

1 **Title**

2 Committed genetic erosion: parallels from climate science for proactive conservation

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4 **Running title**

5 Committed genetic erosion and climate parallels

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52

53 **Abstract**

54 While climate science recognises that past emissions lock in future impacts ("committed climate
55 change"), the idea of lagged losses is less embedded in biodiversity policy. This is particularly
56 true for genetic diversity loss (genetic erosion), the least visible dimension of the biodiversity
57 crisis. To highlight this risk, we introduce the term "committed genetic erosion": future losses set
58 in motion by past habitat loss, fragmentation, and population decline. Using parallels from
59 climate science, we develop four arguments for biodiversity policy: genetic diversity loss is
60 already widespread; biological inertia locks in further loss under business as usual; indicators and
61 forecasting tools can quantify risk and project trajectories; and proactive interventions are more
62 effective than reactive responses or speculative technological fixes. Critically, genetic inertia
63 operates locally, providing a tractable window for restoring connectivity, maintaining viable
64 effective population sizes, and well-planned genetic rescue. Acting now will safeguard the
65 adaptive potential that underpins biodiversity resilience.

66

67 **Keywords**

68 Genetic diversity, extinction debt, genetic rescue, Global Biodiversity Framework, tipping points

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71 **Introduction: Biodiversity and climate, two interlinked crises**

72 A rapidly changing world is increasing the need for resilient species that can adapt to new
73 environments. Climate change and biosphere integrity, encompassing genetic diversity, are
74 recognised as key dimensions of planetary stability, and current trends indicate that both are
75 operating beyond safe limits for humanity (Richardson et al. 2023). Genetic diversity provides
76 the foundation for adaptive capacity, enabling species and the ecosystems they comprise to
77 persist through change. Yet despite being highlighted for decades (Frankel 1974), genetic
78 diversity remains the least visible dimension in biodiversity policy and public discourse, and is
79 the domain most recently integrated into global biodiversity frameworks. For example, the
80 United Nations (UN) Convention on Biological Diversity (CBD) Kunming-Montreal Global
81 Biodiversity Framework (GBF) now explicitly includes maintaining the adaptive potential of all
82 species' populations as part of its goals and targets (Convention on Biological Diversity 2022),
83 marking unprecedented political recognition of genetic processes in nature.

84 Recognition of the importance of genetic diversity has come after decades of addressing global
85 environmental change through separate climate and biodiversity agendas (Pettorelli et al. 2021,
86 Richardson et al. 2023). The UN CBD and the UN Framework Convention on Climate Change
87 (UNFCCC) were signed in 1992 as sister conventions (ratified in 1993 and 1994, respectively,
88 ipbes.net). Yet, their trajectories have diverged. The UNFCCC gained momentum through strong
89 scientific leadership (e.g., Bolin 1991) and a clear global narrative shaped by the
90 Intergovernmental Panel on Climate Change (IPCC). In contrast, scientific input into biodiversity
91 policy evolved more slowly through the CBD Subsidiary Body on Scientific, Technical and
92 Technological Advice (SBSTTA; Koetz et al. 2008; Laikre et al. 2008), and later through the
93 Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services (IPBES;

94 established 2012, first global assessment 2019), which has broadened the information available to
95 the CBD. This imbalance has contributed to less public attention and weaker implementation of
96 biodiversity policy than of climate policy (Mace et al. 2018, Pörtner et al. 2021). Today, calls to
97 enhance coherence among these frameworks, and recent collaborations between IPBES and the
98 IPCC, signal renewed recognition that the solutions to climate and biodiversity loss are
99 inseparable (Pörtner et al. 2021, IPBES 2024).

100 Climate science offers important lessons for biodiversity policy, including the mechanistic
101 concept of commitment: the idea that past emissions and the inertia of the Earth system
102 predetermine aspects of future change. This does not mean that global temperature will keep
103 rising indefinitely if emissions cease; rather, it refers to impacts already locked in by past
104 warming, such as ongoing sea level rise, ice-sheet melt, and thawing permafrost (Schuur et al.
105 2015, IPCC 2023, Abram et al. 2025). Components of the climate system are also thought to
106 include critical thresholds that, if breached, would trigger large, irreversible changes (Armstrong
107 McKay et al. 2022). Whether “committed climate change” will force the Earth system beyond
108 these tipping points is the subject of intense concern and scholarship (Steffen et al. 2018).

109 A clear parallel exists for biodiversity. Habitat loss, fragmentation, and over-exploitation have
110 driven elevated extinction rates and reduced many remaining species to small, isolated
111 populations (Pimm et al. 2014, Ralls et al. 2018). Persistent small population sizes set in motion
112 an extinction vortex: a reinforcing cycle of genetic loss and population decline (Figure 1; Gilpin
113 and Soulé 1986). In this feedback loop, stochastic effects (genetic drift) and inbreeding can result
114 in an accumulation of harmful mutations and lowered individual reproductive fitness and survival
115 (inbreeding depression), in turn driving further declines in population size. Genetic diversity can
116 continue to erode even after threats are mitigated or population numbers appear stable (Figure 1 –

117 maintenance trajectory; Gargiulo et al. 2025; Pinto et al. 2024). Indeed, even if a population
118 grows in size, some life-history traits (e.g., long generation length) create biological inertia,
119 producing a lag in which past impacts continue to define the future state of populations long after
120 disturbance (Kurland et al. 2023, Gargiulo et al. 2025). Over generations, these processes can
121 intensify, leading to irreversible loss of genetic variation and, ultimately, species extinction
122 (Frankham et al. 2017). This “genetic extinction debt” represents the first stage of a broader
123 extinction trajectory, reflecting similar time-lagged dynamics that underlie “extinction debt”:
124 situations in which species are likely to go extinct in the future because environmental conditions
125 can no longer support long-term persistence (Tilman et al. 1994, Hanski and Ovaskainen 2002).

126 The concept of committed climate change has mobilised consensus and action by underscoring a
127 critical insight: delayed action locks in future impacts. Therefore, to motivate policy discussions
128 and communicate urgency in the biodiversity sphere, we propose the term “committed genetic
129 erosion” to describe the future loss of genetic diversity already set in motion by past demographic
130 and landscape change. Like committed climate change, committed genetic erosion arises from
131 inertia and time lags, rather than inevitable global outcomes (Gargiulo et al. 2025); however, it
132 differs from the climate concept in two critical respects. Climate inertia locks in regional to
133 global impacts that are largely irreversible on human timescales (King et al. 2024, Perkins-
134 Kirkpatrick et al. 2025), whereas genetic inertia operates at the local scale and can, to a certain
135 degree, still be altered through timely intervention (Box 1). These attributes offer a window of
136 opportunity: targeted interventions can often attenuate the realisation of committed genetic
137 erosion. In this sense, the outcomes of ambitious action for biodiversity can be more immediate
138 and tangible than in climate systems. This tractability is a strategic advantage that conservation
139 policy and planning must now embrace.

140 Here we use climate science as a scaffold to re-frame conservation of genetic diversity and
141 explore the idea of committed genetic erosion through four parallels: (1) multiple lines of
142 evidence show change is already occurring; (2) time lags mean that further change is committed
143 under business as usual; (3) we can measure, track, and forecast risk to guide early action; and (4)
144 proactive measures are needed that address root causes. The need for action is supported by
145 strong evidence, broad consensus, and increasingly sobering predictions for the future (Hoban et
146 al. 2021, Exposito-Alonso et al. 2022, Gargiulo et al. 2025, Shaw et al. 2025). Business as usual,
147 primarily reactive conservation (acting only after loss is observed; Figure 1), could lead to a
148 future of ongoing genetic loss that cascades into species extinction, triggering broader
149 biodiversity collapse. By recognising the scale of this challenge and the importance of early
150 action, we hope to catalyse the long-called-for paradigm shift in conservation practice (Ralls et
151 al. 2018), placing genetic diversity at the core of biodiversity conservation strategies.

152

153 **Parallel 1: Change is happening now**

154 Human-caused climate change is now undeniable. Human activities have unequivocally warmed
155 the planet, with the long-term global surface temperature trend now estimated at 1.34–1.41 °C
156 above the 1850–1900 baseline (WMO 2025). Continued greenhouse gas emissions are driving
157 widespread and rapid changes across the atmosphere, ocean, cryosphere, and biosphere, and
158 climate change is influencing weather and climate extremes in every region of the world (IPCC
159 2023).

160 Like climate change, multiple lines of evidence show that major genetic erosion of species is
161 underway. This is true for a wide range of taxa and regions, even over just a few generations, and

162 even for species not currently considered threatened (Schmidt et al. 2023, Mastretta-Yanes et al.
163 2024, Shaw et al. 2025). This trend aligns with recent studies on North American birds and
164 mammals (Schmidt et al. 2020), global plant populations (González et al. 2020), and global
165 predictions suggesting more than 10 % of genetic diversity may already be lost (Exposito-Alonso
166 et al. 2022). The widespread signal of genetic erosion, even over short time frames, suggests that
167 substantial population size declines are driving this loss.

168 Shaw et al. (2025) found that genetic diversity loss was most pronounced where land use change,
169 disease, abiotic disturbances (e.g., wildfire, floods), and harvesting or harassment (e.g., hunting,
170 logging, stressing) occurred. Many species now persist in small, fragmented populations
171 (Frankham 2022). These conditions are expected to erode genome-wide diversity and increase
172 inbreeding (Soulé 1980); predictions well-supported by experimental data (Rich et al. 1979,
173 Montgomery et al. 2000), and real-world observations (e.g., Figure 2; Hoelzel et al. 2024; Ralls
174 et al. 2018). A recent global assessment of genetic indicators across 5,271 populations also found
175 that while the majority of species maintain most of their historic populations (for now), for 58 %
176 of species assessed, all populations were too small to sustain genetic diversity over time; thus
177 reducing their adaptive capacity (Mastretta-Yanes et al. 2024). Massive population losses could
178 follow in the near-term, even triggering extinction vortices for many taxa (Figure 1; Gilpin and
179 Soulé 1986).

180 Human-induced changes can create conditions for rapid evolutionary responses (IPBES 2019),
181 but the capacity for such adaptation depends largely on existing (standing) genetic variation
182 (Allendorf et al. 2024, Mergeay 2024). The kelp (*Ecklonia radiata*) forests of Western Australia,
183 where the interdisciplinary study of marine heatwaves emerged (e.g., Hobday et al. 2016),
184 provides a well-documented example. In 2011, a marine heatwave caused widespread kelp

185 canopy loss. Populations with higher genetic diversity were more likely to persist through the
186 event, while low-diversity stands collapsed (Wernberg et al. 2018). In the following years,
187 recovery was aided by recolonisation from nearby forests, with high connectivity enabling the
188 gene flow needed to maintain diversity and support adaptation to warming conditions (Vranken
189 et al. 2025). These studies show that genetic diversity and gene flow are essential for persistence
190 and resilience under accelerating change. Genetic diversity thus comprises the within-species
191 component of “response diversity”, a key aspect of ecological resilience that reflects the capacity
192 of populations to respond differently to environmental change (Nadeau and Urban 2019, Walker
193 et al. 2023). Managing diversity and connectivity is central to maintaining such flexibility and is
194 a cornerstone of “resilience thinking” in ecosystem management (Biggs et al. 2012). Thus,
195 conserving genetic diversity is fundamental to conserving species and ecosystems, and ensuring
196 their long-term resilience (Figure 1). Because the recovery of diversity through mutation often
197 requires hundreds or thousands of generations (Lande and Barrowclough 1987), preventing
198 genetic erosion is critical (Figure 2).

199

200 **Parallel 2: Change will continue under business as usual**

201 Many human-driven climate change impacts are locked in for decades to millennia, even under
202 rapid emissions reduction scenarios (Abrams et al. 2023, IPCC 2023). Because of the thermal
203 inertia of the climate system, global temperatures are expected to continue rising for several
204 decades even if greenhouse gas concentrations were stabilised today (Abrams et al. 2023). Some
205 scenarios even involve a temporary “overshoot”, where temperatures exceed 1.5–2 °C before
206 later declining through mitigation, yet many ecosystems will not return to their prior states once

207 thresholds are crossed (Armstrong McKay et al. 2022, Abrams et al. 2023). Under current
208 national commitments (business as usual), critical thresholds for several major tipping elements
209 are likely to be crossed this century, including the loss of Greenland and West Antarctic ice
210 sheets, the Amazon rainforest, permafrost carbon stores, and tropical coral reefs (Steffen et al.
211 2018, Armstrong McKay et al. 2022, Abrams et al. 2023, IPCC 2023, Abram et al. 2025). Once
212 these thresholds are transgressed, self-reinforcing feedbacks could drive cascading and
213 irreversible impacts (Schuur et al. 2015, Armstrong McKay et al. 2022, IPCC 2023, Wunderling
214 et al. 2023).

215 Similar lagged responses also occur in biological systems, including in population-level genetic
216 processes. The effective population size (N_e) is the genetic analogue of census size (N_c); where
217 N_c shapes the ecological dynamics of a population, N_e captures its evolutionary dimensions,
218 determining the strength of genetic drift and the resulting rates of inbreeding accumulation and
219 genetic diversity loss (Waples 2025). Theoretical models predict that the time it takes for genetic
220 diversity loss to manifest fully after a demographic decline is on the order of $2N_e$ generations
221 (Crow and Aoki 1984). For many populations and species, this corresponds to hundreds to
222 thousands of generations. This means that genetic erosion may go undetected long after
223 populations begin to decline, potentially giving the false impression that genetic diversity is
224 stable (Lande 1988). Global genetic diversity losses of 13–22 % have been estimated relative to
225 recent baselines, and long-term losses are forecast to reach 42–48 % even if population numbers
226 stabilise, with greater losses under high habitat fragmentation (Mualim et al. 2025 [preprint]
227 doi:10.1101/2024.10.21.619096).

228 Species with large populations and broad distributions can hold a reservoir of genetic diversity
229 that masks local losses, and seed banks and dormant life stages can retain genetic signatures of

230 past conditions (Plue et al. 2017, Staude et al. 2020). For example, genetic diversity was
231 apparently maintained over time in taxa with historically large and/or widely distributed
232 populations, such as insects (including bees and flies), ray-finned fishes (pelagic and migratory),
233 and dicotyledonous plants (Shaw et al. 2025). But these groups may still be at risk - now or in the
234 future - if large demographic changes occur. Similarly, in long-lived species, even substantial risk
235 of genetic erosion may not be detectable over relatively short-term sampling periods (Gargiulo et
236 al. 2025), potentially leading to underestimates of extinction risk.

237 These lagged genetic responses mirror climate dynamics, where gradual change can culminate in
238 abrupt, often irreversible shifts. As resilience declines, stressors may tip an already vulnerable
239 system into collapse. For example, long-term climatic “presses” (e.g., gradual warming or drying)
240 interacting with short-term “pulses” (e.g., droughts or fires) can trigger abrupt transitions, as seen
241 in population collapses and regime shifts worldwide (Biggs et al. 2018, Harris et al. 2018).

242 Similar dynamics may have contributed to the extinction of the woolly mammoth (*Mammuthus*
243 *primigenius*) on Wrangel Island, which lacked human occupation (Dehasque et al. 2024). Ancient
244 DNA revealed a severe population bottleneck followed by demographic recovery, but eventual
245 collapse. Hundreds of generations of inbreeding and the accumulation of mildly deleterious
246 mutations may have left the population close to a tipping point (Dehasque et al. 2024). Such
247 processes also occur in agriculture, where selective breeding reduces genetic diversity and
248 heightens disease vulnerability, as seen in the devastating effects of southern corn leaf blight on
249 maize in the USA (Strange and Scott 2005).

250 The consequences of genetic erosion can also feed back into, and exacerbate, climate and
251 ecosystem processes. Increasing evidence shows that transgressing key planetary boundaries,
252 whether related to climate change or to the wider biosphere, can amplify risks across domains

253 (Richardson et al. 2023). Genetic diversity is a key component of biosphere function, and its
254 erosion can weaken both ecological and climate resilience. For example, black spruce (*Picea*
255 *mariana*) is projected to experience slower growth due to warming, threatening its carbon
256 sequestration capacity. Genetic diversity may enable local adaptation to these conditions (Robert
257 et al. 2024), but loss of this adaptive capacity may have cascading effects for carbon
258 sequestration and climate dynamics.

259 While some committed climate change impacts will play out over centuries to millennia, giving
260 society the opportunity to plan and adapt, others are likely in the coming decades (Abrams et al.
261 2023, IPCC 2023). Similarly, committed genetic erosion represents an imminent and
262 compounding hazard (Exposito-Alonso et al. 2022, Shaw et al. 2025). Acting early and within
263 this lag period, can prevent irreversible losses of adaptive potential before they cascade into
264 species extinctions, disruptions to agriculture, ecosystem collapse, and broader climate
265 instability.

266

267 **Parallel 3: We can forecast risk to guide early action**

268 By the time genetic erosion is measurable, species are already at heightened risk of extinction,
269 even if demographic recovery occurs (Figure 2). Once genetic variants (alleles) are lost from all
270 populations, they are effectively irrecoverable on management timescales and long-term adaptive
271 potential is permanently reduced (Lande and Barrowclough 1987, Allendorf et al. 2024).

272 Preventing committed genetic erosion means identifying at-risk species *before* loss occurs, and
273 intervening while genetic diversity can still be restored. Preventing further erosion once it has
274 begun also remains an important and often necessary goal (Figure 1).

275 Detecting the warning signs of committed genetic erosion requires monitoring both genetic and
276 demographic indicators. Ecological evidence, such as population decline, habitat fragmentation,
277 or reduced dispersal, signals that genetic extinction debt is accumulating (O’Brien et al. 2022).
278 Such trends are already reflected in major extinction-risk frameworks, including the IUCN Red
279 List. However, a population’s genetic diversity does not always align with Red List status
280 (Schmidt et al. 2023, Mastretta-Yanes et al. 2024, McLaughlin et al. 2025). This mismatch may
281 be consistent with the expected lag between population decline and detectable genetic erosion.
282 Where genetic or genomic data exist, genetic metrics (Box 1) can reveal reduced gene flow,
283 bottlenecks, recent inbreeding or other early signals that genetic erosion is already underway,
284 even when overall diversity appears superficially stable (Gargiulo et al. 2025, McLaughlin et al.
285 2025).

286 Correctly interpreting genetic statistics as rates of loss, and thus measures of risk, requires clear
287 baselines. Without this, we risk accepting current, degraded conditions as the new norm
288 (“shifting-baseline syndrome”; Lotze and Worm 2009; Pauly 1995). Establishing consistent,
289 globally comparable reference points is therefore critical for detecting the consequences of
290 demographic stress early and enabling more proactive, targeted responses. Inspired by Essential
291 Climate Variables (ECVs; GCOS 2010), biodiversity scientists developed Essential Biodiversity
292 Variables (EBVs; Pereira et al. 2013) to quantify and report biological change worldwide. In the
293 genetic domain, EBVs formalise key population genetic measures (Hoban et al. 2022), translating
294 heterogeneous datasets into comparable metrics and providing a common language for
295 monitoring across species and regions. As with ECVs, some EBVs have been translated into
296 simplified indicators, with clear thresholds for decision-makers. Genetic diversity indicators,
297 such as the proportion of populations within a species that are sufficiently large to maintain

298 genetic diversity ($N_e > 500$), and the proportion of populations maintained (Hoban et al. 2021),
299 provide a simple yet powerful way to track genetic extinction debt and the extent to which
300 populations have already fallen below levels needed to sustain adaptive potential. These
301 indicators are already being implemented in policy under the CBD (Convention on Biological
302 Diversity 2022) and the EU Habitats Directive (O'Brien et al. 2025), although genetic diversity
303 remains absent from major restoration frameworks (Ngeve 2025). Although thresholds in other
304 EBVs such as heterozygosity and habitat fragmentation have been proposed and are the focus of
305 active discussion (Andersson et al. 2022, Kurland et al. 2024), these are not yet established as
306 critical tipping points and remain less operationalised than N_e thresholds. Further work to
307 formalise these thresholds is urgently needed.

308 Forecasting future genetic change is equally vital for guiding proactive interventions. Simulations
309 underpinned by physical climate models have been critical in climate science, projecting likely
310 outcomes under different emissions and socio-economic scenarios (IPCC 2023). In population
311 genetics, computational simulations serve a similar purpose: they help characterise tipping points,
312 lags, and thresholds for genetic erosion, and forecast how demographic trends and landscape
313 change will influence future diversity (Hoban 2014, Pinto et al. 2024). These tools also enable
314 biodiversity managers to evaluate the consequences of various strategic options (e.g., population
315 translocations) before losses occur, enabling them to identify when populations are likely to fall
316 below viable thresholds, where gene flow must be maintained, and where early intervention can
317 prevent committed genetic erosion.

318 Forecasting is critical in landscapes where economic development and biodiversity conservation
319 are in direct tension, facilitating decisions on whether natural populations remain viable or are
320 pushed toward irreversible genetic decline through disturbance. Examples include the Pilbara

321 region of Western Australia, where mining removes cave roosts of threatened bat populations and
322 fragments the roost network needed to maintain gene flow (Umbrello et al. 2022), and Brazil’s
323 Atlantic Forest, where agriculture, logging, and urban expansion fragment habitats (Galán-Acedo
324 et al. 2023). Simulations built to model such systems can help determine which areas or corridors
325 must remain intact to avoid committed genetic erosion, and where limited development may still
326 be compatible with maintaining connectivity (e.g., Dutcher et al. 2023; Haller and Messer 2023).
327 In many cases, avoiding genetic decline will require leaving some land or resources unexploited,
328 analogous to the climate concept of “unburnable carbon”, in which meeting climate targets
329 requires leaving fossil-fuel reserves untapped (Griffin et al. 2015, Pye et al. 2020). Like these fuel
330 reserves, land set aside for biodiversity objectives would become what climate governance terms
331 a “stranded asset”: economically valuable on paper, but unusable in practice without committing
332 populations to long-term genetic decline (Ploeg and Rezai 2020, Caldecott et al. 2021, Österblom
333 and Blasiak 2021). As in climate governance, implementing these limits depends on political will
334 and overcoming resistance from actors who stand to lose economically, while explicitly
335 addressing trade-offs to prevent conflict and ensure equitable outcomes (Ferns and Amaeshi
336 2021, Meyfroidt et al. 2022, Wright and Nyberg 2024).

337

338 **Parallel 4: We need proactive and ambitious action now**

339 Just as climate and biological systems exhibit inertia, so too do political, institutional, and
340 economic systems. Overcoming societal inertia requires systemic change in governance and
341 finance, not just scientific clarity, as climate governance has shown. Mechanisms such as the
342 Task Force on Climate-Related Financial Disclosures (TCFD) and the Science-Based Targets

343 Initiative (SBTi) evolved from voluntary commitments into widely adopted frameworks that now
344 inform mandatory reporting and incentivise corporate leadership (Ben-Amar et al. 2024).
345 Biodiversity reporting is beginning to follow suit through initiatives such as the Nature Positive
346 Initiative, the Task Force on Nature-related Financial Disclosures (TNFD), and the Global
347 Reporting Initiative’s biodiversity standard, but genetic diversity remains largely absent from
348 these frameworks (O’Brien et al. 2025). Nevertheless, promising examples exist in the private
349 sector of companies recognising the value of assessing and maintaining genetic diversity (e.g.,
350 commercial wild salmon fisheries; Connors et al. 2022).

351 Challenges do, however, remain. Climate change is perceived as a global phenomenon with
352 standardised metrics (e.g., tonnes of CO₂), whereas biodiversity loss is inherently local, context-
353 dependent, and lacks a universally accepted measurement unit (Wauchope et al. 2024). This
354 imbalance contributes to “carbon tunnel vision”, where emissions dominate sustainability
355 reporting while biodiversity metrics are overlooked (Jouffray et al. 2025), despite emerging
356 quantitative thresholds for ecosystem resilience. For example, when semi-natural habitat area
357 falls below roughly 20–25 % (at a 1 km² scale), critical ecosystem functions collapse.

358 Alarmingly, about two-thirds of agricultural and urban areas globally fall below this threshold
359 (Mohamed et al. 2024). Genetic diversity offers a complementary and quantifiable dimension of
360 resilience, providing standardised, scalable measures that link local biological processes to global
361 monitoring and forecasting frameworks (Mastretta-Yanes et al. 2024, Henry 2025). Embedding
362 these metrics within global monitoring, finance, and policy systems is increasingly feasible, and
363 supported by a growing number of tools and case studies (Hoban et al. 2021, Mastretta-Yanes et
364 al. 2024).

365 Genetic diversity is declining across both threatened and non-threatened species (Mastretta-
366 Yanes et al. 2024, Shaw et al. 2025), yet conservation policies and subsequent actions are often
367 triggered only after a species' total population has declined to critical levels, typically in response
368 to legal obligations or political pressure (Drechsler et al. 2011). This reflects a reactive mindset,
369 intervening only once risk is clear, and far more difficult (or impossible) to reverse. Resilience
370 science stresses that acting early, before thresholds are crossed, is far more effective (Folke et al.
371 2004). Furthermore, the costs associated with re-establishing depleted populations can be much
372 higher than those required to maintain them above critical thresholds, despite the latter being
373 needed over longer periods (Drechsler et al. 2011). Proactive management is therefore cost-
374 effective and more likely to prevent irreversible loss.

375 The belief that emerging technologies can compensate for inaction can reinforce a reactive
376 mindset and entrench the status quo, diverting attention from root causes and delaying
377 meaningful change (Anderson and Peters 2016, Gifford et al. 2018). Despite decades of
378 warnings, the world continues to rely on technological promises such as large-scale carbon
379 capture, which the IPCC (2022) judges unlikely to deliver substantial emission reductions by
380 2030. If these technologies fail to perform as expected, society will be locked into a high-
381 temperature trajectory (Anderson and Peters 2016). Biodiversity conservation faces similar risks.
382 Advances such as CRISPR and cloning cannot recreate the evolutionary potential lost through
383 extinction, which relies on large, diverse, self-sustaining populations. De-extinction efforts
384 inevitably begin with extremely low genetic diversity, exposing "resurrected" populations to a re-
385 extinction vortex (Steeves et al. 2017). While these technologies may have legitimate
386 applications in targeted management, they remain limited. The urgent need to reduce emissions

387 and to protect and restore genetic diversity in existing species, will rest primarily on investment
388 in existing, proven actions.

389 Time lags offer a critical opportunity: the chance to intervene before losses become irreversible,
390 and proactive management offers reason for hope (Figure 1). Because genetic diversity often
391 declines more slowly than population size, detecting demographic problems and restoring
392 populations before genetic erosion becomes committed, offers a window for action. For example,
393 the Eastern North Pacific population of fin whales (*Balaenoptera physalus*) suffered a 99 %
394 reduction in N_e during twentieth century whaling, yet retained relatively high diversity. This was
395 likely because the bottleneck lasted about 70 years, a short period relative to the species' long
396 generation time (~25.9 years), corresponding to three generations (Nigenda-Morales et al. 2023).
397 A moratorium on commercial whaling was implemented within this window for action, enabling
398 demographic recovery before genetic erosion was committed.

399 Species and populations face many different kinds of genetic threats, and discriminating among
400 these helps determine the most effective management actions (O'Brien et al. 2022). A global
401 meta-analysis suggests that several interventions can stabilise or improve genetic diversity,
402 including supplementation (Shaw et al. 2025). Supplementation spans a continuum from
403 facilitating natural gene flow via restored connectivity to more intensive translocations among
404 isolated or captive populations to effect genetic rescue. Given how widespread habitat
405 fragmentation has become, restoring gene flow is now essential in many systems (Frankham et al.
406 2017). Once approached with hesitancy, well-planned genetic rescue is increasingly recognised
407 as a powerful tool for reversing inbreeding depression and restoring adaptive potential (Frankham
408 et al. 2017, Ralls et al. 2018, Hoffmann et al. 2021). Such approaches also facilitate climate-
409 resilient assisted gene flow, identifying and moving adaptive variation to help populations track

410 rapidly changing environments (Meek et al. 2025). When implemented early in recovery
411 planning and conservation management, these approaches are low-risk, relatively inexpensive,
412 and capable of sustaining adaptive potential before genetic erosion becomes committed
413 (Frankham et al. 2017, Ralls et al. 2018, Fitzpatrick et al. 2023, da Silva et al. 2026).

414

415 **Conclusion**

416 Conserving genetic diversity is not a niche concern; it is central to long-term ecological resilience
417 and must become a core component of biodiversity planning. Genetic erosion is widespread, can
418 lag behind demographic losses, and becomes committed unless proactive steps are taken to
419 maintain and restore the evolutionary processes that sustain life on Earth. Doing so requires a
420 shift in thinking: from reacting only after genetic erosion is detected, to also proactively
421 safeguarding the raw material of resilience before it disappears.

422 Drawing from climate science, the concept of committed genetic erosion underscores the urgency
423 of action. Just as past greenhouse gas emissions have committed the planet to future climate
424 change impacts and potential tipping points, past and ongoing habitat loss, fragmentation, and
425 other consequent impacts on population decline, have committed many species to future genetic
426 erosion. The critical difference is that, unlike climate inertia, genetic inertia is tractable: erosion
427 unfolds population by population and can be prevented if action occurs during the lag period. We
428 cannot afford to miss this window for strategic intervention. Integrating genetic considerations
429 into planning and policy early maximises the capacity for adaptation and recovery. To help
430 achieve this, scientists, policymakers, practitioners, the private sector, and broader society all

431 play a role in ensuring that genetic considerations are communicated clearly, embedded early in
432 decisions, and underpinned by rigorous thresholds and terminology.

433 The solutions are well tested. In addition to monitoring, we must also apply genetic principles to
434 management: ecological strategies that achieve genetic outcomes. Established conservation
435 strategies, such as maintaining or improving connectivity, effecting carefully designed genetic
436 rescue, and supporting large, viable populations, sustain the evolutionary processes that maintain
437 genetic diversity. As such, they carry explicit genetic benefits that are too often overlooked.
438 Recognising these genetic outcomes gives renewed urgency to implementing these strategies
439 effectively. In much of the world, this will require applying genetic principles without direct
440 genetic data, and prioritising large-scale actions such as habitat protection and restoration. By
441 doing so, we intervene earlier in the extinction trajectory, acting before species are on the brink
442 of extinction. These approaches align with major international biodiversity and development
443 goals, supported by technical indicators that identify where losses are occurring, and which
444 species are most at risk. Now, bold, resilience-focused biodiversity action that embeds genetic
445 considerations into governance, finance, planning, and on-ground management is essential for
446 maintaining adaptive potential in a rapidly changing world. The choices made today will shape
447 the evolutionary futures of species and the ecological systems on which we depend, for
448 generations to come.

449

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458

459 **Conflicts of Interest**

460 The authors declare no conflict of interest.

461

462 **Data Availability Statement**

463 No new data were generated or analysed in support of this article.

464

465 **Biographical narrative**

466 The co-authors form an interdisciplinary team spanning conservation genetics, climate science,
467 resilience research, and biodiversity policy, based at institutions across four continents
468 encompassing government agencies, universities, botanical gardens, NGOs, and international
469 research organisations.

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Box 1: Signals of committed genetic erosion

Measurable changes in a population's genetic diversity are rarely the earliest warning of risk. The first indications that genetic extinction debt is accumulating are habitat loss, fragmentation and demographic decline, detectable through field monitoring or even remotely via satellite imagery (Hoban et al. 2024). These ecological signals indicate that declines in **effective population size (N_e)** and subsequent **genetic erosion** are likely.

Genetic diversity analyses reveal erosion already underway. Because components of genetic diversity respond to perturbation at different rates (Nei et al. 1975), comprehensive analysis can help quantify the severity, timing and trajectory of committed genetic erosion and highlight where intervention is most urgent (Frankham et al. 2010, Hoban et al. 2024).

Like species richness and diversity in ecology (Hu et al. 2006), alpha-level genetic diversity is commonly summarised using **richness** (e.g., **allelic richness, AR**; **Watterson's theta, θ_w**) and **evenness** (**gene diversity**, also called **expected heterozygosity, H_e** ; **nucleotide diversity, π**) metrics. Richness captures how many genetic variants (alleles) are present and is informative for long-term adaptive potential, while evenness reflects how allele frequencies are distributed and reflects short-term evolutionary responses (Caballero and García-Dorado 2013, Allendorf et al. 2024, Mergeay 2024).

While ongoing genetic monitoring is the gold standard, genetic erosion can also be detected from a single genetic snapshot. For example, following a population bottleneck, rare alleles are lost first, so richness declines rapidly. Evenness declines more slowly because it is

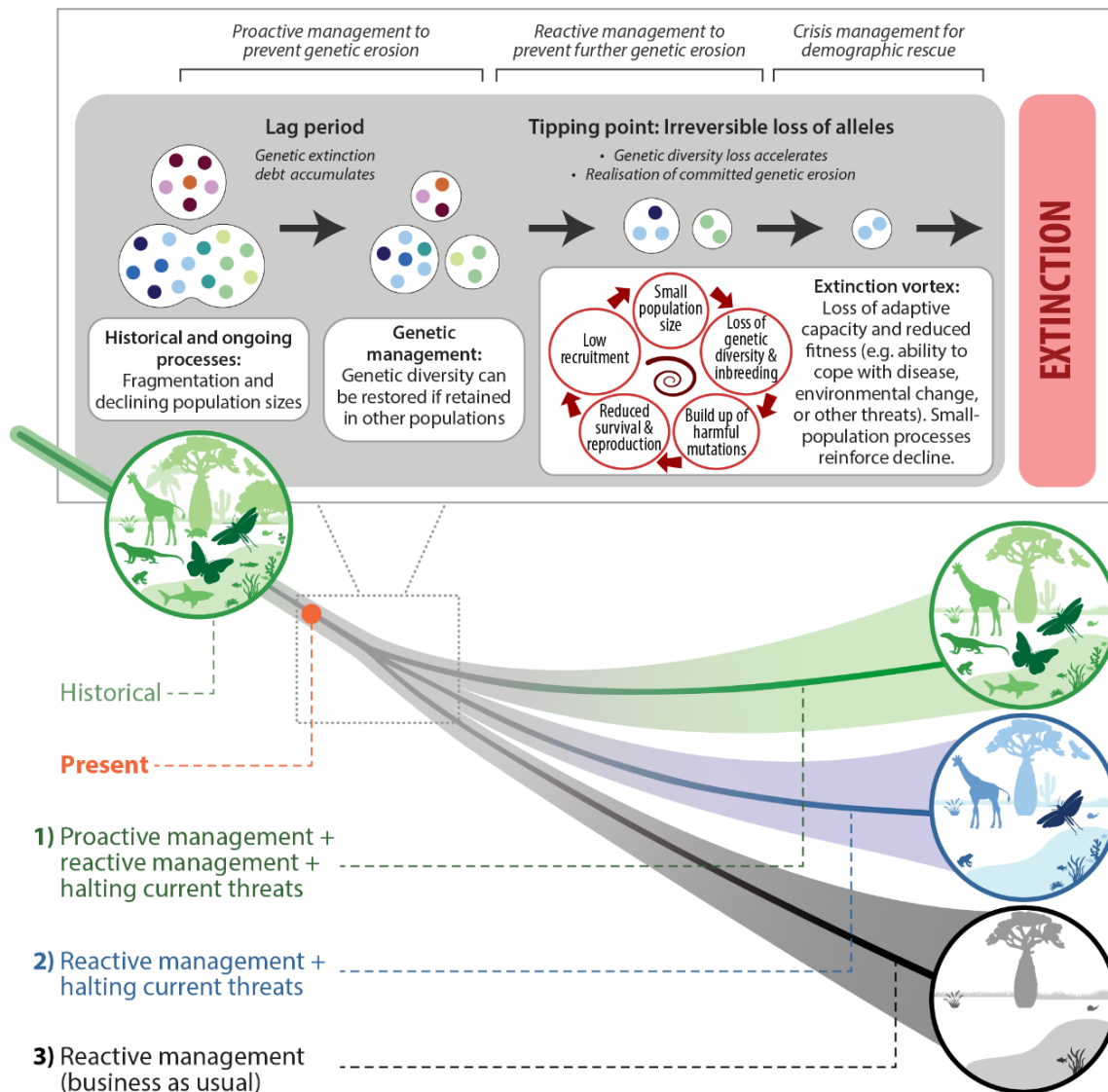
predominantly determined by common alleles. During or shortly after demographic decline, these properties manifest as a mismatch between metrics, indicating that the population is not at mutation–drift equilibrium (Cornuet and Luikart 1996).

Reduced connectivity can generate similar mismatches: when gene flow drops below roughly one effective migrant per generation, subpopulations rapidly lose rare alleles even without detectable demographic decline (Broquet et al. 2010). This mismatch indicates a genetic extinction debt for evenness, as further loss of H_e or π is inevitable unless population size or connectivity increases. More generally, reduced gene flow increases genetic drift and inbreeding within each subpopulation, lowering N_e at the species level (Ryman et al. 2019). These non-equilibrium dynamics can unfold unevenly across a species' range, especially if subpopulations are permanently lost (Kurland et al. 2023).

Tests for identifying genetic erosion from a single snapshot include Bottleneck (Cornuet and Luikart 1996), Tajima's D (Tajima 1989), runs of homozygosity (ROH; Kardos et al. 2018), linkage disequilibrium (Santiago et al. 2025), or site-frequency spectrum analyses (Gutenkunst et al. 2009).

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844 **Figure 1. Conceptual diagram of biodiversity trajectories under alternative genetic**

845 **management pathways.** The top panel illustrates the population-level timeline occurring within

846 the strategic window for action (grey dotted box): the period in which we can intervene before

847 the realisation of committed genetic erosion pushes many species toward extinction.

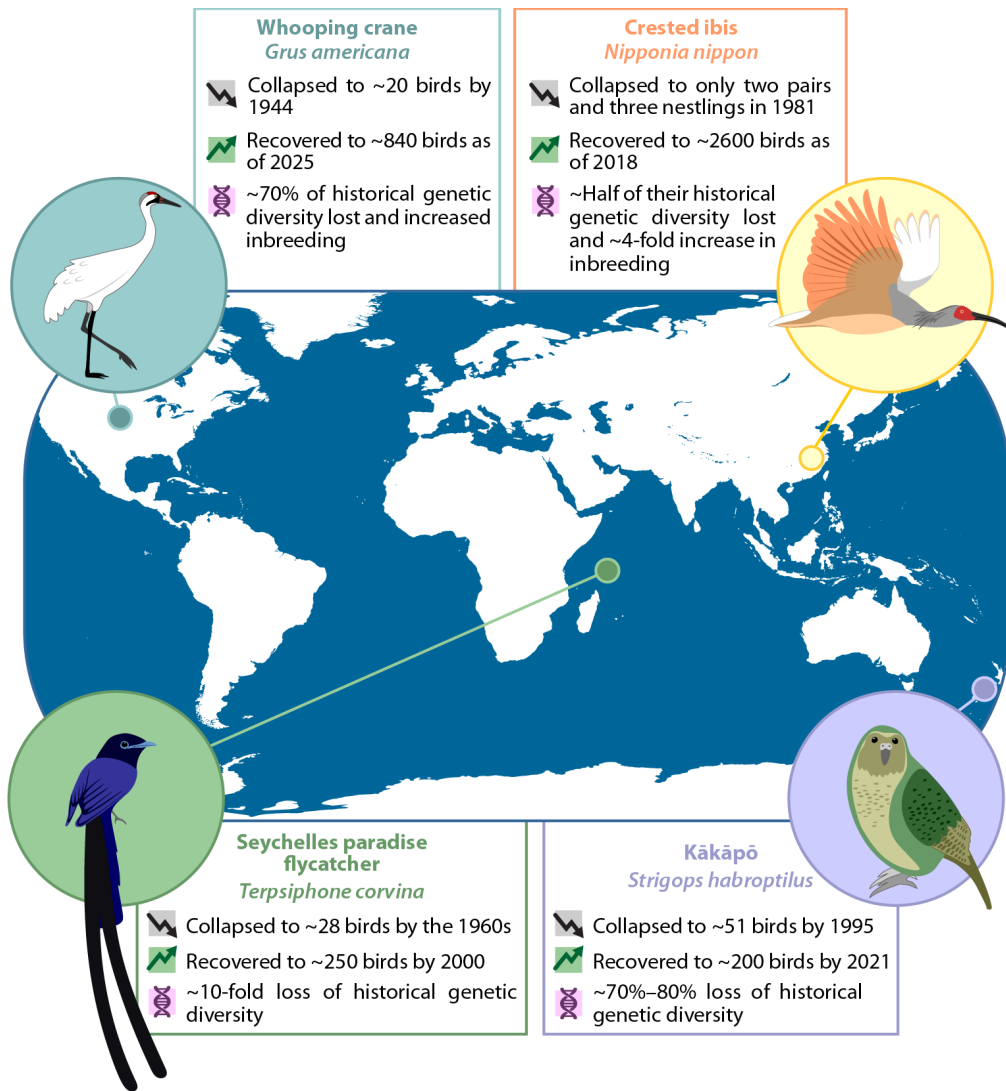
848 Fragmentation and population decline initiate genetic erosion, causing some immediate loss, and

849 resulting in a lag period in which genetic extinction debt accumulates. If genetic variants persist

850 in nearby populations, diversity at the species level can be restored through early intervention,

851 such as by mixing populations (supplementation). After a tipping point, however, genetic erosion

852 accelerates and becomes irreversible, reducing adaptive capacity and fitness, and reinforcing
853 small-population processes in a feedback loop (the “extinction vortex”). As systems move
854 through these stages, management shifts from proactive (preventing loss) to reactive (preventing
855 further loss) and finally to crisis management. At this stage, saving the species becomes
856 extremely difficult and/or costly, and extinction risk is elevated due to severely depleted genetic
857 diversity, even if demographic recovery is achieved. These often unseen genetic processes
858 underpin three potential biodiversity futures (bottom panel: represented by species diversity on
859 the y-axis and time on the x-axis), including: (1, green) recovery: proactive intervention to
860 maintain adaptive capacity before genetic erosion is measurable (e.g., restoring connectivity/gene
861 flow and sustaining large, viable population sizes), combined with threat abatement and reactive
862 responses where needed; (2, blue) maintenance: halting current threats (e.g., no further habitat
863 loss) alongside reactive management, reducing, but not preventing, future biodiversity loss; and
864 (3, black) losses under business as usual, where action is delayed until genetic erosion is
865 measurable (reactive management), resulting in biodiversity declining towards extinction, with
866 cascading losses across ecosystems.



867

868 **Figure 2. Global examples of realised committed genetic erosion in bird species - in all four**

869 **cases substantial genetic erosion has occurred despite demographic recovery.** Historical

870 baselines derived from 19th-century specimens reveal that four well-studied bird species (the

871 whooping crane, crested ibis, Seychelles paradise flycatcher and kākāpō) have each lost

872 substantial proportions of their historical genetic diversity following severe population declines.

873 Conservation actions such as translocations, captive breeding, predator control and habitat

874 protection were essential for securing or maintaining demographic recovery and, in several cases,

875 preventing extinction of the species. However, these efforts occurred after committed genetic

876 erosion had been realised. The irreversible loss of genetic diversity means that these species now
877 persist with reduced adaptive capacity and elevated long-term extinction risk. Demographic
878 recovery is a genuine conservation success, but genetic erosion persists even after numbers
879 rebound. *Data sources:* whooping crane: Fontseré et al. 2025; crested ibis: Feng et al. 2019;
880 Seychelles paradise flycatcher: Femerling et al. 2023; kākāpō: Bergner et al. 2016; Dussex et al.
881 2021.