#### 1 The relative roles of in situ diversification and lineage dispersal underlying diversity 2 patterns at the assemblage level Arthur Vinicius Rodrigues<sup>1,2</sup>, Gabriel Nakamura<sup>2,3</sup>, and Leandro Duarte<sup>2</sup> 3 4 <sup>1</sup>Research Centre for Ecological Change, Faculty of Biological and Environmental Sciences, 5 University of Helsinki, Helsinki, Finland. <sup>2</sup> Programa de Pós-Graduação em Ecologia, Departamento de Ecologia, Universidade Federal do 6 7 Rio Grande do Sul, Brazil 8 <sup>3</sup> Universidade de São Paulo, Instituto de Biociencias, Brazil 9 Corresponding author: Arthur Vinicius Rodrigues (rodrigues.arthur.v@gmail.com) 10 Abstract 11 Speciation, extinction, and dispersal are the historical processes influencing the spatial 12 distribution of lineages and strongly influence diversity patterns. Here, we apply a recently 13 developed methodological approach to quantify the relative legacies in situ diversification 14 history (i.e. diversification occurring in the biogeographical region) and historical dispersal 15 (inferred from ex-situ diversification) on current diversity patterns of the plant genus Myrcia 16 (Myrtaceae) in assemblages distributed across the Brazilian Atlantic Forest. To distinguish 17 among these processes, we proposed a framework that characterized the assemblages based 18 on the mean colonization age of the assemblages, phylogenetic structure, species richness 19 and proportion of in situ diversification. Our results suggest that environmental dynamics 20 have played important role in shaping of diversity. In the Southern Atlantic Forest, 21 environmentally stable areas appear to have served as cradles for cold-adapted lineages. In 22 the central region, environmental instability in this montane area seems to be a unstable 23 area acting as cradle, with assemblages showing moderate to high in situ diversification and 24 intermediate colonization ages. In the Northern Atlantic Forest, older and richer 25 assemblages with high phylogenetic diversity suggest climatic stability and a refugium role. 26 In contrast, the Central-West region, with younger colonization ages and a higher 27 proportion of lineage dispersal, appears to act as an unstable sink for lineages. By 28 combining community phylogenetics and diversification metrics, we infer evolutionary 29 legacies at the assemblage level and disentangle the roles of in situ diversification and 30 lineage dispersal. In some regions, particularly outside Evoregion A, lineage dispersal 31 appears more relevant, possibly due to weaker selection pressures. In contrast, colder areas 32 may have favored lineages with broader climatic tolerance. Our findings reveal distinct 33 biogeographic dynamics across the Atlantic Forest, including areas acting as cradles, sinks, 34 and refugia, and demonstrate the value of assemblage-level diversification metrics in

understanding within-biome evolutionary processes and their influence on current diversity patterns.

#### Introduction

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The historical processes driving the diversity and distribution of biological diversity across the geographic space are, ultimately, speciation, extinction, and dispersal (Ricklefs 1987; Wiens and Donoghue 2004). Environmentally-driven ecological dynamics might also influence each of these processes by promoting ecological drift and selection-driven speciation along the evolutionary time, causing extinction (local or globally), allowing dispersal (colonization) and secondary sympatry (Mittelbach and Schemske 2015). Altogether, these processes contribute to determine the distribution of species and the spatial diversity patterns, placing the evolutionary history of a clade as an important piece to explain its current geographic distribution (Cavender-Bares et al. 2016; Gerhold et al. 2018).

In the context of historical processes shaping biodiversity, the stability (or instability) of a region can influence diversification and lineage dispersal, thereby affecting present-day diversity patterns. In environmentally unstable regions with heterogeneous landscapes, populations are more likely to have their distributions fragmented across space, which increases the chances of both speciation and extinction. Mountainous regions are good examples of such areas, and it can act as both cradles and museums of biodiversity (Rahbek et al. 2019; Vasconcelos et al. 2020; Vasconcelos, O'Meara, and Beaulieu 2022). Cradle areas can be harder to detect because environmental instability drives both high rates of speciation and extinction, leading to high species turnover leaving fragmentary evidence of past dynamics (Vasconcelos et al. 2022). Museums areas, in turn, are easier to recognize since they environmental stability allows the persistence of ancient lineages, though this same stability can also give the false impression that they were centers of origin (Vasconcelos et al. 2022). As stability allows species dispersing from to adjacent areas to prevent extinction, speciation rates are expected to be lower as well as extinction rates. Over time, this source-sink dynamics between unstable and stable areas could generate present-day assemblages with high phylogenetic diversity, moderate to low proportion of in situ diversification, consequently, higher lineage dispersal and old colonization age of assemblages on stable areas (hereafter age of assemblages, for simplicity), the so called museums. On the other hand, unstable areas should present more recent colonization events (low assemblage ages), due to high levels of both speciation and extinction rates,

moderate to high proportion of in situ diversification and high species richness where environmental heterogeneity is high. Stable areas can also act as sources of lineages. In these regions, we can expect high in situ diversification, with varying levels of phylogenetic relatedness and relatively recent colonization events (i.e., older assemblages). Therefore, a comprehensive investigation of historical processes shaping present-day diversity patterns should integrate community phylogenetics with diversification analyses and colonization/dispersal timing. These expectations are summarized in Table 1.

**Table 1**. Scenarios and the expected patterns that can be detected using assemblage-level metrics to assess historical biogeographical processes and current phylogenetic (expressed as MPD) and taxonomic alpha diversity (expressed as species richness).

Scenario	Biogeographical process		Current diversity	
	Proportion of in situ  Diversification	Colonization age of assemblage	MPD	Species richness
Stable area acting as refugium	Moderate to low	Old	High	High
Unstable area acting as a cradle	Moderate to high	Moderate to young	Low	High
Stable area acting as cradle	High	Old	Either low or high	Intermediate
Unstable area acting as sink of lineages	Low	Young	Either low or high	Low

Note: MPD is mean pairwise phylogenetic distance.

Species-level diversification metrics, species age estimates (or colonization time estimates), and ancestral area reconstruction have been widely used to uncover the historical biogeographic processes underlying current diversity patterns (Pinto-Ledezma et al. 2019; Velasco and Pinto-Ledezma 2022; Villalobos et al. 2020; Nakamura et al. 2024). Recently, a new method was proposed to integrate ancestral area reconstruction (Matzke 2013; Ree and Smith 2008), which estimates the biogeographical history of a clade across regions, with diversification and colonization age estimation, commonly used to describe evolutionary patterns at the assemblage level (McGill et al. 2019; Velasco and Pinto-Ledezma 2022). Combining these approaches allows quantifying the relative role of in situ diversification (diversification occurred within a biogeographical region) and lineage

dispersal (diversification occurred in another biogeographical region, or ex situ) for the species assemblage (Nakamura et al. 2024; see Box 1 and Fig. 1).

The original method relies on estimating the most likely ancestral range at the nodes of the phylogeny and does not account for uncertainty in ancestral estimates or for range shifts along branches (anagenetic range evolution) (Nakamura et al. 2024). In this study, we address these limitations by using Biogeographical Stochastic Mapping (BSM) to estimate ancestral area (Dupin et al. 2017). BSM allows propagation of uncertainty in ancestral range area estimates into the calculation of in situ diversification and colonization age of assemblages. Additionally, both anagenetic and cladogenetic range-shift processes are incorporated, unlike the original method, which considered only cladogenetic shifts, resulting in a more reliable assemblage-level metric estimates.

Our aim is to investigate the roles of in situ diversification and lineage dispersal in shaping the species assemblages in the Atlantic Forest in South America (Figure 2). Atlantic Forest harbors high species diversity and present turnover in taxonomic and phylogenetic composition that are consistent among different taxonomic groups (Carnaval and Moritz 2008; Costa 2003; Fiaschi and Pirani 2009). The major composition turnover distinction is between Northern and Southern portions of the Atlantic Forest (but see Brown et al., 2020), which has been attributed to current and historical differences in climatic conditions between regions (Carnaval et al. 2014; Saiter et al. 2016). Climatic stability since the last interglacial period (ca. 120 ky) seems to better explain the phylogenetic endemism in the northern portion, while the current climate seems to better explain the phylogenetic endemism of the southern portion (Carnaval et al. 2014).

The combination of high species diversity with the legacy of climatic influence makes Atlantic Forest a good system to investigate the interplay of historical processes acting on current macroecological and diversity patterns. As a model group we used *Myrcia* (Myrtaceae), which is a genus of trees and shrubs. *Myrcia* present several characteristics that make it suitable as a model group for the Atlantic Forest (Lucas and Bünger 2015; Murray-Smith et al. 2009). *Myrcia* has high diversity in Atlantic Forest (Lucas et al. 2018; Rodrigues and Duarte 2024) and is an important component of tree assemblages across the Atlantic Forest (Bergamin et al. 2021; Lucas et al. 2018; Oliveira-Filho and Fontes 2000; Rodrigues and Duarte 2024). Furthermore, the evolutionary history of *Myrcia* is linked to the Atlantic Forest history (Amorim et al. 2019; Rodrigues and Duarte 2024; Santos et al. 2017). Phylogenetic evidence suggests that *Myrcia* originated in the Atlantic Forest, probably in the

Serra do Mar mountains range (Amorim et al. 2019; Santos et al. 2017). However, a recent study (Rodrigues and Duarte 2024), presents a probable origin of the group in the northern Atlantic Forest, while the southern Atlantic Forest and adjacent highlands harbor the most recent diversification of several *Myrcia* lineages within the biome. Our study builds upon the work of Rodrigues and Duarte (2024) and Nakamura et al. (2024) by proposing a general framework that uses both methods from historical biogeography (Nakamura et al, 2024) with phylogenetic metrics from community ecology to shed light on historical processes acting on present-day assemblage diversity.

#### BOX 1

Species-level phylogenetic metrics, such as diversification and species colonization age, are calculated using only phylogenetic information (topology and divergence ages). However, when these metrics are summarized at assemblage-level, they fail to account for the biogeographical context of species origination and dispersal. One way to address this limitation is to integrate these metrics with ancestral area estimates. By incorporating information about the geography of speciation, we can refine phylogenetic metrics to reflect the biogeographical history of a lineage.

Using this approach, we can quantify the portion of tip's diversification rate that occurred within a region (in situ diversification). Since our focus is on historical process affecting local assemblages, we attribute the *ex situ* diversification (when the ancestors of a species have diversified in another region) to the lineage dispersal process because the lineage had to disperse at some point to the assemblage before starting to diversify in situ. In addition, we can quantify for how long in the evolutionary history the species is present in a region (colonization age). This framework allows us to disentangle the contributions of in situ diversification (when a lineage diversified within the assemblage's biogeographical region) and lineage dispersal (ex situ diversification). A key advantage of this approach is that tip-based metrics can be aggregated at the assemblage level, enabling the spatial quantification of in situ diversification and its impact on phylogenetic metrics, such as diversification rates or phylogenetic endemism, across large spatial scales.

Figure 1 illustrates our approach, showing how a phylogenetic tree can be decomposed into in situ diversification and lineage dispersal components using a hypothetical example. In this example, a clade with 7 species occurs exclusively across two regions, West and East regions (Figure 1A), where there are four sites with species composition. The evolutionary history of the clade was reconstructed using

biogeographical stochastic mapping and DEC model. For simplicity, we considered only one region occupied by each species currently, the region with more sites occupied and show one realization of all possible stochastic maps (Figure 1B). The phylogenetic tree in Figure 1C shows the evolutionary history only for the species occurring in the assemblage. To separate the importance of the evolutionary history for the local assembly into two components - in situ diversification and lineage dispersal - we trace the path from each tip to the root of the phylogeny. In situ diversification is defined as the continuous portion of the phylogenetic tree connecting a tip species to its most recent ancestor occurring in the same region as the site in which the species is found (occupied sites are represented by a color filled square in Figure 1B). In the site 1, all species had part of their history in situ. In the site 2, however, two species had no in situ component of its history in West region. These make recent historical dispersal more important for the assemblage of site 2 than for that of site 1. It is important to note that the interpretation of evolutionary process for a species is dependent on the site's biogeographic region. In other words, a species has higher diversification rate (or colonization age) in a site within the region where it diversified. Consider sites 2 and 3, as an example. Site 2 is in the West region, where species b and c had part of their evolutionary history in the region, while species a and gdiversified entirely in the East region. Site 3 contains the same species as site 2 but it is located in a different region. As a result, the partitioning of species history changes: species a and g are now classified as having in situ diversification, whereas species b and c are considered as recent dispersers, since their lineages evolved in the West region. Thus, in situ diversification plays a greater role in the assemblage of site 3 than in that of 2. Finally, in site 4 (East region) all species shared the evolutionary history within the region, making in situ diversification the sole evolutionary process determining assemblage.

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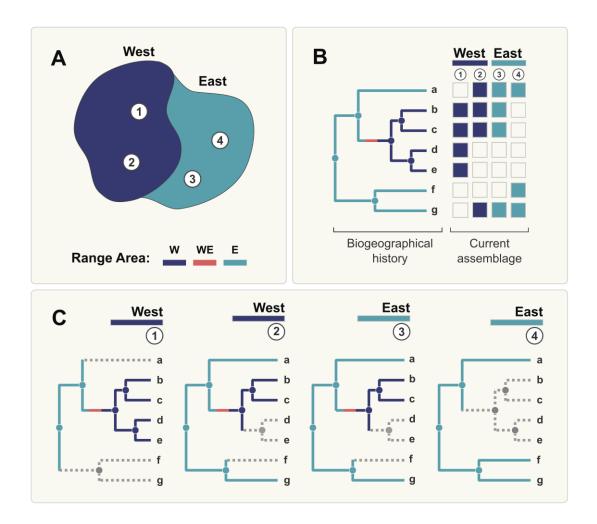


Figure 1. Schematic representation of the framework to quantify the relative role of in situ diversification processes and lineage dispersal in current assemblages. Panel A shows the assemblages at the West and East regions (dark blue and light blue, respectively). Panel B shows the biogeographical history of the clade, represented by the colors (regions) along the branches and nodes of the phylogenetic tree; and the current occurrence of species in assemblages is represented in the matrix of squares in which filled squares represent the presence of a species in the assemblage. Finally, in panel C, we highlight the biogeographical history of the species present in the assemblage; dotted branches are species not occurring in the assemblage. The delimitation of in situ processes is done following the path of the lineage from the tip to the root of the phylogeny, while the area estimated for an ancestral node is required to be within the same region of the current site of the assemblage. Note that the species could have different timing of in situ diversification depending on the site's region.

#### Material and Methods

We used a methodological framework that integrates biogeographical history of the lineages estimated by ancestral area reconstruction and assemblage-level phylogenetic metrics (in situ diversification, colonization age of assemblages and mean phylogenetic

196	pairwise distance) to provide assemblage-level descriptors that captures the biogeographic
197	history of current assemblages (Nakamura et al. 2024). Here we extend the original
198	framework described by Nakamura et al. (2024) by using biogeographic stochastic mapping
199	(Dupin et al. 2017). Biogeographic Stochastic Mapping (BSM) is a method that samples
200	from the distribution of possible biogeographical histories (100 samples in this study),
201	estimated by a model and conditioned on model parameters, the phylogenetic tree, and the
202	data. This approach allows quantification of the uncertainty associated with reconstructing
203	the biogeographical history of species. Importantly, it enables us to propagate uncertainty
204	in ancestral area estimates and the assemblage-level phylogenetic metrics we use.
205	We applied BSM to reconstruct the biogeographical history of the genus Myrcia using a
206	phylogenetic tree. We also use the information on current species distribution and the
207	biogeographical region in which the assemblages occurs. These data were combined to
208	calculate the proportional in situ diversification rate (DR <sub>prop</sub> ), colonization age of
209	assemblage, mean phylogenetic distance and species richness. Using these metrics, we
210	investigated the relative roles of in situ diversification and lineage dispersal in shaping
211	Myrcia assemblages in the Atlantic Forest, following the framework illustrated in Table 1.
212	All analyses were conducted in R version 4.2.2 (R Core Team 2022) and the main packages
213	used are mentioned in italic. To define the boundary of the Atlantic Forest (our study area),
214	we used the integrative limit proposed by Muylaert et al. (2018).
215	Genus Myrcia and phylogenetic tree
216	The genus <i>Myrcia</i> comprise 794 species distributed across the Neotropics, subdivided in 10
217	taxonomic clades (hereafter sections) (Lucas et al. 2018; POWO 2025). A total of 264
218	Myrcia species are found in the Atlantic Forest, the region with the highest Myrcia richness
219	(Amorim et al. 2019; Lucas et al. 2018). The phylogeny was produced by Amorim et al.
220	(2019) using internal and external transcribed spacer (ITS and ETS) of the ribosomal
221	nuclear region and seven plastid markers. It presents strong statistical support for the
222	topology of the 10 sections (Amorim et al. 2019). Time-calibration used two calibration
223	points one to constrain the maximum age (the age of Neotropical tribe Myrteae) and
224	another to the Myrcia crown node (Vasconcelos et al. 2020). Here, we used the consensus
225	time-calibrated phylogenetic tree available in Vasconcelos et al. (2020). The available
226	phylogenetic tree has 195 taxons (24,5% of Myrcia's species) (Amorim et al. 2019;
227	Vasconcelos et al. 2020).

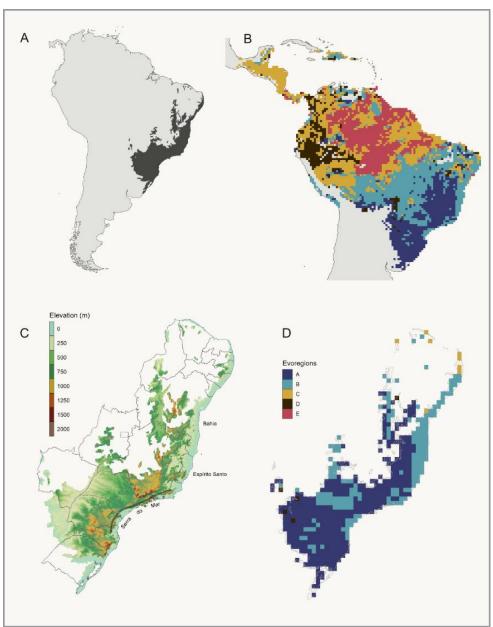
228 Species assemblages

The species assemblages used here are based on binary predictions obtained from models of stacked species distribution across the Neotropics produced by Rodrigues and Duarte (2024). We used raster for 307 Myrcia species available (38.6% of Myrcia's species) with resolution of 0.5 decimal degrees. Species distributions were constructed using buffers (50 km radius for n <4 records, convex-hull with 50 km buffer for 4-5 records) and species distribution models using Maxent algorithm (n < 5 records) (Phillips et al. 2017). Maxent models were fitted using a tuning approach and six environmental variables: four climatic variables (annual mean temperature, temperature seasonality, precipitation of wettest and driest quarters) and two soil variables (cation exchange capacity and clay percentage). The binarization process employed a site-specific threshold approach (Scherrer et al. 2018), where model predictions were converted to presence/absence by first estimating site richness through summing species probabilities, then ranking species by probability and selecting the highest-ranked species up to the estimated richness (Rodrigues and Duarte 2024). Here we use species distribution predictions for 96 species (12% of Myrcia's Species, representing the 10 sections) that are also present in the phylogenetic tree. Although the number of species is low related to the genus richness, a comparison of the richness using all the species with distribution maps and the ones present in the phylogeny shows a high correlation (rho = 0.958) and the visual pattern is very similar across the Neotropical region, especially in the Atlantic Forest, which is the focus of this study (Figure S1).

### Biogeographic regions and ancestral area reconstructions of Myrcia

Biogeographical regionalization was based on evoregions, a regionalization method that uses phylogenetic composition of *Myrcia* across the Neotropics (Rodrigues and Duarte 2024). This regionalization divides the area occupied by the genus in the Neotropics in five regions. In the Atlantic Forest there are 2 main regions (Figure 2D), the evoregion A covering the southern part of Atlantic Forest and the mountainous area, and the evoregion B covering the northern part of Atlantic Forest and low land areas. In the extreme north of Atlantic Forest, some grid cells are defined as part of the evoregion C. This region extends mostly outside the Atlantic Forest, Neotropics, ranging from Central America, Andes and Eastern Amazonia (Figure 2B). Rodrigues and Duarte (2024) previously reconstructed the ancestral ranges of *Myrcia* using six biogeographic models in BioGeoBEARS, selecting DEC+J as the best-fitting model based on AICc. This model allowed ancestral nodes to occupy up to three areas and suggested Evoregion B as the most likely origin of *Myrcia*, with clades in Evoregion A associated with colder environments.

In this study, we build on that framework and use Biogeographical Stochastic Mapping (BSM), generating 100 realizations based on the DEC+J parameters, to infer the biogeographical history of the lineages. This approach incorporates both cladogenetic and anagenetic events and allows us to estimate colonization ages and in situ diversification metrics while accounting for uncertainty in ancestral histories. Details on the implementation of the BSM in the *Herodotools* R package are described in the Suplemmentary Material (Text S1.1).



**Figure 2.** Atlantic Forest spatial context. A) Spatial distribution of Atlantic Forest (dark grey) in the South America. B) Biogeographic regions defined from phylogenetic composition (evoregions) for the genus *Myrcia* (Rodrigues and Duarte 2024). C) Elevation variation across Atlantic Forest. D) Biogeographic regions in the Atlantic Forest.

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274 We calculated a modified version of the Diversification Rate metric (DR; Jetz et al. 275 2012), to separate it into two components: in situ diversification (hereafter called DR<sub>in situ</sub>) and historical dispersal., DRin situ represents the proportion of the speciation events 276 277 (measured as DR) that occurred exclusively within the region where a species currently 278 occurs. DR metric is the inverse of the equal-splits measure (ES; Redding and Mooers 2006) and is calculated as  $DR_{Jetz} = 1/ES$ . On a fully binary tree, the ES metric for a species 279 *i* is the sum of the edge lengths in the path from tip to root of the tree, consecutively 280 discounted by a factor of 0.5: 281

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$$ES_i = \sum_{j=1}^{N_i} l_j \frac{1}{2^{j-1}}$$

where  $l_i$  is the length of the edge j, j=1 is the edge of the tip of the focal species, and  $N_i$  is the number of nodes in the path from the tip to the root of the tree. We used the same equation to derive the  $ES_{in \, situ}$  but with a modified definition of  $N_i$ . In  $ES_{in \, situ}$  (i.e. considering only the speciation events occurring within the focal region),  $N_i$  represents the number of nodes from the tip to the earliest internal node that continuously occupied the region of the focal assemblage where the species occurs. The information on node occupancy along the lineage path was derived from an ancestral reconstruction model. From this, DR<sub>in situ</sub> was calculated as 1/ES<sub>in situ</sub>. To calculate the community average, we use the harmonic mean for DR<sub>letz</sub> and DR<sub>in situ</sub> (ignoring zeros for DR<sub>in situ</sub>). To calculate a proportion of the DR that have occurred in situ, we calculate species level as  $DR_{prop} =$  $DR_{Jetz} \times (ES_{in \, situ}/ES_{total})$  and averaged for each site as Mean  $DR_{prop} = \sum DR_{prop} / \sum DR_{Jetz}$ . This proportional DR<sub>prop</sub> (hereafter simply DR<sub>prop</sub>) ranges from 0 to 1, where 0 indicates that the entire evolutionary history of the species in an assemblage occurred in another region (lineage dispersal) and 1 indicates that the entire the evolutionary history of the species occurred within the assemblage's region (in situ diversification process). Consequently, Mean DR<sub>prop</sub> capture the relative role of lineage dispersal (higher role when Mean  $DR_{prop}$  < 0.5) and in situ diversification history (higher role when Mean  $DR_{prop}$  > 0.5) for the assemblage. Diversification metrics were computed with the function 'calc\_insitu\_diversification' from *Herodotools* version 2.0 (Nakamura et al. 2024). Since we are using 100 BSM maps to infer biogeographical history, we compute average and coefficient of variation of mean DR<sub>prop</sub> and DR<sub>in situ</sub> across the BSM maps.

Colonization Age of Assemblages

We used the same framework described in Box 1 to calculate the colonization age of each species in the region of the assemblage, using the function 'calc\_age\_arrival' from the *Herodotools* package. We calculated the mean colonization age for each assemblage based on all species' colonization ages in the assemblages they currently occur. Using results from the 100 BSM maps, we compute average and coefficient of variation of community average age across the BSM maps.

### Diversity metrics

We quantified two assemblage-level diversity metrics: species richness and phylogenetic divergence (Tucker et al. 2017). Species richness was calculated as the number of species in each grid. We quantified phylogenetic divergence of assemblages using the standardized effect size of mean pairwise distance (SES MPD) among the species in a site. The effect size was calculated by using a null model that randomizes the taxa labels in the phylogeny with 1000 repetitions of MPD calculation. To compute the SES MPD we used the package *picante* (Kembel et al. 2010). Low values of SES MPD indicate that the assemblage is assembled by phylogenetically closed-related species and high values of SES MPD indicate that the assemblage is assembled by phylogenetically distant-related species.

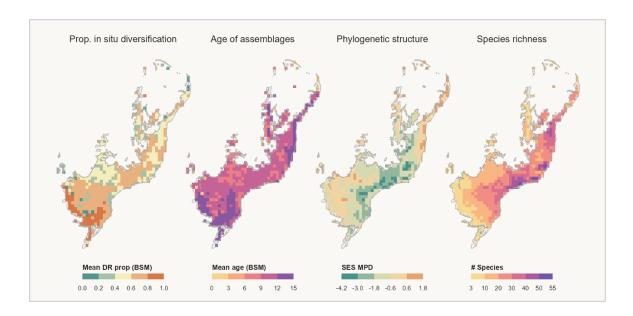
# Identyfying legacies in Myrtaceae in Atlantic Forest

The last step in our methodological framework consists in providing a general overview of the different processes affecting the assemblages across the Atlantic Forest (see Table 1). We did this by performing a cluster analysis where we classified the assemblages based on four metrics: DR<sub>prop</sub>, colonization age of assemblages, SES MPD (phylogenetic divergence) and species richness. We used Euclidian distance of the metrics' z-scores with ward clustering algorithm. In addition, to interpret the differences among groups we performed a Principal Component Analysis (PCA) on the metrics. We choose the number of groups (three to five) by visually assessing the spatial arrangement of the cluster analysis, resulting in four groups.

## Results

The  $DR_{prop}$  values were higher in the evoregion A sites than in the evoregion B sites (Figure 3A). High  $DR_{prop}$  means that evoregion A showed higher contribution of in situ diversification than lineage dispersal process to the assemblage composition. On the other hand, evoregion B showed an overall even contribution of both processes ( $DR_{prop}$  around 0.5), but the NE coast shows higher  $DR_{prop}$  (values between 0.6 and 0.8). Most of the sites

with low DR<sub>prop</sub> values were located in evoregion B and close to the border of the region. DR<sub>jetz</sub> and DR<sub>in situ</sub> showed similar spatial patters between them (Figure S2) in which the higher diversification rates are in the central Atlantic Forest. This is a mountainous region that connects with the Cerrado biome in central Brazil. Spatial patterns of assemblage colonization ages (Figure 3) showed older colonization ages in assemblages at the southernmost Atlantic Forest, northeastern coast and in highlands of the northeast continental area. Assemblage age varied between 12-15 Ma in these areas (purple sites in Figure 3B). Most of the assemblages in the Central Atlantic Forest had ages between 9-12 Ma. It is important to highlight that *Myrcia* has an crown node estimated around 27 Ma (24.55 – 31.19 Ma; Vasconcelos et al. 2020).

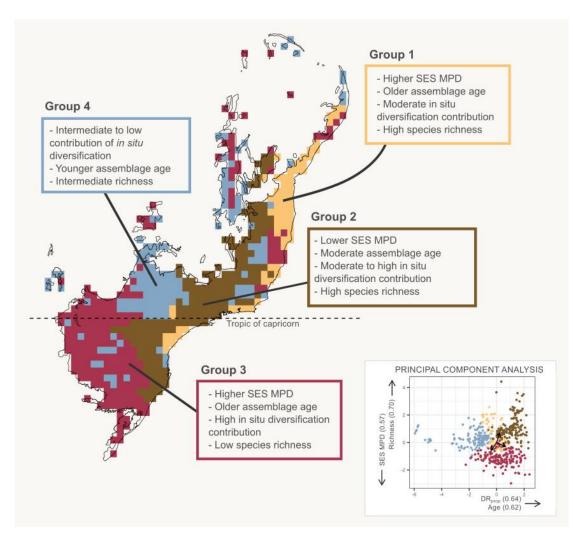


**Figure 3.** Spatial patterns of diversity and biogeographical processes influencing the assembly of *Myrcia* species in the Atlantic Forest. From left to right, proportion of in situ diversification quantifies the role of in situ diversification (high values) vs dispersal (low values) processes for the assemblage; mean age of assemblages, Phylogenetic structure, represent clustered (negative values) or overdispersed (positive values) assemblages; Species richness (D) is the number of species in each site.

The ancestral reconstruction affects the  $DR_{prop}$ ,  $DR_{in \, situ}$  and age of assemblages estimates since these metrics assumes the ancestral reconstruction is known. Using the BSM approach we were able to propagate the uncertainty in the ancestral reconstruction estimates to the assemblage level metrics. Our results show low uncertainty for  $DR_{prop}$  and  $DR_{in \, situ}$  which median coefficient of variation across the sites were 0.05 and 0.07, respectively. However, the age of assemblages showed higher uncertainty with median

coefficient of variation across the sites of 0.35. The spatial variation of the uncertainty is quite uniform in all the metrics (Figure S3)

Species richness, using only the species present in the phylogeny, showed at least three peaks along the coastal area of the Atlantic Forest (Figure 3). From South to North, the largest area with high species richness is in the Northern portion of the Serra do Mar, the second peak is in the Espirito Santo State, and the last one is in the Bahia State. The former two peaks in species richness also had low SES MPD values, but that in the Bahia state had high values of SES MPD. SES MPD metric revealed clustered assemblages of *Myrcia* along the Serra do Mar range (greener area in Figure 3B). By contrast, phylogenetically overdispersed assemblages predominated in the northeastern Atlantic Forest (orange region in Figure 3B). The southwestern the Atlantic Forest displayed a phylogenetic structure consistent with random expectation.



**Figure 4.** Schematic figure showing distinct historical structure of lineages for *Myrcia* assemblages in the Atlantic Forest. Groups were defined using Ward method of hierarchical clustering based on

4 metrics: DR<sub>prop</sub>, age of assemblage, SES MPD and richness. The inset graphic represents a Principal Component Analysis (PCA) of assemblages using SES MPD, DR<sub>prop</sub>, and richenss. The values in parenthesis represent the correlation between each variable and the PCA axis. Cluster analysis and Principal Component Analysis used to interpret differences between groups are presented in the Supplementary Material.

**Table 2**. Principal Component Analysis summary showing the relationship of the variables with the Principal Components (PC) and the variation explained by each PC axis.

	PC1	PC2	PC3	PC4
Mean Colonization Age	0.623	-0.303	0.334	0.639
Mean Drprop	0.644	-0.291	-0.136	-0.694
SES MPD	-0.385	-0.572	0.680	-0.251
Richness	0.221	0.705	0.639	-0.215
Proportion of Variance	0.495	0.336	0.141	0.028
Cumulative Proportion	0.495	0.831	0.972	1.000

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By aggregating the information on DR<sub>prop</sub>, age of colonization, SES MPD and richness we found four groups of sites with distinct legacies from biogeographical historical processes (Figures 4 and S4). The Principal Component Analysis using the same variables, the first tree axes explaining 97% of the variation (Table 2 and Figure S5). PC1 explains 49.5% of the variation and are positively related to mean DR<sub>prop</sub> and mean colonization age. PC2 explains 33.6% of the variation and are positively related to richness and negatively related to SES MPD. In contrast, PC3 (14.1% of explained variation) is positively related to both richness and SES MPD. This means that DR<sub>prop</sub> and colonization age are corelated but independent to SES MPD. The four groups are well separated in the PCA space (Figure 4 and S5). Following the expectations summarized in Table 1, we can classify the Group 1 (yellow in Figure 4) as a stable area acting as refugium since sites in this group present moderate history of in situ diversification, old colonization age of assemblages, overdispersed phylogenetic structure and high species richness. Group 2 (brown in Figure 4) seems to be an unstable area acting as cradle since it has moderate to high proportion of in situ diversification, moderate to high colonization ages, clustered phylogenetic structure and intermediate to high species richness. Group 3 (magenta in Figure 4) can be classified as stable area acting as cradle since it has high proportion of in situ diversification, old age of colonization, high SES MPD but with low species richness. Finally, we classify Group 4

(blue in Figure 4) as a unstable area acting as sink of lineages since it has lower proportion of in situ diversification, younger age of colonization, neutral SES MPD (similar to expected by chance) and intermediate species richness.

#### Discussion

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Our study examined how evolutionary and biogeographical history shape present-day patterns of phylogenetic composition. We quantified the relative roles of in situ diversification and lineage dispersal in assembling communities, along with the colonization age and phylogenetic structure of these assemblages. Our results show that in situ diversification is proportionally higher in Evoregion A (groups 2 and 3 in Figure 4), especially in the subtropical zone, whereas lineage dispersal plays a larger role in shaping assemblages in Evoregion B. Furthermore, by integrating DR<sub>prop</sub>, colonization age, SES MPD, and species richness, we identified four regions with distinct historical biogeographical dynamics. This demonstrates the potential of our framework to uncover historical processes at the assemblage level.

High levels of in situ diversification may provide evolutionary advantages for species in assemblages subjected to stronger environmental selection pressure. In Evoregion A, particularly in the subtropical Southern Atlantic Forest (Group 3; Figure 4), assemblages are largely composed by species that diversified within this region. In contrast, assemblages in the evoregion B, are more evenly shaped by both in situ diversification and lineage dispersal. That suggests that lineages originating in Evoregion A have likely accumulated adaptations that enhanced their success during community assembly in these areas. Since evoregion A has colder climate, which are more frequent and intense at higher altitudes and latitudes (Rodrigues and Duarte 2024), we argue that these adaptations might be associated with the ability to cope with colder climatic conditions. For instance, in Southwest Atlantic Forest (Group 3, Figure 4), assemblages are dominated by some of the oldest lineages evolving within the Evoregion A, with only a few lineages having recently dispersed into the region. This indicates that in situ diversification plays a key role in shaping the phylogenetic composition of this area, despite its relative low species richness. Supporting this pattern, previous studies provided evidence of selection for cold-adapted lineages in tree assemblages in the subtropical Atlantic Forest, where Myrtaceae species are particularly important at high altitude sites (Bergamin et al. 2021; Duarte et al. 2014; Giehl and Jarenkow 2012). In addition, the subtropical Atlantic Forest appears to have remained under stable conditions since the last ~120 kyr (Carnaval et al. 2014). Together, these

patterns suggest that Group 3 area has have provided stable environmental condition for cold adapted species, and that the long-term persistence of the lineages in this region has promoted diversification, likely driven by lower extinction rates on cold adapted lineages.

Lineage dispersal appears to play a greater role in shaping tropical assemblages, particularly outside the Evoregion A. In these areas, evolutionary history seems to be less important for successful colonization, likely due to weaker selection pressures. Within the Atlantic Rainforest, species inhabiting warmer and wetter environmental conditions tend to exhibit a narrower climatic niche, whereas those in colder and drier environments exhibit broader niche (Klipel et al. 2022). These contrasting niche characteristics likely favored the selection of lineages tolerant or adapted to colder conditions in the Evoregion A. Conversely, in areas where the selection pressure is weaker, the history of diversification is less determinant for the assembly process. The Central-West Atlantic Forest (Group 4, Figure 4) exemplify such a case: here, in situ diversification and lineage dispersal contributes equally to assemblages with young age of colonization. This suggests that the region functions as an unstable sink for lineages originating elsewhere. In such context, contemporary ecological processes, such as competition and priority effects, are likely more important in shaping community assembly than geographical diversification history (Gerhold et al. 2018; Mittelbach and Schemske 2015; Ricklefs 1987).

Northeastern Atlantic Forest is characterized by an old colonization history, intermediate proportion of in situ diversification, an overdispersed phylogenetic structure, and high species richness (Group 1, Figure 4). These patterns are consistent with the Bahia refugium hypothesis, which proposes that this region has maintained relatively stable climatic conditions since at least the last 21 ky (Carnaval et al. 2014; Carnaval and Moritz 2008; Staggemeier et al. 2015). The stable climatic condition potentially enabled the persistence of old and distant-related lineages, which in turn contributed to the high species richness found in this area. In a study focused on *Myrcia* section *Aulomyrcia*, Staggemeier et al. (2015) found a low extinction rate in the Bahia refugium and high speciation rates outside the refugium area. Therefore, this region seems to be stable for long time and act as a refugium for *Myrcia* lineages.

Central Atlantic Forest (Group 2, Figure 4) probably act as an unstable cradle area. This is a mountainous region with the higher altitude in Atlantic Forest. Mountainous regions can promote ecological speciation due to high environmental and topographic heterogeneity (Pyron et al. 2015; Rahbek et al. 2019) which can also be a source of higher

levels of extinction (Rahbek et al. 2019; Vasconcelos et al. 2022). The climate in this area has been shown to have higher levels of paleoinstability since last interglacial maximum (Carnaval et al. 2014; Carnaval and Moritz 2008). The climatic instability associated to the topographic heterogeneity of the area has resulted in phylogenetic clustered assemblages with moderate to high levels of in situ diversification, but with intermediate colonization ages in comparison to other areas. We speculate that the diversity in this area results from highly dynamic evolutionary and biogeographical processes, characterized by continuous speciation driven by the creation of novel habitats and the fragmentation of species distribution, as well as range expansion and contractions in species range in response to environmental changes, alongside the extinction of some lineages.

By integrating diversity, in situ diversification history, and phylogenetic structure of assemblages, we were able to identify areas with distinct historical dynamics that surpass biogeographical borders (defined here as evoregions) and explain compositional variation within biogeographical areas. Typically, the influence of biogeographical history and its dynamics on biodiversity patterns is examined at courser resolutions using biogeographical areas as the unity of analysis (Antonelli et al. 2018; Matos-Maraví et al. 2021). Here we increase the resolution of analysis and address historical biogeographical dynamics at local assemblages, which is an important step towards understanding the effects of past dynamics on local communities (Gerhold et al. 2018; Mittelbach and Schemske 2015). This methodological advance was possible because we incorporate biogeographical information (from ancestral area estimates models) into tip-based diversification metrics and combining these with diversity (richness) and phylogenetic metrics (MPD). Therefore, our approach allows the detection of the influence of past source-sink dynamics among evolutionary regions on present-day local assemblages that could be applied in other systems and taxonomic groups.

Our study shed light on the legacies of the biogeographical history on the current assembly of *Myrcia* species in the Atlantic Forest. We were able to capture distinct role of in situ diversification and dispersal across the Atlantic Forest and identify areas with distinct biogeographical dynamics. In line with other studies, we detected patterns of a stable area acting as a refugium in the Northeastern Atlantic Forest (Carnaval et al. 2014; Carnaval and Moritz 2008). In addition, we showed a large area acting as cradle and areas acting as sink of lineages. Finally, our findings demonstrate the ability of the new proposed metrics of in situ diversification to shed light on within biome diversification dynamics and its legacies at the local assemblages.

501	Data availability statement
502	Code and data necessary to reproduce the analysis are available at
503	https://doi.org/10.5281/zenodo.14171268
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# Supplementary Material

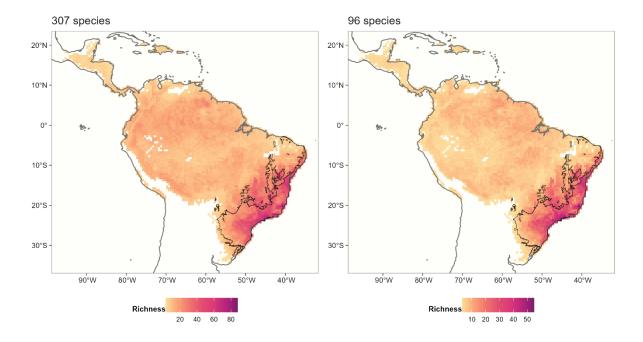


Figure S1. *Myrcia*'s species richness patterns in Neotropics. The figures compare the spatial pattern between two datasets, one with all species with distribution map available (left panel) and one with only the species that are present in the phylogenetic tree used in this study (right panel). Pearson coefficient of correlation is 0.958. The outer lines show the border of America's continent and inner lines show the border of the Atlantic Forest.

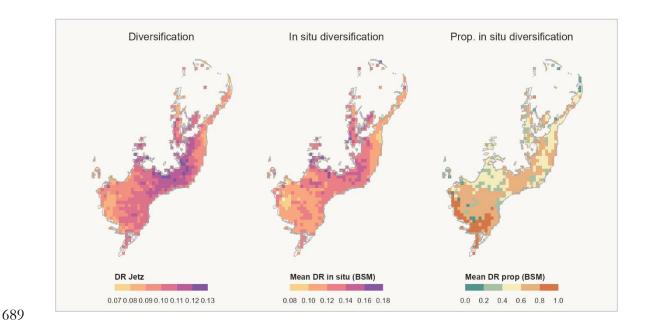


Figure S2. Comparison among the diversification rate (DR) metrics. Community average DR as the inverse of the equal splits as proposed by Jetz et al. 2012 is shown in the left panel (Diversification). Central panel shows DR in situ (In situ diversification) and the right panel shows DR proportional (Prop. In situ diversification). DR in situ calculates the inverse of the equal splits only for species in the community that have diversified in situ and for the time it has occurred in situ since colonization. DR proportional calculates the mean proportion of the diversification history that has occurred in situ. For details on the calculation of DR metrics refer to the Methods section of the main text.

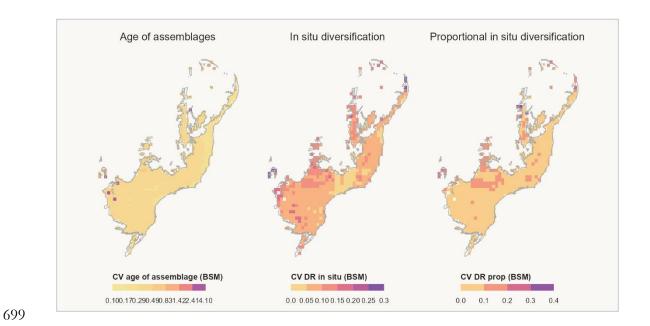


Figure S3. Uncertainty (calculated as the coefficient of variation) of colonization age of assemblages, DR in situ (In situ diversification) and DR prop (Proportional in situ diversification) based on 100 biogeographical stochastic mappings using DEC model of ancestral state reconstruction with BioGeoBEARS.

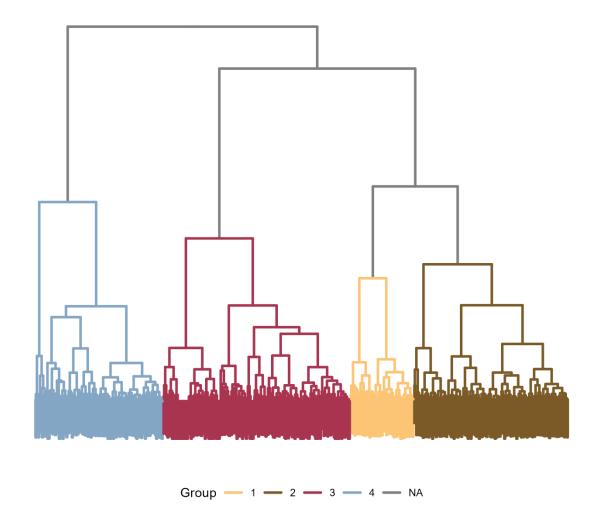


Figure S4. Dendrogram of ward clustering of sites based on *in situ* diversification,
 colonization age of assemblages, SES MPD and richness. Each color represents one group.

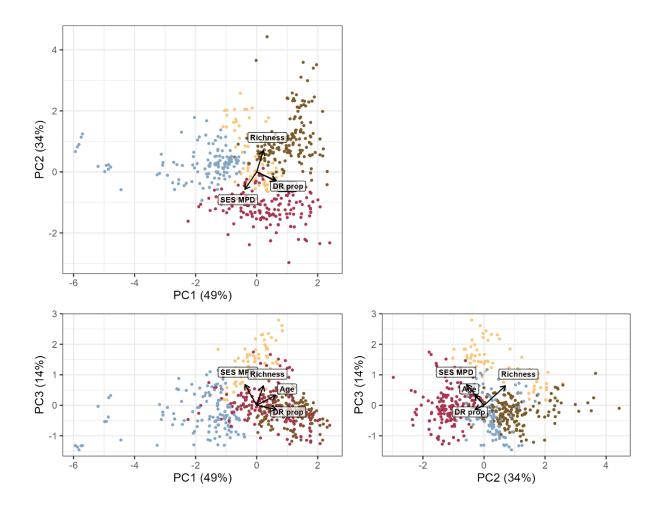


Figure S5. Principal Component Analysis of the site's variability on *in situ* diversification (represented by DR prop), colonization age of assemblages (Age), SES MPD and richness.

Each color represents one group of sites obtained using ward clustering method (Figure S4).

# **Text S1 - Methods description:**

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716 Text S1.1 - Implementation of the Biogeographic Stochastic Mapping (BSM) in 717 Herodotools R package 718 719 The Herodotools package was developed to integrate biogeographical history methods of 720 ancestral range reconstruction with phylogenetic metrics at the assemblage level. This 721 integration uses ancestral area estimates as the basis for disentangling a species' history 722 between the region where it currently occurs (the assemblage) and other regions (Box 1, 723 main text). The method accepts biogeographic reconstructions from any of the models 724 implemented in the R package BioGeoBEARS (Matzke, 2013). 725 In the first implementation (Nakamura et al., 2024), Herodotools, used the most likely range 726 state estimated at each phylogenetic node to track species' biogeographical histories. A key 727 limitation of this approach is that it cannot account for range-state changes along branches, 728 since only node-based changes are considered. Moreover, by relying on a single most likely 729 ancestral state, this approach does not incorporate the uncertainty inherent in ancestral 730 range estimates. 731 To overcome these issues, we extended the *Herodotools* to incorporate Biogeographic 732 Stochastic Mapping (BSM) outputs from BioGeoBEARS into its workflow. BSM 733 generates a set of equally probable histories consistent with the phylogenetic tree, model 734 parameters, and data. Each stochastic map represents one possible biogeographic history, 735 assigning range changes to both nodes and branches by random sampling based on a fitted 736 biogeographic model (e.g., the Dispersal-Extinction-Cladogenesis model, DEC). Because 737 each map is a random draw, multiple reconstructions (e.g., 100) must be generated to 738 capture the uncertainty in range change estimates. The advantage of using the BSM is that it tracks the range state transition along the entire phylogeny (nodes and branches) and by 739

producing many maps, allows the quantification of uncertainty in ancestral range
reconstructions. Within the *Herodotools* framework, assemblage-level phylogenetic metrics
are computed for each BSM map, thereby propagating uncertainty in biogeographic
reconstruction into the phylogenetic metrics. After metrics are computed across all maps,
summary statistics such as the mean and standard deviation can be derived.

# Text S1.2 - Obtaining Diversification Rates and importance of in situ diversification

# at assemblage scale

Based on the results of biogeographical ancestral range reconstruction with BSM, we modified the diversification rate (DR) metric proposed by Jetz et al., (2012) to account for in situ diversification. By in situ diversification, we refer to the speciation rate that occurs for a set of lineages within a given region (in this study, an evoregion). Here, we detail how we obtained the proportion of diversification attributable to in situ speciation, which we call  $DR_{prop}$ .

We first calculated  $DR_{in \, situ}$ —the diversification rate of each lineage within the region it currently occupies. This value is defined as the inverse of the in situ evolutionary distinctness,  $ED_{in \, situ}$ , for each lineage. For a given region,  $ED_{in \, situ}$  is computed by tracing each lineage only along the branches (from tip to root) where the ancestral range includes that region. By definition,  $ED_{in \, situ} \leq ED_{total}$ . EDtotal is the ED calculation that do not account for the in situ diversification, as originally proposed (Redding et al., 2008; Redding & Mooers, 2006).

Finally, based on  $ED_{toalt}$ ,  $ED_{in \, situ}$ , and  $DR_{Jetz}$ , we derived the proportion of diversification due to in situ speciation, denoted as  $DR_{prop}$  in situ per lineage i.

$$DR_i^{prop} = DR_i \cdot \frac{ED_i^{insitu}}{ED_i^{tot}} = \frac{ED_i^{insitu}}{(ED_i^{tot})^2}$$

DR<sub>prop</sub> was calculated for each lineage i, which in this study corresponds to each tip of the phylogenetic tree. We then scaled up lineage-level DR<sub>prop</sub> in situ to the assemblage level by computing the community-weighted average of per-species in situ proportions, denoted as mean DR<sub>prop</sub>, as follows:

$$Mean DR_{comm} = \frac{\sum_{i} DR_{i}^{prop}}{\sum_{i} DR_{i}^{Jetz}}$$

This equation corresponds to a community-weighted average of per-lineage in situ

proportions, where the weights are proportional to the diversification rate of each lineage.

It can be interpreted as the average fraction of diversification occurring in situ across all

lineages present in the assemblage, with species exhibiting higher in situ diversification

rates contributing more strongly to the overall value. In the Herodotools package, the

function calc\_insitu\_diversification is used to calculate DR<sub>Jetz</sub> DR<sub>in situ</sub>, DR<sub>prop</sub>, and their

correspondents community mean values.

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