moderate	At least 2 time points	≥ 1 (often 2)	Medium
Fast-slow continuum	Small / moderate	No TS needed	≥ 2	Low
Temporal covariation	Moderate	Long-term TS	≥ 2	High
Demographic buffering-lability	Moderate	Long-term TS	≥ 2	High
Individual Heterogeneity	Large	No TS needed (but better to have repeated measurements)	≥ 1	Medium
Transient dynamics	Moderate	Medium-length TS (some years during and after the disturbance)	Multiple	Medium

Table S8.2 Modelling complexity for each concept was determined based on the type of models required to quantify the demographic outcomes (regression models, capture-mark-recapture models, population models) and the number of analytical steps (one- or two-steps). A two-step process involves quantifying time- and (st-)age-specific demographic parameters, and then integrating them into population models. In Figure 2 and throughout the manuscript, we focus on models of minimal complexity used to quantify demographic outcomes. For each concept, more sophisticated approaches are possible (e.g., models with latent variables, multivariate mixed models, or complex structured population models). These approaches are indicated in grey. CMR models is for capture-mark-recapture models.

<i>Demographic concepts</i>	<i>Type of models needed (regression models, CMR models, population models)</i>	<i>One or two-step process</i>	<i>Score used in Fig. 2 (low, medium, high)</i>
Life-history trade-off	Regression models and/or CMR models are required, or population models if assessing trade-offs from temporal covariance among multiple parameters. A more sophisticated approach to estimating trade-offs involves multivariate CMR or integrated population models, which explicitly propagate uncertainty in demographic estimates, or mechanistic frameworks such as Dynamic Energy Budget models.	1	Low
Fast-slow continuum	Simple regression or CMR models required to quantify some demographic parameters, but no extra modelling needed to categorize population/species as fast or slow-living organisms.	0-1	Low
Temporal covariation	Mixed-effect models and/or CMR models and covariance/correlation estimates. A more sophisticated approach involves multivariate CMR or integrated population models to account for uncertainty.	1	Medium
Demographic buffering-lability	i) Regression/CMR models to quantify time and stage-specific demographic parameters, and ii) structured population models to quantify buffering and lability. It is possible to quantify demographic parameters directly into population models (integrated population models) or from multivariate, hierarchical models.	2	Medium
Individual heterogeneity	Regression models and/or CMR with individual covariates required. More complex approaches can be applied to consider individual heterogeneity as a latent variable (e.g., finite mixture models or mixed effect models quantifying random individual effects). Individual-based models are also used.	1	Low
Transient dynamics	i) Regression/CMR models to quantify time and stage-specific demographic parameters, and ii) structured population models to quantify resilience and transient metrics	2	Medium

S9 – Full search queries and template of the table summarizing the screening criteria

Life-history trade-offs

(ALL=((life-history OR life history) AND (trade-off* OR tradeoff*)) OR (survival AND (trade-off* OR tradeoff*)) OR (reproduction AND (trade-off* OR tradeoff*)) OR (growth AND (trade-off* OR tradeoff*)) OR (fitness AND (trade-off* OR tradeoff*)))

AND

SO=("conservation letters" OR "nature ecology & evolution" OR "Frontiers in Ecology and the Environment" OR "Conservation biology" OR "Biological conservation" OR "One earth" OR "Ambio" OR "People and nature" OR "Ecological applications" OR "Ecology and society" OR "Animal conservation" OR "Biodiversity and Conservation" OR "Biodiversity & Conservation" OR "Basic and applied ecology" OR "Biological invasions" OR "Endangered species research" OR "Oryx" OR "Conservation physiology" OR "Conservation science and practice" OR "Aquatic conservation: Marine and Freshwater Ecosystems" OR "Environmental conservation" OR "Conservation genetics" OR "Journal of insect conservation" OR "Bird conservation international" OR "Conservation & society" OR "Ecological Management & Restoration" OR "Journal for Nature Conservation" OR "Tropical Conservation Science" OR "Journal of applied ecology" OR "Emu – Austral Ornithology" OR "Emu" OR "Global Change Biology" OR "The Journal of Wildlife Management" OR "Insect Conservation and Diversity" OR "Restoration ecology")

Fast-slow continuum

ALL=("slow-fast continuum") OR ALL=("fast-slow continuum") OR ALL=("fast-slow life-history continuum") OR ALL=("variation in life history strategies") OR ALL=("Variation in life history traits") OR ALL=("life-history strategies") OR ALL=("life history strategies") OR ALL=("life-history strategy") OR ALL=("life history strategy") OR ALL=("axis of life history strategies") OR ALL=("r-/K-Selected") OR ALL=("r-K Strategies") OR ALL=("K-selected species") OR ALL=("r-selected species") OR ALL=("r-Strategy") OR ALL=("K-Strategy") OR ALL=("fast-living species") OR ALL=("slow-living species") OR ALL=("fast life histories") OR ALL=("slow life histories") OR ALL=("fast life history") OR ALL=("slow life history") OR ALL=("fast-slow life history") OR ALL=("slow-fast life history") OR ALL=("pace of life continuum") OR ALL=("pace of life") OR ALL=("fast-slow life history spectrum")

AND

SO=("conservation letters" OR "nature ecology & evolution" OR "Frontiers in Ecology and the Environment" OR "Conservation biology" OR "Biological conservation" OR "One earth" OR "Ambio" OR "People and nature" OR "Ecological applications" OR "Ecology and society" OR "Animal conservation" OR "Biodiversity and Conservation" OR "Biodiversity & Conservation" OR "Basic and applied ecology" OR "Biological invasions" OR "Endangered species research" OR "Oryx" OR "Conservation physiology" OR "Conservation science and practice" OR "Aquatic conservation: Marine and Freshwater Ecosystems" OR "Environmental conservation" OR "Conservation genetics" OR "Journal of insect conservation" OR "Bird conservation international" OR "Conservation & society" OR "Ecological Management & Restoration" OR "Journal for Nature Conservation" OR "Tropical Conservation Science" OR "Journal of applied ecology" OR "Emu – Austral Ornithology" OR "Emu" OR "Global Change Biology" OR "The Journal of Wildlife Management" OR "Insect Conservation and Diversity" OR "Restoration ecology")

Temporal covariation

(ALL=("temporal correlation\$" AND "demographic parameter\$") OR ALL=("temporal correlation\$" AND "demographic rate\$") OR ALL=("temporal correlation\$" AND "vital rate\$") OR ALL=("correlation\$" AND "demographic parameter\$") OR ALL=("correlation\$" AND "parameter") OR ALL=("correlation\$" AND "vital rate\$") OR ALL=("demographic correlation\$") OR ALL=("vital rate correlation\$") OR ALL=("temporal covariation\$" AND "demographic parameter\$") OR ALL=("temporal covariation\$" AND "demographic rate\$") OR ALL=("temporal covariation\$" AND "vital rate\$") OR ALL=("temporal covariation\$" AND "parameter\$") OR ALL=("covariation\$" AND "demographic parameter\$") OR ALL=("covariation\$" AND "vital rate\$") OR ALL=("demographic covariation\$") OR ALL=("vital rate covariation\$") OR ALL=("temporal covariance\$" AND "demographic parameter\$") OR ALL=("temporal covariance\$" AND "demographic rate\$") OR ALL=("temporal covariance\$" AND "vital rate\$") OR ALL=("temporal covariance\$" AND "parameter\$") OR ALL=("covariance\$" AND "demographic parameter\$") OR ALL=("covariance\$" AND "vital rate\$") OR ALL=("demographic covariance\$") OR ALL=("vital rate covariance\$") OR ALL=("demographic compensation"))

AND

SO=("conservation letters" OR "nature ecology & evolution" OR "Frontiers in Ecology and the Environment" OR "Conservation biology" OR "Biological conservation" OR "One earth" OR "Ambio" OR "People and nature" OR "Ecological applications" OR "Ecology and society" OR "Animal conservation" OR "Biodiversity and

Conservation" OR "Biodiversity & Conservation" OR "Basic and applied ecology" OR "Biological invasions" OR "Endangered species research" OR "Oryx" OR "Conservation physiology" OR "Conservation science and practice" OR "Aquatic conservation: Marine and Freshwater Ecosystems" OR "Environmental conservation" OR "Conservation genetics" OR "Journal of insect conservation" OR "Bird conservation international" OR "Conservation & society" OR "Ecological Management & Restoration" OR "Journal for Nature Conservation" OR "Tropical Conservation Science" OR "Journal of applied ecology" OR "Emu – Austral Ornithology" OR "Emu" OR "Global Change Biology" OR "The Journal of Wildlife Management" OR "Insect Conservation and Diversity" OR "Restoration ecology")

Demographic buffering and lability

(ALL=("Environmental canalization" OR "Canalized fitness component" OR "Canalization of survival" OR "Canalization of fertility" OR "Life-history buffering" OR "Life history buffering" OR "Demographic buffering" OR "Buffering hypothesis" OR "Buffering of demographic rates" OR "Buffering of demographic parameters" OR "Buffering in the vital rates" OR "Buffered against environmental variability" OR "temporal variation in vital rates" OR "temporal variation in demographic parameters" OR "Demographic lability" OR "life history lability" OR "life-history lability" OR ("lability" AND "vital rates") OR ("lability" AND "demographic parameters") OR "Labile fertility" OR "Labile demography"))

AND

SO=("conservation letters" OR "nature ecology & evolution" OR "Frontiers in Ecology and the Environment" OR "Conservation biology" OR "Biological conservation" OR "One earth" OR "Ambio" OR "People and nature" OR "Ecological applications" OR "Ecology and society" OR "Animal conservation" OR "Biodiversity and Conservation" OR "Biodiversity & Conservation" OR "Basic and applied ecology" OR "Biological invasions" OR "Endangered species research" OR "Oryx" OR "Conservation physiology" OR "Conservation science and practice" OR "Aquatic conservation: Marine and Freshwater Ecosystems" OR "Environmental conservation" OR "Conservation genetics" OR "Journal of insect conservation" OR "Bird conservation international" OR "Conservation & society" OR "Ecological Management & Restoration" OR "Journal for Nature Conservation" OR "Tropical Conservation Science" OR "Journal of applied ecology" OR "Emu – Austral Ornithology" OR "Emu" OR "Global Change Biology" OR "The Journal of Wildlife Management" OR "Insect Conservation and Diversity" OR "Restoration ecology")

Individual heterogeneity

ALL=("individual heterogeneity" OR "individual quality" OR "frailty" OR "among individual variation" OR "personality" OR "individual behavior" OR "individual behaviour" OR "temperament")

AND

SO=("conservation letters" OR "nature ecology & evolution" OR "Frontiers in Ecology and the Environment" OR "Conservation biology" OR "Biological conservation" OR "One earth" OR "Ambio" OR "People and nature" OR "Ecological applications" OR "Ecology and society" OR "Animal conservation" OR "Biodiversity and Conservation" OR "Biodiversity & Conservation" OR "Basic and applied ecology" OR "Biological invasions" OR "Endangered species research" OR "Oryx" OR "Conservation physiology" OR "Conservation science and practice" OR "Aquatic conservation: Marine and Freshwater Ecosystems" OR "Environmental conservation" OR "Conservation genetics" OR "Journal of insect conservation" OR "Bird conservation international" OR "Conservation & society" OR "Ecological Management & Restoration" OR "Journal for Nature Conservation" OR "Tropical Conservation Science" OR "Journal of applied ecology" OR "Emu – Austral Ornithology" OR "Emu" OR "Global Change Biology" OR "The Journal of Wildlife Management" OR "Insect Conservation and Diversity" OR "Restoration ecology")

Transient dynamics

(ALL=("transient demography" OR "transient dynamics" OR "demographic resilience" OR "short-term population dynamics"))

AND

SO=("conservation letters" OR "nature ecology & evolution" OR "Frontiers in Ecology and the Environment" OR "Conservation biology" OR "Biological conservation" OR "One earth" OR "Ambio" OR "People and nature" OR "Ecological applications" OR "Ecology and society" OR "Animal conservation" OR "Biodiversity and Conservation" OR "Biodiversity & Conservation" OR "Basic and applied ecology" OR "Biological invasions" OR "Endangered species research" OR "Oryx" OR "Conservation physiology" OR "Conservation science and practice" OR "Aquatic conservation: Marine and Freshwater Ecosystems" OR "Environmental conservation" OR "Conservation genetics" OR "Journal of insect conservation" OR "Bird conservation international" OR "Conservation & society" OR "Ecological Management & Restoration" OR "Journal for Nature Conservation" OR "Tropical Conservation Science" OR "Journal of applied ecology" OR "Emu – Austral Ornithology" OR "Emu" OR "Global Change Biology" OR "The Journal of Wildlife Management" OR "Insect Conservation and Diversity" OR "Restoration ecology")

Template of the table summarizing the screening criteria used for the six concepts.

Additional columns specific to each concept were added (e.g., definition use [Fitness or Non-Fitness related trait(s)] and type of heterogeneity [latent or measured] for individual heterogeneity; Life-history traits involved in trade-offs for LHTO). See details in Appendix S2-S7. Y or N: Yes or No

Author names	Year of publication	Journal	Relevant abstract for the concept? [Y or N]	Keywords used in the article (related to the concept)	Is the article a review? [Y or N]	Population type [wild, captive, experimental]	Is the main purpose of the article conservation-related? [Y or N]	Is the concept discussed or quantified?	Organismal groups [mammalia, reptilia, actinopterygii, amphibia, magnoliopsida, etc]	Species name	Species status [invasive, exploited, threatened or other conservation interest]	Conservation application category	Short description of the study