

1 **Acorn availability reduces agricultural**  
2 **damage by ungulates**

3 Maria Bogdańska\*<sup>1</sup>, Valentin Journé<sup>2</sup>, Michał Bogdziewicz<sup>1</sup>

4 <sup>1</sup>Forest Biology Center, Institute of Environmental Biology, Faculty of Biology, Adam Mickiewicz  
5 University, Poznań, Poland

6 <sup>2</sup>Department of Biology, Faculty of Science, Kyushu University, Fukuoka, Japan

7 \*Corresponding author: maria.a.bogdanska@gmail.com

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9 human-wildlife conflict

## Abstract

11 Human-wildlife conflicts, particularly the damage to agricultural crops caused by  
12 ungulates, pose significant ecological and economic challenges. Understanding the  
13 role of natural food availability in driving these conflicts is important for developing  
14 effective management strategies. We investigated how the pulsed availability of for-  
15 est tree seeds, i.e., mast seeding, influences the extent of agricultural crop damage in  
16 Poland. Using a 19-year national dataset (2001–2020), we analyzed the relationship  
17 between oak (*Quercus* spp.) and European beech (*Fagus sylvatica*) seed production,  
18 the abundance of wild boar (*Sus scrofa*) and red deer (*Cervus elaphus*), and the  
19 area of damaged agricultural crops. We found a negative relationship between oak  
20 seed production and the level of crop damage, with estimated damage decreasing  
21 by 30% from years of seed failure to years of abundant seed production, supporting  
22 the hypothesis that a diet shift occurs in ungulates during years of seed abundance  
23 that averts ungulates from damaging the crop. In contrast, beech seed production  
24 showed no significant effect on crop damage. Our findings demonstrate that pulsed  
25 resource dynamics in forests are an important driver of human-wildlife conflict in  
26 adjacent agricultural landscapes.

# Introduction

28 Anthropogenic change has led many taxa to supplement natural foods with agricul-  
29 tural crops, including insects, birds, and especially ungulates (Bereś 2015; Vergin  
30 et al. 2025; Jacob et al. 2014; Blount et al. 2021; Montràs-Janer et al. 2019; Brown  
31 et al. 2007). In wild boar, a synthesis of 11 studies across eight European countries  
32 showed frequent and often large consumption of crops with strong seasonal and ge-  
33 ographic variation (Schley and Roper 2003). Crop damage is now a major form of  
34 human–wildlife conflict with ecological and socioeconomic costs. Habitat conver-  
35 sion increases animal–human contact, and large-bodied species commonly damage  
36 fields, livestock, orchards, and other property, fueling financial losses and negative  
37 attitudes (McKee et al. 2021; Hill 2018; Hulme et al. 2020; Distefano 2005; Basak  
38 et al. 2023). Because farmland provides accessible, energy-rich food, wildlife turns  
39 to anthropogenic sources when natural resources are limited. For example, a 15-year  
40 study in Italy reported fewer bear attacks on livestock in years of abundant beech  
41 masting but 67% higher attacks in poor mast years (Tattoni et al. 2025). These  
42 patterns show how fluctuations in natural food modulate the severity of conflict,  
43 highlighting the importance of identifying the ecological drivers of crop damage for  
44 better prediction and management. Financial compensation schemes have been in-  
45 troduced in many countries to support those affected (Ravenelle and Nyhus 2017),  
46 reaching tens of millions of euros annually (Coordination Rurale 2024; Bleier et al.  
47 2012; Schley, Dufrêne, et al. 2008). Such losses burden farmers, institutions, and  
48 taxpayers, and mitigation strategies such as fencing, repellents, regulated hunting  
49 (Geisser and Reyer 2004), and supplementary feeding (Calenge et al. 2004) require  
50 further investment. Compensation records illustrate the scale of the problem: in  
51 2000/2001 these amounted to nearly 26 million PLN in Poland ( $\sim 6$  mln EUR), ris-  
52 ing to 104 million PLN ( $\sim 24$  mln EUR) in 2019/2020 (Główny Urząd Statystyczny  
53 2006; Główny Urząd Statystyczny 2021). Analysis of compensation records from the  
54 2005/2006–2019/2020 hunting seasons showed that, on average, 1.72 thousand PLN  
55 was paid per hectare of damaged cropland nationwide (Figure S1 in Supplementary  
56 Materials). Pinpointing the environmental drivers of damage is thus important for  
57 both ecological understanding, forecasting, and cost-effective management.

58 Ungulates are major contributors to crop damage due to their large body size,  
59 high energy demands, social foraging, and mobility. Wild boar (*Sus scrofa*) and red

60 deer (*Cervus elaphus*) trample fields and consume large amounts of maize, cereals,  
61 potatoes, and rapeseed (Massei, Kindberg, et al. 2015; Dzieciółowski 1979; Schley  
62 and Roper 2003; Picard et al. 1991). Behaviors such as grubbing alter soil structure  
63 (Mohr et al. 2005; Risch et al. 2010), while group foraging increases the spatial extent  
64 of damage (Maselli et al. 2014). Yet, agricultural plants supplement rather than  
65 replace forest foods. Red deer are herbivorous, shifting seasonally among grasses,  
66 shrubs, and conifer needles, and in autumn consume acorns and beechnuts (Gebert  
67 and Verheyden 2008; Barrere et al. 2020). Wild boar are omnivorous generalists but  
68 depend on energy-dense resources such as acorns, turning to crops or alternative  
69 foods when mast is scarce (Massei, Kindberg, et al. 2015). Thus, both species  
70 combine forest and agricultural foods, with mast availability shaping their autumn  
71 diets (Picard et al. 1991; Schley and Roper 2003).

72 Mast, the highly interannually variable and synchronized production of large  
73 seed crops, is a widespread strategy among European trees, including large-seeded  
74 oaks (*Quercus* spp.) and European beech (*Fagus sylvatica*) (Kelly 1994; Bogdziewicz,  
75 Kelly, et al. 2024; Szymkowiak et al. 2024). By concentrating reproduction in high-  
76 seeding years, trees enhance pollination and reduce seed predation (Zwolak et al.  
77 2022). The resulting pulses of acorns and beechnuts provide critical resources for  
78 insects, birds, rodents, and ungulates (Thomas 2008; Myczko et al. 2014; Ruscoe et  
79 al. 2005; Schley and Roper 2003). While abundance of invertebrate seed predators  
80 is strongly regulated by mast cycles, large-bodied vertebrates such as ungulates  
81 can compensate for poor seed years by shifting to alternative foods (Bogdziewicz,  
82 Kuijper, et al. 2022; Curran and Leighton 2000). Nevertheless, fluctuations in seed  
83 abundance are expected to alter their foraging behavior, space use, and the degree to  
84 which they exploit agricultural crops (Zwolak et al. 2022; Pucek et al. 1993; Touzot  
85 et al. 2020; Gamelon, Focardi, et al. 2017).

86 Mast-generated resource pulses also affect ungulate population dynamics. Wild  
87 boar depend on at least one energy-rich food source throughout the year, whether  
88 mast or crops (Schley and Roper 2003). In high-seeding years, acorn and beechnut  
89 availability improves body condition, advances reproduction (Drimaj et al. 2019;  
90 Touzot et al. 2020; Cachelou et al. 2022), increases the proportion of breeding females  
91 (Gamelon, Touzot, et al. 2021), and leads to larger litters (Touzot et al. 2020),  
92 thereby promoting population growth (Massei, Genov, et al. 1996). Access to crops  
93 alongside mast can further enhance body size and condition (Merta et al. 2014). In  
94 contrast, mortality may increase when mast resources are unavailable, as observed in  
95 wild boar from Białowieża Forest, whereas red deer showed no such pattern (Okarma  
96 et al. 1995). Thus, while mast has particularly strong demographic effects on  
97 wild boar, it also shapes the feeding preferences of both species and modulates their  
98 reliance on agricultural foods.

99 Previous studies show that ungulates preferentially consume mast when available.

100 In France, Picard et al. (1991) found that during oak high-seeding year, acorns  
101 comprised 51% of red deer diet and occurred in 56% of rumen samples, while maize  
102 (12%), twigs (9%), and grasses (6%) were secondary. In the preceding poor seed  
103 production year, diets shifted toward grasses (20%), maize (12%), leaf stalks (13%),  
104 and beechnuts (9%). Long-term data from La Petite Pierre Reserve in northeastern  
105 France support these patterns: over 30 years, both red and roe deer increased acorn  
106 intake in mast years, with red deer consuming up to 52% of diet compared to 34%  
107 in roe deer; in poor years, acorns dropped to 4% and 1% respectively (Barrere et al.  
108 2020). Wild boar show similar responses. A review by Schley and Roper (2003)  
109 highlights mast (e.g., seeds of oaks and beech) as a preferred food, with other items  
110 taken opportunistically. In the Czech Republic, stomach contents of 182 wild boar  
111 collected during a strong oak mast year showed acorns and maize dominating diets,  
112 with acorns reaching up to 95% of intake (Mikulka et al. 2018). As acorns declined  
113 in winter, maize consumption rose, and acorn presence disappeared. Together, these  
114 studies demonstrate that mast abundance strongly shapes ungulate diets. Yet most  
115 work has focused on dietary shifts, while the direct link between seed production  
116 and the extent of crop damage remains poorly quantified, despite its relevance for  
117 anticipating and mitigating human–wildlife conflict.

118 In this study, we tested whether oak and beech masting influences agricultural  
119 damage in Poland by altering ungulate foraging after seedfall. We predicted reduced  
120 crop damage in high-seeding years and elevated damage when seed production failed.  
121 To examine this, we combined 19 years of national data (2001–2019) on oak and  
122 beech seed production, ungulate abundance, and reported crop damage. Our aim  
123 was to clarify whether mast seeding modulates conflict intensity and to assess its po-  
124 tential value as an early-warning indicator for targeted, cost-effective management.

# Material and Methods

## 126 Research area

127 Poland offers a suitable setting to study interactions between masting and crop  
128 damage due to its extensive coverage by both forests and farmland. Oaks (*Quercus*  
129 *robur*, *Q. petraea*) are widespread, while beech (*Fagus sylvatica*) reaches its north-  
130 eastern range limit and is locally scarce (Eaton et al. 2016; Durrant et al. 2016).  
131 Forest cover increased modestly from 28.4% in 2001 to 29.5% in 2020 (Główny Urząd  
132 Statystyczny 2002; Główny Urząd Statystyczny 2020), but connectivity remains low  
133 compared to other European countries (Estreguil et al. 2013). Agricultural land  
134 dominates the landscape, exceeding 50% of the national area, with sown fields de-  
135 clining from 12.3 to 10.7 million ha over the study period.

136 Ungulates are widespread across Poland (Borowik et al. 2013), occupying forests,  
137 farmland, and even urban areas (Kowalewska 2019). The main ungulate species  
138 are wild boar (*Sus scrofa*), red deer (*Cervus elaphus*), roe deer, and fallow deer,  
139 with moose and bison present regionally (Michalska et al. 2023; Mysterud et al.  
140 2007). Wild boar populations increased for decades (Massei, Kindberg, et al. 2015)  
141 but declined recently due to African Swine Fever (Szymańska and Dziwulaki 2022),  
142 whereas deer populations continue to expand across Europe (Burbaité and Csányi  
143 2010). Poland is divided into hunting districts managed by local associations, which  
144 also administer compensation for crop damage. For this study, we used voivodeship-  
145 level (N = 16) population estimates reported by the Polish Hunting Association  
146 (PZŁ).

## 147 Seed production data

148 We extracted seed production data from MASTREE+, a data set that collects an-  
149 nual reproductive time series data from all over the world and makes these data  
150 freely available (Hackett-Pain et al. 2022; Foest et al. 2024). We used data on Eu-  
151 ropean beech (*Fagus sylvatica*) and two types of oaks merged under one category  
152 (*Quercus robur* and *Quercus petraea*, grouped together as *Quercus* spp.) from 2001  
153 to 2019. This particular data set of seed production from Poland is a part of a long-  
154 term monitoring program that began in 1954, covering 17 areas of the Regional  
155 Division of State Forests (RDLP) and presenting seed production as the percentage  
156 of reproducing trees in a population (Pesendorfer et al. 2020).

157 To spatially match the seed production data (collected at the level of RDLP)  
158 to voivodeships (the level at which crop damage and animal abundance data are  
159 collected), we transformed and weighted the seed production data to align with  
160 crop damage and animal abundance data, applying the following equation:

$$SP_{v,t} = \sum(SP_{f,t} * wt_f)$$

161 Seed production in a given voivodship in a given year ( $SP_{v,t}$ ) is calculated as a sum  
162 of seed production in each forest division within that year ( $SP_{f,t}$ ), multiplied by the  
163 proportion of the forest division overlapping with the voivodship ( $wt_f$ ).

164 Seed production exhibited considerable interannual variation across years for both  
165 beech (mean coefficient of variation, CV = 1.02, and standard deviation, SD = 23)  
166 and oak (CV = 0.66, SD = 17) (Figure 1).

## 167 Damaged crops and ungulate count data

168 We obtained data on reduced crop damage area (hectares) and ungulates populations  
169 (number of individuals) spanning 19 hunting seasons from 2000/2001 to 2019/2020  
170 at the voivodeship level, from the Research Station of the Polish Hunting Associa-  
171 tion (PZŁ) in Czempin. Their main objective since 1990 has been to monitor game  
172 species in Poland. Damaged crop data is a set of annual records of reduced area  
173 of damaged agricultural crops for each voivodeship, calculated as the damaged crop  
174 area multiplied by the percentage of its destruction. Damaged crop areas and com-  
175 pensation withdrawals for these damages are documented by the divisions of PZŁ  
176 and aggregated in annual reports. The number of individuals for each ungulate  
177 species is also recorded by hunters across Poland based on year-round observations,  
178 and this data is used to create annual hunting plans for the season ahead. It is there-  
179 fore the estimated number of individuals present in a given area (for our purpose,  
180 grouped by voivodeships). We focused on two ungulate species that are widespread  
181 and known for damaging crops and consuming beech nuts and acorns: wild boars  
182 (*Sus scrofa*) and red deer (*Cervus elaphus*).

183 Crop damage area showed an increasing trend, rising from under 1000 ha in  
184 2001 to approximately 2500 ha in 2020 and was also characterised by relatively  
185 large interannual variation (CV=0.77, SD = 1248, [Figure 1](#)). Regarding animal  
186 populations, ungulate numbers generally increased over the studied period; however,  
187 after 2014, the wild boar population declined sharply due to African Swine Fever  
188 (ASF) ([Figure 1](#)).



## 189 **Statistical analysis**

190 We modeled the relationship between crop damage and seed production, taking  
191 into account ungulate abundance. We used a generalized linear mixed model with  
192 reduced damaged crop area as a response variable, and predictors including two  
193 ungulate species abundance (wild boar and red deer, separately) and seed produc-  
194 tion from two species (beech and oak, separately). All predictors were scaled to  
195 enable direct comparisons between them. After visual inspection of the relation-  
196 ships, animal abundance variables were fitted as cubic splines to allow for non-linear  
197 relationships. We log-transformed the crop damage records to address skewness in  
198 the data (see [Figure S2](#) in Supplementary Materials). The region (voivodeship) was  
199 included as a random intercept to account for spatial variations in crop damage and  
200 repeated sampling.

201 We lagged both the ungulate counts and seed production by one year to match the  
202 timing of the hunting season, which runs from April 1st to March 31st. The reduced  
203 damage area data is referenced to the hunting season (e.g. the 2000/2001 season  
204 is recorded as 2001), while the ungulate counts represent the estimated population  
205 as of the end of March, which is then used to create the annual hunting plans for  
206 the upcoming season (beginning of hunting season e.g. 2000 record is a prediction  
207 for 2000/2001 hunting season). Seed production is estimated in a given year, and  
208 reflects the food availability for ungulates after the seedfall (within the hunting  
209 season, e.g., seeds produced in 2000 are a food source in 2000/2001 hunting season).

210 To assess the potential impact of African Swine Fever (ASF) on crop damage,  
211 we additionally ran models on a shortened time span, using 2015 as a cutoff year  
212 (responding to the 2014/2015 hunting season, within which the first known ASF  
213 outbreak that happened in Poland). Due to insufficient post-2015 data and the  
214 progressive nature of the disease, a classical before–after analysis was not feasible.  
215 Therefore, we limited the analysis to the pre-outbreak period and compared it with  
216 results from the full dataset. To avoid issues of multicollinearity between the two  
217 ungulate species, which were highly correlated ( $r = 0.68$  for the full period;  $r = 0.91$   
218 before ASF), we fitted separate models for wild boar and red deer.

219 All analyses were conducted in R (version 4.4.2). We fitted the models via the  
220 lme4 package (version 1.1.35.5) (Bates et al. 2015). To generate model-based pre-  
221 dictions, we used the ggeffects package (Lüdtke 2018).

222

# Results

224 In both models, oak seed production was negatively related to agricultural crop  
225 damage, while beech seed production showed no significant effect ( $p > 0.1$ ).

226 In the model including wild boar abundance, oak masting significantly reduced  
227 crop damage ( $\beta = -0.07$ ,  $SE = 0.02$ ,  $p < 0.01$ ), while wild boar abundance had a  
228 strong positive effect ( $\beta_1 = 0.647$ ,  $SE = 0.153$ ,  $p < 0.001$ ,  $\beta_2 = 1.820$ ,  $SE = 0.234$ ,  
229  $p < 0.001$ ,  $\beta_3 = 1.226$ ,  $SE = 0.226$ ,  $p < 0.001$ ). Assuming average conditions,  
230 predicted damage decreased from  $\sim 1828$  ha at low oak seed production to  $\sim 1295$   
231 ha at maximum acorn production (Figure 2 A.), representing a  $\sim 29\%$  reduction  
232 in actual area damaged. In contrast, wild boar abundance had a strong positive  
233 effect, where predicted crop damage increased from  $\sim 452$  ha at low abundance to  
234  $\sim 1925$  ha (Figure 2 B.) at high abundance, corresponding to more than a threefold  
235 increase.

236 In the red deer model, oak seed production again had a negative effect ( $\beta = -0.10$ ,  
237  $SE = 0.02$ ,  $p < 0.001$ ), where assuming average conditions predicted crop damage  
238 decreased from  $\sim 2977$  ha at low oak seed production to  $\sim 1828$  ha, which accounted  
239 for a  $\sim 38\%$  reduction. Red deer abundance was a strong positive predictor of crop  
240 damage ( $\beta_1 = 2.133$ ,  $SE = 0.133$ ,  $p < 0.001$ ,  $\beta_2 = 4.483$ ,  $SE = 0.206$ ,  $p < 0.001$ ,  
241  $\beta_3 = 2.553$ ,  $SE = 0.168$ ,  $p < 0.001$ ). Predicted damage increased from  $\sim 224$  ha at  
242 low abundance to  $\sim 6135$  ha at highest values - a more than twentyfold rise (Figure 2  
243 C.).

244 Across both model configurations, oak masting reduced crop damage, while un-  
245 gulate abundance remained the dominant positive driver. The negative effects of  
246 oak masting on crop damage persisted when analyzing only the period before the  
247 ASF outbreak (Table S1 in Supplementary Materials).

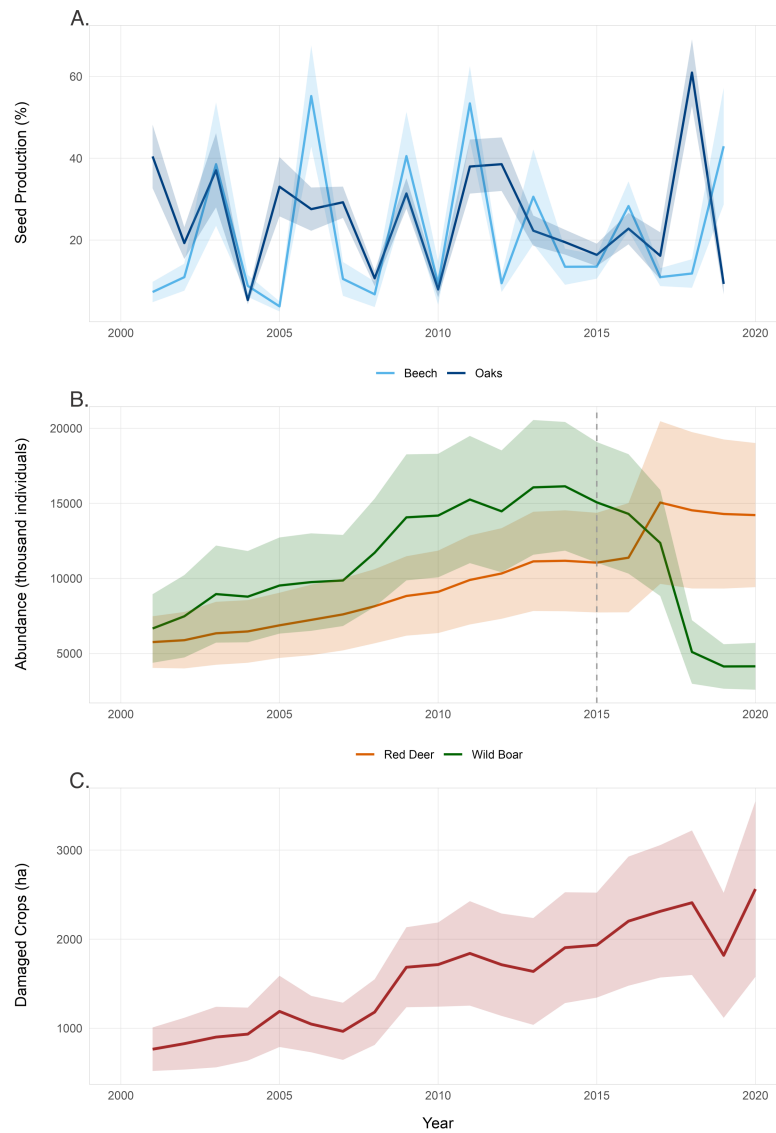


Figure 1: Temporal trends in seed production, ungulate populations and crop damage across all study regions. A) Annual seed production of beech and oaks over time. B) Annual crop damage area across all regions. C) Population trends of red deer and wild boar across all districts. Solid lines represent means and shaded areas indicate one standard deviation around the mean. The vertical dashed line indicates the African Swine Fever (ASF) outbreak (2015).

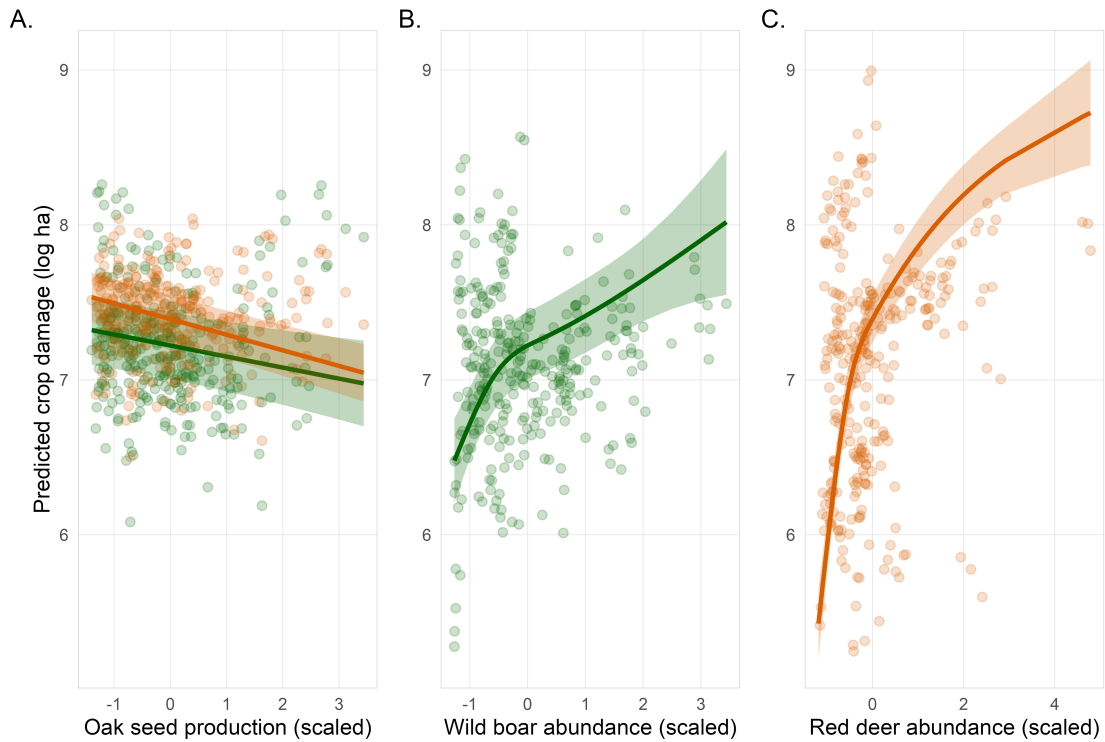


Figure 2: Estimated relationships between crop damage and ecological predictors across the full study period. Panels show estimated crop damage (ha) in relation to (A) oak seed production, (B) wild boar abundance, and (C) red deer abundance. Beech seed production was not a significant predictor of crop damage (Table S1). Lines represent estimated mean values with 95% confidence intervals (shaded areas), and points show partial residuals. Colors show predictions derived from separate models (green - that included wild boar abundance as a predictor; orange - red deer abundance). Predictors were scaled before entering the model. Estimates were derived from linear mixed models with voivodeship included as a random intercept. The conditional  $R^2$  of the model that included wild boar abundance and seed production of both species as predictors was 0.65, while the marginal  $R^2$  was 0.25. The conditional  $R^2$  of the model that included red deer abundance and seed production of both species as predictors was 0.90, while the marginal  $R^2$  was 0.76.

# Discussion

249 Our results show that interannual variation in natural food, specifically oak mastings,  
250 modulates the intensity of human–wildlife conflict. We found a negative relationship  
251 between acorn availability and the area of damaged crops, with predicted damage  
252 decreasing by over 30% between years of lowest and highest oak seed production.  
253 This indicates that high-seeding years buffer conflict by reducing the need for un-  
254 gulates to forage in fields. The pattern suggests that energy-rich acorns act as a  
255 substitute resource, lowering reliance on agricultural foods when abundant, while  
256 poor mast years intensify pressure on farmland. By linking dietary flexibility of un-  
257 gulates to a direct, management-relevant outcome, these results extend earlier work  
258 that focused mainly on diet composition. Practically, acorn availability emerges as a  
259 useful indicator of conflict risk: integrating mast seeding with regional factors such  
260 as forest composition, proximity of farmland to oak stands, and animal movements  
261 across administrative borders could support more targeted deployment of mitigation  
262 strategies. The fact that we detected this effect using broad-scale, non-targeted data  
263 suggests that in oak-rich landscapes with strong fruiting, the local strength of this  
264 relationship may be even greater.

265 In contrast to oak, beech seed production did not show a significant influence  
266 on crop damage in our analyses. This could be due to the lower nutritional value and  
267 smaller size of beechnuts compared to acorns (Rivero et al. 2019). Even during mast  
268 years, beechnuts may not constitute a sufficient resource to become a primary food  
269 source for large ungulates. From a management perspective, our results imply that  
270 monitoring oak flowering intensity can be a valuable forecasting tool, whereas track-  
271 ing beech flowering may be less critical given its negligible impact. Indeed, reports  
272 from other countries indicate that beech masting impacted ungulate population dy-  
273 namics to a lesser extent than oak masting (Gamelon, Focardi, et al. 2017). It is  
274 worth noting, though, that the positive correlation ( $r=0.28$ ) we observed between  
275 oak and beech seed production during our study might have partially masked any  
276 potential, separate effect of the latter species.

277 Our analysis indicates that the abundance of key game species had a substantial  
278 impact on the extent of crop damage, and it has been shown in previous studies  
279 in Poland that agricultural crops are a significant component of their diet (Gebert  
280 and Verheyden 2008; Knizewska and Rekiel 2015). It is important to acknowledge,

281 however, that our analysis of animal abundance, based on hunter observations, is  
282 subject to methodological limitations, as observer expertise and financial incentives  
283 can introduce biases into population estimates (Kamieniarz et al. 2023). These  
284 limitations prevented us from assessing a direct numerical response of ungulates to  
285 seed production, but it remains evident that animal abundance is a critical factor  
286 intensifying human-wildlife conflicts.

287 We observed that recent (since 2015), drastic shifts in the ungulate community  
288 structure, i.e., rising deer numbers and the decline of wild boar due to African Swine  
289 Fever (ASF) did not erase the overall significant influence of oak seed production.  
290 It is, however, important to note that the wild boar population decline was driven  
291 by several interconnected factors: direct mortality from the disease, extensive sani-  
292 tary culling, and intensified hunting pressure related to management policies, with  
293 spatially variable intensity across regions. The subsequent spatiotemporal spread of  
294 the epidemic (Wojewódzki Inspektorat Weterynarii 2019) likely had profound and  
295 regionally varied effects. However, the limited number of observations in the post-  
296 ASF period prevented us from conducting a formal statistical comparison between  
297 the two periods.

298 The species analyzed are game animals, and hunting remains the primary pop-  
299 ulation management tool in Europe. However, high hunting pressure can induce  
300 fear, altering animal behavior—potentially creating "refugee effects" (Amici et al.  
301 2012) where animals avoid certain areas—and influencing their use of agricultural  
302 landscapes (Sütő et al. 2023). Hunting pressure can also drive evolutionary conse-  
303 quences, such as earlier reproduction in wild boars (Gamelon, Besnard, et al. 2011).  
304 Furthermore, hunting pressure might obscure the effects of seed production on pop-  
305 ulation dynamics at local scales, for example, by maintaining animal populations at  
306 low levels. Management recommendations should therefore integrate these environ-  
307 mental factors to better forecast potential crop losses. While the ongoing recovery  
308 of large predators like wolves in Poland (Chapron et al. 2014; Di Bernardi et al.  
309 2025) could theoretically help regulate ungulate populations, habitat fragmentation  
310 and other anthropogenic influences may limit their effectiveness in keeping ungu-  
311 late numbers under control (van Beeck Calkoen et al. 2023). Therefore, it could  
312 be beneficial to align hunting strategies with oak seed production, as masting can  
313 influence not only the extent of wildlife impact on agriculture, but also population  
314 dynamics in the subsequent months. In conclusion, our findings demonstrate that  
315 oak seed production plays a significant role in shaping crop damage by altering un-  
316 gulate foraging behavior. This interaction occurs within a complex socio-ecological  
317 context influenced by climate, habitat structure, and human management, suggest-  
318 ing that integrating regional seed production forecasts with landscape composition  
319 data and adaptive wildlife management practices may provide substantial benefits  
320 for mitigating human-wildlife conflicts.

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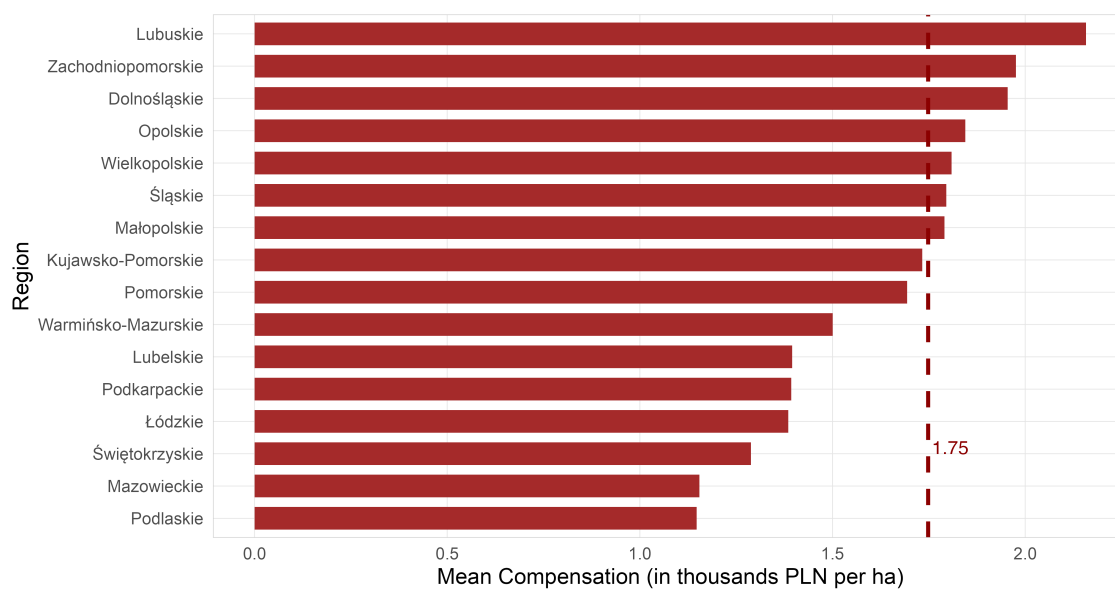


Figure S1: Mean compensations per 1 hectare of damaged crop area for all regions in hunting seasons 2005/2006–2019/2020. The red line represents the calculated mean for the whole of Poland.



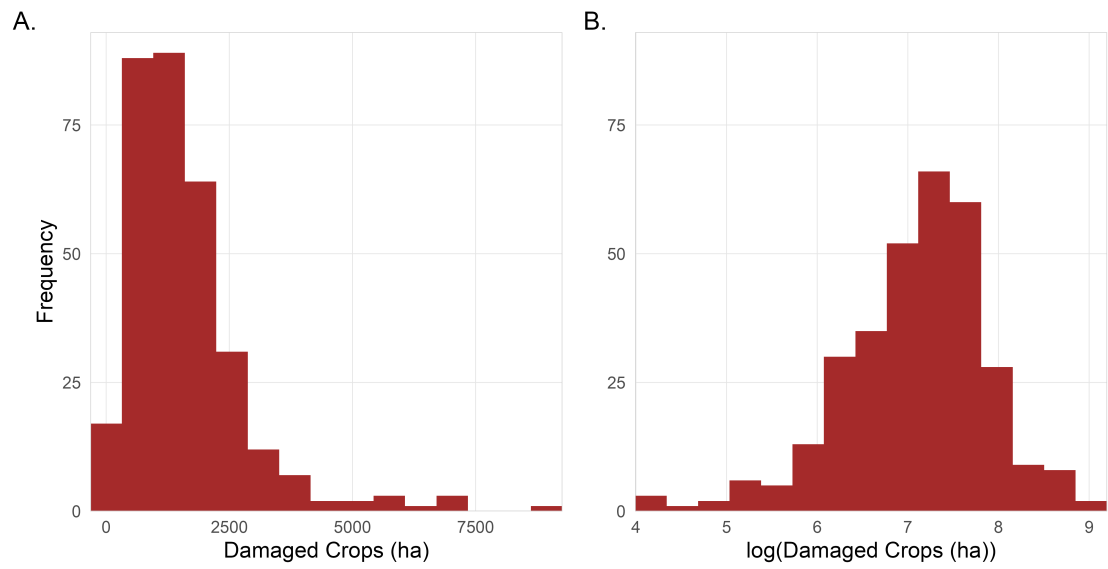


Figure S2: Distribution of damaged crop area (ha) before (A.) and after (B.) log-transformation.

Model	Predictors	Estimates	Std. E.	df	p-value
Wild boar full period	Beech seed production	0.03	0.02	281.73	0.24
	Oak seed production	-0.07	0.02	283.66	0.003
	Wild boar abundance <sub>1</sub>	0.65	0.15	297.81	>0.001
	Wild boar abundance <sub>2</sub>	1.82	0.23	297.87	>0.001
	Wild boar abundance <sub>3</sub>	1.23	0.23	296.27	>0.001
Red deer full period	Beech seed production	0.01	0.02	283.76	0.64
	Oak seed production	-0.10	0.02	284.32	>0.001
	Red deer abundance <sub>1</sub>	2.13	0.13	282.94	>0.001
	Red deer abundance <sub>2</sub>	4.48	0.21	290.99	>0.001
	Red deer abundance <sub>3</sub>	2.55	0.17	297.77	>0.001
Wild boar before ASF	Beech seed production	-0.03	0.01	205.95	0.02
	Oak seed production	-0.05	0.01	205.29	>0.001
	Wild boar abundance <sub>1</sub>	1.60	0.09	217.15	>0.001
	Wild boar abundance <sub>2</sub>	4.33	0.15	216.79	>0.001
	Wild boar abundance <sub>3</sub>	2.48	0.12	217.04	>0.001
Red deer before ASF	Beech seed production	0.01	0.02	204.54	0.72
	Oak seed production	-0.09	0.02	204.23	>0.001
	Red deer abundance <sub>1</sub>	1.95	0.15	207.39	>0.001
	Red deer abundance <sub>2</sub>	4.56	0.27	204.79	>0.001
	Red deer abundance <sub>3</sub>	2.71	0.20	213.41	>0.001

Table S1: Linear Mixed-Effects Regression fixed effects for seed production and ungulate abundance for full period and before the ASF outbreak. Animal abundance variables were fitted as cubic splines to allow for non-linear relationships.

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