

1 Ant mutualists as a biotic interaction filter of flowering plant
2 colonization on islands

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29

30 **Abstract**

31 **Aim:** Oceanic island floras are well-known for their disharmonic assemblages, with certain taxa and
32 functional groups being over- or underrepresented compared to their source pools, due to effects of
33 dispersal, environmental filtering, and biotic interactions. However, the role of biotic interactions in
34 generating this disharmony remains poorly explored.

35 **Location:** Global islands.

36 **Time Period:** Present-day species distribution.

37 **Major Taxa Studied:** Flowering plants bearing extrafloral nectaries (EFNs) and domatia, and their ant
38 partners.

39 **Methods:** We compiled plant distributions from the Global Inventory of Floras and Traits (GIFT) and ant
40 distributions from the Global Ant Biodiversity Informatics database (GABI) for islands worldwide. We then
41 assembled a dataset of ant genera known to interact mutualistically with plants bearing EFNs or domatia.
42 Using this dataset, we quantified the representational disharmony of ant-associated plants and evaluated
43 whether their ant partners act as a biotic interaction filter for the representation of ant-associated plants on
44 islands.

45 **Results:** We found that domatium-bearing plants are generally underrepresented, whereas extrafloral
46 nectary (EFN)-bearing plants are overrepresented, likely reflecting differences in their relationship with ant
47 partners. Besides the effects of island characteristics, the representation of both domatium- and EFN-
48 bearing plants shows a strong yet conditional relationship with the diversity of their interacting ant partners.
49 Specifically, this biotic interaction filter is more pronounced on larger, less isolated islands.

50 **Main Conclusions:** Our findings underscore that mutualistic interactions are a key driver of island plant
51 assembly and that their filtering effect depends on island characteristics. We conclude that explicitly
52 integrating biotic interactions and their context is essential to advance our understanding of island
53 biogeography.

54

55 **Keywords**

56 Domatia, extrafloral nectaries, source pool, island flora, island disharmony

57

58 **Introduction**

59 Understanding the mechanisms that govern the assembly of island biotas is a central topic in ecology and
60 biogeography (Warren *et al.*, 2015; Whittaker *et al.*, 2023). Oceanic islands frequently exhibit island
61 disharmony, where their species assemblages are distinct from mainland source pools due to the over- or
62 underrepresentation of certain taxonomic or functional groups (Carlquist, 1974; Weigelt *et al.*, 2015; Taylor
63 *et al.*, 2019; König *et al.*, 2020). These disharmonic patterns, often found in groups with specific traits, such
64 as long-distance dispersal ability or specific reproductive strategies, reflect the combined consequence of
65 processes including dispersal limitation (Gillespie *et al.*, 2012), environmental filtering (Carvajal-Endara *et*
66 *al.*, 2017), and *in situ* speciation (Weigelt *et al.*, 2015; Burns, 2019). While biotic interactions have been
67 recognized as a potential driver of island assembly (Pannell *et al.*, 2015; Taylor *et al.*, 2019), their specific
68 role in shaping disharmonic assemblages remain poorly examined.

69

70 Biotic interactions, such as competition, predation, and mutualism, significantly affect species distributions
71 and the composition of island plant assemblages (Connor *et al.*, 2013; Wisz *et al.*, 2013; Fowler *et al.*, 2023;
72 Delavaux *et al.*, 2024). The influences of these interaction filters may vary depending on the type of
73 interaction. In the case of mutualisms, interaction partners play an important role in plant dispersal, growth,
74 and reproduction (Bronstein, 2015). Consequently, the absence of a required mutualist on islands can
75 reduce plant fitness, leading to unsuccessful colonization or local extinction. This process contributes to the
76 underrepresentation of plants that depend on those partners. For example, recent studies suggest that
77 plants relying on pollinators for reproduction or mycorrhizal symbioses for nutrition acquisition may be
78 underrepresented on islands due to the lack of appropriate partners (Delavaux *et al.*, 2019, 2024;
79 Razanajatovo *et al.*, 2019). Furthermore, the distribution of interaction partner itself may be constrained by
80 distance from source pools and distinct island environments, leading to a joint influence on island
81 assemblages. For example, plants dispersed by animals (e.g. birds) likely gain an advantages in colonizing
82 isolated islands, such as the Azores (Heleno & Vargas, 2015). Yet, the extent to which this interaction filter
83 interplay with dispersal and environmental filters in affecting the distribution of associated species remain
84 insufficiently understood.

85

86 Island characteristics, particularly area and isolation, are fundamental in shaping the composition of island
87 floras as they directly influences processes like dispersal limitation and environmental filtering (Kreft *et al.*,
88 2007, 2010). Island isolation has long been recognized as a dispersal filter (Carlquist, 1966), which often
89 leads to the overrepresentation of species with long-distance dispersal ability on remote islands (Heleno &
90 Vargas, 2015). However, species lacking specialized dispersal syndromes can also constitute a substantial
91 proportion of island floras (Vargas *et al.*, 2012). Meanwhile, island area primarily constrains species
92 distribution by governing environmental heterogeneity. Larger islands tend to harbour higher habitat
93 heterogeneity (Weigelt *et al.*, 2013) and therefore impose fewer constraints on species colonization, as
94 demonstrated by the importance of climatic suitability for plant colonization on the Galápagos (Carvajal-

95 Endara *et al.*, 2017). Consequently, smaller islands may be associated with more distinctive composition
96 of island assemblages. In contrast, the complex environments of larger islands also are suggested to
97 promote *in situ* diversification, thereby increasing the disharmony of island floras (König *et al.*, 2020).

98

99 Ant-plant mutualisms are widespread and well-studied biotic interactions (Marazzi *et al.*, 2013; Weber &
100 Keeler, 2013; Chomicki & Renner, 2015). Ant-adapted plants have evolved specialised morphological
101 structures such as domatia and extrafloral nectaries (EFNs) that provide nesting sites or food rewards to
102 ant mutualists, who in return protect their plant mutualists from herbivores and plant competitors (Bronstein
103 *et al.*, 2006; Rico-Gray & Oliveira, 2007). By promoting the fitness of ant-associated plants on islands, ant
104 mutualists may contribute to the colonization of these species and ultimately contribute to a higher
105 representation compared to islands without ants. Therefore, ant-associated plants provide an ideal model
106 for studying key processes affecting the colonization and persistence of species in insular systems.

107

108 Here, we test the role of a biotic interaction filter in insular ant-plant mutualisms by analysing whether the
109 diversity of ant mutualists limits the representation of associated plants on islands. Given the strong spatial
110 congruence between ants and ant-associated plants (Luo *et al.*, 2023), we expect that a higher diversity of
111 ants interacting with plants promotes a greater diversity and representation of these plants on islands. To
112 address the general lack of comprehensive biogeographic knowledge for both interaction partners, which
113 is a key limitation in examining biotic interactions, we compiled a novel dataset by integrating recent
114 knowledge from two comprehensive global databases of plants and ants (Guénard *et al.*, 2017; Weigelt *et*
115 *al.*, 2020; Liu *et al.*, 2023). We first evaluate the representational disharmony of ant-associated plants by
116 comparing the observed species richness to null communities drawn from mainland source pools (König *et*
117 *al.*, 2020). Specifically, we ask these following questions: 1) Are domatium- and EFN-bearing plants over-
118 or underrepresented on islands relative to their mainland source floras; 2) Is plant representation affected
119 by the diversity of ant mutualists and island characteristics, and do these factors interact; 3) At the species
120 level, does the colonization probability of ant-associated plants correlate with the diversity of ant mutualists
121 and island characteristics?

122

123 **Materials and Methods**

124 **Plant and ant distributions**

125 We retrieved checklists of angiosperm species with domatia and extrafloral nectaries (EFNs) from
126 previously assembled lists based on comprehensive literature reviews (Weber & Keeler, 2013; Chomicki &
127 Renner, 2015). The original checklists were cleaned from duplicates and unspecified records, and then
128 taxonomically standardized according to the World Checklist of Vascular Plants (WCVP,
129 <https://wcvp.science.kew.org>; Govaerts *et al.*, 2021). Unmatched species were rechecked via the

130 Taxonomic Name Resolution Service (TNRS, <https://tnrs.biendata.org>; Boyle et al., 2021). Overall, we
131 retained 657 domatium-bearing and 3,179 EFN-bearing angiosperm species.

132

133 Plant distributions across islands and mainland regions were derived from checklists of native angiosperms
134 from the Global Inventory of Floras and Traits (GIFT version 2.1), which contains the distribution of 352,232
135 native plant species across 3,088 regions worldwide (Weigelt et al., 2020). As island geological history
136 exerts a strong influence on the biogeographical processes acting on island floras (Whittaker & Fernández-
137 Palacios, 2007), we focused on oceanic islands, defined as uplifted seafloor or volcanic islands (e.g. the
138 Azores). A few islands that once were fully submerged below the ocean were also considered oceanic
139 islands (e.g. New Caledonia). Therefore, we obtained 446 islands with plant information (288 oceanic
140 islands, 158 continental islands).

141

142 We sourced native ant occurrences on islands from the Global Ant Biodiversity Informatics database (GABI),
143 which is a comprehensive compilation of distributional data of 7,010 ant species across 2,678 islands
144 (Guénard et al., 2017; Liu et al., 2023). Due to the lack of a comprehensive list of ant species that interact
145 with domatia and EFNs, we first compiled a list of relevant ant genera based on a comprehensive literature
146 review. We then extracted species richness of these specific ants from the GABI database, separately for
147 domatium- and EFN-bearing plants. We refer to these two distinctive variables collectively as interacting
148 ant richness in subsequent analyses. However, they were treated separately in analyses for their respective
149 interaction types. As a result, we included 338 islands with both plant and ant distribution data available
150 (Fig. S1).

151

152 **Representational disharmony**

153 To infer the effects of biotic interaction filters on island plant assemblages, we adapted the framework by
154 König et al. (2020) to estimate mainland source pools for each island. Using this framework, we modelled
155 species turnover of native angiosperms among mainland regions based on climate conditions and
156 geographic distance. The fitted model was then utilized to predict species turnover among oceanic islands
157 and their potential source regions. For potential source pools, we included non-overlapping adjacent
158 mainland regions, which were primarily administrative units (e.g. countries, provinces, and states). When
159 overlapping regions were present, we prioritized smaller regions with available checklists. Plant checklists
160 for mainland regions were derived from the GIFT database. Additionally, we included 19 large continental
161 islands as potential source pools, as oceanic island floras may colonize from both mainland areas and other
162 large islands (Carlquist, 1967).

163

164 We quantified the representational disharmony of ant-associated plants by comparing the observed species
165 richness of ant-associated plants on each island with 1000 null assemblages randomly sampled from their
166 respective source pools, with sample sizes equal to the angiosperm species richness on the islands. The

167 null distribution of ant-associated plant species richness was assumed to follow a Poisson distribution with
168 a mean equal to the null expectation. We then calculated the probabilistic representational disharmony as
169 the cumulative probability that the null expectation is less than or equal to the observed richness. To further
170 evaluate the magnitude of disharmony, we transformed the probabilistic representational disharmony into
171 the standardized effect size of representational disharmony (SES.RD) using a simulated Poisson
172 distribution. A positive SES.RD value indicates an overrepresentation of ant-associated species on an
173 island, while a negative value indicates an underrepresentation. Islands with an expectation of ant-
174 associated plants less than 1 were excluded from further analyses, as it is statistically difficult to determine
175 whether the absence of ant-associated species deviates from the null expectation when the null expectation
176 approaches 0. In total, we measured the representational disharmony of domatium-bearing plants on 62
177 oceanic islands and 26 continental islands ($n = 88$), and EFN-bearing plants on 254 oceanic islands and
178 136 continental islands ($n = 390$). Different sets of number of islands in analyses are provided in Table S1.
179 For detailed methods on the source pool estimation and representational disharmony of ant-associated
180 plants on islands, see König et al. (2020) and Supplementary Methods.
181

182 **Climatic variables and island characteristics**

183 To evaluate climate conditions in delineating source pools, we chose four climatic variables hypothesized
184 to affect plant distributions on islands (Kreft & Jetz, 2007; Calixto et al., 2021; Cai et al., 2023): mean annual
185 temperature (°C), mean annual precipitation (mm/a), seasonality in temperature and precipitation (standard
186 deviation of the monthly mean temperature or precipitation). All climatic variables were derived from the
187 Climatologies at High Resolution for the Earth's Land Surface Areas (CHELSA; Karger et al., 2017) at a
188 resolution of 30 arc-seconds and measured as mean values for each region. Further, we used three island
189 biogeographic variables to examine the effects of island characteristics on the distribution of ant-associated
190 plants, as island characteristics are known to influence particular ecological processes. Island area (km²)
191 was measured as the total landmass area of each island. As a measure of island isolation, we considered
192 the proportion of surrounding landmass (SLMP), which takes stepping stones and coastlines into account.
193 SLMP has been shown to be a strong determinant of plant diversity on islands and is inversely related to
194 island isolation (Weigelt & Kreft, 2013). All island biogeographic characteristics were extracted from GIFT
195 (see Weigelt et al., 2020 for a detailed description of all workflows).
196

197 **Statistical analyses**

198 To examine the effects of biotic interactions on the representational disharmony of ant-associated plants
199 on islands, we applied linear regressions separately for domatium- and EFN-bearing plants. The
200 standardized effect size of representational disharmony was modelled against island area, isolation, and
201 the species richness of interacting ants. We performed a Pearson correlation analysis to avoid
202 multicollinearity (All Pearson's $r < 0.7$; Table S2). To meet model assumptions and reduce the skewness,

203 island area and interacting ant richness was natural log-transformed, and SLMP was logit-transformed. All
204 variables were then standardized to zero mean and unit variance for comparable results. Because domatia
205 and EFNs are most prevalent in the tropics (Fig. 1) (Weber & Keeler, 2013; Chomicki & Renner, 2015; Luo
206 *et al.*, 2023), we only focused on tropical islands located within 23.4° latitude north and south. Given that
207 biotas on oceanic islands are originally absent and represent an assembly processes from mainland source
208 pools (Whittaker & Fernández-Palacios, 2007), this analysis was restricted to oceanic islands with
209 evaluated representational disharmony and ant information, including 60 islands for domatia and 152
210 islands for EFNs.

211

212 Given that significant spatial autocorrelation was detected in the model residuals, we adopted simultaneous
213 autoregressive error (SAR) models using the 'errorsarlm' function of the R package 'spatialreg' (Dormann
214 *et al.*, 2007; Bivand & Wong, 2018). To define the optimal spatial structure, we constructed the neighbour
215 by varying the upper distance bound from 100km up to the maximum island neighbour distance (i.e. 3,672
216 km) with 11 intervals. Pseudo R², Akaike's information criterion (AIC), and residual spatial autocorrelation
217 were jointly considered to select the ideal distance band. We then applied the row standardization coding
218 scheme to weight the neighbour list of each island. Spatial autocorrelation in model residuals was controlled
219 in SAR models (Table S3). As interaction filters can interact with other processes simultaneously (Cadotte
220 & Tucker, 2017), we included the interaction between island characteristics and interacting ant richness in
221 all models. Both pseudo-R² and AIC significantly improved in models with the interaction term (Table S3).

222

223 To determine the factors influencing the colonization of individual ant-associated plants on islands, we
224 performed presence-absence regressions for each plant species. To ensure sufficient sample sizes and to
225 avoid class imbalance (Steen *et al.*, 2021), we only included species present on at least 30 oceanic islands.
226 This criterion led to 21 EFN-bearing plants for further analyses. All domatium-bearing plants were excluded,
227 as the most prevalent species (*Andira inermis*) occurred on only 15 study islands. We applied generalized
228 linear models (GLMs) of the binomial family for the presence/absence of each of the 21 individual species,
229 using interacting ant richness, island area, and the surrounding landmass proportion as explanatory
230 variables. To account for potential spatial autocorrelation, we corrected these models by adding a spatial
231 autocovariate. This autocovariate was created using the model residual of the non-spatial GLMs, weighted
232 by the neighbour distance defined in the SARs, and implemented using the R package 'spdep' (Bivand &
233 Wong, 2018). To assess the overall effects of ants and island characteristics across all 21 EFN-bearing
234 plants, we also performed a generalized linear mixed model (GLMM). This model incorporated a random
235 slope for both species and islands. Details on model results and performance are provided in Table S4.

236

237 **Results**

238 **Diversity of ant-associated plants on islands**

239 Across the 446 islands in our dataset, we recorded a total of 251 domatium-bearing plants and 1,248 EFN-
240 bearing plants (Fig. 1A, B). The diversity of ant-associated plants on islands showed a significant latitudinal
241 gradient. Both species richness and the proportion of ant-associated plants relative to total angiosperms
242 peaked in the tropics and decreased towards the poles. The Malay Archipelago harbours the greatest
243 number of these species, including 99 domatium-bearing plant species in Borneo and 332 EFN-bearing
244 plants in Singapore. The highest proportions were found on Little Cayman Island (2.7% domatium-bearing
245 plants) and Cocos (Keeling) North Keeling Island (21.2% EFN-bearing plants).

246

247 The representational disharmony (SES.RD) of plants bearing domatia and EFNs varied significantly and
248 showed pronounced geographical patterns on tropical oceanic islands (Fig. 1C, D). Domatium-bearing
249 plants were generally underrepresented on oceanic islands relative to their source floras (mean SES.RD: -
250 0.60 ± 0.17 , $n = 99$). Specifically, 6 islands were significantly overrepresented (SES.RD > 1.96 ; e.g. Little
251 Cayman Island), and 8 islands being significantly underrepresented (SES.RD < -1.96 ; e.g. New Caledonia)
252 (Fig. 1C, E). The disharmony patterns of domatium-bearing plants do not significantly differ between
253 oceanic and continental islands (Wilcoxon signed-rank test, $p = 0.46$). In contrast, we found that EFNs were
254 generally overrepresented in island floras relative to source floras (mean SES.RD: 1.55 ± 0.14 , $n = 390$). A
255 large proportion of studied islands (170 out of 390) was significantly overrepresented, particularly those in
256 Southeast Asia and the Caribbean region. Only 31 islands were significantly underrepresented (e.g. New
257 Caledonia; Fig. 1D, F). The representation of EFN-bearing plants on oceanic islands was more disharmonic
258 than on continental islands (Wilcoxon signed-rank test, $p < 0.001$).

259

260 **Biotic interaction filter among islands**

261 Among the simultaneous autoregressive error models tested, those including interaction terms performed
262 best (Table S3). We found that interacting ant richness significantly shaped the representation patterns of
263 both domatium- and EFN-bearing plants, but its effect depended on specific island characteristics (Fig. 2,
264 Table S3). For domatium-bearing plants, higher interacting ant richness was associated with greater
265 representation, particularly on larger and less-isolated islands (Fig. 3A, Fig. S2). EFN-bearing plants also
266 showed a stronger relationship with interacting ants on islands closer to the mainland (Fig. 3D). On the
267 other hand, the effect of island isolation was more pronounced on islands with higher interacting ant
268 richness for both interaction types (Fig. 3C, F). In contrast, island area showed a non-significant negative
269 relationship with EFN-bearing plants representation (Fig. 3E).

270

271 Of the 21 individual EFN-bearing plant species, 12 showed a significantly positive relationship with the
272 richness of their interacting ant partners (Fig. 4A; Table S4). Consistent with this prevalence of positive

273 effect, the model integrating all 21 species revealed an overall significant effect of interacting ant richness.
274 In contrast, island characteristics showed a more variable effect. Only six species exhibited a positive
275 relationship with island area (Fig. 4B), while results of island isolation was mixed, with eight species showing
276 a negative relationship and eight a positive one (Fig. 4C). Accordingly, neither island area and isolation had
277 a significant overall effect on colonization in the model including all 21 species.

278

279 **Discussion**

280 Our results confirm biotic interactions act as a key filter shaping island biota composition, demonstrating
281 that the richness of ant partners significantly contributes to the biogeography of ant-associated plants on
282 islands worldwide. Furthermore, we find the influence of biotic interactions is not uniform but conditional,
283 being more pronounced on larger, less isolated islands. These results highlight that integrating the mediate
284 influence of biotic interactions is essential to move beyond classical island characteristics and substantially
285 advance our understanding of disharmony in island assemblages.

286

287 **Diversity of ant-associated plants on islands**

288 Domatium-bearing plants were generally underrepresented on islands compared to their source pool floras.
289 Despite this overall pattern, myrmecophytes were notably overrepresented on some islands, e.g. Saibai
290 Island off the coast of New Guinea where many epiphytic *Hydnophytum* species with domatia occur. Indeed,
291 the genera *Hydnophytum*, *Myrmecodia*, and *Neonauclea* collectively account for the majority (58 out of 105
292 species) of all island endemic myrmecophytes and predominantly distributed across Indomalayan and
293 Australasian islands. The high endemism in this region suggests that *in situ* diversification may be a key
294 mechanism driving the disharmonic representation of myrmecophytes (Weigelt *et al.*, 2015). A notable
295 exception to this regional trend is New Caledonia, which hosts low representation of domatium-bearing
296 plants, despite its relatively close distance to source pools of high domatia diversity and its own diversity of
297 Rubiaceae. This discrepancy may reflect the influence of complex evolutionary history of myrmecophytes
298 (Chomicki & Renner, 2016).

299

300 Although EFN-bearing plants can be underrepresented on individual islands or archipelagos (e.g. the
301 Hawaiian Islands; Keeler, 1985), they unexpectedly constitute a high overall representation in island floras
302 at large spatial scales. This may be partly attributed to the broad distributions of dominant EFN-bearing
303 families, such as Euphorbiaceae and Fabaceae, which account for over 30% of all insular EFN-bearing
304 plants (Table S5). Furthermore, our data, though incomplete, indicate that insular EFN-bearing plants
305 exhibit a higher proportion of herbaceous (34% compared to 25%) and hydrochorous (15% compared to
306 7%) species compared to their global distribution (Table S6). This is consistent with general observations
307 that insular floras often favour non-woody species and those with long-distance dispersal ability (Heleno &

308 Vargas, 2015; König *et al.*, 2020), suggesting that the composition of EFN-bearing plants is a response of
309 dispersal limitation and environment filter.

310
311 The distinct difference in the representational disharmony observed between domatium- and EFN-bearing
312 plants may reflect fundamental differences in their interaction characteristics. Domatia are relatively
313 resource cost and mostly obligate, while EFNs are less resource intensive and range from obligate to
314 facultative interactions (Bronstein *et al.*, 2006; Marazzi *et al.*, 2013). While obligate interactions (e.g. ant-
315 domatia) tend to limit plant distributions, facultative interactions (e.g. ant-EFNs) expand them (Stephan *et*
316 *al.*, 2021). This is also supported by our findings that the diversity of ants that may interact with domatia is
317 significantly lower than that for EFNs (Fig. S1). Thus, the disharmony of these two ant-associated plant
318 groups appears to be shaped by a combined influence of dispersal, environments, and interaction filters.
319

320 **Biotic interaction filter among islands**

321 The importance of biotic interactions in shaping species diversity at large scales has been hitherto
322 overlooked, despite a growing body of literature highlighting the role of mutualistic interactions (Delavaux
323 *et al.*, 2019, 2024; Razanajatovo *et al.*, 2019; Taylor *et al.*, 2019). Using a new dataset containing the
324 distribution of plants and ants on islands, we found strong empirical evidence that a higher representation
325 of ant-associated plants on islands is consistently associated with higher diversity of interacting ants, a
326 pattern observed at both the island and species levels. This finding aligns with studies showing that biotic
327 interactions constrain species distributions in various habitats (Wisz *et al.*, 2013; Fowler *et al.*, 2023),
328 including alpine regions (Cavieres *et al.*, 2014), and significantly adds to the macroecological understanding
329 of island assembly (Taylor *et al.*, 2019; Schrader *et al.*, 2021). The absence of interacting ants can expose
330 ant-associated plants to increased herbivory pressure, limiting their success of colonization. Conversely,
331 ants can also benefit from plant-based resources to colonize and persist on islands. For example, extrafloral
332 nectar is a key resource that facilitates the successful colonization of islands by invasive ants (Savage *et*
333 *al.*, 2009; Sugiura, 2010).

334
335 We found that the effect of ants as interaction filters was more pronounced on less isolated islands. This
336 suggests that isolation imposes a primary filter on the distribution of ants and their associated plants, which
337 in turn governs the relatively importance of biotic interaction filters. On remote islands, long distance
338 precludes the establishment of many potential interaction partners, thereby decreasing the signal of biotic
339 interactions. In contrast, on islands closer the mainland sources, where more species could overcome
340 dispersal barriers (Morrison, 2016), the presence of specific ant partners becomes a critical determinant of
341 plant colonization. This interplay of abiotic and biotic filters consequently results in simplified interactions
342 and networks on islands (Traveset *et al.*, 2016).
343

344 At the species level, we observed that island isolation had a non-significant overall effect on the colonization
345 of EFN-bearing plants, yet it showed strong and contrasting effects among individual species. This paradox
346 can be explained by the fact that the geographic isolation of a specific island does not necessarily
347 correspond to its proximity to a particular species' distribution range. For example, *Erythrina variegata*,
348 primarily distributed in Southeast Asia, faces a significant dispersal barrier to the Caribbean islands, despite
349 the archipelago's relatively close distance to the American mainland (SLMP estimate = -1.12 ± 0.26 ; Table
350 S4). However, explicitly delineate potential islands and source pools for each species remains a complex
351 challenge (Carvajal-Endara *et al.*, 2017). Furthermore, our species-level analysis was restricted to widely
352 distributed species (those present on ≥ 30 oceanic islands), which are often near-cosmopolitan (e.g.
353 *Ipomoea pes-caprae*, *Thespesia populnea* etc). For these species, our results confirm that colonization
354 probability is strongly affected by the diversity of interacting ants.

355
356 The effect of island area on the representational disharmony of domatium-bearing plants was dependent
357 on ant diversity, while its effect on EFN-bearing plants was marginal. For domatium-bearing plants, larger
358 islands provides greater habitat heterogeneity, which can support more diverse ants but also lead to higher
359 herbivory pressure (e.g. from mammals; (Barreto *et al.*, 2021)). This can increase the selective advantage
360 of ant defence. Furthermore, large islands can promote *in situ* diversification and thus amplify disharmony.
361 This is reflected in the high proportion of domatium-bearing plants being endemic (16%; Table S7). In
362 contrast, EFN-bearing plants show a low proportion of endemic species (5%), well below the global average
363 for insular angiosperms (21%; (Schrader *et al.*, 2024)). This pattern aligns with EFNs loss in endemic
364 lineages on islands where ant partners are absent, such as the Bonin Islands (Sugiura *et al.*, 2006). Given
365 that EFN is an evolutionarily labile trait (Weber & Keeler, 2013), we hypothesize that in the absence of
366 interacting ants partners, plants may fail to maintain EFN production, leading to its loss in endemic lineages.
367 This may explain why EFN-bearing plants show a weak and negative relationship with island area.
368

369 Conclusion

370 Our study demonstrates that both biotic interactions (ant diversity) and abiotic factors (island area and
371 isolation) are critical in shaping the diversity and representation of ant-associated plants on islands. We
372 found that the effect of biotic interaction filter is more pronounced on larger and less-isolated islands. This
373 key finding reveals that the relative importance of ecological processes can shift across environmental
374 gradients. Moreover, our results show that the representation patterns and underlying drivers for domatium-
375 and EFN-bearing plants are distinct, likely due to the differing characteristics of these two interaction types:
376 domatia are costly and obligate, while EFNs are relatively inexpensive and facultative with ants. By
377 providing a novel empirical example of how ant diversity shapes island plant assemblages, our study
378 highlights the necessity of integrating biotic interactions into island plant biogeography. Furthermore, we
379 suggest that the influence of biotic interactions is not uniform across mutualistic interactions, and we

380 therefore advocate for explicitly incorporating interaction context (e.g. specific interaction characteristics,
381 effect of enemy release, and resource availability) to advance our understanding of island assembly.

382
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388 assistance from AT, PW, BG, and HK. YL wrote the paper with substantial assistance from all authors.

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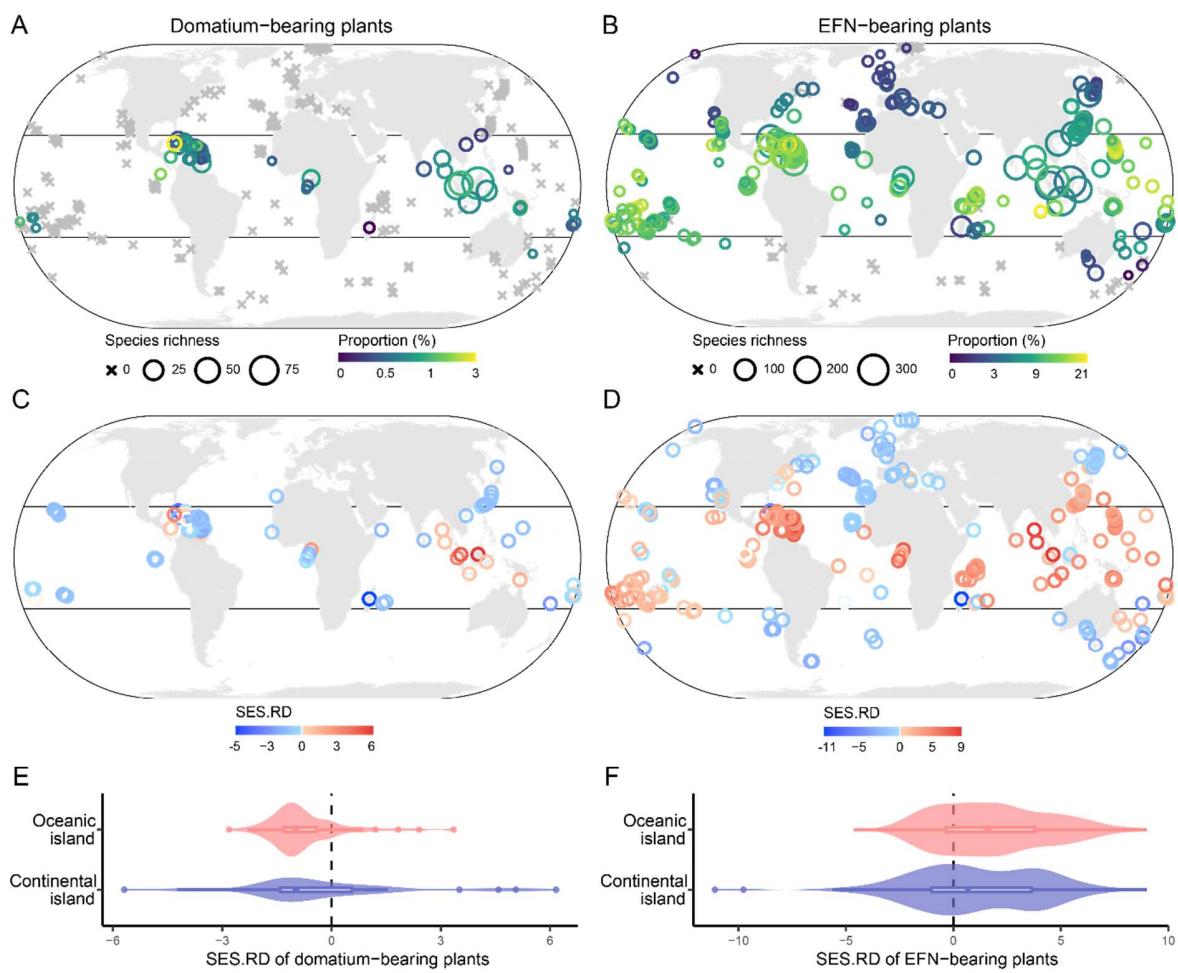
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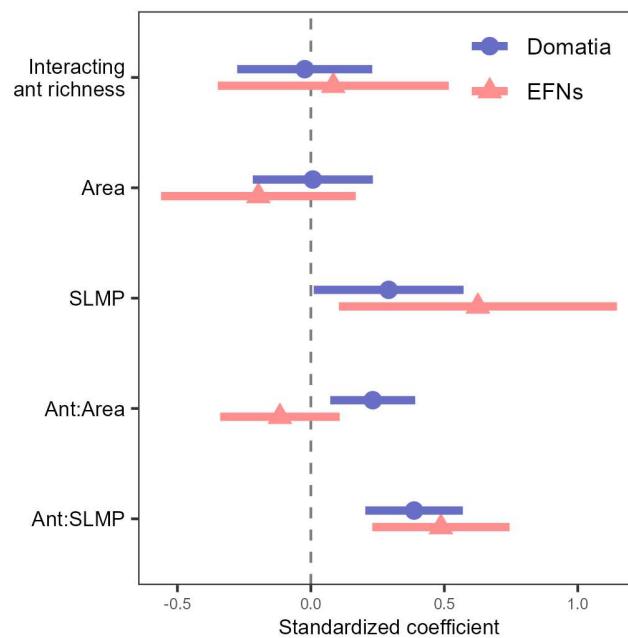
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540 **FIGURE 1 |** (A, B) Species richness and proportion to all angiosperms for plants bearing domatia and
 541 extrafloral nectaries (EFNs) on 446 islands. (C, D) Disharmonic patterns of ant-associated plants on
 542 islands indicated by standardized effect size of representational disharmony (SES.RD; n = 99 for
 543 domatium-bearing plants, n = 390 for EFN-bearing plants). (E, F) Representational disharmony of ant-
 544 associated plants on continental islands and oceanic islands. Differences between continental and
 545 oceanic islands are examined by Wilcoxon rank test (p = 0.46 for domatium-bearing plants, p < 0.001 for
 546 EFN-bearing plants)

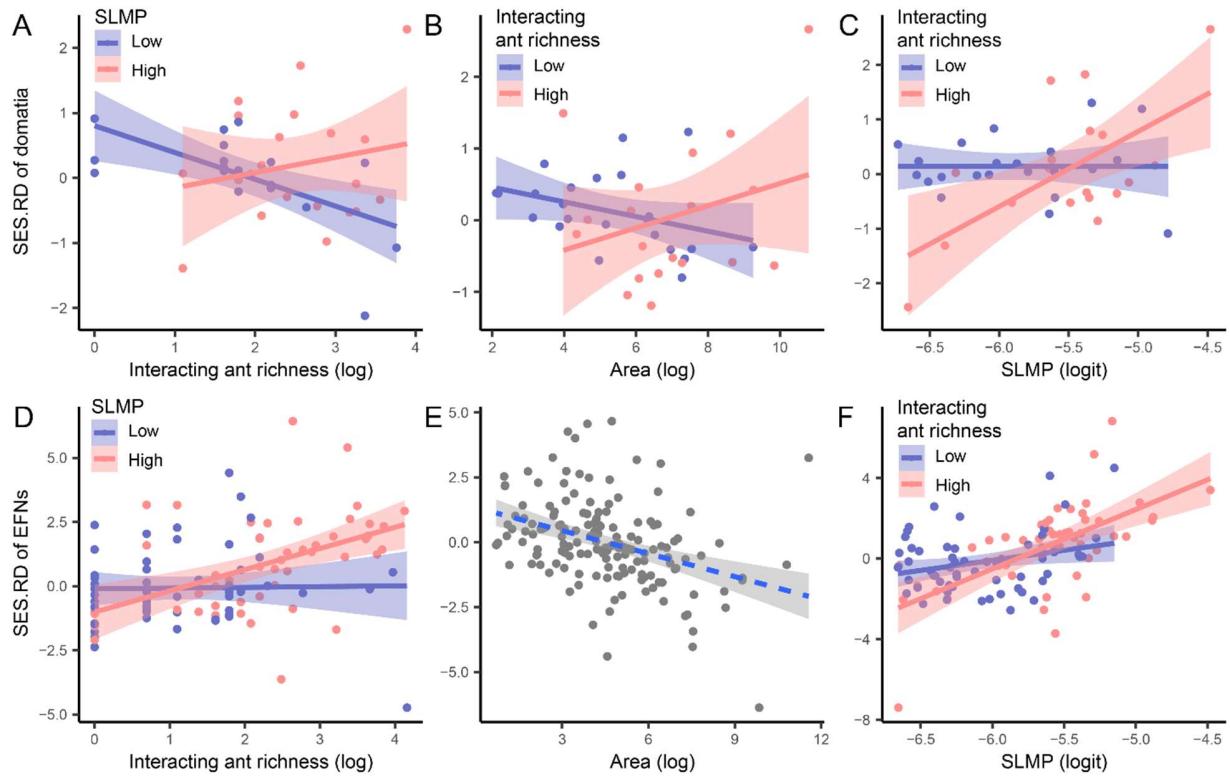
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549 **FIGURE 2** | Standardized coefficients of simultaneous autoregressive models for representational
 550 disharmony of domatium- ($n = 60$) and extrafloral nectary (EFN)-bearing plants ($n = 152$) on tropical oceanic
 551 islands. The figure shows the relative importance of species richness of interacting ants (Ant), island area,
 552 the proportion of surrounding landmass (SLMP), and the interaction between ants and both island area and
 553 SLMP. Horizontal bars indicate 95% confidence intervals.

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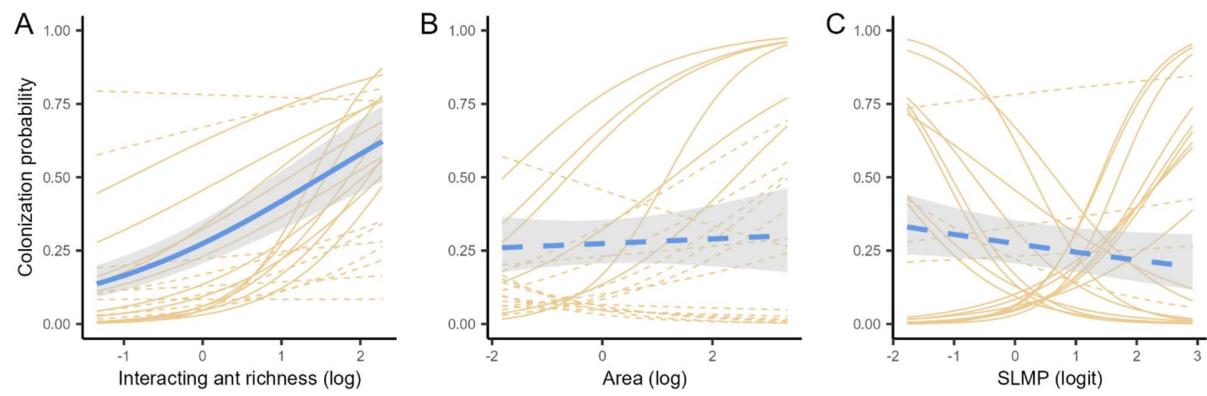


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556 **FIGURE 3** | Partial residuals of interacting ant richness (A, D), island area (B, E), the proportion of
 557 surrounding landmass (SLMP) (C, F), and their interactions on the taxonomic representation of domatium-
 558 (upper panels) and extrafloral nectary (EFN)-bearing plants (bottom panels) on oceanic islands.
 559 Interactions are showed for conditions of low (bottom 33%) and high (top 33%) interacting ant richness,
 560 small (bottom 33%) and large (top 33%) island area, and low (bottom 33%; high isolation) and high (top
 561 33%; low isolation) SLMP, for domatium- and EFN-bearing plants, respectively. Coloured bands represent
 562 95% confidence intervals. Results are illustrated by generalized linear models with spatial autocovariates.
 563 The interaction between interacting ant richness and island area for domatium-bearing plants is showed in
 564 Fig. S2 for simplification.

565

566



567 **FIGURE 4** | Effects of interacting ant richness, island area and the proportion of surrounding landmass
568 (SLMP) on the colonization probability of EFN-bearing plants. Separate generalized linear models of
569 binomial family were applied to 21 EFN-bearing plants across 152 tropical oceanic islands. Yellow lines
570 show the predicted colonization probability for each individual species, while blue lines show the overall
571 predicted colonization probability across all 21 species. Solid lines indicate a statistically significant
572 relationship ($p < 0.05$), and dashed lines indicate a non-significant relationship.