

1   **Changing the narrative: encroached savannas are not forest**

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18   **Author contributions**

19   CLP, RDB, JMB and CEL devised the paper. CLP wrote the first draft. All authors contributed to  
20   subsequent drafts and approved the final version.

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23   **Conflict of Interest Statement**

24   The authors declare no competing interests

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26

27   **Abstract**

28

29   1. Savannas are globally important ecosystems but are often misclassified as forests because  
30   they can support high tree cover, leading to misguided management. This misunderstanding  
31   arises because the presence of grasses, a key defining component of savannas, critical for  
32   their structure and functioning, is overlooked.

33   2. Fundamental tree-based misunderstandings affect the interpretation of woody plant  
34   encroachment, a novel threat facing savannas, which now affects >5 million km<sup>2</sup> of savanna  
35   globally. High tree cover can degrade savannas by reducing grasses, altering fire regimes,  
36   and harming biodiversity and livelihoods.

37   3. We highlight that savannas can naturally vary in woody cover, and therefore high tree cover  
38   does not necessarily equate to an encroached state. We also clarify that even in an  
39   encroached state, high tree cover does not necessarily create forests. We identify three end-  
40   states, determined by rainfall and proximity to true forest, that may emerge as an outcome  
41   of encroachment: encroached, novel savanna, hybrid forest and true forest.

42   4. **Synthesis** Forests should not be viewed as superior to savannas. Savannas are biodiversity  
43   hotspots and vital for human survival. Effective management requires moving beyond  
44   structural definitions, understanding thresholds, and assessing ecosystem value based on  
45   biodiversity and function, not tree cover alone.

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50 **1. Savannas and Woody Plant Encroachment**

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52 Savannas cover more than 20% land area globally, support more than a billion people worldwide  
53 and are important reservoirs for unique biodiversity (Lehmann & Parr, 2016). These environments  
54 are characterized by a near-continuous C4 herbaceous ground-layer (Ratnam et al., 2011), and  
55 although they can support trees, tree cover varies greatly in both space and time (Sankaran et al.,  
56 2005; Ratnam et al., 2011). Fire and herbivores play major roles in savannas, generating important  
57 feedbacks that maintain an open and high light environment critical for the plants and animals these  
58 systems support (Bond & Keeley 2005; Ratnam et al., 2011). In contrast, tropical forests are low-  
59 light environments with a ground layer dominated by litter (not grass) (Ratnam et al., 2011). These  
60 two systems are compositionally, structurally and functionally distinct (Flake et al., 2022), but both  
61 are important and unique. Critically, what distinguishes these systems from each other is the  
62 presence or absence of a grassy understory, not trees (i.e. high tree cover/presence ≠ forest).  
63

64 Relative to forests, from a scientific, policy and lay perspective, savannas have been misunderstood  
65 and overlooked (Bardgett et al., 2021; Hughes et al., 2021; Overbeck et al., 2015; Pillar & Overbeck,  
66 2025; Silveira et al., 2022,). Although both forests and savannas are threatened, because for  
67 centuries trees and forested ecosystems have received disproportionate attention (Pilon et al., 2025;  
68 Silveira et al., 2022,), savannas, like other tropical grassy ecosystems, have been misclassified and  
69 marginalized (Parr et al., 2014). In regions where savannas support high tree number (e.g. miombo  
70 in Africa, savannas of India), they are often mistaken for forest with the important grassy understory  
71 overlooked (Bond et al., 2019; Gopalakrishna et al., 2024); this misclassification as forest can lead  
72 to the potential mismanagement of savannas, for instance through the suppression of fire (e.g.  
73 Kumar et al., 2020).

74 In contrast to well-documented deforestation and tree loss in forests (e.g., GlobalForestWatch.org;  
75 Hoang & Kanemoto, 2021; Li et al., 2022), savannas worldwide are undergoing large increases in  
76 tree cover. Tree cover is increasing via active planting, which is generally linked to afforestation  
77 schemes for restoration and climate mitigation (Parr et al., 2024), but also via a more passive  
78 process of woody thickening (i.e. an increase in tree cover and density) referred to as woody plant  
79 encroachment (WPE; IPCC, 2019). Woody plant encroachment is pervasive, currently estimated to  
80 be occurring over > 5 million km<sup>2</sup>, ~ 7% of global land area (Garcia Criado et al., 2020; Stevens et  
81 al., 2016; Stevens et al., 2022; Venter et al., 2018,). The drivers of the encroachment process include  
82 both local land management practices (e.g. fire suppression, overgrazing) and global drivers  
83 (enhanced atmospheric CO<sub>2</sub>) (Archer et al., 2017; Buitenwerf et al., 2012; Wigley et al., 2010),  
84 although direct attribution at a site level is often difficult. Nevertheless, evidence is mounting that the  
85 increase in trees associated with woody plant encroachment can have detrimental impacts on  
86 savanna biodiversity (White, et al. 2024; Wieczokowski & Lehmann, 2022), water availability (Honda  
87 & Durigan, 2016), fire occurrence (Abades et al., 2014; Archibald et al., 2009) and human livelihoods  
88 (Luvuno et al., 2022; White et al., 2022). Despite this, WPE has received relatively little attention  
89 among scientists and policy makers partly because the notion of trees as a threat runs counter to  
90 most societies' norms and values (e.g., Esperon-Rodriguez, 2025; Jones & Cloke, 2002; O'Brien,  
91 2005; O'Brien et al., 2024; Schroeder, 1995).

92  
93 While increasing tree cover has been widely documented across savannas, as have the drivers  
94 involved (see above), far less is known about the mechanisms underlying this process and its  
95 consequences for biodiversity and ecosystem function. Therefore, to provide a road map to better  
96 understanding of the mechanisms and consequences, here we: (1) unpick and clarify some of the  
97 misunderstandings about savannas and WPE; (2) identify outstanding questions concerning the

99 distribution and consequences of WPE; and (3) offer testable hypotheses to help guide future  
100 research efforts on WPE. In doing so we critically examine the idea that high tree density creates  
101 forests and introduce the novel hypothesis that increasing tree cover in savannas does not lead to  
102 one outcome but can result in three different end states, namely encroached savanna, hybrid forest  
103 and true forest that differ structurally, functionally, and compositionally, and depend on whether there  
104 has been an increase in the density of savanna woody species or forest species invading savanna.  
105 Moreover, we argue that one of these states, 'encroached savanna', represents a novel and  
106 depauperate ecosystem state of reduced biodiversity and functioning.  
107

## 108 **2. Increasing trees doesn't necessarily equate to an encroached state**

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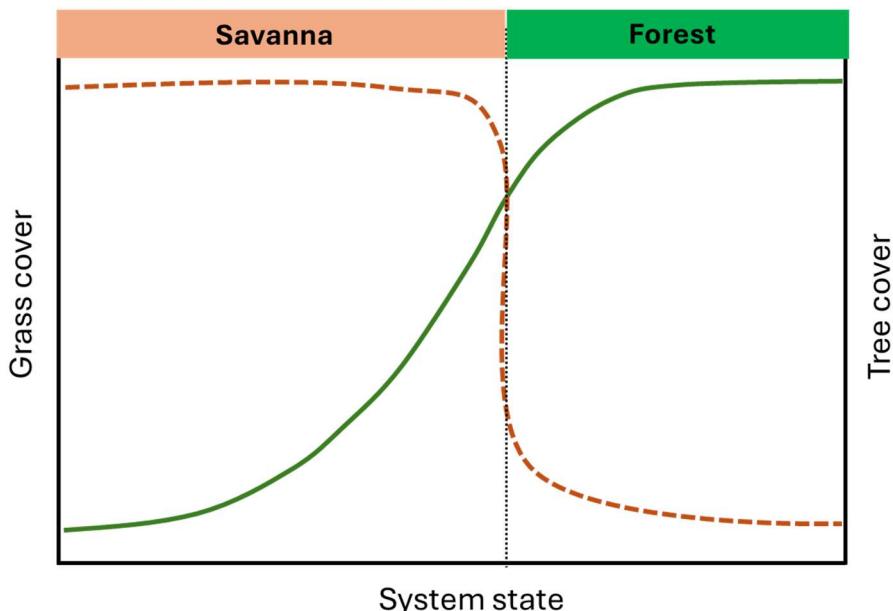
110 In its simplest form, wood thickening is the gain of trees via an increase in density and/or cover  
111 (Archer et al., 2017). The terms thickening and encroached are, however, often used  
112 interchangeably in studies on tree increases in savannas although some studies focus on  
113 documenting increases in tree cover (the process, e.g. Venter et al., 2018), while others focus on  
114 the altered (degraded) state that occurs at high tree cover (end-state, e.g. Sirami et al., 2009;  
115 Wieczorkowski & Lehmann, 2022). We argue that greater consideration and care is needed to avoid  
116 conflation. Distinguishing between the process (woody thickening via increases in tree cover) and  
117 the end-state (encroached) is necessary for interpreting findings.

118 Savannas can exhibit huge natural variation in tree cover both spatially and temporally (for example  
119 from 5 to 80% tree cover, Parr et al., 2014; Ratnam et al., 2011; Sankaran et al., 2005), as a function  
120 of processes that govern tree growth and death, such as fire, animals, drought and rainfall, along  
121 with global changes in climate and atmospheric CO<sub>2</sub> (Aleman et al., 2017; Gopalakrishna et al., 2024;  
122 Sankaran et al., 2005). Therefore, the gain of trees, or woody thickening, is not necessarily  
123 problematic - tree cover can fluctuate as part of natural system flux without any notable impact on  
124 biodiversity and functioning. However, of concern is that woody thickening can result in an  
125 encroached state where some upper threshold has been passed and there is a shift in system state  
126 and associated functioning (i.e. the system is operating outside normal bounds) (Fig 1). Critically, in  
127 savannas, if there are too many trees, C4 grasses decline, precipitating a system state shift with  
128 deleterious consequences for biodiversity and ecosystem function.

129  
130 C4 grasses are highly efficient under high light conditions due to the structure of their photosynthetic  
131 machinery (Wasilewska-Debowska et al., 2022). However, C4 grasses are sensitive to thresholds in  
132 tree cover and associated shading (Abdallah et al., 2016; Charles-Dominique et al., 2018; Pilon et  
133 al., 2021) because photosynthetic efficiency of C4 grasses declines with shading (Tazoe et al. 2008;  
134 Pignon et al. 2017). Consequently, when tree cover increases and shade thresholds are exceeded,  
135 the ground layer qualitatively shifts from grass to litter dominance; C4 grasses lose resilience to  
136 disturbance due to reduced belowground allocation inhibiting resprouting capacity (Pilon et al., 2021)  
137 and decline in competitiveness leading to their loss from the ecosystem (e.g., Archer et al. 1995).  
138

139 The transition from grass to a litter-based state is associated with the loss of key savanna consumers  
140 (fire and herbivory) (Abades et al., 2014; Hoffmann et al., 2011; Ratajczak et al., 2014; Sala &  
141 Maestre, 2014), and detrimental consequences for biodiversity and some ecosystem services  
142 (Bardgett et al., 2021); note that this is in contrast with the situation in litter-based systems (e.g.,  
143 shrublands or forests) where gaining trees is less problematic because there is no change in the  
144 ground layer structure or function, and therefore no state shift. Currently, for savannas, the threshold  
145 in woody cover where an encroached (degraded) state is reached is unknown but likely varies with  
146 environmental and ecological context.

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148  
149 **Figure 1.** Savanna system state switch that results from increasing tree cover and decreasing C4  
150 grass cover (here shown for transition to forest).

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154 **3. Gaining of trees frequently does not result in a forest.**

155

156 Many widely used definitions of forest are structural, focusing on trees and canopy cover (e.g., FAO  
157 and UNFCCC state forests have a min 10% tree cover). There is, therefore, the risk that the process  
158 of woody thickening in savannas will be misconstrued as resulting in forest; in other words that the  
159 encroached state is forest because systems with high tree cover are classified as forest.

160 The misconception that more trees equate to forest is reinforced by the conventional and prevailing  
161 view that across vast swathes of the Tropics, vegetation exists as either forest or savanna (as  
162 alternative stable states or alternative stable biomes) (e.g., Aleman et al., 2021; Pausas & Bond,  
163 2020; Staver et al., 2011). The widespread use of structural measures (i.e. tree cover) combined  
164 with remote sensing to map trees further reinforces this view. However, evidence for the alternative  
165 stable states of forest and savanna is scarce, particularly once species floristics are considered  
166 (Higgins et al., 2025, although see Wieczorkowski et al., 2024). A structural perspective to classifying  
167 vegetation fails to recognize that the key defining parts of ecosystems are not always trees;  
168 characteristics of the ground layer (e.g. presence of a C4 grassy understory) and, therefore, key  
169 processes (e.g. fire) are arguably more important in determining savannas (e.g. IUCN global  
170 ecosystem typology, Keith et al., 2022).

171

172 The ‘more trees = forest’ misunderstanding may also arise because the term ‘encroachment’  
173 suggests movement in, or advance of, trees from elsewhere and therefore forest expansion; for  
174 example, the Merriam-Webster dictionary defines encroachment as “*to advance beyond usual  
175 limits*”. In some regions, forest expansion does occur (examples in: Mitchard & Flintrop, 2013), but  
176 for most areas an encroached state results from an increase in density of pre-occurring savanna  
177 tree species and therefore forest expansion is neither the process, nor is forest the end-state.

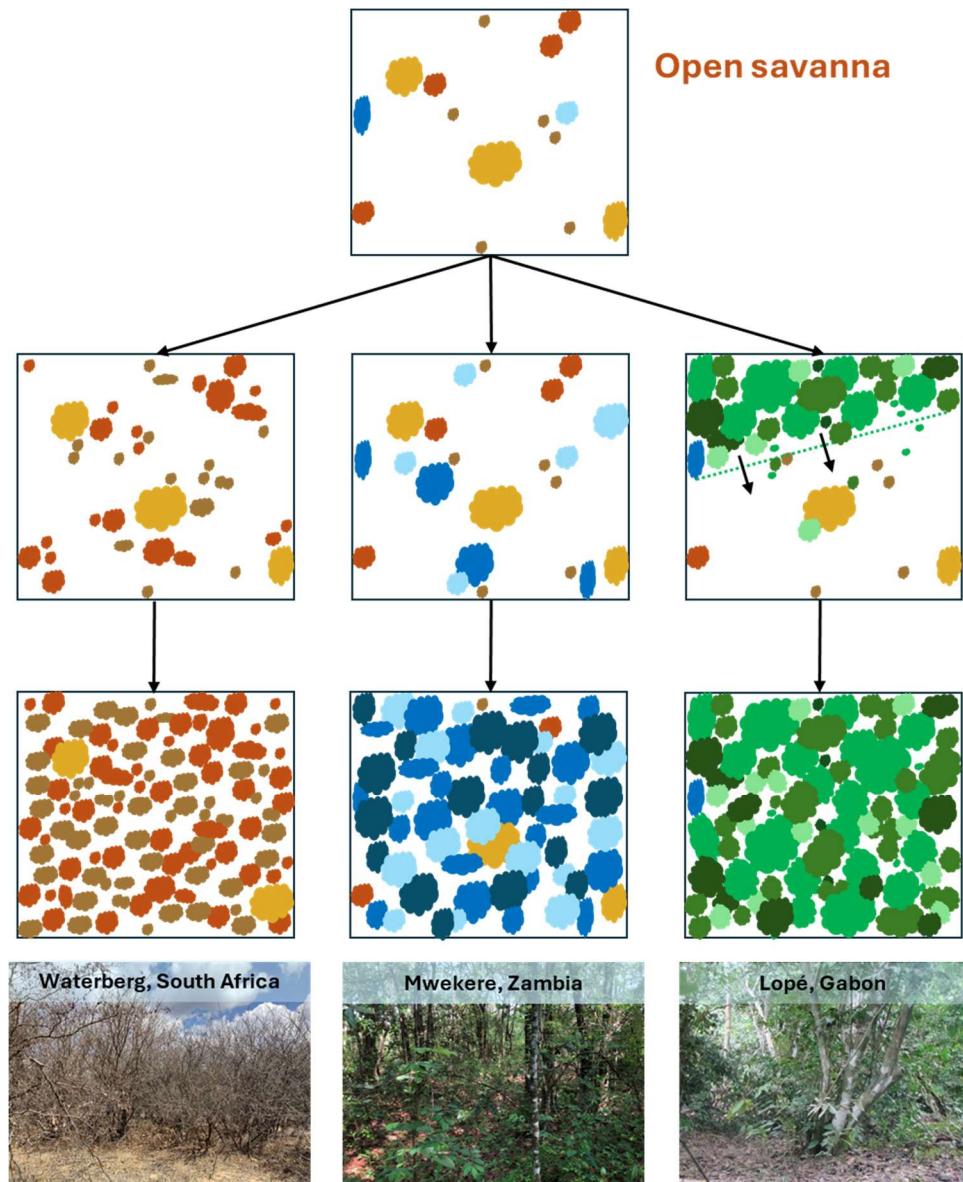
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179 Here we argue that from a compositional and functional perspective, three system states are  
180 possible as a consequence of WPE (or increase in tree cover). Furthermore, with increasing tree  
181 cover there can be quite different end-states depending on whether there has been an increase in

182 the density of savanna woody species, or forest species invading savanna (e.g. expansion of forest).  
183 We propose that with increasing tree cover, the system outcome depends principally on total rainfall  
184 (as a proxy for productivity in these tropical systems) and distance to forest (as a seed source of  
185 forest associated species) (Fig. 2). The drivers can differ with location – for example, grazing is a  
186 more important consumer at lower rainfalls, than at higher rainfalls where the herbaceous layer tends  
187 to be less palatable. We argue that there are three broad end-states, with all representing system  
188 state shifts that differ structurally, functionally and/or compositionally (Fig. 2):  
189

190 State 1: In many regions, encroached savannas are characterized by hyper-dominance of one or  
191 two woody species. These are often shrubs relatively low in stature; for example, dominance of  
192 *Dichrostachys*, *Terminalia* or *Acacia* in many regions in sub-Saharan Africa (e.g., Sirami et al. 2009,  
193 Leitner et al. 2018, Bora et al. 2021). These woody species can grow at very high densities such  
194 that the vegetation can be almost impenetrable. Importantly, this state is not forest because, although  
195 the C4 grassy understory is lost, the tree species are floristically classified as savanna species (i.e.,  
196 not forest-associated) (Fig. 2). Indeed, even at very high tree density and cover, forest tree species  
197 are not found because rainfall is too low, and true forest is too distant geographically (Fig. 2).  
198 Consequently, the biota represents a depauperate, nested subset of the intact savanna. These  
199 savannas are dysfunctional because the lack of a (C4) grass layer means they can no longer support  
200 key processes of fire and large mammalian herbivory (Keith et al., 2022). In sum, they are  
201 functionally, compositionally, and often structurally, distinct from true forest. From a compositional  
202 perspective, we term these novel ecosystems '**encroached savanna**' (Fig. 2). We hypothesize  
203 these regions are found in drier savanna regions (650-700 mm and less) where rainfall is too low to  
204 support forest species (Fig. 3). We predict their extent is likely to be considerable on some continents  
205 where large areas of more arid savanna occur (e.g. Africa).  
206

207 State 2: We identify a depauperate forest-like state in wet savanna regions (>800mm) that are not  
208 near true forest (Fig. 3). This state is intermediate to high rainfall savanna and true forest. In these  
209 regions, the increase in tree cover (e.g. with suppression of fire) can result in nested community with  
210 an increase in forest-associated species that already exist in the wider savanna landscape (e.g. fire  
211 sensitive trees in riparian zones, savanna-forest edge species) and more shade-tolerant savanna  
212 tree species, and a loss of light-loving, fire-resistant savanna species (Fig. 2). Structurally these  
213 systems resemble a forest (e.g., high tree cover, tall trees, shady, absence of C4 grassy understory),  
214 but they lack true forest species because distance from forest is a major dispersal limitation (Fig. 3).  
215 We therefore call this state a '**hybrid forest**' (Fig. 2). This state can occur in wetter savanna regions  
216 such as the miombo savanna in southern Africa (Wieczokowski et al., 2024), or in the Cerrado in  
217 Brazil (Moreira, 2000) where fire is suppressed. Smaller areas of this state may occur naturally linked  
218 to topography and the occurrence of natural barriers (e.g. lower-lying wetter areas, or areas  
219 protected from fire) or as a transition state before succession to forest. This state is likely to cover a  
220 smaller area than States 1 or 3.  
221



222  
223

	<b>State 1: Encroached savanna</b>	<b>State 2: Hybrid forest</b>	<b>State 3: Forest</b>
<b>C4 grass layer</b>	Absent	Absent	Absent
<b>Fire &amp; vertebrate herbivory</b>	Decline	Decline	Fire = absent
<b>Trees: Savanna spp.</b>	1-2 species hyperdominant	Few savanna species	No savanna species
<b>Trees: Forest spp.</b>	None	None?	Common
<b>Similarity to open savanna state:</b>			
<b>Structure</b>	Partly	No	No
<b>Function</b>	No	No	No
<b><math>\beta</math> Diversity</b>	Nested	Nested	Turnover
<b>Driver of change</b>	Over/undergrazing, fire suppression, CO <sub>2</sub>	Fire suppression, CO <sub>2</sub>	Fire suppression, CO <sub>2</sub>

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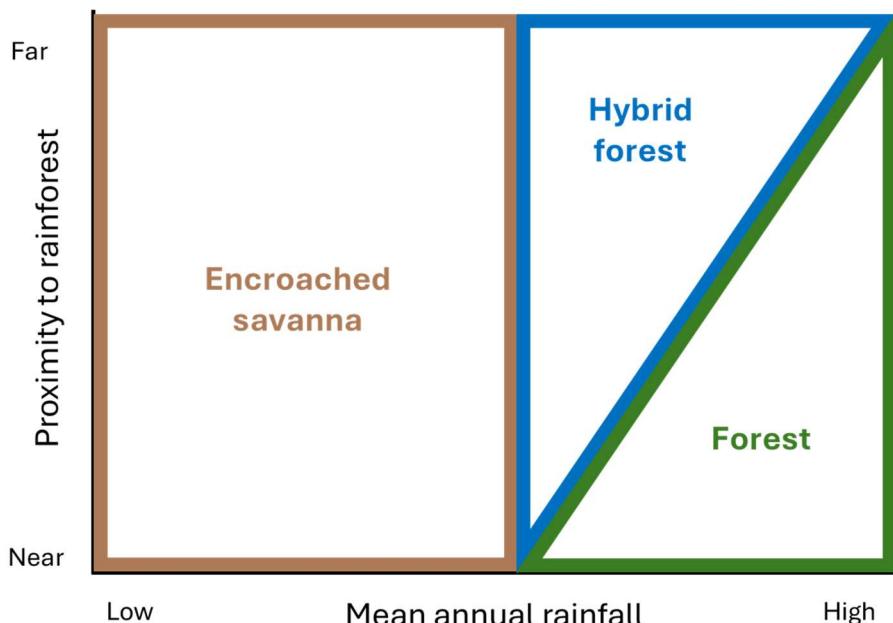
225 **Figure 2.** Potential pathways and state outcomes with increasing tree cover in savannas: 1: Encroached savanna, 2: Hybrid forest, 3: True forest. Brown and yellow colours = savanna tree species, blue colours = forest-associated tree species, green colours = true forest tree species. Only tree species are shown. White = understory, principally either grass or litter. Example locations for different end-states shown in photos.

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234 **Figure 3.** The occurrence of different encroached savanna states is determined by rainfall and 235 proximity to tropical rainforest. The extent of encroached savanna is limited by rainfall, while hybrid 236 forest is limited principally by distance to forest (i.e. reduced opportunity for rainforest propagules to 237 disperse). The occurrence of forest requires high rainfall and close proximity to forest (movement of 238 propagules possible where dispersal community is intact).

239

240

241

242 **State 3:** Finally, in regions where savanna-grasslands occur in mosaics or close proximity with 243 tropical forest and rainfall is high, increasing tree cover in savannas can result in a system switch to 244 **true [tropical rainforest] forest** (Figs. 2 & 3). This is because environmental conditions can support 245 forests (e.g. sufficient rainfall) and there is a pool of forest species that can disperse into the open 246 ecosystems. We predict turnover of tree species and loss of C4 grasses and forbs with the shift to 247 true forest (i.e. with true forest species) (Fig. 2). We anticipate this situation is common where forests 248 expand into savanna (Figs. 2 & 3), particularly along savanna-forest boundaries (e.g., in Brazil 249 [Abreu et al., 2021] and in Gabon [Cardoso et al., 2021], Fig. 3). In these regions, forest is gained, 250 but the important and unique biodiversity of high rainfall savannas is lost. Fire suppression and CO<sub>2</sub> 251 are common drivers of this change in state.

252

#### 253 **4. Forest is not the ideal state**

254 Finally, the lack of understanding about tree gain in savannas is impeded by the general belief that 255 more trees are a good thing *per se*. Trees are particularly valued biota (Dove, 2004; Wall, 2022), 256 with forest often seen as the preferred or ideal state (Duvall et al., 2018; Pillar & Oberbeck, 2025; 257 Stott, 1999; Silveira et al., 2022). These ideas have centuries' long history and have been 258

259 perpetuated by colonialists, politicians, philosophers, and artists (Davis, 2016 and references within,  
260 Ratnam et al., 2016). Indeed, forests were promoted by colonial powers as the desired state; the  
261 low tree cover they encountered in open, grassy ecosystems prompted colonialists to describe these  
262 savanna landscapes as 'wastelands', 'degraded' and even 'wretched' (Bardgett et al., 2021; Davies,  
263 2016). Views from Western scientists in temperate, well-forested regions have also had a powerful  
264 influence on ecology. For example, in ecology, theory on vegetation dynamics has been driven by  
265 observations of change over time primarily derived from temperate ecosystems, or turnover of  
266 ecosystems along elevational gradients (Moret et al., 2019; Vera, 2000). Many of these ecological  
267 ideas were uncritically transferred to tropical regions: Frederick Clement's (1916) theory of  
268 succession promoted the view that the ideal climax community is forest, and savanna was seen as  
269 a non-climax deviation, the 'savanna problem' (Sarmiento, 1984; Sayre, 2017; Veldman, 2016) –  
270 despite the fact that the prevalence of disturbance is supported through the productivity and  
271 seasonality of climates. Collectively, this thinking has led to flawed notions that forested ecosystems  
272 are somehow superior intrinsically, for the services they deliver and the biodiversity they support  
273 (Murphy et al., 2016; Pillar & Overbeck, 2025).

274  
275 Half of the world's biodiversity hotspots are in grassy systems; they support more people than  
276 forested regions, play key roles in earth-atmospheric process and global net primary productivity,  
277 and are the cradle of humankind (Lehmann & Parr, 2016). There is no basis for concluding forests  
278 are 'better' - they are simply one of multiple biomes on Earth, each of which contributes to the Earth  
279 System in all its facets. Encroachment should be assessed relative to the value more trees bring –  
280 for example, in terms of ecosystem services or biodiversity. Although encroached savannas with  
281 their high density of trees are commonly classified as 'forest', these depauperate and dysfunctional  
282 savannas (States 1 and 2 above) are often of less value than intact, functioning savannas when  
283 considered in terms of biodiversity (less biodiverse), carbon (likely less carbon), and ecosystem  
284 services (reduced or altered provision), and considerably less value than natural forest. We therefore  
285 urge that value is not simply a function of tree cover but instead is determined by the extent to which  
286 biodiversity, ecosystem services and other values are supported (Bardgett et al., 2021). It is critical  
287 we consider both what is lost, as well as what is gained.

288  
289 **5. Conclusion**  
290  
291 Here we consider how increased tree cover can result in three structurally, functionally and  
292 compositionally different end-states as opposed to a single end-state of forest (Figs. 2 & 3). Yet, we  
293 do not know the potential distribution of these different end-states and under what specific conditions  
294 there is potential shift to these different states (e.g. rainfall thresholds). Such information is sorely  
295 needed to understand the consequences of WPE more fully and manage landscapes effectively,  
296 especially in regions where wholesale system change could have major consequences for the  
297 biodiversity and people living in these areas (Lehmann & Parr 2016). Space-for-time studies  
298 sampling along gradients of tree cover in different regions would be a major advance to  
299 understanding where there is the potential to switch to forest, but is arguably more important, where  
300 novel, depauperate savanna systems result. While all three 'encroachment' states exist in Africa, we  
301 expect only States 2 and 3 in South America (because savannas there are wetter) and State 3 in  
302 Australia (because narrower trees canopy architecture means less shading and therefore States 1  
303 and 2 are less likely). We predict all three states will be possible in Asia due to similarities with Africa.  
304 To what extent similar states may exist, or be possible, in temperate regions requires further  
305 investigation.

306

307 The persistent assumption that high tree cover equates with forest is problematic for developing  
308 evidence-based conservation and restoration of savanna ecosystems. There is an urgent need to  
309 look beyond structural definitions of vegetation states, otherwise we risk misclassifying and  
310 managing ecosystems. The implications of switches in savanna state to novel, encroached  
311 savannas or hybrid forests are severe given that a large proportion of the global population depends  
312 directly on savannas for their existence/livelihood. Yet, many research challenges remain and will  
313 require future research to address them. Under global change, it is critical we work to understand  
314 the potential ecosystem trajectories better including the potential environmental niches of different  
315 states, the speed of change, the consequences for humans and the biosphere upon which we  
316 depend and the capacity for ecosystem-state reversal.

317

318

## 319 **References**

320

321 Abades, S.R., Gaxiola, A. & Marquet, P.A. (2014). Fire, percolation thresholds and the savanna  
322 forest transition: a neutral model approach. *Journal of Ecology*, **102**, 1386-1393.

323 Abdallah, F., Michalet, R., Maalouf, J.P., Ouled-Dhaou, S., Touzard, B., Noumi, Z. et al.  
324 (2016). Disentangling canopy and soil effects of a savanna tree species on its  
325 understorey. *Journal of Vegetation Science*, **27**, 771-779.

326 Abreu, R.C., Durigan, G., Melo, A.C., Pilon, N.A. & Hoffmann, W.A. (2021). Facilitation by isolated  
327 trees triggers woody encroachment and a biome shift at the savanna–forest transition. *Journal*  
328 *of Applied Ecology*, **58**, 2650–2660

329 Aleman, J., Blarquez, O., Gourlet-Fleury, S. et al. (2017). Tree cover in Central Africa: determinants  
330 and sensitivity under contrasted scenarios of global change. *Scientific Reports*, **7**, 41393.

331 Aleman, J.C., Fayolle, A., Favier, C. et al. (2020). Floristic evidence for alternative biome states in  
332 tropical Africa. *Proceedings of the National Academy of Sciences*, **117**, 28183-28190.

333 Archer, S.R. (1995). Tree-grass dynamics in a *Prosopis*-thornscrub savanna parkland:  
334 reconstructing the past and predicting the future. *Écoscience*, **2**, 83-99.

335 Archer, S.R., Andersen, E.M., Predick, K.I., Schwinning, S., Steidl, R.J., & Woods, S.R. (2017).  
336 Wood plant encroachment: causes and consequences. *Rangeland Systems*, pp. 25-84.  
337 Springer Series on Environmental Management.

338 Archibald, S., Roy, D.P., van Wilgen, B.W. & Scholes, R.J. (2009). What limits fire? An examination  
339 of drivers of burnt area in Southern Africa. *Global Change Biology*, **15**, 613-630.

340 Bardgett, R.D., Bullock, J.M., Lavorel, S., Manning, P. et al. (2021). Combatting global grassland  
341 degradation. *Nature Reviews Earth & Environment*, **2**, 720-735.

342 Bond, W.J., Stevens, N., Midgley, G.F. & Lehmann, C.E.R. (2019). The trouble with trees:  
343 afforestation plans for Africa. *Trends in Ecology & Evolution*, **34**, 963-965.

344 Bond, W. J. & Keeley, J. E. (2005). Fire as a global ‘herbivore’: the ecology and evolution of  
345 flammable ecosystems. *Trends in Ecology & Evolution*, **20**, 387–394.

346 Bora, Z., Wang, Y., Xu, X. Angassa, A. & You, Y. (2021). Effects comparison of co-occurring  
347 *Vachellia* tree species on understory herbaceous vegetation biomass and soil nutrient: case of  
348 semi-arid savanna grasslands in southern Ethiopia. *Journal of Arid Environments*, **190**, 104527.

349 Buitenwerf, R., Bond, W.J., Stevens, N. & Trollope, W.S.W. (2012). Increased trees densities in  
350 South African savannas: >50 years of data suggests CO<sub>2</sub> as a driver. *Global Change Biology*,  
351 **18**, 675-684.

352 Cardoso, A.W., Oliveras, I., Abernethy, K.A., Jeffery, K.J., Glover, S., Lehmann, D., Edzang Ndong,  
353 J., White, L.T., Bond, W.J., & Malhi, Y. (2021). A distinct ecotonal tree community exists at  
354 central African forest-savanna transitions. *Journal of Ecology*, **109**, 1170-1183.

355 Charles-Dominique, T., Midgley, G.F., Tomlinson, K.W. & Bond, W.J. (2018). Steal the light: shade  
356 vs fire adapted vegetation in forest–savanna mosaics. *New Phytologist*, **218**, 1419–1429.

357 Clements, F.E. (1916). *Plant succession: an analysis of the development of vegetation*. Carnegie  
358 Institute of Washington.

359 Dove, M.R. (2004). Anthropogenic grasslands in Southeast Asia: sociology of knowledge and  
360 implications for agroforestry. *Agroforestry Systems*, **61**, 423-435.

361 Davis, D.K. (2016). *The Arid Lands: History, Power, Knowledge*. The MIT Press. Pp. 296.

362 Duvall, C.S., Butt, B. & Neely, A. (2018). The trouble with savanna and other environmental  
363 categories, especially in Africa. *The Palgrave Handbook of Critical Physical Geography* (eds.  
364 Lavem R., Biermann, C., Lane, S.), pp. 107-127. Springer Nature.

365 Esperon-Rodriguez, M. (2025). Trees, society, and the path toward resilient ecosystems. *Plants,  
366 People, Planet*, **7**, 889-893.

367 Flake, S.W., Honda, E.A., Pilon, N.A., Hoffmann, W.A., & Durigan, G. (2022). Not all trees can make  
368 a forest: tree species composition and competition control forest encroachment in a tropical  
369 savanna. *Journal of Ecology*, **110**, 301-312.

370 Garcia Criado, M., Myers-Smith, I.H., Bjorkman, A.D., Lehmann, C.E.R. & Stevens, N. (2020).  
371 Woody plant encroachment intensifies under climate change across tundra and savanna  
372 biomes. *Global Ecology & Biogeography*, **29**, 925-943.

373 Gopalakrishna, T., Rifai, S.W., Ratnam, J. et al. (2024). The distribution and drivers of tree cover in  
374 savannas and forests across India. *Communications Earth & Environment*, **5**, 399.

375 Higgins, S.I., Banerjee, S., Baudena, M., Bowman, D.M.J.S. et al. (2025). Reassessing the  
376 alternative ecosystem states proposition in the African savanna-forest domain. *New Phytologist*,  
377 **243**, 1660-1669.

378 Hoffmann, W.A., Jaconis, S.Y., McKinley, K.L., Geiger, E.L., Gotsch, S.G. & Franco, A.C. (2012).  
379 Fuels or microclimate? Understanding the drivers of fire feedbacks at savanna-forest  
380 boundaries. *Austral Ecology*, **37**, 634-643,

381 Honda, E.A. & Durigan, G. (2016). Woody encroachment and its consequences on hydrological  
382 processes in the savannah. *Philosophical Transactions of the Royal Society London B*, **371**,  
383 20150313.

384 Hoang, N.T. & Kanemoto, K. (2021). Mapping the deforestation footprint of nations reveals growing  
385 threat to tropical forests. *Nature Ecology & Evolution*, **5**, 845–853.

386 Hughes, A.C., Orr, M.C., Ma, K., Costello, M.J., Waller, J. et al. (2021). Sampling biases shape our  
387 view of the natural world. *Ecography*, **44**, 1259-1269.

388 IPCC (2019). *Climate Change and Land: an IPCC special report on climate change, desertification,  
389 land degradation, sustainable land management, food security, and greenhouse gas fluxes in  
390 terrestrial ecosystems* [P.R. Shukla, J. Skea, E. Calvo Buendia, V. Masson-Delmotte, H.-O.  
391 Pörtner, D. C. Roberts, P. Zhai, R. Slade, S. Connors, R. van Diemen, M. Ferrat, E. Haughey,  
392 S. Luz, S. Neogi, M. Pathak, J. Petzold, J. Portugal Pereira, P. Vyas, E. Huntley, K. Kissick, M.  
393 Belkacemi, J. Malley, (eds.)].

394 Jones, O. & Cloke, P. (2002). *Tree Cultures: The Place of Trees and Trees in Their Place*. Taylor &  
395 Francis.

396 Keith, D.A., Ferrer-Paris, J.R., Nicholson, E. et al. (2022). A function-based typology for Earth's  
397 ecosystems. *Nature*, **610**, 513–518.

398 Kumar, D., Pfeiffer, M., Gaillard, C., Langan, L., Martens, C., & Scheiter, S. (2020). Misinterpretation  
399 of Asian savannas as degraded forest can mislead management and conservation policy under  
400 climate change. *Biological Conservation*, **241**, 108293.

401 Lehmann, C.E.R. & Parr, C.L. (2016). Tropical grassy biomes: linking ecology, human use and  
402 conservation. *Philosophical Transactions of the Royal Society London B*, **371**, 20160329.

403 Leitner, M., Davies, A.B., Parr, C.L., Eggleton, P. & Robertson, M.P. (2018). Woody encroachment  
404 slows decomposition and termite activity in an African savanna. *Global Change Biology*, **24**,  
405 2597-2606.

406 Li, Y., Brando, P.M., Morton, D.C. *et al.* (2022). Deforestation-induced climate change reduces  
407 carbon storage in remaining tropical forests. *Nature Communications*, **13**, 1964.

408 Luvuno, L., Biggs, R., Stevens, N. & Esler, K. (2022). Perceived impacts of woody encroachment on  
409 ecosystem services in Hluhluwe, South Africa. *Ecology & Society*, **27**, 4.

410 Mitchard, E.T.A. & Flintrop, C.M. (2013). Woody encroachment and forest degradation in sub-  
411 Saharan Africa's woodlands and savannas 1982-2006. *Philosophical Transactions of the Royal  
412 Society London B*, **368**, 20120406.

413 Moreira, A.G. (2000). Effects of fire protection on savanna structure in Central Brazil. *Journal of  
414 Biogeography*, **27**, 1021-1029.

415 Moret, P., Muriel, P., Jaramillo, R. & Dangles, O. (2019). Humboldt's *Tableau Physique* revisited.  
416 *Proceedings of the National Academy of Sciences*, **116**, 12889-12894.

417 Murphy, B., Andersen, A.N. & Parr, C.L. (2016). The underestimated biodiversity of tropical grassy  
418 biomes. *Philosophical Transactions of the Royal Society B*, **371**, 20150319.

419 O'Brien, E. (2005). *Trees and their impact on the emotional well-being of local residents on two inner  
420 London social housing estates*. Report to Peabody Trust, Trees for Cities and the Forestry  
421 Commission.

422 O'Brien, L., McConnachie, S., Hall, C., Forster, J., Dyke, A., Saraev, V., & Jones, G. (2024).  
423 Exploring the social and cultural values of trees and woodlands in England: a new composite  
424 measure. *People and Nature*, **6**, 1334-1354.

425 Overbeck, G.E., Vélez-Martin, E., Scarano, F.R., Lewinsohn, T.M., Fonseca, C.R. *et al.* (2015).  
426 Conservation in Brazil needs to include non-forest ecosystems. *Diversity & Distributions*, **21**,  
427 1455-1460.

428 Overbeck, G.E., Vélez-Martin, E., da Silva Menezes, L., Anand, M., Baeza, S., Carlucci, M.B. *et al.*  
429 (2022). Placing Brazil's grasslands and savannas on the map of science and conservation.  
430 *Perspectives in Plant Ecology, Evolution & Systematics*, **56**, 125678.

431 Parr, C.L., Lehmann, C.E.R., Bond, W.J., Hoffmann, W.A. & Andersen A.N. (2014). Tropical grassy  
432 biomes: misunderstood, neglected and under threat. *Trends in Ecology & Evolution*, **29**, 205-  
433 213.

434 Parr, C.L., te Beest, M. & Stevens, N. (2024). Conflation of reforestation with restoration is  
435 widespread. *Science*, **383**, 698-701.

436 Pausas, J. & Bond, W.J. (2024). Alternative biome states in terrestrial ecosystems. *Trends in Plant  
437 Science*, **25**, 250-263.

438 Pillar, V.D. & Overbeck, G.E. (2025). Nature conservation policies are biased toward forests and  
439 neglect grassy ecosystems worldwide. *Science*, **388**, eadx7441.

440 Pignon, C.P., Jaiswal, D., McGrath, J.M. & Long, S.P. (2017). Loss of photosynthetic efficiency in  
441 the shade. An Achilles heel for the dense modern stands of our most productive C4 crops?  
442 *Journal of Experimental Biology*, **68**, 335-345.

443 Pilon, N.A.L., Durigan, G., Rickenback, J., Pennington, R.T., Dexter, K.G., Hoffmann, W.A., Abreu,  
444 R.C.R. & Lehmann, C.E.R. (2021). Shade alters savanna grass layer structure and function  
445 along a gradient of canopy cover. *Journal of Vegetation Science*, **32**, e12959.

446 Pilon, N.A.L., Peixoto, F., Olivera, R.S., Oliveira, A.C.C., Alquéres, J. *et al.* (2025). Open letter: There  
447 are more than just trees and forests to be conserved and restored. *Plants, People, Planet*, **7**,  
448 1220-1224

449 Ratajczak, Z., Nippert, J.B., Briggs, J.M. & Blair, J.M. (2014). Fire dynamics distinguish grasslands,  
450 shrublands and woodlands as alternative attractors in the Central Great Plains of North America.  
451 *Journal of Ecology*, **102**, 1374-1385.

452 Ratnam, J, Bond, W.J., Fensham, R.J., Hoffmann, W.A., Archibald, S., Lehmann, C.E.R. *et al.*  
453 (2011). When is a 'forest' a savanna, and why does it matter? *Global Ecology and*  
454 *Biogeography*, **20**, 653–660.

455 Ratnam, J., Tomlinson, K.W., Rasquinha, D.N. & Sankaran, M. (2016). Savannas of Asia: antiquity,  
456 biogeography and an uncertain future. *Philosophical Transactions of the Royal Society B*, **371**,  
457 20150305.

458 Rossatto, D.R., de Araújo, P.E., da Silva, B.H.P. & Franco, A.C. (2018). Photosynthetic responses  
459 of understory savanna plants: Implications for plant persistence in savannas under tree  
460 encroachment. *Flora*, **240**, 34.

461 Sankaran, M., Hanan, N., Scholes, R. *et al.* (2005). Determinants of woody cover in African  
462 savannas. *Nature*, **438**, 846–849.

463 Sala, O.E. & Maestre, F.T. (2014). Grass-woodland transitions: determinants and consequences for  
464 ecosystem functioning and provision of services. *Journal of Ecology*, **102**, 1357–1362.

465 Sarmiento, G. (1984). *The ecology of neotropical savannas*. Harvard University Press. Pp. 235.

466 Sayre, N.F. (2017). *The Politics of Scale: A History of Rangeland Science*. The University of Chicago  
467 Press. Pp. 288.

468 Schroeder, R.A. (1995). Contradictions along the commodity road to environmental stabilisation:  
469 foresting Gambian gardens. *Antipode*, **27**, 325–342.

470 Silveira, F.A.O, Ordóñez-Parra, C.A., Moura, L.C., Schmidt, I.B., Andersen, A.N. *et al.* (2022). Biome  
471 Awareness Disparity is BAD for tropical ecosystem conservation and restoration. *Journal of*  
472 *Applied Ecology*, **59**, 1967–1975.

473 Sirami, C., Seymour, C., Midgley, G. & Barnard, P. (2009). The impact of shrub encroachment on  
474 savanna bird diversity from local to regional scale. *Diversity & Distributions*, **15**, 948–957.

475 Staver, A. C., Archibald, S. & Levin, S. A. (2011). The global extent and determinants of savanna  
476 and forest as alternative biome states. *Science*, **334**, 230–232.

477 Stevens, N., Lehmann, C.E.R., Murphy, B.P. & Durigan, G. (2016). Savanna wood encroachment is  
478 widespread across three continents. *Global Change Biology*, **23**, 235–244.

479 Stevens, N., Bond, W.J., Feurdean, A. & Lehmann, C.E.R. (2022). Grassy ecosystems in the  
480 Anthropocene. *Annual Review of Environment & Resources*, **47**, 261–289.

481 Stott, P. (1999). *Tropical Rainforest: A Political Ecology of Hegemonic Myth Making*. IEA Studies on  
482 the Environment No. 15. The Environment Unit.

483 Tazoe, Y., Hanba, Y.T., Furumoto, T., Noguchi, K. & Terashima, I. (2008). Relationships between  
484 quantum yield for CO<sub>2</sub> assimilation, activity of key enzymes and CO<sub>2</sub> leakiness in *Amaranthus*  
485 *cruentus*, a C4 dicot, grown in high or low light. *Plant & Cell Physiology*, **49**, 19–29.

486 Veldman, J.W. (2016). Clarifying the confusion: Old-growth savannahs and tropical ecosystem  
487 degradation. *Philosophical Transactions of the Royal Society B*, **371**, 20150306.

488 Venter, Z.S., Cramer, M.D., Hawkins, H.J. (2018). Drivers of woody plant encroachment over  
489 Africa. *Nature Communications*, **9**, 2272.

490 Vera, F.W.M. (2000). *Grazing Ecology & Forest History*. CABI.

491 Wall, J. (ed) (2022). *The Cultural Value of Trees: Folk Value and Biocultural Conservation*.  
492 Routledge, London.

493 Wasilewska-Debowska, Zienkiewicz & Drozal, A. (2022). How light reactions of photosynthesis in  
494 C4 plants are optimised and protected under high light conditions. *International Journal of*  
495 *Molecular Science*, **23**, 3626.

496 White, J.D.M., Stevens, N., Fisher, J.T., Archibald, S. & Reynolds, C. (2022). Nature-reliant, low-  
497 income households face the highest rates of woody-plant encroachment in South Africa. *People*  
498 *and Nature*, **4**, 1020–1031.

499 White, J.D.M., Stevens, N., Fisher, J.T. & Reynolds, C. (2024). Woody plant encroachment drives  
500 population declines of 20% of common open ecosystem bird species. *Global Change Biology*,  
501 **30**, e17340.

502 Wieczokowski, J.D. & Lehmann, C.E.R. (2022). Encroachment diminishes herbaceous plant  
503 diversity in grassy ecosystems worldwide. *Global Change Biology*, **28**, 5532-5546.

504 Wieczokowski, J.D., Lehmann, C.E.R., Archibald, S., Banda, S., Goyder, D.J., *et al.* (2024). Fire  
505 facilitates ground layer plant diversity in a miombo ecosystem. *Annals of Botany*, **133**, 743-756.

506 Wigley, B.J., Bond, W.J. & Hoffmann, M.T. (2010). Thicket expansion in a South African savanna  
507 under divergent land use: local vs. global drivers. *Global Change Biology*, **16**, 964-976.