

1 **An open occurrence dataset for European subterranean spiders**

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77 **Abstract**

78 Spiders are remarkably diverse in caves and other subterranean habitats, where they play
79 key ecological roles as generalist predators and strongly influence local food webs. They
80 have been instrumental as model organisms for testing various eco-evolutionary
81 hypotheses. Furthermore, strictly subterranean species exhibiting narrow ranges and high
82 endemism are particularly significant for conservation planning and vulnerability
83 assessments. Although high-quality data are essential for research on and conservation of
84 subterranean spiders, such information remains scarce, especially regarding distribution
85 patterns. To help fill this gap, we screened the literature, unpublished records, and open
86 datasets to compile georeferenced occurrences of subterranean spiders across caves and
87 other subterranean habitats throughout Europe. Based on these data—and to illustrate one
88 potential application of the compiled dataset—we present the first prediction of subterranean
89 spider richness patterns across Europe using stacked species distribution models. The
90 European Subterranean Spider Dataset (ESSD) comprises 31,224 records of 637
91 subterranean-dwelling spider species (including morphospecies under description),
92 covering a range of information including taxonomy, locality details (such as location name,
93 country, geographic coordinates, type of subterranean habitat), and reference information
94 for each record. All variables are coded using the Darwin Core Standard, ensuring
95 interoperability with the Global Biodiversity Information Facility (GBIF) and other biodiversity
96 databases. By enabling integration with trait and phylogenetic resources, the ESSD provides
97 a robust framework to investigate the drivers and processes shaping subterranean
98 biodiversity, assess vulnerability to environmental change and anthropogenic pressures,
99 and guide future sampling to progressively reduce geographic and taxonomic gaps through
100 open data sharing.

101 **Keywords:** Araneae, Darwin Core Standard, Hypogean, Open data, Species distribution
102 modelling (SDM)

103

104 **Introduction**

105 In recent years, there has been an explosion of biogeography and (macro)ecology studies
106 focused on uncovering the patterns and factors that shape biodiversity patterns at
107 increasingly larger spatial scales—continental to global (e.g., Labouyrie et al., 2023;
108 Martínez-Núñez et al., 2023, Sabatini et al., 2022). Building this understanding
109 fundamentally relies on high-quality data, especially distribution records. Over the past few
110 decades, online biodiversity databases have experienced substantial growth, largely due
111 to collaborative efforts that have enhanced data accessibility and sharing. This progress
112 has led to the formation of comprehensive databases covering a wide range of taxa,
113 including the Global Biodiversity Information Facility (GBIF) (GBIF, 2025), LifeWatch ERIC
114 (LifeWatch ERIC, 2025) and BioTIME (Dornelas et al., 2025). These repositories provide
115 extensive taxonomic and distributional information on thousands of taxa across various
116 ecosystem types and time scales. Despite these advances, large gaps and biases in
117 species' known geographic distributions, the so-called Wallacean shortfall, continue to limit
118 the completeness and reliability of macroecological and biogeographic inferences (e.g.,
119 Cardoso et al., 2011; Hortal et al., 2015; Hughes et al., 2021).

120 Primarily due to accessibility challenges (Ficetola et al., 2019; Mammola et al.,
121 2021a), the documentation of biodiversity in subterranean ecosystems (caves,
122 groundwaters, fissural systems, and the like) has historically progressed more slowly than
123 at the surface. However, in recent years, a steady accumulation of knowledge, combined
124 with the funding of specific projects focused on continental biodiversity inventories (e.g.,
125 PASCALIS, Biodiversa+ DarCo), has led many authors to compile this information and
126 publish it in public datasets with varying resolutions and scales. For example, we now
127 have the first global datasets on the distribution of cave-dwelling bats (Tanalgo et al.,
128 2022), cave fish (Bai et al., 2025), asellids (Saclier et al., 2024), and microwhip scorpions
129 (Mammola et al., 2021b). In Europe, Pascalis dataset (Deharveng et al. 2009) and
130 European Groundwater Crustaceans Dataset (Zagmajster et al. 2014) were used for early
131 analyses of continental patterns, but publication of the continental datasets of distribution
132 of subterranean organisms have started only in the last few years, like for bats (Fialas et
133 al. 2025), copepods (Cerasoli et al., 2025) and ostracods (Mori et al., 2025). However, a
134 similar large-scale dataset is still lacking for subterranean spiders.

135 Spiders (Arachnida: Araneae), with over 53,000 species currently described (World
136 Spider Catalog, 2025), and providing numerous essential ecosystem services (Cardoso et
137 al. 2025), are among the most widespread and generalist predators in terrestrial habitats
138 (Turnbull, 1973). Spiders are particularly diversified in caves and other subterranean voids,
139 where they play a key ecological role as predators and strongly structure local food webs.
140 Despite the growing interest in subterranean spiders (Mammola and Isaia, 2017), major
141 knowledge gaps remain, especially regarding their distribution. Limited expertise and lower
142 research interest in certain regions have delayed comprehensive data collection. However,
143 recent efforts are beginning to address these gaps, contributing essential data for
144 advancing our understanding of subterranean spider ecology and biodiversity. These
145 efforts include the publication of trait data for all the species in Europe (Mammola et al.,
146 2022; Patiño-Sauma et al., 2025) and high-resolution distribution data for selected caves
147 (Mammola et al., 2019a; Macías-Hernández et al., 2024) or regions (e.g., Western Alps;
148 Nicolosi et al., 2025; Azores; Crespo et al. 2025).

149 We present a novel dataset comprising occurrence records for all known species
150 and morphospecies of subterranean spiders from Europe, spanning a wide range of
151 ecological affinities to subterranean habitats, ranging from species still able to exploit
152 surface habitats to obligate subterranean dwellers. This dataset is the result of a
153 collaborative effort among multiple partners, including ecologists, conservationists,
154 taxonomists, and biogeographers from various European countries and outermost islands
155 of Europe (Azores, Madeira, Selvagens, and the Canary Islands). Their contributions have
156 significantly enriched the data availability, culminating in a comprehensive, multi-species
157 dataset designed to advance research and conservation efforts for these species. By
158 making these data public, we hope to promote collaborative research on subterranean
159 spider biodiversity, spatial patterns, drivers of distribution patterns, and quantitative
160 conservation efforts.

161

162 **Methods**

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164 **Target species and habitats**

165 We focused on subterranean spiders across continental Europe, including the
166 archipelagos of the Azores, Salvagens, Madeira, and Canary Islands. Following the
167 function-based classification of Earth's ecosystems (Keith et al., 2020, 2022), we focused
168 on ecosystems belonging to the 'Subterranean' (S) domain, which includes diverse
169 terrestrial subterranean systems: i) the 'Subterranean lithic' (S1) biome, namely various
170 type of caves (e.g., aerobic caves, lava tubes, volcanic pits) and other subterranean voids
171 of smaller sizes (e.g., fissure systems, deep scree strata, and the so-called *Milieu
172 Souterrain Superficial* [MSS; reviewed in Mammola et al. (2016)]); and ii) the
173 'Anthropogenic subterranean voids' (S2) biomes, namely all anthropogenic subterranean
174 voids with cave-like environmental conditions, including mines, underground bunkers,
175 blockhouses, tunnels, culverts, and cellars.

176 For the list of target species, we used the latest checklist of European subterranean
177 spiders by Patiño-Sauma et al. (2025), currently listing 637 species across 28 families. Of
178 these, 64 are species under description (hereinafter 'morphospecies'), identified by experts
179 as new taxonomic entities based on morphological and/or genetic information.

180

181 **Data acquisition**

182 The spider dataset is a comprehensive compilation of diverse data sources, created
183 through the collaboration of 40 researchers across Europe. For each of the target species,
184 we compiled occurrence records based on different sources. First, we mined the primary
185 literature for reported localities and georeferenced records. Second, we included
186 unpublished records (e.g., data stored in personal and institutional collections) as provided
187 by each contributor in the authors list. Third, we mined records from the main accessible
188 databases on spiders in Europe, namely ArachnoMap (de Biurrun et al., 2022), Araneae.it
189 (Pantini & Isaia, 2019), the UK Spider and Harvestman Recording Scheme
190 (<https://srs.britishspiders.org.uk/>), Canary Islands Biodiversity Database (Gobierno de
191 Canarias, 2024), the Cave fauna of Greece database (Paragamian et al., 2025) and
192 regional database on subterranean species of the Western Balkans SubBioDB
193 (Zagmajster 2016). Lastly, we mined the GBIF database, which yielded a great number of
194 missing records, especially for the most widespread species. We downloaded the
195 occurrence records from GBIF using the Python package "biodumpy" v.0.1.6 (Cancellario
196 et al., 2025). Specifically, we employed the GBIF module, setting the parameter
197 dataset_key to "d7dddbf4-2cf0-4f39-9b2a-bb099caae36c" and geometry to "POLYGON((-
198 30 25,50 25,50 72,-30 72,-30 25))". The script produced a list of JSON files, which we
199 subsequently converted to CSV format to facilitate handling.

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201 **Format type and data availability**

202 The dataset file is in comma-separated values (csv) format, not compressed. Data are
203 available in Figshare at the following Digital Object Identifier:
204 <https://doi.org/10.6084/m9.figshare.30696173>.

205

206 **Header information**

207 The variables included in the dataset were selected in accordance with the Darwin Core
208 standard (Wieczorek et al., 2012), and the corresponding categories are listed in Table S1.
209 Headers are mostly self-explanatory. The dataset is fully interoperable with the European
210 subterranean spider trait dataset (<https://doi.org/10.6084/m9.figshare.16574255>), allowing
211 the extraction of morphological and ecological trait information for each species (Mammola
212 et al., 2022; Patiño-Sauma et al., 2025).

213

214 **Taxonomic validation**

215 We standardized taxonomy to the species level when feasible, following the latest
216 nomenclature of the World Spider Catalog (2025). Furthermore, we incorporated genus-
217 level records with uncertain specific attribution (e.g., *Meta* sp., *Troglohyphantes* cf.
218 *lucifuga*), as well as morphospecies under description (labelled as Genus + an
219 alphanumeric code [e.g., *Meta* sp. 1]) to ensure maximal dataset breadth. Please refer to
220 the column “acceptedNameUsage” for the most up-to-date, validated taxonomic attribution
221 for each record (note that taxonomy will be updated with any new release of the dataset).
222 Eventual remarks on taxonomic decisions are provided in column “identificationRemarks”.

223

224 **Spatial validation**

225 We validated geographic coordinates (based on the WGS84 datum) through cross-
226 referencing with online resources (e.g., speleological cadastres) and, where available,
227 species-specific reference materials. We subsequently projected and visually inspected
228 localities using both R (R Core Team, 2025) and QGIS (QGIS.org, 2025). We harmonized
229 cave names and their corresponding coordinates, obtained from various sources, through
230 additional verification with national speleological cadastres whenever possible.
231 Notwithstanding these quality checks, due to missing information (especially for old
232 records), the precision of 1,990 records remains low (e.g., georeferenced using the
233 centroid of the municipality) and 108 records lack coordinates. Uncertainty in the precision
234 of coordinates is provided in the column georeferenceRemarks.

235

236 **Prediction of species richness**

237 We illustrate a potential usage of the dataset by predicting species richness patterns in
238 Europe. We achieved this by using stack species distribution modelling (SSDM) to
239 calculate species potential distribution across the continent. A ODMAP (Overview, Data,
240 Model, Assessment and Prediction) (Zurell et al., 2020) protocol for the model, detailing
241 the main analytical steps, is available in the supplementary materials (Appendix 1).

242 We included in the modelling all species with at least 10 independent records,
243 defined as occurrences from different localities separated by at least 10 km (i.e., the
244 spatial resolution of our environmental predictors), with a total of 99 different spider
245 species distributions modelled in the present study. Note that species with fewer than 10
246 independent records were only later included in the final richness prediction (see below).
247 Since some localities in the species' distribution data have highly precise geo-localization,

248 multiple points may fall within the same cell of the downloaded abiotic layers, either due to
249 the accuracy of the coordinates or because specimens were collected in nearby caves. To
250 avoid redundant distribution points, we adjusted all coordinates to match the centroid of
251 the corresponding cell. We then applied spatial thinning using the *thin* function from the R
252 package "spThin" (Aiello-Lammens et al. 2015).

253 Previous research has shown the importance of present (Mammola and Leroy
254 2018) and past climatic factors (Hewitt 1999; Mammola et al. 2018, 2019b; Knüsel et al.
255 2024), as well as soil composition (Pavlek and Mammola 2021) in shaping subterranean
256 species distributions. To include these variables in our model, we downloaded climatic and
257 elevation data from the WorldClim 2 database (Fick and Hijmans 2017), specifically annual
258 mean temperature (BIO 1), temperature seasonality (BIO 4), maximum temperature of the
259 warmest month (BIO 5), annual precipitation (BIO 12), precipitation seasonality (BIO 15),
260 precipitation of the warmest quarter (BIO 18) and precipitation of the coldest quarter (BIO
261 19). All these variables have been shown to be good proxies for subterranean climatic
262 conditions (Mammola and Leroy 2018). In addition, we included the differences between
263 the present and past precipitation and temperature compared during the last glacial
264 maximum (LGM). For soil composition, we downloaded layers from the SoilGrids database
265 (Poggio et al. 2021), specifically the percentage of coarse fragments and the gravimetric
266 content of sand and clay in the soil. Finally, we included layers regarding
267 evapotranspiration (Muñoz Sabater, J. (2019), normalized difference vegetation index
268 (NDVI) from Li et al. (2023), and the soil organic carbon (SOC) and organic carbon
269 detection (OCD) downloaded using the *soil_world* function from the "geodata" R package
270 (Hijmans et al. 2024). All these variables are potential proxies for energy availability within
271 the subterranean domain. All layers had a resolution of approximately 10km. We
272 calculated pairwise Pearson's *r* correlation coefficients among these variables, and
273 excluded those with high correlation ($|r| \geq 0.7$), retaining only one variable from each
274 correlated group based on ecological relevance, data quality, and consistency with
275 previous subterranean ecology studies (Mammola & Leroy, 2018). The final list of
276 predictors included: temperature seasonality, precipitation of the warmest quarter,
277 evapotranspiration, percentage of coarse fragments and content of clay in the soil.

278 We generated individual species models using the *modelling* function from the R
279 package "SSDM" (Schmitt et al. 2017) with the MAXENT algorithm (Phillips et al., 2004,
280 2006; Elith et al., 2011). To estimate species' potential distributions and reduce
281 overprediction, thresholds on environmental suitability were applied using Cohen's Kappa
282 and True Skill Statistic (TSS) values via the *ecospat.max.kappa* and *ecospat.max.tss*
283 functions from the R package "ecospat" (Broennimann et al. 2025), with the more
284 restrictive threshold being selected. We then stacked all individual species distributions,
285 and included species with fewer than 10 records as single-cell localities in the map.

286

287 **Results**

288 The dataset contains 31,224 records of subterranean spiders, accounting for 637 species,
289 comprising all available georeferenced data on 31,116 records up to 2025, spanning 40
290 countries also including Azerbaijan, Georgia, Turkey, and Russia.

291 The spatial extent of the dataset ranges from -28.80°W to 50.02°E in longitude, and from
292 27.65°N to 67.94°N in latitude.

293 Occurrence densities are particularly high in several European regions. Northern
294 Italy, especially the northwestern and northeastern Alps, shows the greatest number of
295 records. Elevated densities also characterize northern Spain along the Atlantic coast, as
296 well as Slovenia and Croatia within the Dinaric karst (Fig. 1A). Observed species richness
297 closely matches these patterns. The highest values occur in Croatia (Dinaric karst) and
298 Slovenia, where multiple taxa overlap, followed by parts of the Alps and northern Spain's
299 Atlantic region (Fig. 1B).

300 A country-level summary highlights marked geographic disparities. Italy stands out for both
301 richness and number of records, followed by Croatia, Spain, France, and Slovenia (Fig.
302 1C). Most other European countries display comparatively low values, underscoring the
303 strong imbalance in sampling effort.

304 Predicted species richness (Fig. 1D) confirms the Alps and adjacent mountain systems,
305 including the Dinaric Arc, as major hotspots, with high values spanning northern Italy,
306 Slovenia, Croatia, and the Atlantic coast of northern Spain. Moderate richness also
307 extends into parts of central and southeastern Europe, including Austria, Germany,
308 France, and Bosnia and Herzegovina.

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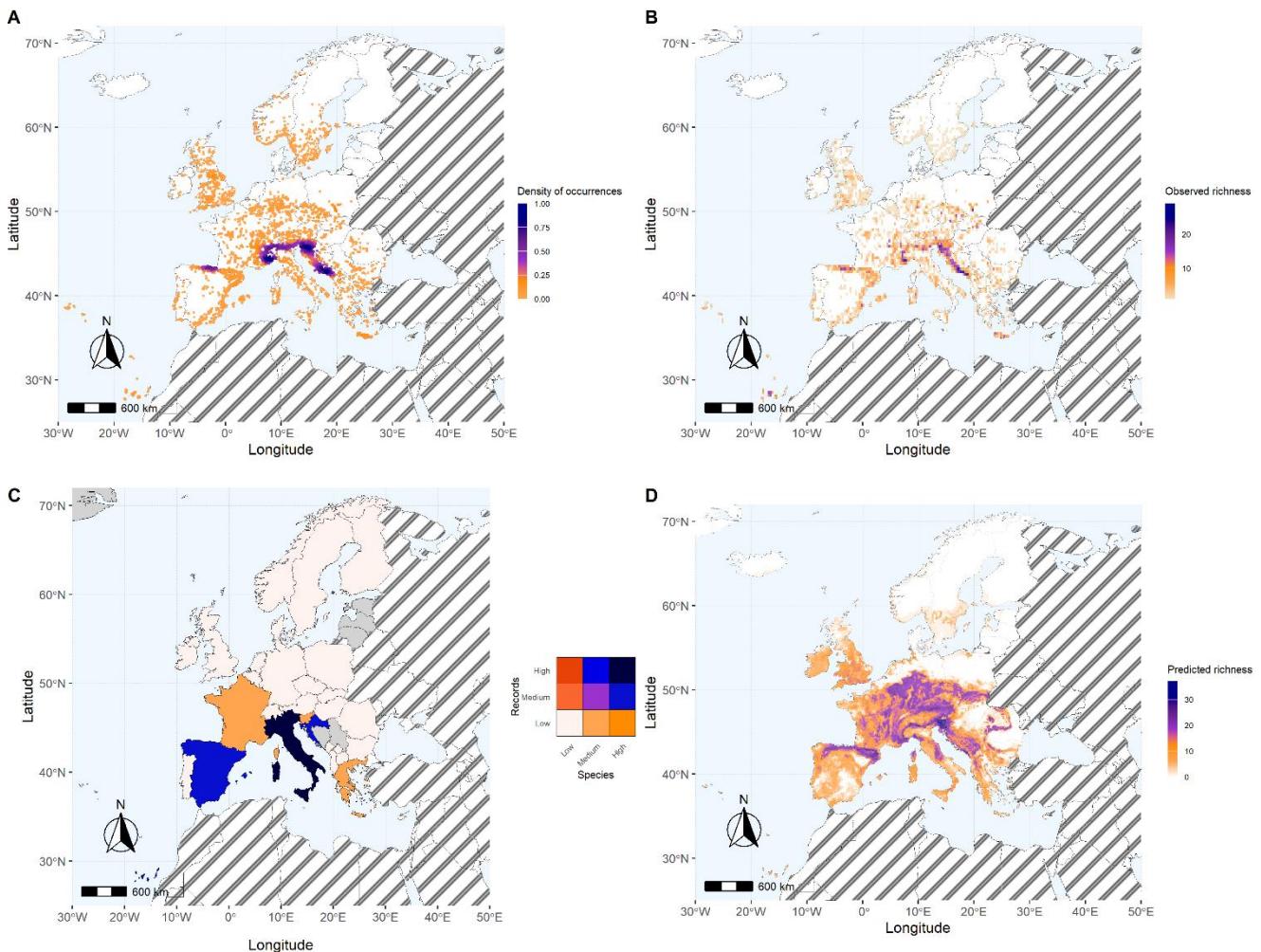
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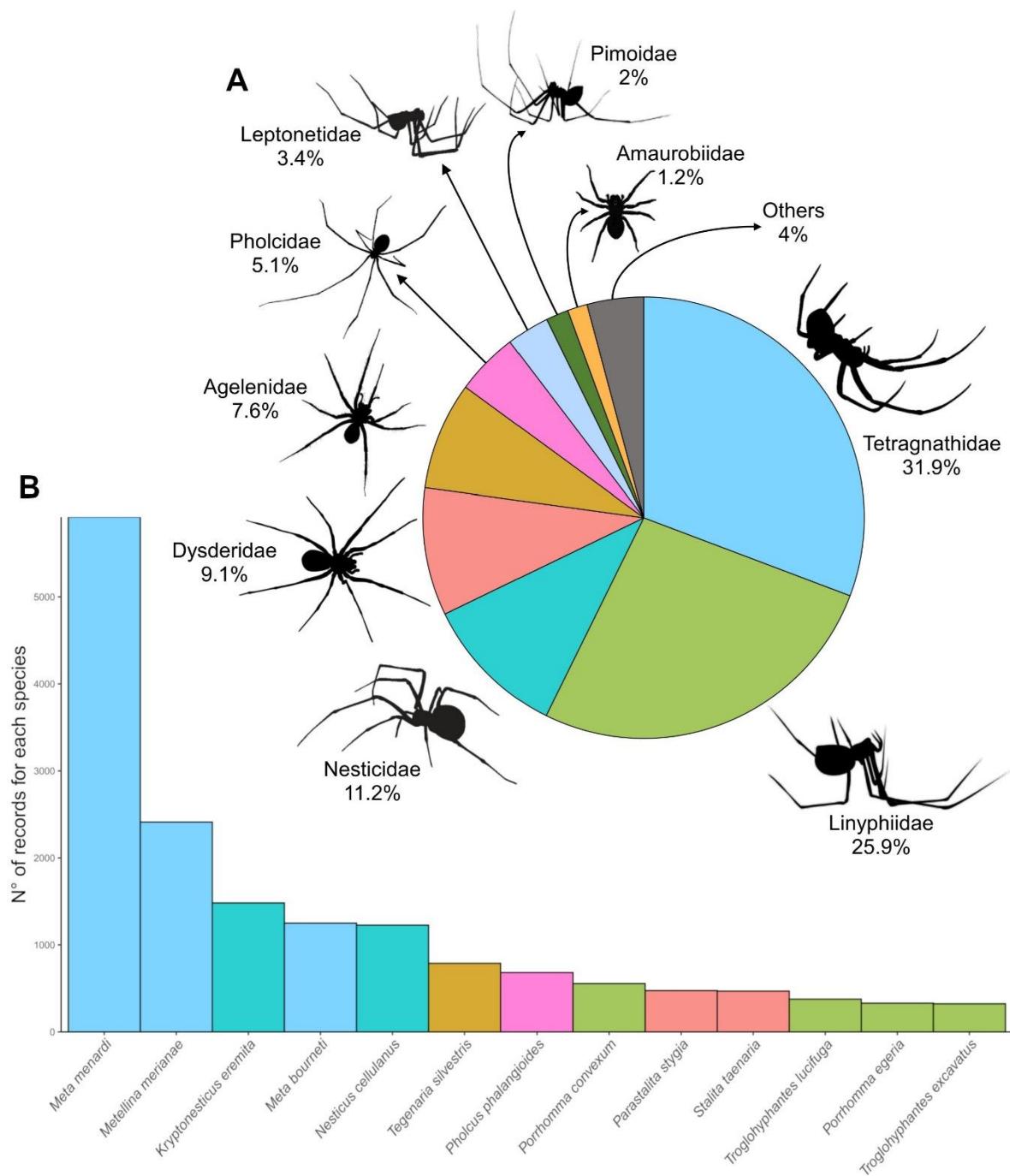
329 **Figure 1. (A)** Density of species occurrences across Europe. Points represent individual records,
 330 with color intensity reflecting local occurrence density; **(B)** Observed richness shows the number of
 331 species recorded per cell across the continent. **(C)** Geographic distribution of spider records in
 332 Europe. The map shows the combined effect of the number of records and the species richness for
 333 each country, classified into tertiles (Low, Medium, High). The 3x3 legend indicates the intersection
 334 between record abundance (rows) and species richness (columns). Countries without records are
 335 shown in grey. **(D)** Predicted richness shows the modelled species richness across Europe
 336 obtained through stacked species distribution modelling, with colours indicating the predicted
 337 number of species per grid cell. Mapping is restricted to countries with sufficient data availability;
 338 countries with very few records (e.g. Russia) are therefore not represented.

340 Most records were collected from natural caves, totalling 25,947 (83.10%) of all
 341 occurrences. These were followed by artificial habitats (e.g., mines, bunkers, blockhouses,
 342 cellars), which accounted for 1,929 records (6.17%), and shallow subterranean habitats
 343 (SSH) with 323 records (1.03%).

344 In terms of taxonomic distribution, most records belong to the family Tetragnathidae
 345 (9,951; 31.87%), followed by Linyphiidae (8,094; 25.93%), Nesticidae (3,483; 11.16%),
 346 Dysderidae (2,822; 9.04%), Agelenidae (2,378; 7.62%), Pholcidae (1,584; 5.07%),
 347 Leptonetidae (1,057; 3.39%), Pimoidae (618; 1.98%), and Amaurobiidae (362; 1.16%).

348 Other families, including Sicariidae, Theridiidae, and Cicurinidae, were each represented
349 by less than 1% of the total records (Fig. 2A).

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351

352 **Figure 2. (A)** Pie chart showing the relative percentage of occurrences records for each spider
353 family in the dataset. Families representing less than 1% of records were grouped into the category
354 “Others”. **(B)** Barplot displaying the number of records for the most common species in the spiders
355 dataset. Only species with at least 300 records are shown.

356

357 Among the 637 species in the dataset, the most frequently recorded species is *Meta*
358 *menardi* (Latreille, 1804), with 5,914 occurrences. This is followed by *Metellina merianae*
359 (Scopoli, 1763) with 2,412 records, *Kryptonesticus eremita* (Simon, 1880) with 1,482
360 records, *Meta bourneti* (Simon, 1922) with 1,251 records, and *Nesticus cellulanus* (Clerck,
361 1757) with 1,226 records. All other species are represented by fewer than 1,000 records
362 (average number of records \pm s.d. = 20 ± 38) (Fig. 2B).

363

364 Discussion

365 The dataset provides the most comprehensive distribution data for subterranean spiders
366 across Europe. It brings together data from multiple countries and research groups,
367 consolidating previously scattered information into a single, standardized, and harmonized
368 resource.

369 The taxonomic composition of the dataset reveals meaningful ecological and
370 sampling-related patterns. In particular, some families are overrepresented in terms of
371 occurrence records relative to their species richness. For example, Tetragnathidae account
372 for a high number of records despite including few subterranean-associated species (*Meta*
373 spp. and *Metellina merianae*; Fig. 2A). This likely reflects the large body size, high
374 conspicuousness, and often high local densities of these spiders near cave entrances
375 (Smithers 2005; Novak et al., 2010; Mammola & Isaia, 2014), all resulting in high
376 detectability, as well as their broad geographic distributions in Europe (Mammola et al.,
377 2019a, 2021c).

378 The predicted richness patterns (Fig. 1D) also illustrate the potential of the dataset to
379 support macroecological inference. By integrating species distribution models for tens of
380 taxa, the resulting map highlights broad-scale biogeographic structures that are not always
381 evident from raw data alone. In particular, the Alps, the Dinaric Arc, and parts of northern
382 Spain emerge as major hotspots of subterranean diversity, consistent with the long-term
383 persistence of stable microclimatic refugia and the complex geomorphological history of
384 these regions (Culver & Sket, 2000; Deharveng et al., 2024). The smoother gradients
385 revealed by the prediction also indicate that true richness likely extends beyond areas with
386 dense sampling, highlighting regions such as Austria, Germany, France, and Bosnia and
387 Herzegovina as potentially important, yet comparatively understudied. This underscores
388 both the ecological value of the modelling approach and the role of the dataset in identifying
389 priority areas where additional sampling would substantially improve knowledge of
390 subterranean biodiversity. Integrating predicted patterns with conservation planning may
391 therefore help guide efforts toward mountain systems and karstic landscapes that harbour
392 disproportionate levels of subterranean diversity.

393 The initiative reflects the growing recognition that open data sharing at continental
394 and global scales is essential to advance (macro)ecological research, improve biodiversity
395 monitoring, and inform conservation strategies under increasing environmental threats
396 (Urbano et al., 2024). This goal has been greatly facilitated in recent years by the
397 development of international research infrastructures such as the Global Biodiversity
398 Information Facility (GBIF; GBIF.org, 2025) and LifeWatch ERIC (Basset & Los, 2012), which
399 promote standardized, interoperable, and openly accessible biodiversity data.

400 Although focused on distributional records, its interoperability with other data sources
401 makes it particularly valuable. For example, it can be combined with complementary
402 resources, such as datasets on spider morphological and ecological traits (Mammola et al.,
403 2022; Patiño-Sauma et al., 2025) or phylogenetic information from other sources, thereby
404 enabling comparative analyses, functional diversity assessments, and large-scale
405 ecological modelling. By linking distributional data with traits and phylogenies, researchers
406 can further explore questions related to the processes shaping subterranean biodiversity
407 and identify species and regions most vulnerable to environmental change and
408 anthropogenic pressures.

409 The geographical and taxonomic breadth of the dataset makes it a valuable resource
410 for addressing key questions in subterranean ecology, from species distributions and
411 environmental drivers to long-standing gaps in conservation status, functional diversity, and
412 the ecological factors structuring subterranean spider assemblages. While the dataset
413 marks a substantial step forward, some regions and taxa remain underrepresented,
414 reflecting historical biases in research effort and data availability. Open data sharing,
415 however, provides the basis for progressively improving coverage and quality, fostering new
416 sampling initiatives, comparative analyses, and multi-scale investigations of subterranean
417 biodiversity.

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633

634 **Conflict of Interest Statement**

635 The authors declare no conflicts of interest.

636

637 **Data and Code Availability Statement**

638 The dataset and associated R codes are available from the Figshare repository under
639 <https://doi.org/10.6084/m9.figshare.30696173>

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668 **Appendix 1.** Summary of the modelling pipeline according to the
669 ODMAP (Zurell et al., 2020) reporting protocol.

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671

672 **OVERVIEW**

673 **Title**

674 An open occurrence dataset for European subterranean spider

675 **Model objective**

676 **Model objective:** Mapping and interpolation.

677 **Target output:** Obtaining a map of species richness based on stacked predicted distribution ranges of each
678 species.

679 **Focal Taxon**

680 **Focal Taxon:** Spiders (Arachnida: Araneae).

681 **Location**

682 **Location:** Europe.

683 **Scale of Analysis**

684 **Spatial extent:** -30, 50, 25, 72 (xmin, xmax, ymin, ymax).
Spatial resolution: 10 km.
Temporal extent: Present.
Temporal resolution: NA.
Boundary: political.

685 **Biodiversity data**

686 **Observation type:** field survey, GPS tracking.
Response data type: presence-only.

687 **Predictors**

688 **Predictor types:** climatic, habitat, edaphic.

689 **Hypotheses**

690 **Hypotheses:** No specific hypotheses were made regarding the species-environment relationships.

691 **Assumptions**

692 **Model assumptions:** Unlimited dispersal.

693 **Algorithms**

94 **Modelling techniques:**

maxent.

Model complexity:

Only ecologically interpretable predictors were included, and interactions were not considered to maintain biological interpretability.

Model averaging:

NA.

694

695 **Workflow**

696 Environmental predictor layers were standardized prior to modelling.

697 For each species, a MaxEnt model was fitted using presence locations and environmental predictors to
698 generate continuous habitat suitability predictions.

699 Predictions were spatially constrained using a buffer-based mask derived from the mean and maximum
700 inter-point distances among occurrence records, and suitability values were smoothed as a function of
701 distance to the accessible area boundary.

702 Model performance was assessed by extracting predicted suitability values at observed presences and
703 randomly sampled background points, and optimal thresholds were determined using both Kappa and True
704 Skill Statistic (TSS).

705 Binary presence–absence maps were generated using the most conservative threshold and combined
706 across species to produce a stacked species distribution model representing predicted species richness.

707 **Software**

708 **Software:** R version 4.5.0. SSDM package version 0.2.9.

Code availability: The code to run this analysis is available in the Figshare repository
(<https://doi.org/10.6084/m9.figshare.30696173>).

Data availability: Distribution data are available in the Figshare repository
(<https://doi.org/10.6084/m9.figshare.30696173>). Environmental predictors are available from
different sources (details in section: Predictor variables).

709 **DATA**

710 **Biodiversity data**

711 **Taxon names:** 99 different spider species (including some under description) were modelled. Taxonomy is
712 according to the World Spider Catalog version 26 (<https://wsc.nmbe.ch/>) and the version 3 of the Checklist
713 of European Subterranean spider (<https://doi.org/10.6084/m9.figshare.16574255>). See supplementary
714 material of the manuscript for a detailed list of all species analyzed.

715 **Taxonomic reference system**

716 **Ecological level:** species.

717 **Data sources**

718 **Sampling design:** Random.

719 **Sample size:** 7000 different observations for 99 species (average number of records per species \pm s.d.: 70.7 \pm 180.7).

721 **Clipping:** Europe.

722 **Scaling:** We used one record per cell for each species.

723 **Cleaning:** Species with less than 10 records were discarded for the modelling analyses. For species with less than 10 records, the observed distribution was used to generate the stacked prediction of species richness. To address spatial sampling bias and spatial autocorrelation, occurrence records were spatially thinned separately for each species using a minimum nearest-neighbour distance of 10 km. The thinning procedure was repeated 100 times and a single thinned realization was retained for model fitting. Environmental predictor layers and model predictions were processed on a common raster grid and spatial resolution to ensure spatial consistency in subsequent thresholding and stacking.

730 **Absence data:** No true absence data were available.

731 **Background data:** Background data were derived separately for each species by defining a species-specific accessible area based on the spatial configuration of occurrence records. The mean and maximum pairwise distances among occurrence locations were used to generate a buffered spatial mask representing the area available for dispersal. Random background points were then sampled within this mask at approximately twice the number of presence records and used for threshold selection. This spatially constrained background sampling reduced the influence of environmentally unrealistic or geographically inaccessible areas on model assessment.

738 **Errors and biases:** See main text for discussion.

739 **Data partitioning**

Training data: NA.

Validation data: NA.

Test data: Expert-based assessment.

740

741 **Predictor variables**

742 **Predictor variables:** Temperature seasonality, precipitation of the warmest quarter, evapotranspiration, percentage of coarse fragments and content of clay in the soil.

744 **Data sources:** Evapotranspiration: <https://cds.climate.copernicus.eu/datasets/reanalysis-era5-land-monthly-means?tab=overview>; Bioclimatic data: <https://www.worldclim.org/data/worldclim21.html>; Soil data: <https://soilgrids.org/>.

747 **Spatial extent:** -30, 50, 25, 72 (xmin, xmax, ymin, ymax).

748 **Spatial resolution:** res: 0.08333333 (around 10km at the equator).

749 **Coordinate reference system:** WGS84 decimal degrees.

750 **Temporal extent:** 1970–2000 for Bioclimatic variables, 2020 for soil data, and 1950–2025 for
751 evapotranspiration.

752 **Temporal resolution:** NA.

753 **Data processing:** Environmental predictor layers were standardized using z-score normalization (mean-
754 centering and scaling by standard deviation) prior to model fitting to ensure comparability among variables.

755 **Errors and biases:** NA.

756 **Dimension reduction:** Expert-based assessment.

757 **Transfer data**

Data sources:	NA.
Spatial extent:	
Spatial resolution:	NA.
Temporal extent:	NA.
Temporal resolution:	NA.
Models and scenarios:	NA.
Data processing:	NA.
Quantification of Novelty:	NA.

758 **MODEL**

759 **Variable pre-selection**

760 **Variable pre-selection:** NA.

761 **Multicollinearity**

762 **Multicollinearity:** We tested the correlation between all variables by calculating pairwise Pearson's r
763 correlations, setting a threshold for collinearity at $|r| > 0.7$, and then selected the variables with higher
764 ecological meaning based on expert based assessment.

765 **Model settings**

766 **maxent:** featureSet (Default), featureRule (Default), regularizationMultiplierSet (1), regularizationRule
767 (Default), convergenceThresholdSet (Default (0.00001)), samplingBiasRule (NA), samplingBiasNotes (NA),
768 targetGroupSampleSize (NA), offsetSet (NA), offsetRule (NA), expertMapProbSet (NA), expertMapProbRule
769 (NA), expertMapRateSet (NA), expertMapRateRule (NA), expertMapSkewSet (NA), expertMapSkewRule
770 (NA), expertMapShiftSet (NA), expertMapShiftRule (NA), notes (NA).

771 **Model settings (extrapolation):** Default.

772 **Model estimates**

773 **Coefficients:** Default.

774 **Parameter uncertainty:** Parameter uncertainty was not explicitly quantified in this workflow.

775 **Variable importance:** NA.

776 **Model selection - model averaging - ensembles**

777 **Model selection:** No applicable due to single model used.

778 **Model averaging:** NA.

779 **Model ensembles:** NA.

780 **Analysis and Correction of non-independence**

781 **Spatial autocorrelation:** NA.

782 **Temporal autocorrelation:** NA.

783 **Nested data:** NA.

784 **Threshold selection**

785 **Threshold selection:** Continuous habitat suitability predictions from MaxEnt were converted to binary presence-absence maps using species-specific thresholds. Thresholds were determined by maximizing both Cohen's Kappa and the True Skill Statistic (TSS) based on model predictions at observed presences and randomly sampled background points. The most conservative threshold among the two was applied to generate final binary maps.

790 **ASSESSMENT**

791 **Performance statistics**

792 **Performance on training data:** Expert-based assessment on the stacked richness map as we were not interested in single species distributions performance.

794 **Performance on validation data:** NA.

795 **Performance on test data:** NA.

796 **Plausibility check**

797 **Response shapes:** NA..

798 **Expert judgement:** Yes

799 **PREDICTION**

800 **Prediction output**

801 **Prediction unit:** Same as extent.

802 **Post-processing:** NA.

803 **Uncertainty quantification**

804 **Algorithmic uncertainty:** NA.

805 **Input data uncertainty:** NA.

806 **Parameter uncertainty:** NA.

807 **Scenario uncertainty:** NA.

808 **Novel environments:** NA.