

Life-history variation mediates the importance of population structure to the short-term dynamics of plant populations worldwide

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24 **Abstract**

25 A population's structure (*i.e.*, the distribution of individuals across the life cycle of a species)
26 influences how it responds to environmental changes and recurrent disturbances, and shapes its
27 vulnerability to extinction. Yet, despite compelling evidence that population structures are rarely
28 stationary over time, assessments of population viability have only recently begun to consider how
29 a population's structure interacts with its vital rates of survival, development, and fecundity to
30 determine its demographic performance. Using 268 comparisons of population dynamics in 56
31 plant species across differing time periods, locations, and/or experimental treatments, we use
32 transient Life Table Response Experiments (tLTREs) to quantify how changes in population
33 structure and vital rates contribute to observed variation in demographic performance. We
34 illustrate how changes in population structure are major contributors to short-term (*i.e.* transient)
35 demographic performance, and are just as important as the effects of, more routinely evaluated,
36 changes in survival and fecundity rates for informing population viability assessments. Moreover,
37 these contribution patterns challenge prevailing understanding of the links between plant growth
38 form and population viability. Accordingly, we emphasise how quantifying the effects of
39 population structure on transient population dynamics can reveal key mechanistic insights
40 necessary for accurately predicting the responses of natural populations worldwide to ongoing
41 global change. Indeed, linking the structure and dynamics of natural populations is essential for
42 the effective management and conservation of global biodiversity.

43 **Significance Statement**

44 Using 268 comparisons of population dynamics in 56 plant species across different time periods,
45 locations, and/or experimental treatments, we show how changes in population structure are just
46 as important as the effects of changing survival and fecundity rates in determining population
47 viability; particularly in shorter-lived species. Current assessments of population viability largely
48 focus on evaluating how changing vital rates affect population dynamics. However, the critical
49 role of population structure in mediating population dynamics evidences how current efforts to
50 forecast plant populations and communities are incomplete. As such, we emphasise the need for
51 explicitly including population structure in assessments of demographic performance and
52 resilience for the effective management of natural populations.

Introduction

Historically, assessments of population viability typically evaluate how individual-level patterns in survival, development, and fecundity respond to disturbances and environmental fluctuations (*e.g.*, seasonal cycles, drought, forest forest etc.) (1). Indeed, transient (*i.e.*, short-term; although some transient cycles can persist over long periods (2)) variation in these vital rates underpins the resistance and recovery potential of natural populations (3). However, demographic performance (*i.e.*, a population's growth rate) arises from an interaction between a population's vital rates and its structure (4). Survival, development, and fecundity rates shape the structure of populations to directly inform their realised performance (5, 6) (Fig. 1A). Unless a population possesses a stable structure, whereby it will change in size at a constant rate over time (its long-term population growth rate, λ), its vital rates will change the age- (7), size- (8), developmental stage- (9), or other state-distribution (*e.g.*, sex ratio (10)) of the individuals within the population. Accordingly, population trajectories comprise the direct and indirect effects of stochastic and disturbance-induced variability across both the vital rates and structure of populations (11, 12). Thus, accounting for population structure, a feature that is rarely done in demography (5, 13), can provide greater mechanistic insight into demographic performance and resilience (14), though to what extent remains unquantified.

Periodic disturbances and/or environmental shifts prevent populations from attaining their stable population structure or stationary growth trajectory (*i.e.*, λ) (15). Instead, populations are maintained within a transient state during which their realised growth rates ($\lambda_{transient}$) change through time and/or space (16). The transient Life Table Response Experiment framework (4) (*tLTRE*, hereafter) was introduced as a way to decompose this variation in $\lambda_{transient}$ into the contributions provided by underlying shifts in observed vital rates and population structure (4, 11)

(Fig. 1B). tLTREs have since been implemented in the case-specific assessment of various mammal (*e.g.*, (5, 6, 17)) and bird species (*e.g.*, (14, 18–21)). However, while theory predicts a link between life-history variation and the contributions of population structure to variation in $\lambda_{transient}$ (4), how population structure influences realised demographic performance across natural populations and life-history strategies remains untested.

Here, to assess how the contribution of population structure to variation in demographic performance corresponds with established axes of life-history variation, we apply tLTREs to comparisons of population dynamics in 56 plant species across differing time periods, locations, and/or experimental treatments (22). Specifically, we test the following hypotheses: (*H1*) Observed variation in population structure will elicit comparable contributions to variance in $\lambda_{transient}$ to those from observed fluctuations in population vital rates. Our rationale for this expectation is that changes to population vital rates do not occur independently of changes to population structure, and instead comprise the indirect effects of changing population structures (4, 11, 23). However, we expect that (*H2*) the relative importance of contributions from population structure will be greater in species with faster life-histories (*e.g.*, shorter generation times (24)). Indeed, the low survival associated with shorter-lived species renders individuals more susceptible to abiotic anomalies, compared to longer-lived species (25), and thus their population structures are likely more variable through time. Concurrently, we predict that (*H3*) greater increases in survival across the life cycle of a population will be associated with increased contributions from variation in population structure to variance in $\lambda_{transient}$. Our rationale here is that low survival early in the life cycle suggests that species invest little into offspring survival (*i.e.*, smaller seed size), a characteristic associated with faster-lived plant species (26).

We also expect that (*H4*) increasing iteroparity will be associated with decreased contributions from variation in population structure to variance in $\lambda_{transient}$. Iteroparous species typically have lengthened reproductive windows and thus likely more opportunities to contribute new individuals to their populations (27). As such, demographic performance in these species is likely less sensitive to changes in the distribution of individuals across their life cycle. However, we anticipate that (*H5*) observed relationships between sources of variance in demographic performance and life-history strategies will be strongly mediated by plant growth form, with contributions from changes in population structure becoming increasingly important in species whose anatomy does not allow for the seasonal loss of aboveground biomass (*e.g.*, American beech, *Fagus grandifolia*) compared to those that do (*e.g.*, Cowslip, *Primula veris*). Plant species whose individuals physically recede during overwintering periods likely experience more frequent changes in their population structure, and so the dynamics of these species should be less sensitive to changes in population structure.

Results

Population structure cannot be overlooked

Fluctuations in population structure have similar impacts on variance in demographic performance as changes in the population vital rates (Fig. 2A). We used tLTREs to test how (*H1*) contributions from changes in population structure (*n*) to observed variance in $\lambda_{transient}$ compared to the contributions from changes in survival (σ), development (γ), and fecundity (ϕ). To do so, we partitioned observed variance in $\lambda_{transient}$ across our 268 comparisons of plant population dynamics into the contributions from associated changes in population vital rates, structure, and their interactions. Changes in any one of the vital rates or population structure increase observed

variance in demographic performance (Fig. 2A). Meanwhile, the effects of the interactions between changes across the vital rates ($\sigma:\gamma:\phi$) and between the vital rates and population structure ($\sigma:\gamma:\phi:n$) can both dampen or amplify variance observed in demographic performance (Fig. 2A). Selective pressures determine how organisms allocate finite energetic resources between survival and reproduction (28). Accordingly, the impacts of changing survival and fecundity patterns on demographic performance routinely form the crux of assessments into population viability (29, 30) and coexistence theory (31). However, we show that the impact of observed variation in population structure on demographic performance is comparable to that of shifts in survival and fecundity (Fig. 2A); a pattern that is insensitive to data imputations carried out during our analyses (Fig. S9A). Thus, overlooking the effect of a population's structure on its demographic performance can hinder the effective management of natural populations (32).

Like the effects of survival, development, and fecundity, the contribution of population structure to variance in $\lambda_{transient}$ exhibits a strong phylogenetic signal (Fig. 2A). To further evaluate (*HI*) how the effect of population structure on demographic performance compares to the impacts of the vital rates, we tested the strength of phylogenetic signal across our estimates of the contributions of changes in population vital rates and structure to observed variance in $\lambda_{transient}$. We estimated phylogenetic signal using Pagel's λ (33). We highlight the difference between λ , which represents the long-term growth rate of a population (34), and Pagel's λ , a proxy of phylogenetic inertia that ranges between 0 and 1. High phylogenetic signal (Pagel's $\lambda \approx 1$) suggests that the value of a trait of interest is likely more similar in more closely related species (35). As such, patterns in the contribution of population structure to variance in demographic performance appear shaped by species ancestry (Pagel's $\lambda = 0.65$ [95%

CI: 0.37, 0.83]; Fig. 2A). However, this signal is not as strong as that observed across the contributions of survival (σ), development (γ), fecundity (ϕ) and their interactions ($\sigma:\gamma:\phi$) (Pagel's $\lambda > 0.80$; Fig. 2A). Meanwhile, the contribution of interactions between population structure and vital rates ($\sigma:\gamma:\phi:n$) displays a low phylogenetic signal (Pagel's $\lambda = 0.25$ [0.03, 0.61]), suggesting that the effects of these interactions on short-term demographic performance are shaped by the environment, rather than phylogenetic ancestry.

Contributions of population structure to demographic performance align with the fast-slow continuum

We found support for our hypothesis ($H2$) that the relative contributions of variation in population structure to variance in demographic performance would increase with increasing life-history speed (Fig. 2B). We carried out a phylogenetically informed partial least squares analysis (pPLS) to assess the relationship of our measures of the contributions of changes in survival, development, fecundity, and population structure, with size-scaled measures of generation time (T), age at reproductive maturity (L_α), mean life expectancy (η_e), and reproductive window ($L_{\alpha-\omega}$; Table 1) computed from the mean model represented across each population comparison. These four life-history measures form a well-defined trait space that can be used to position species and their populations along two dominant axes of life-history variation: the fast-slow and reproductive strategy continua (27, 36).

Changes in population structure (n) typically have the greatest contributions to variance in short-term demographic performance across population comparisons in species with shorter generation times (Fig. 2B). In contrast, the population comparisons of species with longer generation times show greater contributions to variance in demographic performance from changes

in survival (σ) and development (γ) (Fig. 2B). Meanwhile, population comparisons exhibiting higher variance in $\lambda_{transient}$ align along an axis associated with the contributions of the interaction between population structure and vital rates ($\sigma:\gamma:\phi:n$) (Fig 2B). Thus, our findings suggest that accommodating population structure is important when assessing demographic performance (Fig. 2A), but that including population structure will not change the key role survival has been demonstrated to play in driving the short-term dynamics of populations with slow life-histories (e.g., *Malva nut*, *Scaphium borneense* (37)). However, to accurately forecast the demographic performance of short-lived species, we show that quantifying changes in population structure is essential.

We found evidence supporting our hypothesis (*H3*) that species with life cycle survival profiles ($\Delta\sigma$) associated with low offspring survival experience higher contributions from population structure to variance in demographic performance (Fig. 3A). We used phylogenetically informed Bayesian multilevel modelling to test the relationship between measures of contribution ratio (C_{ratio}) and the additive effects of size-scaled measures of mean lifetime survival and life cycle survival profile ($\underline{\sigma}$ & $\Delta\sigma$; Table 1) obtained across our 268 population comparisons. We estimated C_{ratio} as the sum of the main-effect and interaction contributions involving population structure (n & $\sigma:\gamma:\phi:n$) divided by the sum of the main-effect and interaction contributions involving vital rates only (σ , γ , ϕ and $\sigma:\gamma:\phi$). Thus, contribution ratio provides a measure of the relative contributions from population structure vs. vital rates to variance in demographic performance, with $C_{ratio} > 1$ indicating a greater influence of population structure. This observed relationship between $\Delta\sigma$ and C_{ratio} , although consistent across plant growth forms, becomes steeper in species who retain more above-ground biomass during unfavourable growth intervals

(Fig. 3B); represented here by the transition from hemicryptophytes (*i.e.*, plants whose shoot meristems recede to ground level during overwintering periods), to chamaephytes (*i.e.*, woody plants whose shoot meristems remain above ground, but lower than ~ 25 cm, during overwintering periods), and phanerophytes (*i.e.*, woody plants whose shoot meristems remain higher than 25 cm aboveground during overwintering periods; Fig. 3B). We note here, that although represented in our dataset, we do not show the pattern observed for epiphytic plant populations due to them only representing a limited range along our $\Delta\sigma$ scale (Fig. S2).

Modelling the relationship between degree of iteroparity (S) and contribution ratio (C_{ratio} ; Fig. 4A) offers little support to our hypothesis ($H4$) that the importance of contributions of population structure to demographic performance would decrease with increasing iteroparity. However, including Raunkiær growth form as a fixed effect reveals important heterogeneity across plant types, thus partially supporting $H5$, with C_{ratio} decreasing with increasing iteroparity in both hemicryptophytes (*e.g.* Hoary plantain, *Plantago media*, and Sticklewort, *Agrimonia eupatoria*; Fig. 4B, left) and phanerophytes (*e.g.*, shrubs like Palor palm, *Chamaedorea elegans*, or trees like the Sugar maple, *Acer saccharum*; Fig. 4B, right). Iteroparity affords species increased opportunity for contributing offspring throughout their life cycle, thus reducing their sensitivity to the loss of juvenile life stages (38). Iteroparous species are also better able to take advantage of reproductive strategies maximising reproductive output during periods of favourable conditions (*e.g.*, masting), potentially buffering their demographic performance against increased stochasticity (39). However, that a negative relationship between iteroparity (S) and the contribution of population structure to demographic performance is not ubiquitous across plant species emphasizes the importance of evaluating the contextual role population structure plays in influencing population dynamics.

Before interpreting patterns in the relationship between contribution ratio (C_{ratio}) and our measures of life cycle survival profile ($\Delta\sigma$) and degree of iteroparity (S) across plant growth forms, it is necessary to consider the analytical pipeline of a tLTRE. Indeed, contributions to variance in short-term demographic performance derived from tLTREs comprise both the sensitivity of $\lambda_{transient}$ to population parameters as well as the magnitude by which these parameters changed empirically in our dataset (4). Summarising the contribution ratios (C_{ratio}) exhibited across epiphytes (*i.e.*, plants who non-parasitically grow on other plants, like Bird's nest Fern, *Asplenium nidus*), hemicryptophytes, chamaephytes (*e.g.*, Periwinkle, *Vinca minor*), and phanerophytes, it becomes apparent that contributions from population structure are not only more important in species that exhibit seasonal individual-level shrinkage (*e.g.*, hemicryptophytes like *P. veris*), but also in phanerophytes, slow-growing trees and shrubs that invest more into woody tissues (40) (*e.g.*, Giant sequoia, *Sequoiadendron giganteum*; Fig 5). As such, demographic performance in phanerophytes is likely more sensitive to the loss of individuals. Alternatively, hemicryptophytes typically undergo extreme annual shrinkage, and so their population structures typically change frequently over time (41). Accordingly, higher contribution ratios across hemicryptophytes and phanerophytes likely reflects the amount their population structure changes *vs.* their sensitivity to even just small changes in population structure.

Discussion

Linking demographic tools and conservation science

The stochastic nature of ecological systems (42), coupled with the rising prevalence of recurrent disturbance regimes (43), means that transient dynamics play a key role in determining near-term population forecasts (23, 44). Here, we emphasise how population structure has a substantial

impact on the short-term dynamics of natural populations, and must therefore be explicitly accommodated within assessments of population viability (45). Specifically, the impacts of changes in population structure on demographic performance are of comparable magnitude to the effects of changing the fitness components of survival and fecundity rates, which underpin population viability (24, 28).

Documenting trends in population abundances is a key part of conservation science. Indeed, serving as the foundation for robust predictions of local (46) and global extinction risks (47), assessments of population size underpin three of the five criteria used in determining the IUCN Red List conservation status of species worldwide: ‘*High Decline Rate*’, ‘*Small Population Size and Decline*’ & ‘*Very Small Population Size*’ (48). A fourth criterion, ‘*Unfavourable Quantitative Analysis*’, then relates to estimating measures of demographic performance, which the IUCN recommends be done via population viability analyses (PVAs) (49). However, the utility of PVA frameworks for accurately predicting the trajectories of real-world systems is increasingly disputed (50). Moreover, these approaches frequently overlook transient dynamics and, crucially, the role of population structure in mediating population viability (Fig. 2A) (4, 51). Thus, although monitoring programs integrating population structure into assessments of species viability do exist (e.g., NOAA’s Pacific salmon [*Oncorhynchus spp.*] monitoring service (52)), the inclusion of empirical records of population structure is not commonplace in biodiversity management and conservation. Hence, demonstrating the link between population structure and demographic performance, as we do here, offers a valuable platform for translating demographic modelling outputs into comprehensible applied insights for enhancing management predictions (11). Equally, highlighting the relevance of contributions of population structure to demographic performance,

emphasises the value of adopting adaptive strategies to manage population structure in order to elicit its cascading effects on the dynamics of populations (32).

Linking population structure and life-history variation

We find that the value of population structure increases in species with faster life-history strategies. Specifically, contributions from changes in population structure to variance in short-term demographic performance are greater in populations with shorter generation times and/or whose investments into survival during earlier life stages are lower. These characteristics are indicative of faster life-history strategies (53, 54), thus, our findings highlight the particularly high value of population structure for informing the near-term forecasts and conservation of faster-lived species. With generation time and lifespan also linked to extinction risk (48, 55) and adaptability (56), the relationship between the importance of changes in population structure to short-term population dynamics and generation time compounds the importance of considering population structure as part of ongoing biodiversity conservation efforts.

Our findings also suggest that temporal, spatial, and/or experimental comparisons of populations possessing intermediate generation times report the highest variance in short-term demographic performance; a pattern that persists irrespective of the populations position along the second component axis (Fig. 2B). Although, short-lived species are expected to perform well in environments characterised by random stochastic variability (i.i.d.) (57), their fitness is expected to decline relative to that of long-lived species as stochastic variability becomes increasingly autocorrelated over time (58). Subsequently, although mediated by the type of stochasticity imposed, populations with both high and low generation times display an apparent capacity for buffering their dynamics against increased environmental stochasticity. As such, we might expect

demographic performance in natural populations positioned more centrally on the generation time scale to be more labile in response to environmental variability. However, faster-lived species (*i.e.*, shorter-generation times) typically display more labile responses to environmental stochasticity, with the dynamics of slower-lived species (*i.e.*, shorter-generation times) more buffered against stochasticity (59). Accordingly, our findings here hint at how it is the cumulative effect of a species or populations suite of life-history strategies alongside its relative generation time that informs its overall response to environmental stochasticity. Interestingly, we report the population comparisons exhibiting greater variation in demographic performance to align along an axis associated with contributions from the interaction between population structure and vital rates to demographic performance ($\sigma:\gamma:\phi:n$). This axis describes a gradient from negative contributions of $\sigma:\gamma:\phi:n$ acting to diminish observed variance in short-term demographic performance, to positive contributions of $\sigma:\gamma:\phi:n$ that amplify observed variance. Regardless of these effects, populations that possess intermediate generation times appear less able to buffer their demographic performance against temporal, spatial, or experimentally imposed variability.

We also show how differences in plant growth form offer a nuanced understanding of the relationships between life-history traits and drivers of short-term demographic performance. For instance, we show how populations of slow growing woody phanerophytes (*e.g.*, Sugar maple, *A. saccharum*), who are sensitive to the loss of individuals (60), and populations of hemicryptophytes (*e.g.*, Silky oat-grass, *Danthonia sericea*), whose individuals annually recede back to ground level over winter, both exhibit greater contributions from their population structure to their demographic performance. One argument to explain these nuances relates to the mathematical pipeline underlying tLTRE analyses (4). Specifically a tLTRE is concerned with the sensitivity of short-term demographic performance to underlying population characteristics as well as the magnitude

of any observed changes in said underlying characteristics (4). However, similar patterns have also been reported in the contributions of transient dynamics to the realised demographic performance of plant populations, with greater contributions from transient dynamics observed in plant species positioned at the extremities of life-history complexity (23, 51). Thus transient potential is larger in populations of trees and monocarpic plants (*i.e.*, plants that flower and produce seeds only once in their lifetime) than perennial herbs and shrubs (51). Although not wholly aligned with Raunkiaer's growth form classifications, many monocarpic species are also considered hemicryptophytes (61). Thus, our observation here of congruent patterns in the demographic profiles of phanerophyte and hemicryptophyte plant species aligns with this previous evidence (51). The community contexts of many phanerophyte and hemicryptophyte species places a strong selective pressure on them producing many viable offspring to maximise their colonisation potential. A process that in turn elevates their transient potential (51). Hemicryptophytes include many early-successional plant species, who benefit from being able to anticipate and quickly monopolise available resources (62, 63). Meanwhile, although late-successional, many tree species exist within dense forest ecosystems, and need to take advantage of erratic canopy gap dynamics (64). Thus, the short-term demographic performance of both phanerophyte and hemicryptophyte species may indeed be similarly reliant on the contributions of changes in their population structure.

Caveats and cautions

The life-history trait space represented by our study sample is not exempt of spatial and taxonomic biases, with the majority of species represented comprising herbaceous perennials (*i.e.*, plants that do not develop woody tissues and typically reach reproductive maturity within a few years or less

before losing their aboveground biomass during overwintering) studied in North America and Europe (65) (Fig. S1). Having shown how growth form and life-history variation correlate with the contributions of population structure to demographic performance, it is important to consider how the traits and growth forms of species not included could affect the patterns we observed. For instance, our sample lacks species possessing both long generation times and delayed ages at maturity (*e.g.*, Hoop pine, *Araucaria cunninghamii*) (27). Disturbance events associated with impacts on the structures and/or vital rates of plant populations are also not distributed equally across the globe. For instance, shifts in climatic regimes associated with changes to plant vital rates will likely be most acute in already highly arid and humid environments (66). Meanwhile, hotspots of deforestation, that directly affect the structure of affected tree populations, are found in South America and Asia (67), two areas for which we have limited data linking population structure and demographic performance (Fig. S1). Thus, continued efforts to understand how the interactions between population structure and vital rates informs demographic performance should focus on regions where the impacts of disturbances select for differing responses in population vital rates *vs.* population structure.

The findings we present here are not sensitive to the pooling of temporal, spatial and/or experimental comparisons across our data sample (Appendix S4). However, space-for-time substitutions have been criticised in ecology, as species do not necessarily respond to spatial climatic variation in the same way as they do to temporal shifts (68). One distinction in responses is likely due to the accumulation of impacts accrued whilst responding to temporal shifts over time (69), and the notion that future climate conditions will differ from those contemporary populations have experienced before (70). Alternatively, spatial comparisons likely involve comparisons among different populations who may have acclimatized to subtly different conditions over time.

Our dataset offers the largest scale of demographic data available to date linking associated changes in population vital rates and structure across time, space, and/or experimental treatments. Past synthesis work has demonstrated how COMPADRE (22), from which we sourced our data, and COMADRE (71) together contain over 1,600 temporal, spatial, and/or experimental comparisons of animal and plant population dynamics (72). Further explorations of these relationships are currently hindered by the lack of data explicitly detailing population structure across those studies (73). Considering the importance of population structure in determining short-term population dynamics, as we have shown here, we encourage authors to make population structure data publicly available, and for future efforts to quantify and assess population viability to capitalise on this key information.

Conclusions

Using 268 comparisons of population dynamics in 56 plant species across differing time periods, locations, and/or experimental treatments, we comprehensively assessed how population structure informs demographic performance across axes of life-history variation. Past theoretical work based on simplified population simulations evidenced a link between life-history variation and the contributions of population structure to variation in demographic performance (4). Yet, until now, we have lacked a robust empirical understanding of how population structure shapes population viability (13). Here, we demonstrate that the contributions of population structure to realised demographic performance are comparable to the effects of more routinely assessed patterns in population vital rates. As such, we illustrate the limitations of widespread practices assigning conservation statuses or offering forecasts of population dynamics in the absence of data describing population structure. We also highlight how life-history strategies and plant growth

forms can modulate the impacts of population structure on demographic performance. With robust population forecasts requiring an appreciation for the nuanced interplay between a population's structure and its vital rates (11, 74), we encourage monitoring efforts to adopt approaches facilitating the streamlined integration of empirical records of population structure into viability assessments, or at the very least, appropriate usage of data integration to estimate latent population structure and its impact on population viability (11, 75). For instance, combining remote sensing technologies with semi-automated image segmentation can provide extensive records linking the structure and vital rates of plant (76, 77) and animal populations (78). Indeed, effectively managing and conserving global biodiversity warrants continued efforts to understand how population structure mediates demographic performance.

Methods

We developed a comparative framework using transient Life Table Response Experiments (tLTREs) (*sensu* (4)) to test our five hypotheses. Briefly, our approach comprised the following steps: (a) sourcing suitable structured population models, (b) quantifying life-history traits, (c) implementing tLTREs, (d) determining the strength of phylogenetic ancestry across sources of variance in population growth rates, and (e) assessing the association between axes of life-history variation and the relative contribution of population structure and vital rates to non-stable population growth rates. We carried out all analyses in *R* (79), and our fully commented scripts are publicly available on Zenodo (<https://doi.org/10.5281/zenodo.15600668>).

Sourcing population models

We tested our hypotheses using a sample of structured population models representing the dynamics of different plant populations sourced from the COMPADRE Plant Matrix Database (version 6.23.5..0; (22)). COMPADRE contains almost 9,000 matrix population models (MPMs) detailing survival, development (growth & shrinkage), and fecundity patterns across the life cycles of ~790 plant species. This information has been compiled from mostly (>98%) peer-reviewed publications addressing various research questions (65).

We derived our sample of MPMs from an initial collection of 2,217 MPMs representing 212 plant species. This initial collection comprised sets of structured population models originally screened for LTRE testing used to illustrate the exactLTRE methodology (72). Briefly, this screening entailed a series of selection criteria. First MPMs were confirmed as ergodic, primitive, and irreducible, to ensure they contain direct or indirect transition pathways between all life cycle stages and thus possessed a single dominant eigenvalue (80). Second, MPMs comprising fewer than three life stages were rejected, as lower-dimensional MPMs provide unreliable representations of population dynamics (81). Third, accepted MPMs described annual census periodicities and were from studies that parameterised multiple MPMs for their focal species/population. This step ensured consistent time intervals and that we could compare between MPMs representing the dynamics of a species or population compared over different time periods (*e.g.*, (82)), sites (*e.g.*, (83)), or experimental treatments (*e.g.*, (84)). Fourth, accepted MPMs contain no records of clonal reproduction because inconsistencies in the distinction between genets and ramets (*i.e.*, colonies *vs.* clones) across studies can lead to unreliable comparisons of population growth rates. Fifth, accepted MPMs also contained no life-cycle transitions that had not been observed empirically or that represented post-hoc manipulations of observed data, such that no transitions had been observed once and assumed to be constant across all replicates (time,

sites, or treatments). Each accepted MPM was also meticulously screened for common life-cycle errors identified in Kendal *et al.* (85). Finally, accepted MPMs were required to have been digitised into COMPADRE alongside three sub-matrices (U , F , and C), allowing for the life cycle transitions it describes to be decomposed into the processes of survival (U), reproduction (F), and asexual propagation (C) (Eq. 1).

$$A = U + F + C \quad (1)$$

Here, from this initial collection of 2,217 MPMs, we specifically retained only models for which corresponding population vectors were available, either via publication in the original sources or because we obtained them via direct requests to authors. These population vectors detail the observed distributions of individuals across each life-stage (*i.e.*, state distribution). This step allowed us to link population vectors documented during demographic census to the corresponding MPM describing the associated stage-specific vital rates. As such, we partitioned sources of variance in observed population growth rates into contributions from population structure and vital rates. Following this second screening phase, we retained 845 MPMs representing 56 plant species across 268 populations suitable for tLTRE decomposition (Table S1). This final sample comprised sets of MPMs originating from various comparisons of plant population dynamics across time, space, or experimental treatments, or combinations of these cases (Table S1). However, our comparisons across these different dimensions are justified, as our findings are not sensitive to the pooling of comparisons of population dynamics made across time, space, or experimental treatments (Appendix S4).

Estimating tLTRE contributions

Testing each of our hypotheses necessitated quantifying the relative contribution of changes in population structure to observed variation in demographic performance across comparisons of plant population dynamics. For this step, we implemented tLTRE analyses using a modified version of the recently developed exactLTRE framework (72). Classical LTRE analysis quantifies the contributions to any observed variation in long-term (*i.e.* asymptotic) population growth rates (λ) affected by changes in particular MPM elements across a collection of models (86). However, due to their basis in Taylor Series approximation, classical LTRE approaches provide only an approximation of the contribution of vital rate changes to variation in λ (72). The exactLTRE framework uses a functional analysis of variance (fANOVA; (87, 88)) approach to avoid this approximation, accurately partitioning observed variance in demographic performance across both the main effects of survival, development, and fecundity, and the interactions between these vital rates (72). Here, to assess the contribution of changes in population structure, alongside changes in survival, development, and fecundity, to variation in realised (*i.e.* transient) population growth rates ($\lambda_{\text{transient}}$), we modified the functionality of the *exactLTRE* R package (89). Specifically, this modification allowed for partitioning observed variance in $\lambda_{\text{transient}}$ across the main and interaction effects of changes in population vital rates and age/stage structure. We note, that efforts are ongoing to add these modified functions to future versions of *exactLTRE*, to promote tLTRE tools.

We ran tLTRE analyses across our sample of 268 population comparisons. First, across each comparison, we computed $\lambda_{\text{transient}}$ for each population replicate (x) using their associated MPM (A_x) and population vectors (n_x) (Eq. 2). This metric indicates retrospectively whether the size of the population grew, remained constant, or declined ($\lambda_{\text{transient}} > 1, = 1, < 1$, respectively) over a single time interval.

$$\lambda_{transient} = \frac{\sum A_x n_x}{\sum n_x} \quad (2)$$

Next, we calculated the variance across the sequence of $\lambda_{transient}$ estimates obtained for each population replicate [$var(\lambda_{transient})$]. Typically, variance is calculated as the sum of the squared deviations divided by $N - 1$, where N is the sample size. However, this approach assumes an incomplete sample. Instead, when computing $var(\lambda_{transient})$, our use of the fANOVA framework assumes that we possess the complete sample of $\lambda_{transient}$ estimates. As such, our estimates of $var(\lambda_{transient})$ are equal to the mean of the squared deviations across each population.

After calculating $var(\lambda_{transient})$ for each population comparison, we decomposed the source of this variance into the relative contributions of any changes observed across their associated sequences of A and n products. By way of a walkthrough our exact decomposition approach, consider the following population, whose dynamics and population structure were censused across two separate scenarios (x and $x+1$):

$$A_x = \begin{bmatrix} 0.2 & 2.0 \\ 0.3 & 0.5 \end{bmatrix} \ \& \ n_x = \begin{bmatrix} 0.6 \\ 0.4 \end{bmatrix} \quad A_{x+1} = \begin{bmatrix} 0.2 & 1.5 \\ 0.7 & 0.5 \end{bmatrix} \ \& \ n_{x+1} = \begin{bmatrix} 0.2 \\ 0.8 \end{bmatrix}. \quad (3)$$

Accordingly, we possess a sequence of two A and n products for this population, with $var(\lambda_{transient})$ across these replicates estimated to be 0.0576 ($\lambda_{transient}$: $x = 1.3$ & $x+1 = 1.78$). We then compute the mean MPM (\underline{A}) and population vector (\underline{n}) (arithmetic element-by-element means, in each case) represented by our observed A and n products:

$$\underline{A} = \begin{bmatrix} 0.2 & 1.75 \\ 0.5 & 0.5 \end{bmatrix} \ \& \ \underline{n} = \begin{bmatrix} 0.4 \\ 0.6 \end{bmatrix}. \quad (4)$$

Next, holding all matrix elements and/or n at their mean condition, we systematically replace the value of each matrix element, a_{ij} , and the population vector, n , with their observed conditions across our sequence of A and n products. We then recalculate the $\lambda_{transient}$ estimates

associated with these perturbed A and n products (A^{pert} and n^{pert}) before recalculating the variance across these modified $\lambda_{transient}$ estimates. In our hypothetical population, matrix elements $a_{2,1}$ and $a_{1,2}$, and n exhibit different conditions over time. Subsequently, we are only interested in how the changes in these three entities contribute to $var(\lambda_{transient})$. To quantify the main (first-order) effects of each entity, we isolate the independent changes observed in each entity whilst holding all other values at their mean condition (Eqs. 5-7), before then computing the variance associated with only these imposed modifications.

$$A_x^{pert} = \begin{bmatrix} 0.2 & 2.0 \\ 0.5 & 0.5 \end{bmatrix} \ \& \ \underline{n} = \begin{bmatrix} 0.4 \\ 0.6 \end{bmatrix} \quad A_{x+1}^{pert} = \begin{bmatrix} 0.2 & 1.5 \\ 0.5 & 0.5 \end{bmatrix} \ \& \ \underline{n} = \begin{bmatrix} 0.4 \\ 0.6 \end{bmatrix} \quad (5)$$

$$A_x^{pert} = \begin{bmatrix} 0.2 & 1.75 \\ 0.3 & 0.5 \end{bmatrix} \ \& \ \underline{n} = \begin{bmatrix} 0.4 \\ 0.6 \end{bmatrix} \quad A_{x+1}^{pert} = \begin{bmatrix} 0.2 & 1.75 \\ 0.7 & 0.5 \end{bmatrix} \ \& \ \underline{n} = \begin{bmatrix} 0.4 \\ 0.6 \end{bmatrix} \quad (6)$$

$$\underline{A} = \begin{bmatrix} 0.2 & 1.75 \\ 0.5 & 0.5 \end{bmatrix} \ \& \ n_x^{pert} = \begin{bmatrix} 0.6 \\ 0.4 \end{bmatrix} \quad \underline{A} = \begin{bmatrix} 0.2 & 1.75 \\ 0.5 & 0.5 \end{bmatrix} \ \& \ n_x^{pert} = \begin{bmatrix} 0.2 \\ 0.8 \end{bmatrix} \quad (7)$$

In equation 5, we focus on the effect of changes in matrix element $a_{2,1}$. Here, the $\lambda_{transient}$ values for the two scenarios presented are 1.78 & 1.48, respectively, such that the recorded changes in element $a_{2,1}$ contribute to an observed $var(\lambda_{transient})$ of 0.0225. Meanwhile, the recorded changes in element $a_{1,2}$ (equation 6) contribute an observed $var(\lambda_{transient})$ of 0.0064 ($\lambda_{transient}$: $x = 1.55$ & $x+1 = 1.71$), while changes in n (equation 7) contribute an observed $var(\lambda_{transient})$ of 0.0961 ($\lambda_{transient}$: $x = 1.32$ & $x+1 = 1.94$). Repeating the above steps for different combinations of the focal entities then allows for quantifying the contributions of interactions between multiple changing entities to $var(\lambda_{transient})$ (72); for instance, the effects of simultaneous changes to the population vector (n) and element $a_{1,2}$ of matrix A on $var(\lambda_{transient})$ constitutes an interaction effect.

After identifying the contributions of different matrix elements, to compare contributions from population structure to those of the vital rates of survival (σ), development (γ), and fecundity

(ϕ) (HI), we then determined which vital rate each matrix element corresponded to. For fecundity, this step simply involved summing the contributions of matrix elements only detailing per-capita fecundity rates, which we identified using each population's reproductive sub-matrix (F ; Eq. 1). For survival and development, we first determined the matrix elements corresponding with stasis (*i.e.*, individuals remaining within a life-stage from t to $t+1$), stage progression (*i.e.*, growth/development), and stage retrogression (*i.e.* shrinkage/de-development). Next, for matrix elements describing progression and retrogression, we determined the proportion of each element associated with survival (P_σ) and development (P_γ ; growth and shrinkage pooled). Dividing each matrix element by its corresponding stage-specific survival, the column-sum of the population's survival sub-matrix (U), we extracted the population's survival-independent progression/retrogression matrix (U'). We then quantified survival contributions to $var(\lambda_{transient})$ by totalling the contributions of matrix elements associated with stasis and the contributions of matrix elements associated with development weighted by P_σ . For the contributions of development to $var(\lambda_{transient})$, we summed the contributions of matrix elements associated with development weighted by P_γ . Finally, we summed the contributions to $var(\lambda_{transient})$ arising from second-order interactions between different entities, categorising only whether they involved interactions between just vital rates or between vital rates and population structure.

Finally, we standardised our contribution estimates to ensure comparability across populations displaying a range of magnitudes in $var(\lambda_{transient})$. Specifically, we translated our estimates of contribution from survival (σ), development (γ), fecundity (ϕ), population structure (n), and their interaction effects ($\sigma:\gamma:\phi:n$ & $\sigma:\gamma:\phi$) to observed variance in $\lambda_{transient}$ into proportional effects. For each population set, combining the absolute values of contributions of

survival, development, fecundity, and population structure $var(\lambda_{transient})$ provided us with a total contribution estimate (C_{total}) whilst accommodating for both negative and positive contribution patterns. Subsequently, we separately quantified the directional proportional effects of survival, development, fecundity, population structure, and their interaction effects (C_i) as C_i/C_{total} . We then calculated a contribution ratio for each population comparison (C_{ratio}), equal to the summed contributions of population structure (n) and the interaction between population structure and vital rates ($\sigma:\gamma:\phi:n$) divided by the summed contributions of survival (σ), development (γ), fecundity (ϕ), and their interactions ($\sigma:\gamma:\phi$). Accordingly, a ratio of one implies a balanced contribution of population vital rates vs. population structure, with higher estimates inferring a greater influence from the population structure. Finally, following our computation of these proportional effects, to mitigate for skewed distributions we omitted all estimates outside the 90th percentiles of our $var(\lambda_{transient})$ and contribution variables.

Quantifying life-history variation

Testing our hypotheses required quantifying life-history trait variation across our population comparison sample. For each population comparison, we estimated the mean MPM (*i.e.*, the element-by-element arithmetic mean) from the associated set of MPMs describing the focal species/population dynamics across different time intervals, locations or experimental treatments. Next, with life-history trait estimates obtained from MPMs sensitive to matrix dimensionality (90), we collapsed each mean MPM into a 3×3 matrix to ensure comparability across populations and species. Using the matrix collapsing function *mpm_collapse* of the *Rcompadre* R package (91), we condensed together any life stages beyond the second life stage into a single terminal stage while keeping the first two stages unaltered. This approach retains the original eigen-

structure of the original MPM (92); thereby minimising the impact on any subsequent measures derived from the collapsed MPMs. Using age-from-stage approaches (34, 93) on our collapsed mean MPMs, we then computed life-history traits routinely used to describe how individuals invest differentially into survival, development, and reproduction (24, 28).

Whilst many life-history traits can be obtained from MPMs (91), our hypotheses focused on selected trait measures (Table 1). To evaluate how contributions from changes in population structure vary across the fast-slow continuum of life-history strategies (*H2*), we focused on the selected traits of generation time (T), age at reproductive maturity (L_{α}), mean life expectancy (η_e), and reproductive window ($L_{\alpha-\omega}$). Our rationale for selecting these metrics is that they all possess the same units (time) and thus allow for cross-species comparisons while avoiding artifactual patterns that can emerge from mixing life-history traits with different units (36). Meanwhile, to explore the relationship between contributions from variation in population structure and patterns in population life cycle survival profile ($\Delta\sigma$) (*H3*), we calculated the slope coefficient of a linear regression model fitted to the sequence of mean stage-specific survival probabilities obtained for each population comparison (*e.g.*, slope of survival across life stages; see Appendix S3 for details on the validity of this approach). Consequently, positive values reflect an increasing likelihood of survival as individuals progress through their life cycle. To ensure this assessment accounted for the effect of lifetime survival (*i.e.*, fast *vs.* slow) on the relationship between sources of variance in demographic performance and life cycle survival profile, we also computed a measure of mean lifetime survival ($\underline{\sigma}$) associated with each population comparison. Mean lifetime survival was equal to the grand mean of mean stage-specific survival probabilities obtained for each population comparison. Finally, to assess how contributions from changes in population structure align with reproductive trait variation (*H4*), we used the measure of degree of iteroparity (S). This

dimensionless measure of entropy (94) reflects the reproductive parity continuum, delineating species based on the timing and frequency of their reproductive investments (27). After obtaining estimates for each selected life-history trait, we omitted all estimates outside the 90th percentiles of each variable (estimates which were later replaced using phylogenetic imputation, see *Establishing phylogenetic relationships*).

Establishing phylogenetic relationships

Comparative assessments of demographic and life-history trait variation need to consider non-random patterns of variance-covariance across populations based on shared phylogenetic ancestry (95). To further compare patterns in the contribution of population structure to $var(\lambda_{transient})$ vs. those of the vital rates of survival, development, and fecundity (HI), we sought to test the strength of phylogenetic signal across our contributonal effect variables. Thus, we computed the phylogenetic distances between the 56 species represented across our LTRE comparison sample. We sourced unique identifier codes for each species from the Open Tree of Life (OTL) (taxonomy version 3.6; (96)) to construct a species-level phylogenetic subtree using the *phytools* (97) and *ape* R packages (98). We then computed branch lengths for our species-level phylogenetic subtree using Grafen's computation method (99), ensured our subtree contained a single common ancestor, and removed any polytomies (≥ 3 species emerging from a single origin).

Next, we used our species-level phylogenetic subtree to impute any missing or omitted estimates across our vital rate and life-history measures, enabling us to retain details from all populations across our sample. We used the *phylopars* function from the *Rphylopars* package

(100) to impute missing values (assuming a Brownian Motion evolutionary model) across our demographic and life-history measures. 2.7% of our final data sample was generated using phylogenetic imputation, and we ran sensitivity tests confirming that our results are insensitive to this imputation (Appendix S5).

Finally, to further assess how the contribution of population structure to $var(\lambda_{transient})$ compares to those of the vital rates of survival, development, and fecundity (HI), we tested the strength of phylogenetic signal across our contributory effect variables. After using our species-level subtree to impute missing data, we expanded subtree branch tips for species replicates across our population sample (48 of 56 species), generating a population-level phylogenetic subtree. Carrying out this branch expansion involved rooting replicate instances of the same species onto the species' corresponding branch tip in our species-level subtree. When rooting replicate entries, we ensured each replicate was separated by branch lengths of infinitesimal distances ($\epsilon = 0.0000001$ units). This step ensured that populations of the same species were considered very closely related, preventing the introduction of polytomies. Recent work using the same demographic data has demonstrated that the order in which population replicates are added to population-level phylogenetic trees does not affect transient metrics (101). The exact population-level subtree we used in this study is available on Zenodo (<https://doi.org/10.5281/zenodo.15600668>). Using this population-level subtree, we estimated phylogenetic signal using Pagel's λ (33) via the *caper* package (102). Pagel's λ ranges between 0 to 1 depending on whether attribute estimates vary randomly (Pagel's $\lambda \approx 0$) or are more similar across more closely related species (Pagel's $\lambda \approx 1$) (35).

Linking *t*LTRE contributions and life-histories

To test whether the contributions to $\text{var}(\lambda_{\text{transient}})$ from changes in population structure are greater in species with faster life-histories (*H2*), we carried out a phylogenetically weighted partial least squares analysis (pPLS). Specifically, we used pPLS to assess the relationship between our proportional effects variables and the life-history traits of generation time (T), age at reproductive maturity (L_a), mean life expectancy (η_e), and reproductive window ($L_{a-\omega}$). Prior to this pPLS we transformed our life-history trait estimates to ensure normal distributions, applying log-transformations to our estimates of generation time (T), age at reproductive maturity (L_a) (both $\log(x)$), and reproductive window ($L_{a-\omega}$) ($\log(x+5)$), and an inverted power-transformation to our mean life expectancy estimates (η_e) ($-x^{-0.275}$). To account for the influence of body size on the development of life-history traits (24), we then scaled our transformed life-history variables according to a measure of adult plant height assigned to each species (27). We attributed each plant species a height class corresponding with their Raunkiær growth form classification (40), which catalogues plant species according to the extent to which their apical meristems (*i.e.*, renewal buds) recede to ground level during periods of unfavourable growing conditions. Synthesising details of the Raunkiær growth form classification of different plant species from the TRY database (103) and Plants of the World Online (104), we grouped our plant species into one of seven Raunkiær growth form classes that subsequently defined their associated height (m): (i) geophytes [-0.25m]; (ii) hemicryptophytes [0m]; (iii) chamaephytes [0.25m]; (iv) epiphytes [4m]; (v) nanophanerophytes [2m]; (vi) mesophanerophytes [8m]; and (vii) megaphanerophytes [30m] (See (27) for further details). To scale our life-history traits according to size, we retained the residuals from linear models of the relationship between each life-history trait and the height class estimates attributed to each species. Next, following (105), we transformed all our pPLS variables

according to their phylogenetic weighting, using a phylogenetic variance-covariance matrix generated from our population-level phylogenetic subtree. We then implemented our pPLS using the *plsdepot* R package (106) before imposing a varimax rotation upon the multivariate parameter space using the *GPArotation* package (107). We performed varimax rotation to ensure that our life-history variables were better aligned with the main axes of the parameter space (108); a common approach used to improve the interpretation of multivariate analysis outputs (109).

To test (*H3*) whether the importance of contributions from variation in population structure to $var(\lambda_{transient})$ correlates with the life cycle survival profiles of natural populations, we used phylogenetically-corrected Bayesian multilevel modelling. To do so, and following the approach described above, we first scaled our measure of life cycle survival profiles ($\Delta\sigma$) according to the plant size estimates associated with assigned Raunkiær growth form classifications. Using the *brms* R package (110), we then modelled the relationship between our log-transformed C_{ratio} ($\log(x+0.01)$) and the additive effects of our scaled $\Delta\sigma$ and $\underline{\sigma}$ estimates using a model fitted with default priors. When extracting outputs from this model we set $\underline{\sigma}$ to its median value observed across our sample, allowing us to eliminate the time-dimension from each survival profile, to focus instead on the shape of survival through the life-cycle (111). To test (*H5*) how growth form mediates any observed patterns between sources of variance in $\lambda_{transient}$ and their life-history traits, this model also included assigned Raunkiær growth form classification as a fixed effect variable. Note, due to limited sample sizes in our geophyte ($n = 4$), nanophanerophyte (28), mesophanerophyte (28), and megaphanerophyte (6) Rankiær classes, we pooled geophytes with hemicryptophytes and collapsed nano-, meso-, and megaphanerophytes together as phanerophytes. Thus, growth form was a fixed effect variable comprised of four classes (chamaephytes [$n = 45$], hemicryptophytes [144], epiphytes [17], and phanerophytes [62]). Our *brms* model also included

a grouping term to account for variance-covariance among our 268 populations across 56 species according to the branch lengths outlined in our phylogenetic subtree. To ensure we selected the most appropriate model, we initially applied both linear and polynomial functions of explanatory variables to our model before determining a linear function to be the best model fit using leave-one-out comparison (LOO) (LOO weights: Linear = 0.999; Polynomial < 0.001). We ran our model over four Markov Chains, each consisting of 3,000 iterations, extracting the predicted effects of $\Delta\sigma$ on C_{ratio} across growth forms, when $\underline{\sigma}$ is held at its median value.

To test (*H4*) whether the relative importance of contributions from population structure to $var(\lambda_{transient})$ decreases with increasing iteroparity, we again used a phylogenetically-corrected Bayesian multilevel model fitted using default priors. Specifically, we modelled the relationship between our log-transformed C_{ratio} and power-transformed ($x^{0.35}$) degree of iteroparity (S) variables. As before, we scaled degree of iteroparity according to plant size and included a phylogenetic grouping term within our model to account for variance-covariance between populations/species. We also, again, included Raunkiaer growth form classification as a fixed effect variable, to test (*H5*) how growth form mediates any observed patterns between sources of variance in $\lambda_{transient}$ and their life-history traits. We preliminarily fitted both linear and polynomial functions to assess model fit before using LOO comparisons to identify that the linear model provided the best representation of our data (LOO weights: Linear = 0.718; Polynomial = 0.282). We once again implemented our model over four Markov Chains (3,000 iterations each) and extracted the predicted effects of S on C_{ratio} .

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References

1. W. F. Morris, D. F. Doak, *Quantitative conservation biology: Theory and practice of population viability analysis* (Sinauer Associates, Inc., 2002).
2. A. Hastings, *et al.*, Transient phenomena in ecology. *Science* **361**, 1–9 (2018).
3. P. Capdevila, I. Stott, M. Beger, R. Salguero-Gómez, Towards a comparative framework of demographic resilience. *Trends Ecol. Evol.* **35**, 776–786 (2020).
4. D. N. Koons, D. T. Iles, M. Schaub, H. Caswell, A life-history perspective on the demographic drivers of structured population dynamics in changing environments. *Ecol. Lett.* **19**, 1023–1031 (2016).
5. A. A. Maldonado-Chaparro, D. T. Blumstein, K. B. Armitage, D. Z. Childs, Transient LTRE analysis reveals the demographic and trait-mediated processes that buffer population growth. *Ecol. Lett.* **21**, 1693–1703 (2018).
6. F. S. Dobson, *et al.*, The Demographic Basis of Population Growth: A 32-Year Transient Life Table Response Experiment. *Ecol. Lett.* **27**, e14512 (2024).

- 704 7. S. R. Hoy, *et al.*, Fluctuations in age structure and their variable influence on population
705 growth. *Funct. Ecol.* **34**, 203–216 (2020).
- 706 8. B. Sommer, *et al.*, Decadal demographic shifts and size-dependent disturbance responses of
707 corals in a subtropical warming hotspot. *Sci. Rep.* **14** (2024).
- 708 9. V. Radchuk, C. Turlure, N. Schtickzelle, Each life stage matters: the importance of
709 assessing the response to climate change over the complete life cycle in butterflies. *J. Anim.*
710 *Ecol.* **82**, 275–285 (2013).
- 711 10. T. Coulson, *et al.*, Age, Sex, Density, Winter Weather, and Population Crashes in Soay
712 Sheep. *Science* **292**, 1528–1531 (2001).
- 713 11. D. N. Koons, T. W. Arnold, M. Schaub, Understanding the demographic drivers of realized
714 population growth rates. *Ecol. Appl.* **27**, 2102–2115 (2017).
- 715 12. S. Jenouvrier, Impacts of climate change on avian populations. *Glob. Change Biol.* **19**,
716 2036–2057 (2013).
- 717 13. E. E. Crone, *et al.*, How do plant ecologists use matrix population models? *Ecol. Lett.* **14**,
718 1–8 (2011).
- 719 14. D. Gibson, T. W. Arnold, F. E. Buderman, D. N. Koons, Explaining the divergence of
720 population trajectories for two interacting waterfowl species. *Ecol. Monogr.* **95**, e1642
721 (2025).
- 722 15. A. Hastings, Transients: The key to long-term ecological understanding? *Trends Ecol. Evol.*
723 **19**, 39–45 (2004).

- 724 16. T. H. G. Ezard, *et al.*, Matrix models for a changeable world: The importance of transient
725 dynamics in population management. *J. Appl. Ecol.* **47**, 515–523 (2010).
- 726 17. M. Christian, W. C. Oosthuizen, M. N. Bester, P. J. N. de Bruyn, Robustly estimating the
727 demographic contribution of immigration: Simulation, sensitivity analysis and seals. *J.*
728 *Anim. Ecol.* **93**, 632–645 (2024).
- 729 18. J. Summers, E. J. Cosgrove, R. Bowman, J. W. Fitzpatrick, N. Chen, Density dependence
730 maintains long-term stability despite increased isolation and inbreeding in the Florida
731 Scrub-Jay. *Ecol. Lett.* **27**, e14483 (2024).
- 732 19. C. R. Nater, *et al.*, Spatial consistency in drivers of population dynamics of a declining
733 migratory bird. *J. Anim. Ecol.* **92**, 97–111 (2023).
- 734 20. S. Tenan, D. Becker, D. Tolkmitt, M. Schaub, Decomposing fecundity and evaluating
735 demographic influence of multiple broods in a migratory bird. *J. Anim. Ecol.* **90**,
736 1071–1084 (2021).
- 737 21. K. Layton-Matthews, *et al.*, Environmental change reduces body condition, but not
738 population growth, in a high-arctic herbivore. *Ecol. Lett.* **24**, 227–238 (2021).
- 739 22. R. Salguero-Gómez, *et al.*, The COMPADRE Plant Matrix Database: An open online
740 repository for plant demography. *J. Ecol.* **103**, 202–218 (2015).
- 741 23. J. L. McDonald, I. Stott, S. Townley, D. J. Hodgson, Transients drive the demographic
742 dynamics of plant populations in variable environments. *J. Ecol.* **104**, 306–314 (2016).
- 743 24. S. C. Stearns, *The Evolution of Life Histories* (Oxford University Press, 1992).

- 744 25. A. Compagnoni, *et al.*, Herbaceous perennial plants with short generation time have
745 stronger responses to climate anomalies than those with longer generation time. *Nat.*
746 *Commun.* **12**, 1–8 (2021).
- 747 26. Leishman, M. R., Wright, I. J., Moles, A. T., Westoby, M., “The Evolutionary Ecology of
748 Seed Size” in *Seeds: The Ecology of Regeneration in Plant Communities*, 2nd Edition,
749 (CAB International, 2000).
- 750 27. R. Salguero-Gómez, *et al.*, Fast–slow continuum and reproductive strategies structure plant
751 life-history variation worldwide. *Proc. Natl. Acad. Sci.* **113**, 220–235 (2016).
- 752 28. S. C. Stearns, Trade-Offs in Life-History Evolution. *Funct. Ecol.* **3**, 259–268 (1989).
- 753 29. J.-P. Desforges, *et al.*, Environment and physiology shape Arctic ungulate population
754 dynamics. *Glob. Change Biol.* **27**, 1755–1771 (2021).
- 755 30. C. Stuligross, N. M. Williams, Past insecticide exposure reduces bee reproduction and
756 population growth rate. *Proc. Natl. Acad. Sci.* **118**, e2109909118 (2021).
- 757 31. M. C. Allen, M. Clinchy, L. Y. Zannette, Fear of predators in free-living wildlife reduces
758 population growth over generations. *Proc. Natl. Acad. Sci.* **119**, e2112404119 (2022).
- 759 32. M. Gamelon, *et al.*, Efficient use of harvest data: a size-class-structured integrated
760 population model for exploited populations. *Ecography* **44**, 1296–1310 (2021).
- 761 33. M. Pagel, Inferring the historical patterns of biological evolution. *Nature* **401**, 877–884
762 (1999).
- 763 34. H. Caswell, *Matrix population models: Construction, analysis and interpretation* (Oxford

- 764 University Press Inc, 2001).
- 765 35. R. P. Freckleton, P. H. Harvey, M. Pagel, Phylogenetic analysis and comparative data: A
766 test and review of evidence. *Am. Nat.* **160**, 712–726 (2002).
- 767 36. I. Stott, *et al.*, Life histories are not just fast or slow. *Trends Ecol. Evol.* **39**, 830–840
768 (2024).
- 769 37. T. Yamada, *et al.*, Strong habitat preference of a tropical rain forest tree does not imply
770 large differences in population dynamics across habitats. *J. Ecol.* **95**, 332–342 (2007).
- 771 38. A. Jonsson, B. Ebenman, Are Certain Life Histories Particularly Prone to Local Extinction?
772 *J. Theor. Biol.* **209**, 455–463 (2001).
- 773 39. M. Bogdziewicz, *et al.*, Evolutionary ecology of masting: mechanisms, models, and climate
774 change. *Trends Ecol. Evol.* **39**, 851–862 (2024).
- 775 40. C. Raunkiaer, *The life forms of plants and statistical plant geography; being the collected*
776 *papers of C. Raunkiaer.* (1934).
- 777 41. R. Salguero-Gómez, B. B. Casper, Keeping plant shrinkage in the demographic loop. *J.*
778 *Ecol.* **98**, 312–323 (2010).
- 779 42. L. G. Shoemaker, *et al.*, Integrating the underlying structure of stochasticity into
780 community ecology. *Ecology* **101**, e02922 (2020).
- 781 43. M. J. Emslie, *et al.*, Increasing disturbance frequency undermines coral reef recovery. *Ecol.*
782 *Monogr.* **94**, e1619 (2024).

- 783 44. M. M. Ellis, E. E. Crone, The role of transient dynamics in stochastic population growth for
784 nine perennial plants. *Ecology* **94**, 1681–1686 (2013).
- 785 45. M. Gamelon, *et al.*, Making use of harvest information to examine alternative management
786 scenarios: a body weight-structured model for wild boar. *J. Appl. Ecol.* **49**, 833–841 (2012).
- 787 46. R. A. F. De Lima, *et al.*, Comprehensive conservation assessments reveal high extinction
788 risks across Atlantic Forest trees. *Science* **383**, 219–225 (2024).
- 789 47. A. Dietzel, M. Bode, S. R. Connolly, T. P. Hughes, The population sizes and global
790 extinction risk of reef-building coral species at biogeographic scales. *Nat. Ecol. Evol.* **5**,
791 663–669 (2021).
- 792 48. G. M. Mace, *et al.*, Quantification of extinction risk: IUCN’s system for classifying
793 threatened species. *Conserv. Biol.* **22**, 1424–1442 (2008).
- 794 49. IUCN Standards and Petitions Committee, Guidelines for Using the IUCN Red List
795 Categories and Criteria. Version 16. *Prep. Stand. Petitions Comm.* (2024).
- 796 50. T. Coulson, G. M. Mace, E. Hudson, H. Possingham, The use and abuse of population
797 viability analysis. *Trends Ecol. Evol.* **16**, 219–221 (2001).
- 798 51. I. Stott, M. Franco, D. Carslake, S. Townley, D. Hodgson, Boom or bust? A comparative
799 analysis of transient population dynamics in plants. *J. Ecol.* **98**, 302–311 (2010).
- 800 52. E. R. Buhle, *et al.*, Using integrated population models to evaluate fishery and
801 environmental impacts on Pacific salmon viability.
- 802 53. S. Hamel, *et al.*, Fitness costs of reproduction depend on life speed: empirical evidence

- 803 from mammalian populations. *Ecol. Lett.* **13**, 915–935 (2010).
- 804 54. J. Gaillard, *et al.*, Generation Time: A reliable metric to measure life-history variation
805 among mammalian populations. *Am. Nat.* **166**, 119–123 (2005).
- 806 55. J. Staerk, *et al.*, Performance of generation time approximations for extinction risk
807 assessments. *J. Appl. Ecol.* **56**, 1436–1446 (2019).
- 808 56. Pearson, R. G., *et al.*, Life history and spatial traits predict extinction risk due to climate
809 change. *Nat. Clim. Change* **4**, 217–221 (2014).
- 810 57. S. D. Tuljapurkar, Population dynamics in variable environments. III. Evolutionary
811 dynamics of *r*-selection. *Theor. Popul. Biol.* **21**, 141–165 (1982).
- 812 58. M. Paniw, A. Ozgul, R. Salguero-Gómez, Interactive life-history traits predict sensitivity of
813 plants and animals to temporal autocorrelation. *Ecol. Lett.* **21**, 275–286 (2018).
- 814 59. Le Coeur, C., Yoccoz, N. G., Salguero-Gómez R., Vindenes, Y., Life history adaptations to
815 fluctuating environments: Combined effects of demographic buffering and lability. *Ecol.*
816 *Lett.* **25**, 2107–2119 (2022).
- 817 60. W. A. Hoffmann, Fire and population dynamics of woody plants in a neotropical savanna:
818 Matrix model projections. *Ecology* **80**, 1354–1369 (1999).
- 819 61. K. E. Frankiewicz, *et al.*, Derived woodiness and annual habit evolved in African
820 umbellifers as alternative solutions for coping with drought. *BMC Plant Biol.* **21**, 383
821 (2021).
- 822 62. M. Huston, T. Smith, Plant Succession: Life History and Competition. *Am. Nat.* **130**,

- 823 168–198 (1987).
- 824 63. D. Tilman, Constraints and Tradeoffs: Toward a Predictive Theory of Competition and
825 Succession. *Oikos* **58**, 3–15 (1990).
- 826 64. T. C. Whitmore, Canopy Gaps and the Two Major Groups of Forest Trees. *Ecology* **70**,
827 536–538 (1989).
- 828 65. G. Römer, J. P. Dahlgren, R. Salguero-Gómez, I. M. Stott, O. R. Jones, Plant demographic
829 knowledge is biased towards short-term studies of temperate-region herbaceous perennials.
830 *Oikos* **2024**, e10250 (2024).
- 831 66. E. M. Fischer, U. Beyerle, R. Knutti, Robust spatially aggregated projections of climate
832 extremes. *Nat. Clim. Change* **3**, 1033–1038 (2013).
- 833 67. S. L. Lewis, D. P. Edwards, D. Galbraith, Increasing human dominance of tropical forests.
834 *Science* **349**, 827–832 (2015).
- 835 68. D. L. Perret, M. E. K. Evans, D. F. Sax, A species’ response to spatial climatic variation
836 does not predict its response to climate change. *Proc. Natl. Acad. Sci.* **121**, e2304404120
837 (2024).
- 838 69. J. Cant, *et al.*, Current estimates of population resilience do not predict resilience to
839 directional environmental shifts. *BioRxiv* (2025).
- 840 70. G. Bell, Evolutionary rescue and the limits of adaptation. *Philos. Trans. R. Soc. B Biol. Sci.*
841 **368**, 1–6 (2013).
- 842 71. R. Salguero-Gómez, *et al.*, COMADRE: A global data base of animal demography. *J.*

843 *Anim. Ecol.* **85**, 371–384 (2016).

844 72. C. M. Hernández, S. P. Ellner, P. B. Adler, G. Hooker, R. E. Snyder, An exact version of
845 Life Table Response Experiment analysis, and the R package exactLTRE. *Methods Ecol.*
846 *Evol.* **14**, 939–951 (2023).

847 73. S. J. L. Gascoigne, *et al.*, A standard protocol to report discrete stage-structured
848 demographic information. *Methods Ecol. Evol.* **14**, 2065–2083 (2023).

849 74. E. White, *et al.*, Resilience of a long-lived mammal: time and demographic structure matter.
850 [Preprint] (2025). Available at: [https://www.authorea.com/users/879023/articles/1258323-](https://www.authorea.com/users/879023/articles/1258323-resilience-of-a-long-lived-mammal-time-and-demographic-structure-matter?commit=402f9b0e343cea73d8dc9f9303885c8628ddb418)
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853 2025].

854 75. Oppel, S., *et al.*, Assessing population viability while accounting for demographic and
855 environmental uncertainty. *Ecology* **95**, 1809–1818 (2014).

856 76. P. J. Olsoy, *et al.*, Demography with drones: detecting growth and survival of shrubs with
857 unoccupied aerial systems. *Restor. Ecol.* **32**, e14106 (2024).

858 77. A. Rosen, *et al.*, Modelling forest dynamics using integral projection models (IPMs) and
859 repeat LiDAR. [Preprint] (2025). Available at:
860 <http://biorxiv.org/lookup/doi/10.1101/2025.01.06.631514> [Accessed 4 June 2025].

861 78. T. C. Stone, K. J. Davis, Using unmanned aerial vehicles to estimate body volume at scale
862 for ecological monitoring. *Methods Ecol. Evol.* **16**, 317–331 (2025).

79. R Core Team, R: A language and environment for statistical computing. (2022). Deposited 2022.
80. I. Stott, S. Townley, D. Carslake, D. J. Hodgson, On reducibility and ergodicity of population projection matrix models. *Methods Ecol. Evol.* **1**, 242–252 (2010).
81. D. N. Koons, J. B. Grand, B. Zinner, R. F. Rockwell, Transient population dynamics: Relations to life history and initial population state. *Ecol. Model.* **185**, 283–297 (2005).
82. T. Ticktin, R. Ganesan, M. Paramesha, S. Setty, Disentangling the effects of multiple anthropogenic drivers on the decline of two tropical dry forest trees. *J. Appl. Ecol.* **49**, 774–784 (2012).
83. S. Pisanu, E. Farris, R. Filigheddu, M. B. García, Demographic effects of large, introduced herbivores on a long-lived endemic plant. *Plant Ecol.* **213**, 1543–1553 (2012).
84. S. K. Ghimire, O. Gimenez, R. Pradel, D. McKey, Y. Aumeeruddy-Thomas, Demographic variation and population viability in a threatened Himalayan medicinal and aromatic herb *Nardostachys grandiflora*: Matrix modelling of harvesting effects in two contrasting habitats. *J. Appl. Ecol.* **45**, 41–51 (2008).
85. B. E. Kendall, *et al.*, Persistent problems in the construction of matrix population models. *Ecol. Model.* **406**, 33–43 (2019).
86. H. Caswell, Analysis of Life Table Response Experiments I. Decomposition of effects on population growth rate. *Ecol. Model.* **46**, 221–237 (1989).
87. S. P. Ellner, R. E. Snyder, P. B. Adler, G. Hooker, An expanded modern coexistence theory

883 for empirical applications. *Ecol. Lett.* **22**, 3–18 (2019).

884 88. G. Hooker, Generalized Functional ANOVA Diagnostics for High-Dimensional Functions
885 of Dependent Variables. *J. Comput. Graph. Stat.* **16**, 709–732 (2007).

886 89. C. Hernandez, exactLTRE: An Exact Method for Life Table Response Experiment (LTRE)
887 Analysis. (2022). Deposited 17 November 2022.

888 90. N. J. Enright, M. Franco, J. Silvertown, Comparing plant life histories using elasticity
889 analysis: the importance of life span and the number of life-cycle stages. *Oecologia* **104**,
890 79–84 (1995).

891 91. O. R. Jones, *et al.*, Rcompadre and Rage - two R packages to facilitate the use of the
892 COMPADRE and COMADRE databases and calculation of life history traits from matrix
893 population models. *Methods Ecol. Evol.* **13**, 770–781 (2022).

894 92. R. Salguero-Gómez, J. B. Plotkin, Matrix dimensions bias demographic inferences:
895 Implications for comparative plant demography. *Am. Nat.* **176**, 710–722 (2010).

896 93. M. E. Cochran, S. Ellner, Simple Methods for Calculating Age-Based Life History
897 Parameters for Stage-Structured Populations. *Ecol. Monogr.* **62**, 345–364 (1992).

898 94. Demetrius, L., Entropy and the Life Table. *Naturwissenschaften* **65**, 435–436 (1978).

899 95. Capdevila, P., Walker, T. W. N., Schrod, F., Rodriguez Caro, R. C., Salguero-Gomez, R.,
900 Global patterns of plant form and function are strongly determined by evolutionary
901 relationships. *BioRxiv* (2023).

902 96. OpenTreeOfLife, *et al.*, Open Tree of Life Synthetic Tree. (2019).

- 903 97. L. J. Revell, phytools 2.0: an updated R ecosystem for phylogenetic comparative methods
904 (and other things). *PeerJ* **12**, e16505 (2024).
- 905 98. E. Paradis, K. Schliep, ape 5.0: an environment for modern phylogenetics and evolutionary
906 analyses in R. *Bioinformatics* **35** (2018).
- 907 99. A. Grafen, The phylogenetic regression. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.* **326**, 119-
908 157. (1989).
- 909 100. E. W. Goolsby, J. Bruggeman, C. Ané, Rphylopars: fast multivariate phylogenetic
910 comparative methods for missing data and within-species variation. *Methods Ecol. Evol.* **8**,
911 22–27 (2017).
- 912 101. T. Merrien, K. Davis, M. D. Marco, P. Capdevila, R. Salguero-Gómez, Human pressures
913 filter out the less resilient demographic strategies in natural populations of plants and
914 animals worldwide. *BioRxiv* 1–36 (2022). <https://doi.org/10.1101/2021.09.29.462372>.
- 915 102. Orme, D., *et al.*, caper: Comparative Analyses of Phylogenetics and Evolution in R. (2023).
916 Deposited 2023.
- 917 103. J. Kattge, *et al.*, TRY – a global database of plant traits. *Glob. Change Biol.* **17**, 2905–2935
918 (2011).
- 919 104. POWO, Plants of the World Online. Facilitated by the Royal Botanic Gardens, Kew.
920 Published on the Internet. (2025). Available at: <https://powo.science.kew.org/>.
- 921 105. D. C. Adams, R. N. Felice, Assessing trait covariation and morphological integration on
922 phylogenies using evolutionary covariance matrices. *PLoS ONE* **9** (2014).

- 923 106. G. Sanchez, plsdepot: Partial least squares (PLS) data analysis methods. (2012). Deposited
924 2012.
- 925 107. C. A. Bernaards, R. I. Jennrich, Gradient Projection Algorithms and Software for Arbitrary
926 Rotation Criteria in Factor Analysis. *Educ. Psychol. Meas.* **65**, 676–696 (2005).
- 927 108. I. T. Jolliffe, J. Cadima, Principal component analysis: a review and recent developments.
928 *Philos. Trans. R. Soc. Math. Phys. Eng. Sci.* **374**, 20150202 (2016).
- 929 109. C. P. Carmona, *et al.*, Fine-root traits in the global spectrum of plant form and function.
930 *Nature* **597**, 683–687 (2021).
- 931 110. Bürkner, P., brms: An R Package for Bayesian Multilevel Models Using Stan. *J. Stat.*
932 *Softw.* **80**, 1–28 (2017).
- 933 111. T. F. Wrycza, T. I. Missov, A. Baudisch, Quantifying the Shape of Aging. *PLOS ONE* **10**,
934 e0119163 (2015).
- 935

937 Figure 1

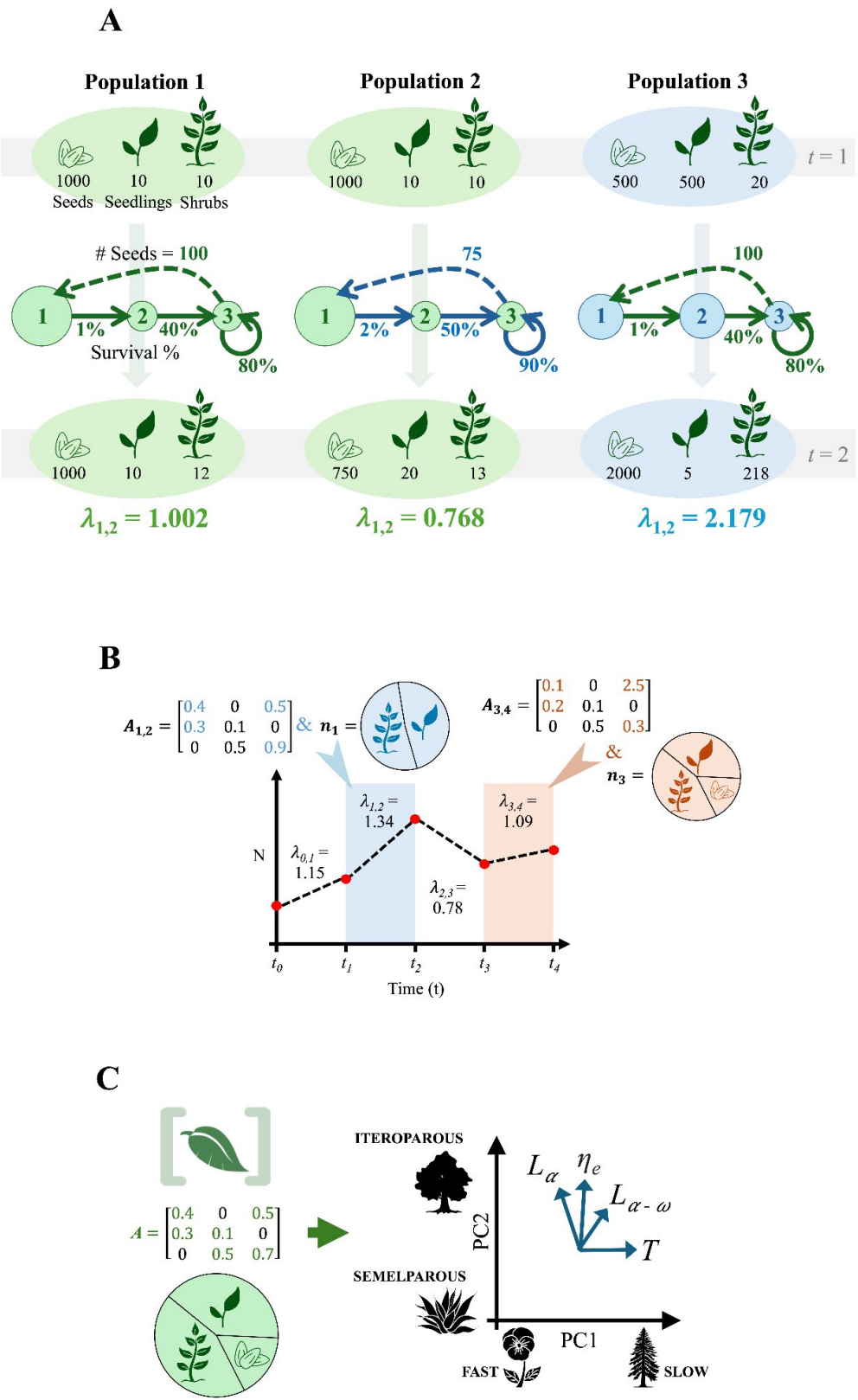
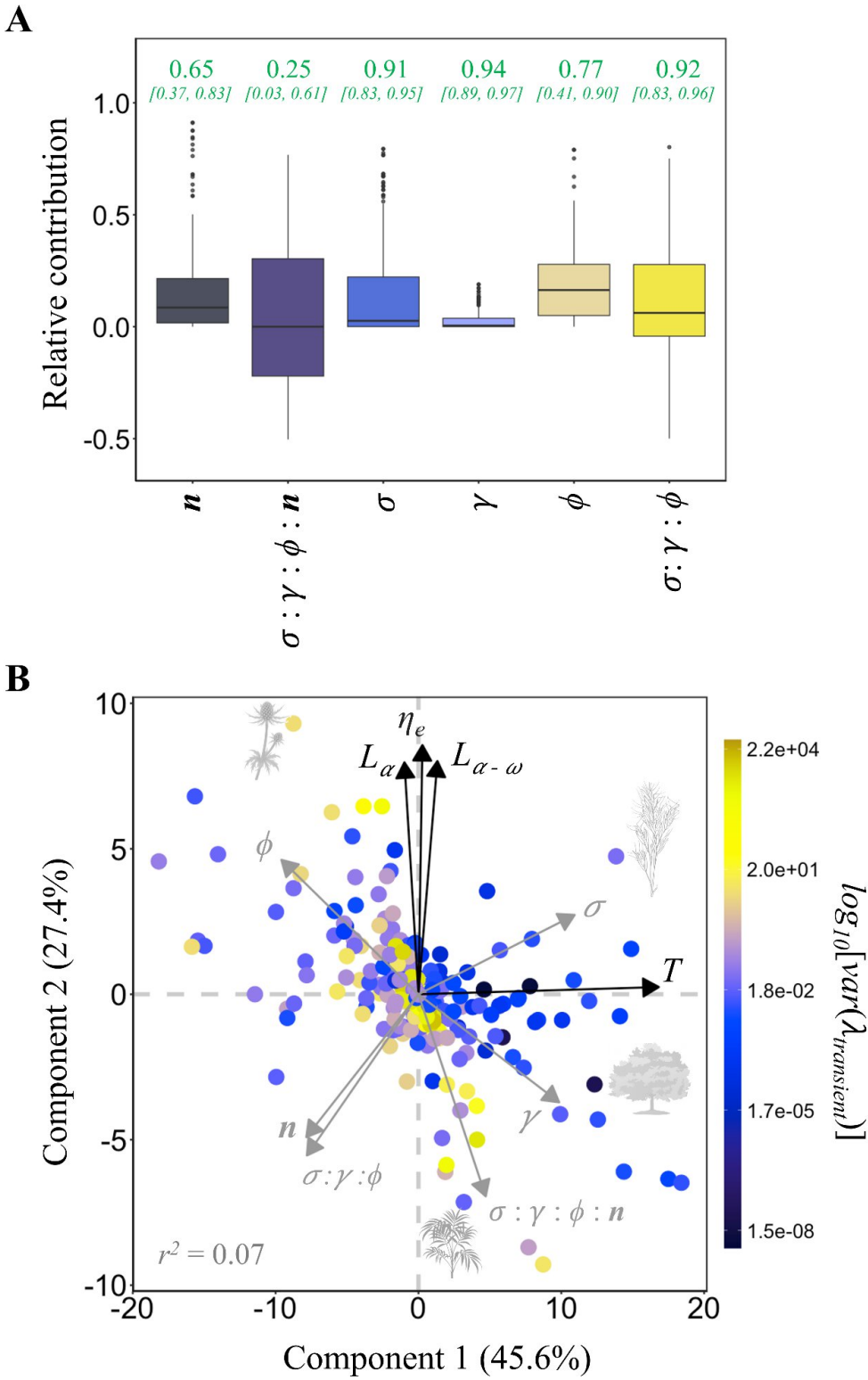
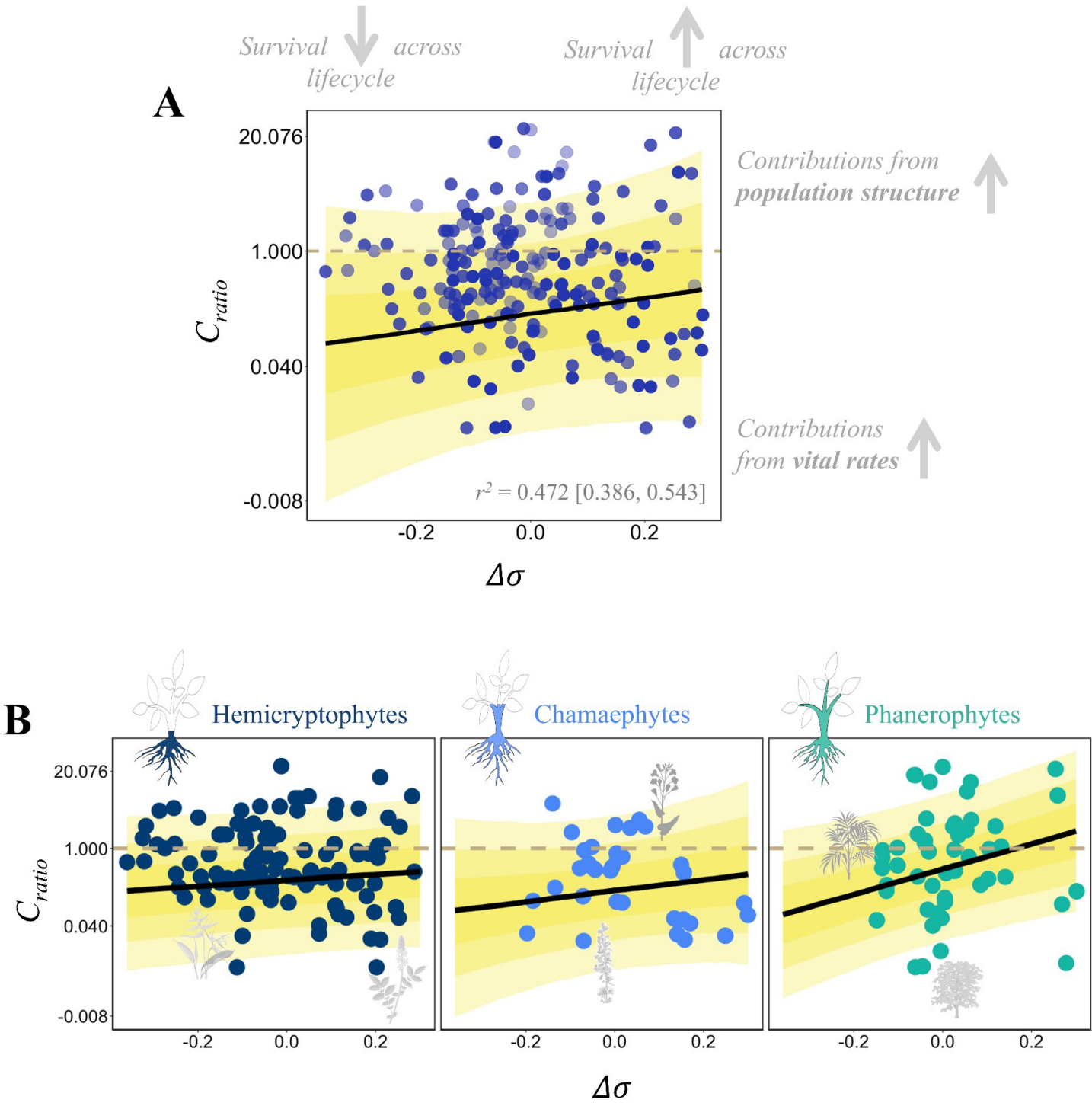


Figure 1. Our analytical pipeline to evaluate how the relative contributions of the vital rates and structure of a population to observed variation in demographic performance align with established life-history theory. (A) Variation across the realised growth rates expressed by natural populations are underpinned by changes in both their vital rates (*e.g.*, survival & fecundity) and structure (*i.e.*, the distribution of individuals across their life cycle). Consider three populations, each comprising the exact same population size ($N_t = 1,020$ individuals) at time $t = 1$. Two of these populations (Populations 1 & 2) have an identical structure but exhibit differing survival and fecundity rates, while two (Populations 1 & 3) have different structures but possess the same survival and fecundity rates. Despite these paired similarities, each population displays very different growth trajectories between times $t = 1$ and $t = 2$ owing to their differing structure and vital rate combinations, with Population 1 remaining stable ($\lambda_{1,2} \approx 1$), Population 2 declining ($\lambda_{1,2} < 1$), and Population 3 growing rapidly ($\lambda_{1,2} \gg 1$). **(B)** Transient life table response experiments (tLTREs) decompose variation in short-term (*i.e.*, transient) population growth rates, observed across time, space, and experimental treatments, into the proportional contributions of changes in vital rates (**A**) and population structure (**n**). **(C)** We identified a sample of matrix population models (MPMs) describing temporal, spatial or experimental variation in the dynamics of different plant populations and corresponding census records describing their population structure from the COMPADRE database (22). Using these series of paired MPMs and population structures, we then evaluated how the proportional effects of vital rates and population structure on variance in $\lambda_{transient}$ corresponds with patterns across key life-history traits: generation time (T), age at reproductive maturity (L_α), mean life expectancy (η_e), and reproductive window ($L_{\alpha-\omega}$) (note the loadings shown are hypothetical representations).



961 **Figure 2. Variation in population structure has a comparable influence on observed variation**
 962 **in realised population growth rates ($\lambda_{transient}$) as fluctuations in survival and fecundity rates;**
 963 **particularly in species with faster life-history strategies. (A)** The relative contributions
 964 observed across different sources of variance in $\lambda_{transient}$: Population structure (n), the interactions
 965 between population structure and vital rates ($\sigma : \gamma : \phi : n$), survival (σ), development (γ),
 966 fecundity (ϕ), and the interactions between these vital rates ($\sigma : \gamma : \phi$). The inset values at the top
 967 illustrate the phylogenetic signal (Pagels I) associated with each contribution source, with error
 968 shown as 95% CIs. **(B)** Partial least squares biplot showing the relationship between sources of
 969 variation in $\lambda_{transient}$ and life-history traits: generation time (T), age at reproductive maturity (L_{α}),
 970 mean life expectancy (η_e), and reproductive window ($L_{\alpha-\omega}$). The colour scale depicts the log-
 971 transformed variance in $\lambda_{transient}$ recorded for population comparison. Representative species icons
 972 included adjacent to corresponding data point showing (top to bottom): *Eryngium alpinum* (Alpine
 973 sea holly), *Danthonia sericea* (Silky oat-grass), *Acer saccharum* (Sugar maple), and *Chamaedorea*
 974 *elegans* (Parlor palm).



977 **Figure 3. Populations with life cycle survival profiles increasingly weighted towards survival**
 978 **in later life experience greater contributions from their population structure to variation in**
 979 **realised growth rates ($\lambda_{transient}$), a pattern that manifests consistently across plant growth**
 980 **strategies. (A)** The relationship between $\Delta\sigma$ (*i.e.*, rate of change in stage-specific survival across
 981 the lifecycle) and contribution ratio estimates (C_{ratio}) across our 268 plant population comparisons.
 982 C_{ratio} estimates reflect the relative balance of contributions of population structure (n) and vital
 983 rates (A) toward observed variance in demographic performance, with the dashed line at $C_{ratio} =$
 984 1 indicating equal contributions. Point shading differentiates between tLTRE assessments based
 985 on temporal, spatial, or experimental comparisons, or a combination of these cases (n: Time = 120,
 986 Space = 63, Experimental = 34, Combination = 51). Error displayed using 95% CIs. **(B)** The
 987 relationship between $\Delta\sigma$ and contribution ratio estimates (C_{ratio}) shown separately for
 988 Hemicryphytes, Chamaephytes, and Phanerophytes. Error is displayed as 95% CI with the
 989 horizontal dashed line indicating the 1:1 threshold between contributions from vital rates and
 990 population structure. Representative species icons included adjacent to corresponding data point
 991 showing (left to right): *Heliconia acuminata* (Rich heliconia), *Agrimonia eupatoria* (Sticklewort),
 992 *Ambrosia dumosa* (White bursage), *Brassica insularis* (Indian mustard), *Chamaedorea elegans*
 993 (Parlor palm), and *Fagus grandifolia* (American beech).

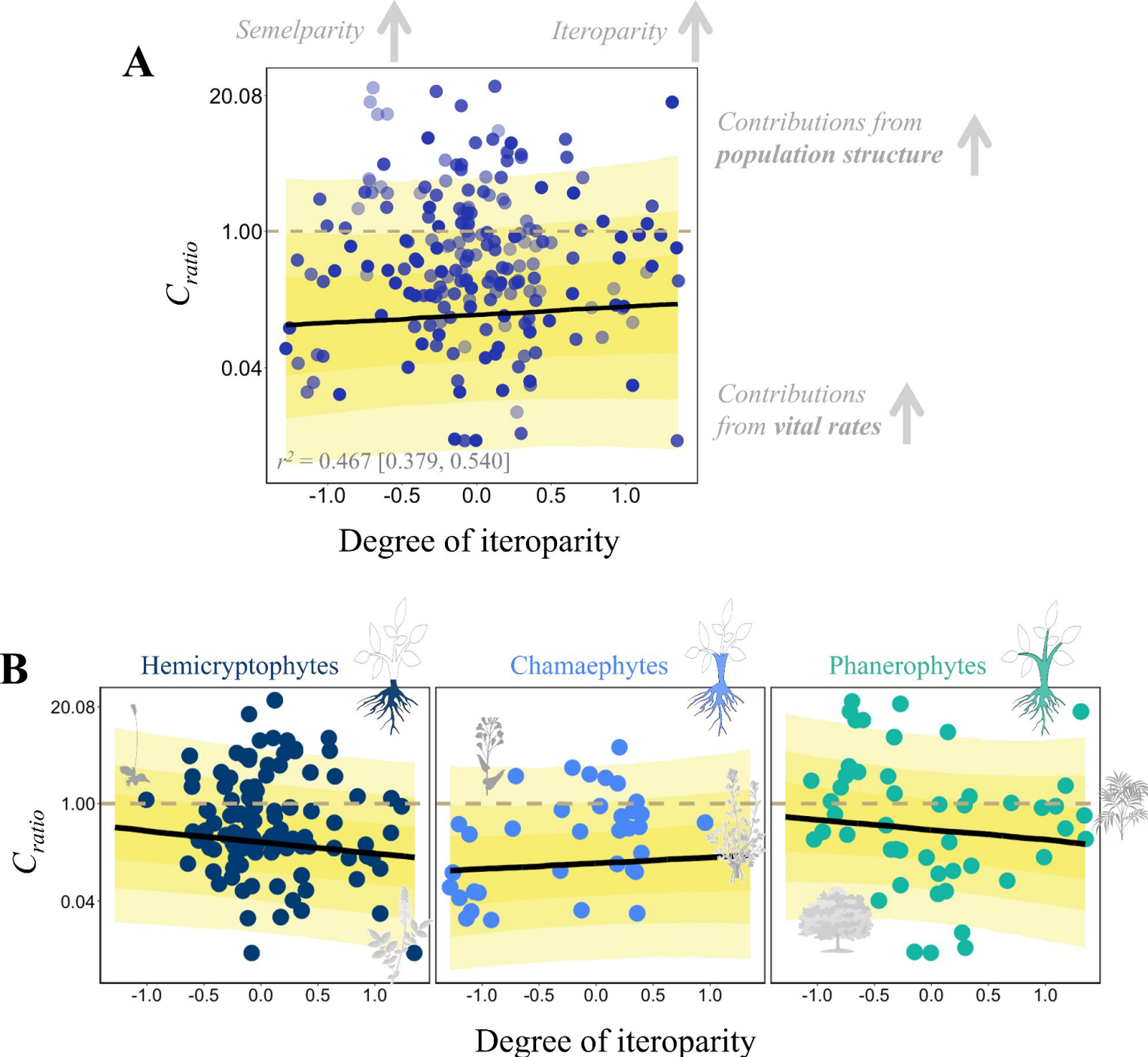
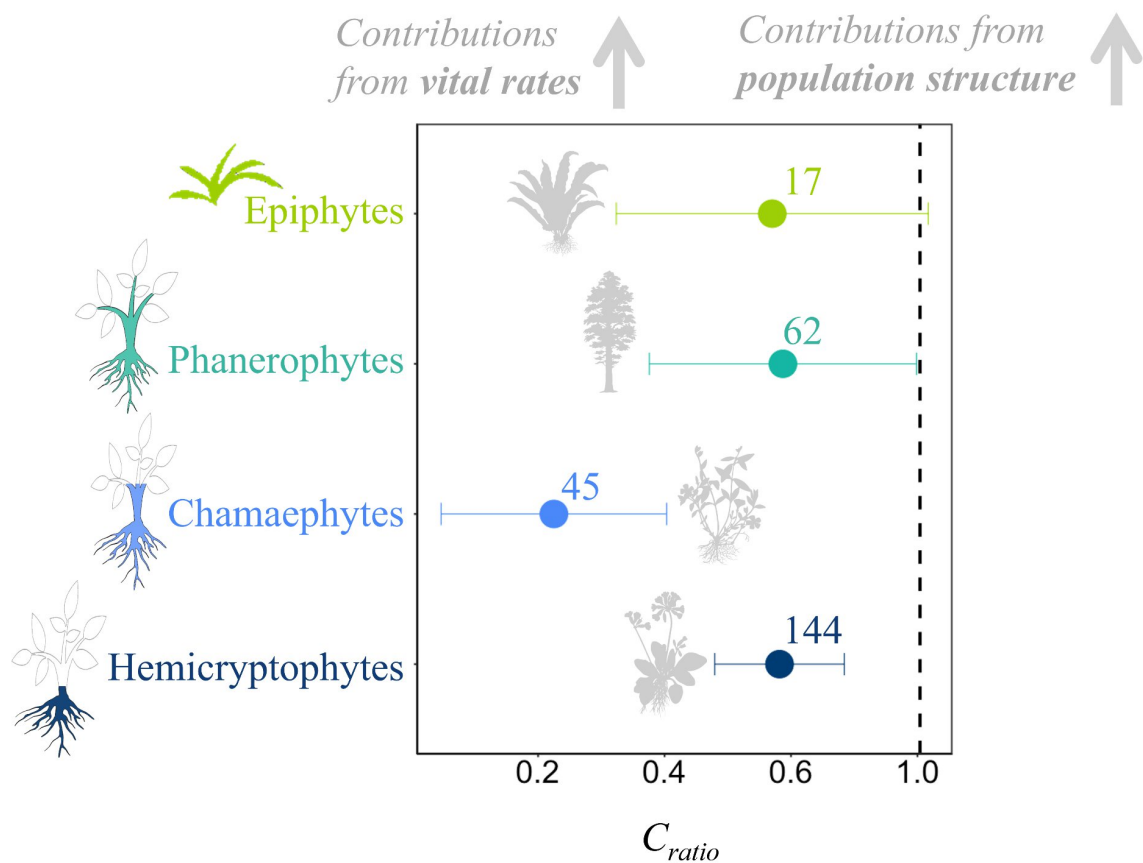


Figure 4. The relationship between degree of iteroparity and the relative value of contributions from population structure to variation in realised growth rates ($\lambda_{transient}$) varies across plant growth forms. (A) The relationship between degree of iteroparity (*i.e.*, spread of reproduction through the life cycle) and contribution ratio estimates (C_{ratio}) across our 268 plant population comparisons. C_{ratio} estimates reflect the relative balance of contributions of population structure (***n***) and vital rates (***A***) toward observed variance in demographic performance, with the dashed line at $C_{ratio} = 1$ indicating equal contributions. Point shading differentiates between tLTRE assessments based on temporal, spatial, or experimental comparisons, or a combination of these cases (n: Time = 120, Space = 63, Experimental = 34, Combination = 51). Error displayed using 95% CIs. **(B)** The relationship between degree of iteroparity and contribution ratio estimates (C_{ratio}) shown separately for Hemicrypophytes, Chamaephytes, and Phanerophytes. Error is displayed as 95% CI with the horizontal dashed line indicating the 1:1 threshold between contributions from vital rates and population structure. Representative species icons included adjacent to corresponding data point showing (left to right): *Plantago media* (Hoary plantain), *Agrimonia eupatoria* (Sticklewort), *Brassica insularis* (Indian mustard), *Artemisia genipi* (Genipi), *Acer saccharum* (Sugar maple), and *Chamaedorea elegans* (Parlor palm).

1012 **Figure 5**



1013 **Figure 5. The relative importance of contributions of population structure vs. vital rates to**
1014 **variation in realised growth rates ($\lambda_{transient}$) varies across plant growth forms.** Mean
1015 contribution ratio estimates (C_{ratio}) obtained for populations of plants classified as epiphytes (e.g.,
1016 Bird's nest Fern, *Asplenium nidus*), phanerophytes (e.g., Giant sequoia, *Sequoiadendron*
1017 *giganteum*), chamaephytes (e.g., Periwinkle, *Vinca minor*), and hemicryptophytes (e.g., Cowslip,
1018 *Primula veris*). The vertical dashed line indicates equal contributions from vital rates and
1019 population structure to realised growth rates, while $C_{ratio} < 1$ implies a greater contribution from
1020 vital rates. Error bars display 95% CIs. Inset values indicate respective sample sizes obtained for
1021 each plant growth form.

Table 1. Definitions of the selected life-history traits used in this study.

Trait	Definition
Generation time (T)	Time steps required for individuals within a population to be replaced.
Mean life expectancy (η_e)	The average lifespan of individuals within a population.
Age at sexual maturity (L_α)	The average number of time steps taken by individuals to reach reproductive maturity.
Reproductive window ($L_{\alpha-\omega}$)	The average duration of reproduction in individuals across a population.
Life cycle survival profile ($\Delta\sigma$)	The rate of change in stage-specific survival across the lifecycle of a population. Higher values correspond with faster rates of change in stage-specific survival. Positive/negative values imply increases/decreases in survival throughout the lifecycle.
Mean lifetime survival ($\underline{\sigma}$)	Average survival probability of individuals across their lifecycle, irrespective of life-stage.
Degree of iteroparity (S)	Spread of reproduction through the lifespan of individuals. Higher/lower values correspond with greater iteroparity/semelparity.

1045 ***Supporting Information for:***
1046 **Life-history variation mediates the importance of population structure to the**
1047 **short-term dynamics of plant populations worldwide**

1048

1049 James Cant, Christina M. Hernández, David N. Koons, Dave J. Hodgson, Man Qi, Andrew
1050 Hector, Iain Stott, Roberto Salguero-Gómez

1051

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1056 **This PDF file includes:**

1057 Supporting text

1058 Figures S1 to S12

1059 Table S1

1060 SI References

Supporting Information Text:

S1. Evaluating the spatial representation of our tLTRE sample.

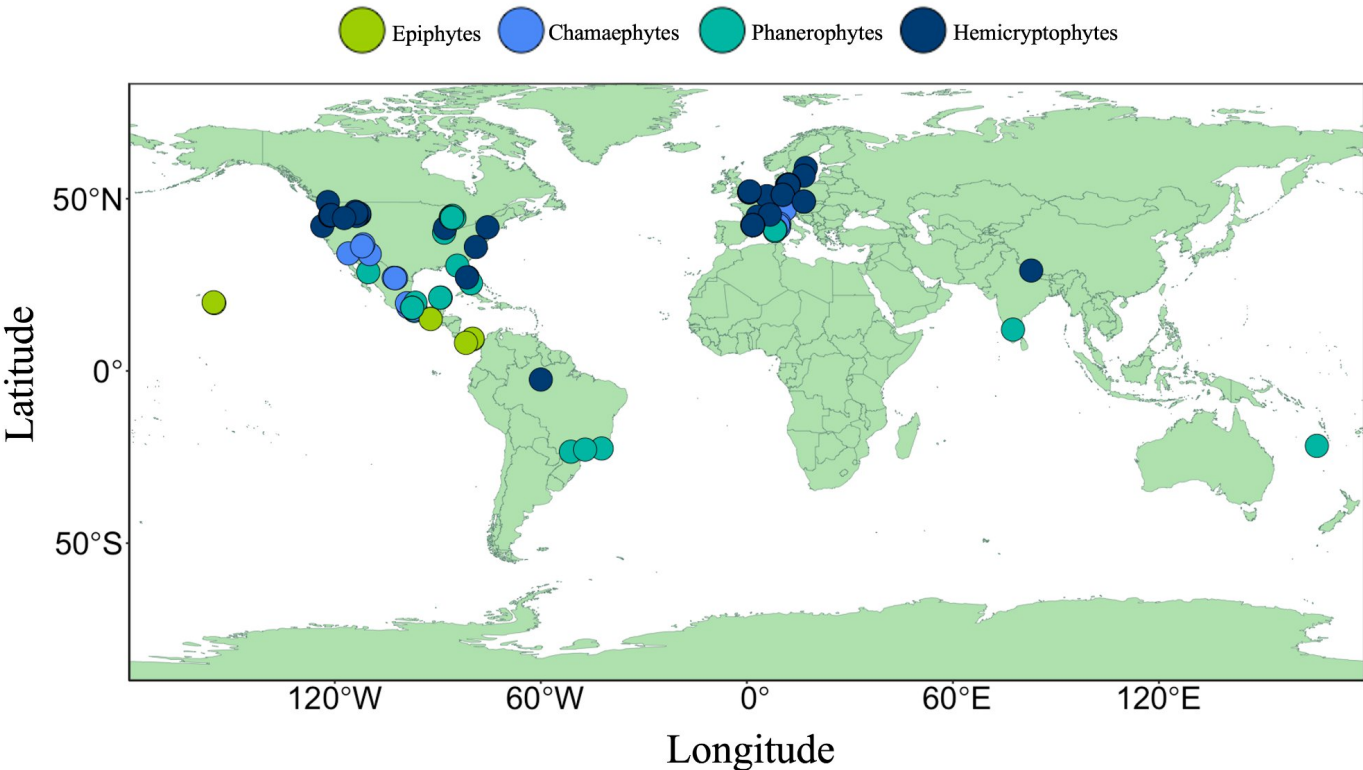
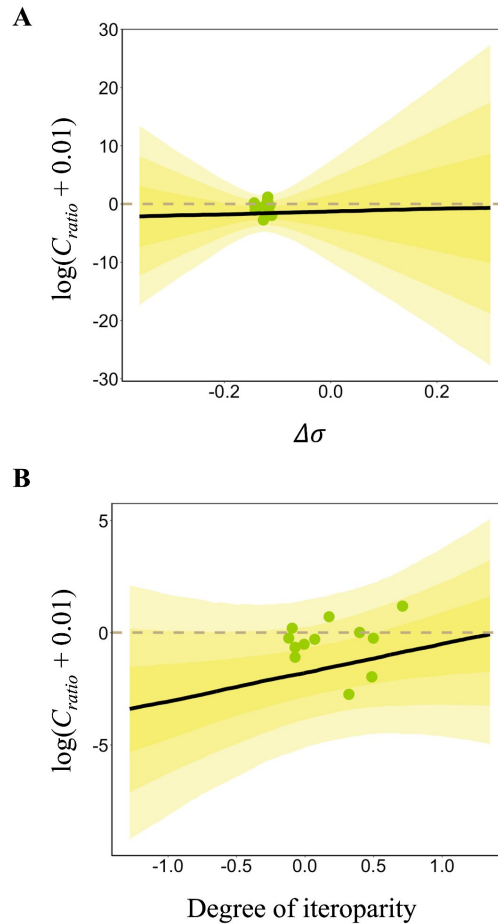


Figure S1. Recorded locations of all the unique populations represented by our 268 population comparisons across time, space, and experimental treatments. Points are coloured based on the Raunkiaer growth form classification assigned to each population, with geophytes ($n = 1$) and hemicryptophytes (43) grouped as hemicryptophytes, and mega- (3), meso- (15), and nanophanerophytes (10) grouped as phanerophytes. Epiphytes and chamaephytes are then represented by 6 and 16 populations, respectively.

1069 **S2. Excluding epiphytes from growth form comparisons.**



1070 **Figure S2. The epiphytic plant population comparisons included in our sample covered a limited**
 1071 **range of life-cycle survival profile ($\Delta\sigma$) and degree of iteroparity (S) estimates, undermining the**
 1072 **observed relationships between these life-history traits and contribution ratio estimates (C_{ratio})**
 1073 **reported for epiphytes. (A)** The relationship between $\Delta\sigma$ (*i.e.*, rate of change in stage-specific survival
 1074 across the lifecycle) and contribution ratio estimates (C_{ratio}) obtained for 17 comparisons of epiphyte
 1075 populations across time, space, and/or experimental treatments. **(B)** The relationship between degree of
 1076 iteroparity and contribution ratio estimates (C_{ratio}) for the same 17 comparisons. C_{ratio} estimates reflect the
 1077 relative balance of contributions of population structure (n) and vital rates (A) toward observed variance in
 1078 demographic performance, with the dashed lines at $C_{ratio} = 1$ across both panels indicating equal
 1079 contributions. Error displayed using 95% CIs.

S3. Quantifying life cycle survival profile ($\Delta\sigma$).

As part of this study, we introduced the measure of life cycle survival profile ($\Delta\sigma$). Briefly, this measure describes the rate of change in stage-specific survival across the life cycle of a population. Higher absolute values of $|\Delta\sigma|$ correspond with more abrupt rates of change in stage-specific survival along the life cycle of the examined population, while positive/negative values imply increases/decreases in survival throughout its life cycle. Across our sample of 268 population comparisons, we estimated the element-by-element arithmetic mean matrix population model (MPM) from each associated set of MPMs describing the focal species/population dynamics across different time intervals, locations, or experimental treatments. We then collapsed each mean MPM into a 3×3 matrix, resulting in all comparisons comprising three life cycle stages to ensure comparability across populations and species. We carried out this step using the function *mpm_collapse* of the *Rcompadre* R package (1) to collapse together all life stages beyond the second life cycle stage into a single terminal stage while keeping the first two stages unaltered. This approach ensured we retained the eigen-structure of the original MPMs (2). We then computed $\Delta\sigma$ as the slope coefficient of a linear regression model fitted to the sequence of mean stage-specific survival probabilities obtained for each mean population/species MPM (*e.g.*, the slope of survival across life cycle stages). Below, we visually demonstrate the suitability of assuming that the gradient in stage-specific survival across three-stage life cycles can be appropriately modelled using a linear relationship. For this demonstration, we present the stage-specific survival (σ) patterns obtained for 30 randomly selected mean MPMs across our 268 population comparisons (Fig. S3), illustrating how, following the collapsing of each population's dynamics into a three-stage model, survival typically follows a linear rate of change across life stages.

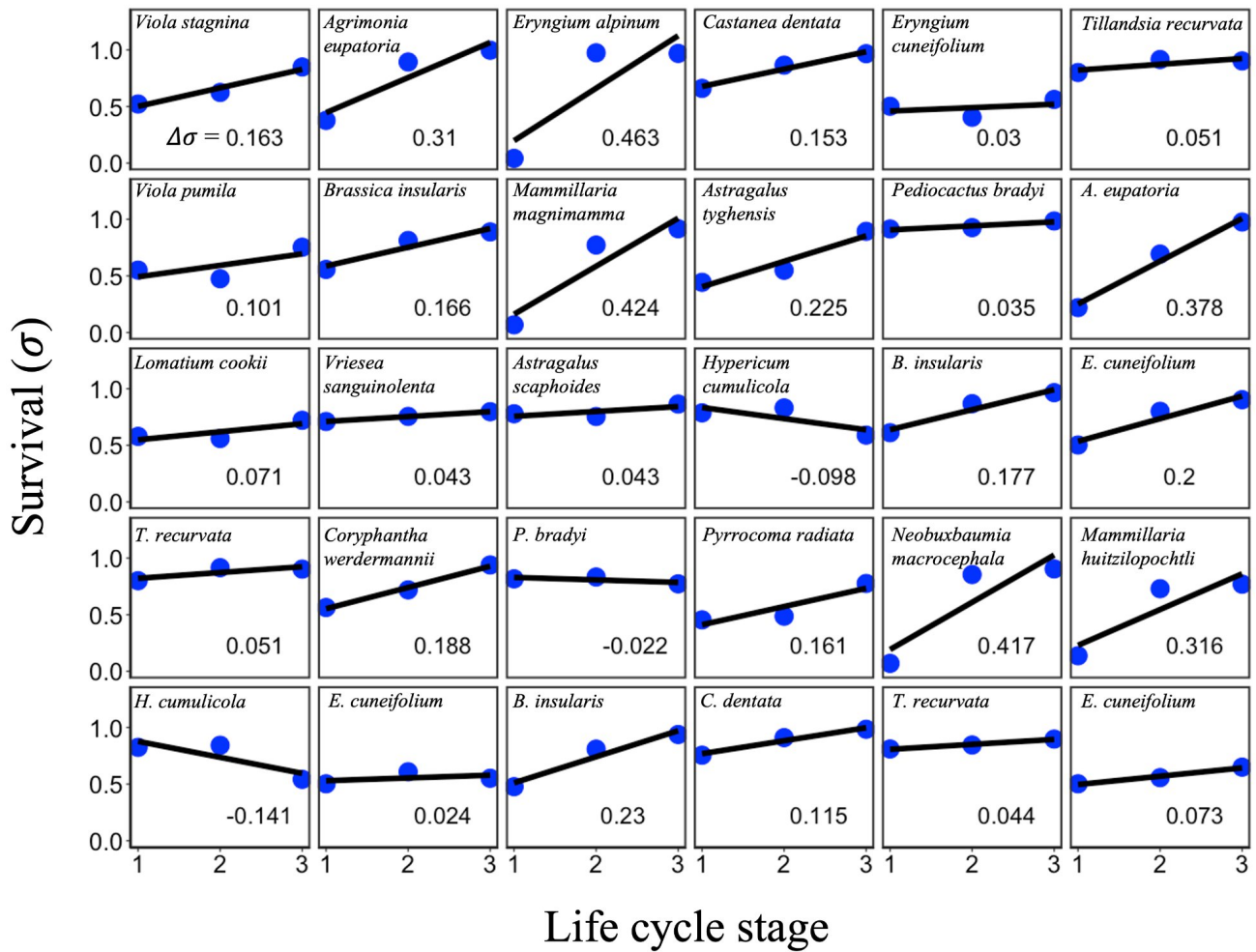


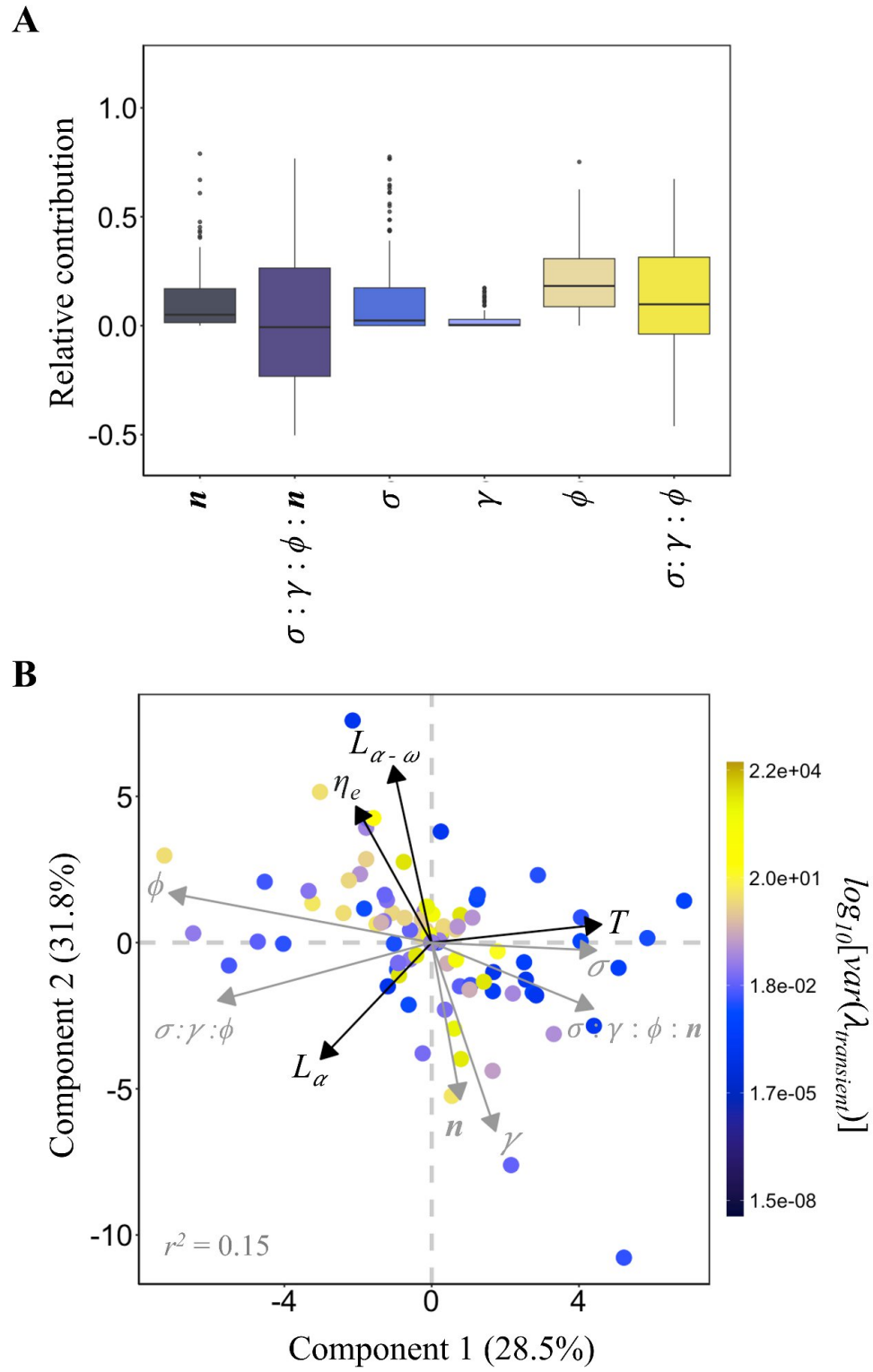
Figure S3. The mean stage-specific survival (σ) patterns across the life cycles of 30 plant populations randomly selected from our 268 population comparisons can be approximated via a linear model, whose slope is summarised by our metric: the life cycle survival profile ($\Delta\sigma$). Across each panel, the solid lines represent the linear relationship fit to each population's corresponding pattern in stage-specific survival (σ), with the inset text indicating the associated slope coefficient. Subsequently, these slope coefficients reflect the estimates of life cycle survival profile ($\Delta\sigma$) assigned to the selected populations.

S4. Testing sensitivity to pooling temporal, spatial, and experimental comparisons.

Our sample of 268 population comparisons comprises comparisons of population dynamics across time, space, and experimental treatments, or a combination of these cases (Time = 120 comparisons, Space = 63, Experimental = 34, Combination = 51). To test the sensitivity of our findings to our assumption that the short-term population growth rates in our plant dataset vary in the same fashion when comparisons are made across time, space, and experimental treatments, we repeated our analyses using only population comparisons across time. To do so, we repeated our analyses, implementing the same methodological pipeline as described in the main manuscript, but focused only on data sourced from population comparisons across time.

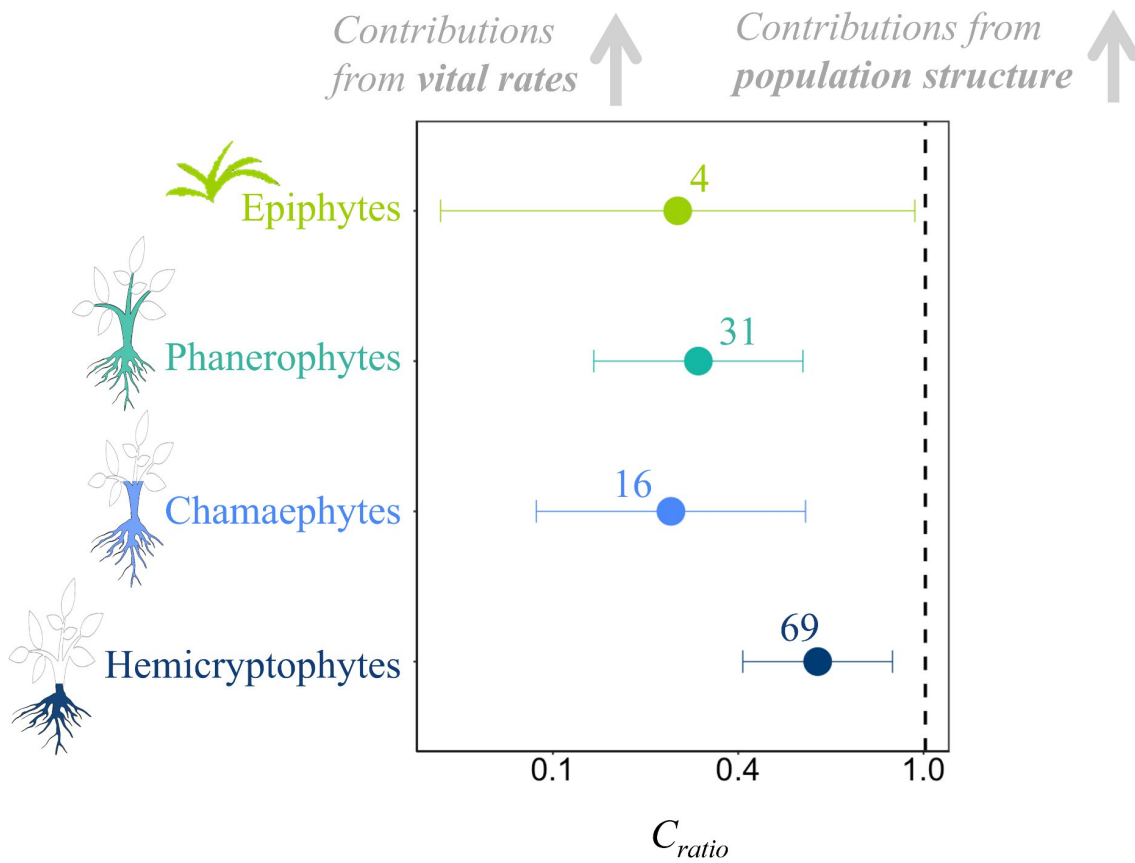
This brute-force sensitivity approach demonstrates how changes in population structure (n) over time contribute to temporal variance in demographic performance to a similar extent as temporal shifts in survival (σ) and fecundity (ϕ) (Fig. S4A). Next, we reassessed the relationship between sources of variance in demographic performance and the life-history traits of generation time (T), age at reproductive maturity (L_a), mean life expectancy (η_e), and reproductive window ($L_{a-\omega}$) (Fig. S4B). This analysis shows that the contributions from temporal shifts in survival to demographic performance correlate with generation time, while contributions from temporal shifts in population structure are greater in populations with shorter lifespans. We do note, however, that when only considering comparisons across time, the association between contributions from population structure to variance in demographic performance and shorter generation times becomes weaker (compare to Fig. 2).

1129 **Figure S4. The contribution of population structure to observed temporal variation in realised**
 1130 **growth rates ($\lambda_{transient}$) is greater in species with faster life-history strategies. (A)** The proportional
 1131 effect distributions observed across different sources of temporal variance in $\lambda_{transient}$: Population structure
 1132 (n), the interaction between population structure and vital rates ($\sigma : \gamma : \phi : n$), survival (σ), development
 1133 (γ), fecundity (ϕ), and the interactions between these vital rates ($\sigma : \gamma : \phi$). **(B)** Partial least squares biplot
 1134 showing the relationship between sources of temporal variation in $\lambda_{transient}$ and life-history traits: generation
 1135 time (T), mean life expectancy (η_e), age at reproductive maturity (L_a), and reproductive window ($L_{a-\omega}$). The



1136 colour scale depicts the log-transformed variance in $\lambda_{\text{transient}}$ recorded for population comparison.

Pooling temporal, spatial, and experimental comparisons did have a slight influence on patterns between life-history traits and sources of variance in demographic performance across growth forms. Again, we computed estimates of contribution ratio (C_{ratio}) for each of our population comparisons, providing a measure of the balance between contributions from population structure *vs.* vital rates towards observed variance in demographic performance. Focusing only on temporal population comparisons, we observe how, although the relative patterns across C_{ratio} estimates obtained for hemicryptophytes, chamaephytes, and phanerophytes are maintained, the difference between estimates observed for chamaephytes and phanerophytes is diminished slightly (Fig. S5) compared to when spatial, temporal, and experimental comparisons are pooled (Fig. 5). We also observe how, when focusing on only temporal comparisons of population performance, C_{ratio} displays a negative relationship with the life cycle survival profile ($\Delta\sigma$) in both hemicryptophytes and chamaephytes (Fig. S6). This pattern is actually the reverse of the one obtained when temporal, spatial, and experimental comparisons are pooled (compare to Fig. 3). The pattern between $\Delta\sigma$ and C_{ratio} is, however, insensitive in phanerophytes across pooled and time-only comparisons. Finally, although stronger in time-only comparisons, the positive relationship observed between degree of iteroparity and C_{ratio} is maintained across pooled and time-only comparisons (Figs. S7, compared to Fig. 4). Thus, the primary findings we report are largely insensitive to our choice to pool temporal, spatial, and experimental comparisons of population dynamics in our analyses.



1157

1158 **Figure S5. The relative importance of contributions of population structure vs. vital rates to observed**

1159 **temporal variation in realised growth rates ($\lambda_{transient}$) across plant growth forms.** Mean contribution

1160 ratio estimates (C_{ratio}) obtained for populations of plants classified as epiphytes, phanerophytes,

1161 chamaephytes, and hemicryptophytes. The vertical dashed line indicates equal contributions from vital

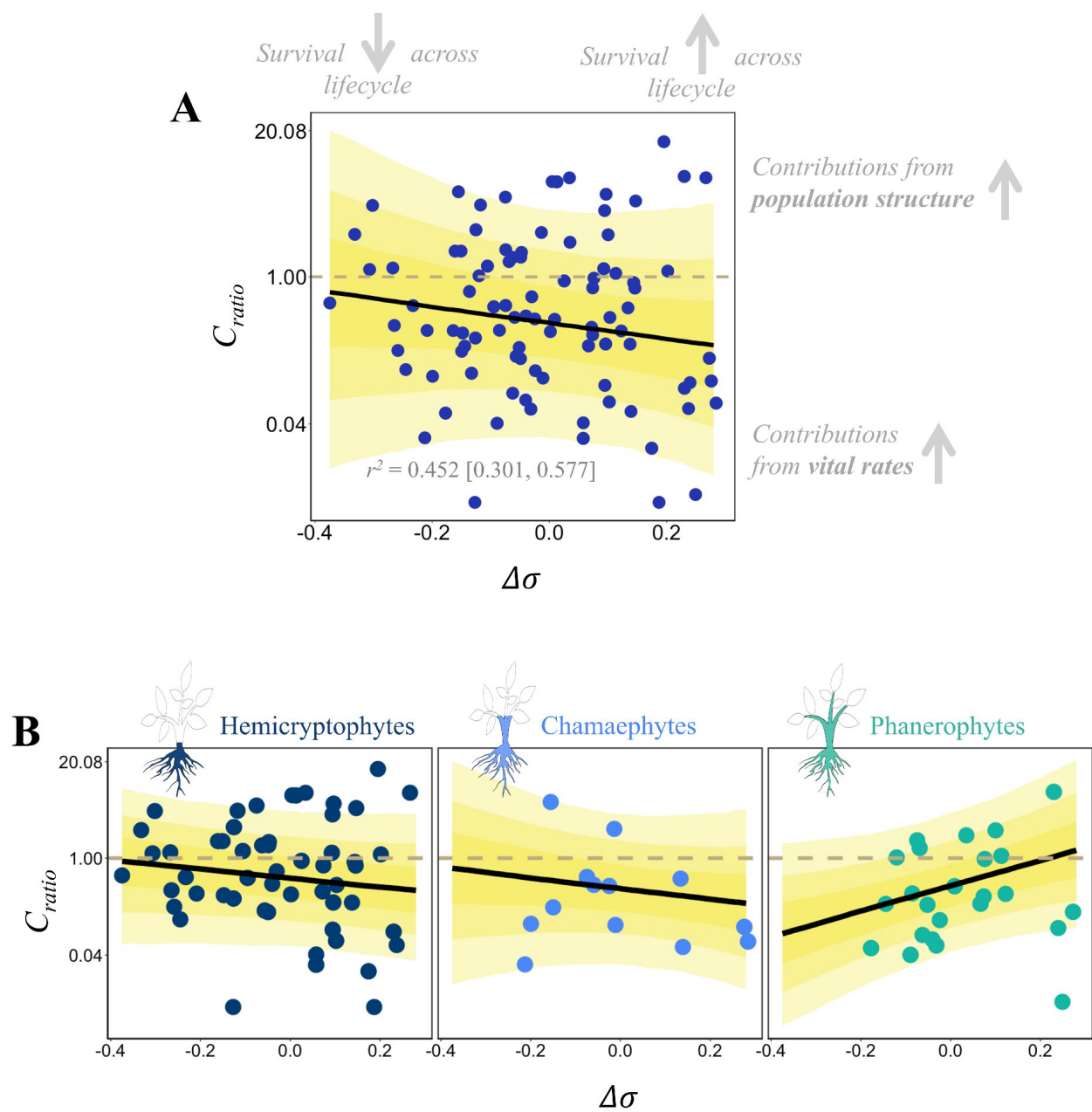
1162 rates and population structure to realised growth rates, while $C_{ratio} < 1$ implies a greater contribution from

1163 vital rates. Error bars display 95% CIs. Inset values indicate respective sample sizes used for each plant

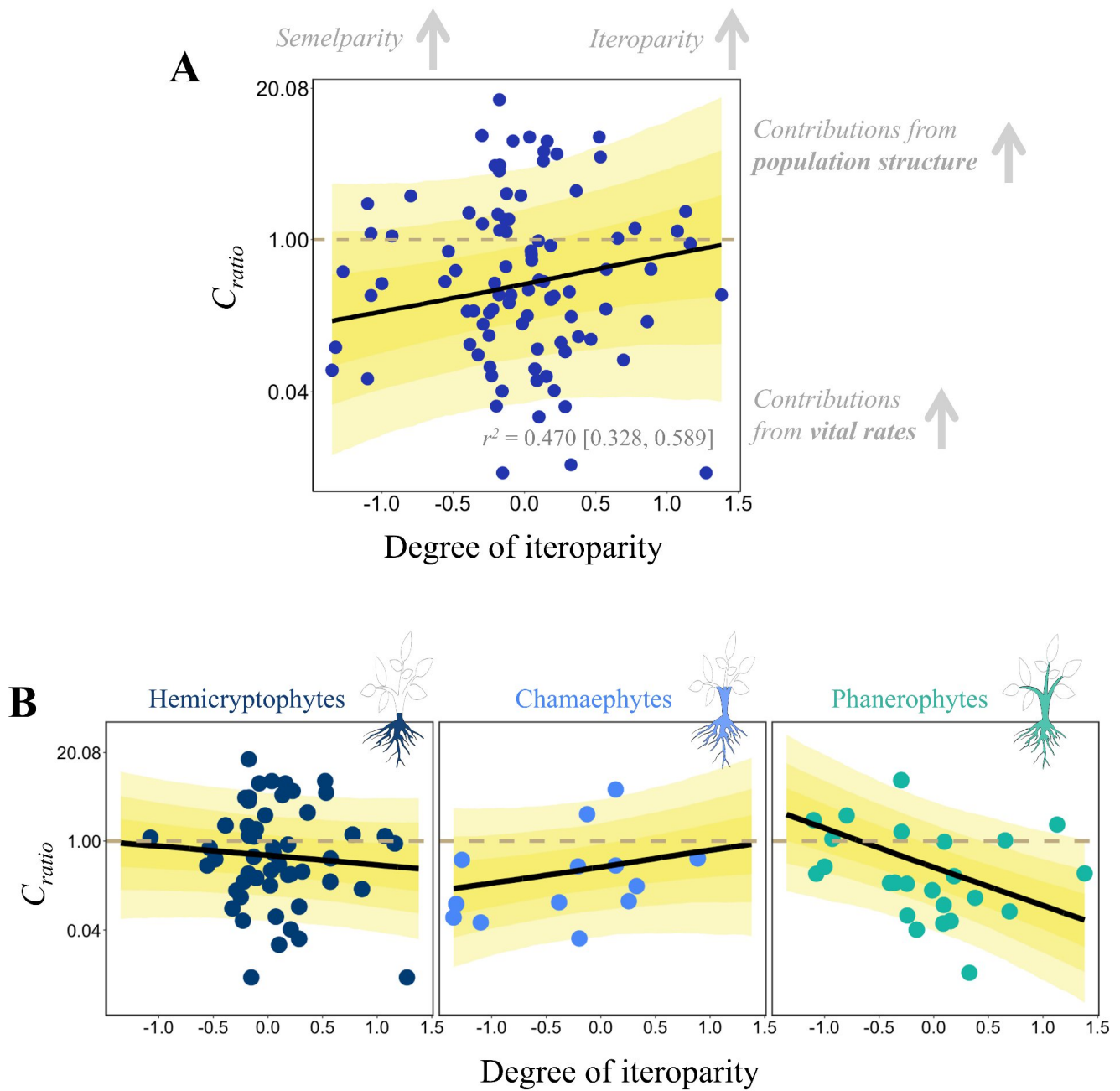
1164 growth form.

Figure S6. Populations with life cycle survival profiles increasingly weighted to survival in later life experience greater contributions from their population structure towards temporal variation in realised growth rates ($\lambda_{transient}$); a pattern that manifests consistently across plant growth strategies.

(A) The relationship between $\Delta\sigma$ (*i.e.*, rate of change in stage-specific survival across the lifecycle) and contribution ratio estimates (C_{ratio}) across 120 comparisons of plant population dynamics through time. C_{ratio} estimates reflect the relative balance of contributions of population structure (n) and vital rates (A) toward observed variance in demographic performance, with the dashed line at $C_{ratio} = 1$ indicating equal contributions. Error displayed using 95% CIs. **(B)** The relationship between $\Delta\sigma$ and contribution ratio estimates (C_{ratio}) shown separately for Hemicrypophytes, Chamaephytes, and Phanerophytes. Error is displayed as 95% CI with the horizontal dashed line indicating the 1:1 threshold between contributions



1175 from vital rates and population structure.



1176

1177 **Figure S7. The relationship between degree of iteroparity and the relative value of contributions**

1178 **from population structure to temporal variation in realised growth rates ($\lambda_{transient}$) varies across plant**

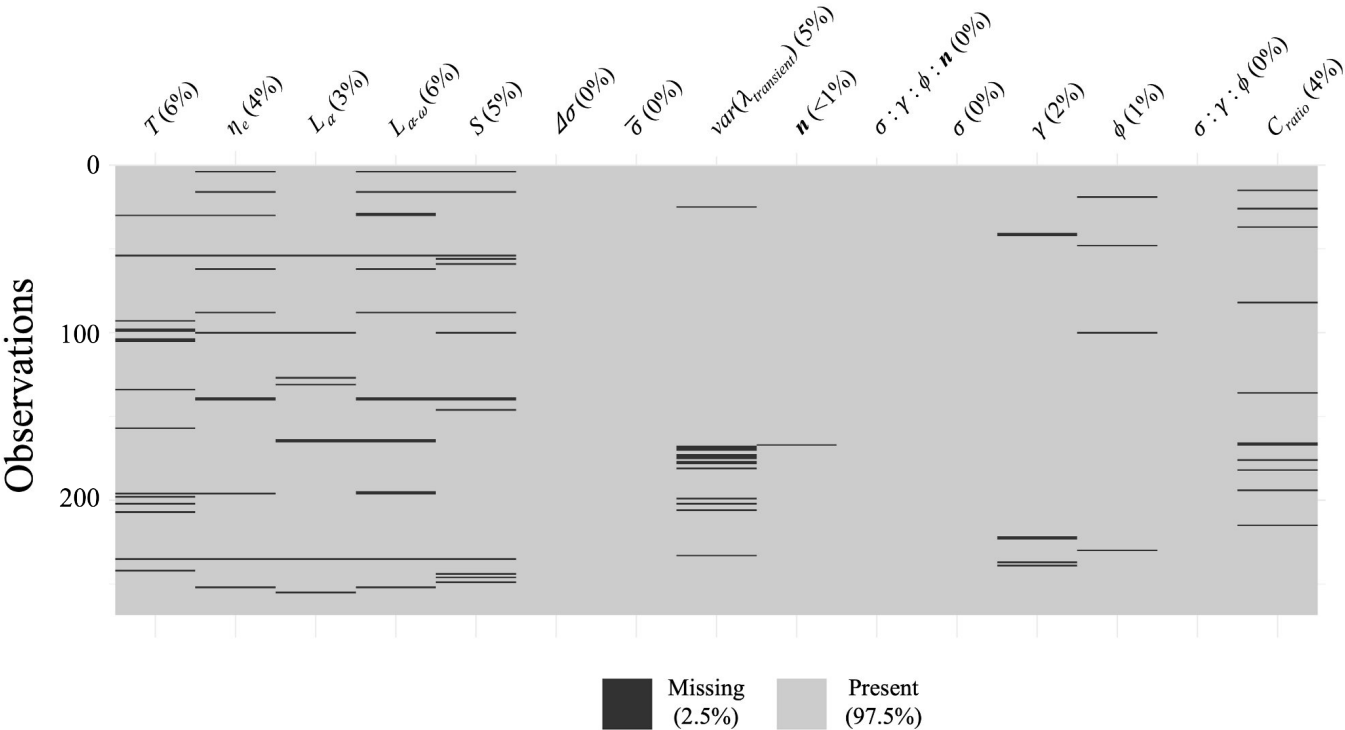
1179 **growth forms. (A) The relationship between degree of iteroparity (*i.e.*, spread of reproduction through the**

1180 life cycle) and contribution ratio estimates (C_{ratio}) across 120 comparisons of plant population dynamics
1181 through time. C_{ratio} estimates reflect the relative balance of contributions of population structure (n) and
1182 vital rates (A) toward observed variance in demographic performance, with the dashed line at $C_{ratio} = 1$
1183 indicating equal contributions. Error displayed using 95% CIs. **(B)** The relationship between degree of
1184 iteroparity and contribution ratio estimates (C_{ratio}) is shown separately for Hemicrypophytes,
1185 Chamaephytes, and Phanerophytes. Error is displayed as 95% CI with the horizontal dashed line indicating
1186 the 1:1 threshold between contributions from vital rates and population structure.

1187 **S5. Testing sensitivity to data imputation.**

1188 During data extraction, we used phylogenetic imputations to estimate missing values across our
1189 demographic and life-history variables (Fig. S8). This step was necessary to allow us to retain
1190 details from all population comparisons within our finalised data sample, to limit the effect of data
1191 gaps on our demographic inferences (5). As such, we imputed estimates for one or more variables
1192 across 85 population comparisons. Combined, just 2.5% of our demographic and life-history trait
1193 data required imputing (Fig. S8). The majority of this missing data comprised estimates of the life-
1194 history traits: generation time (T), age at reproductive maturity (L_a), mean life expectancy (η_e),
1195 reproductive window (L_{a-w}) and degree of iteroparity (S) (Fig. S8).

1196



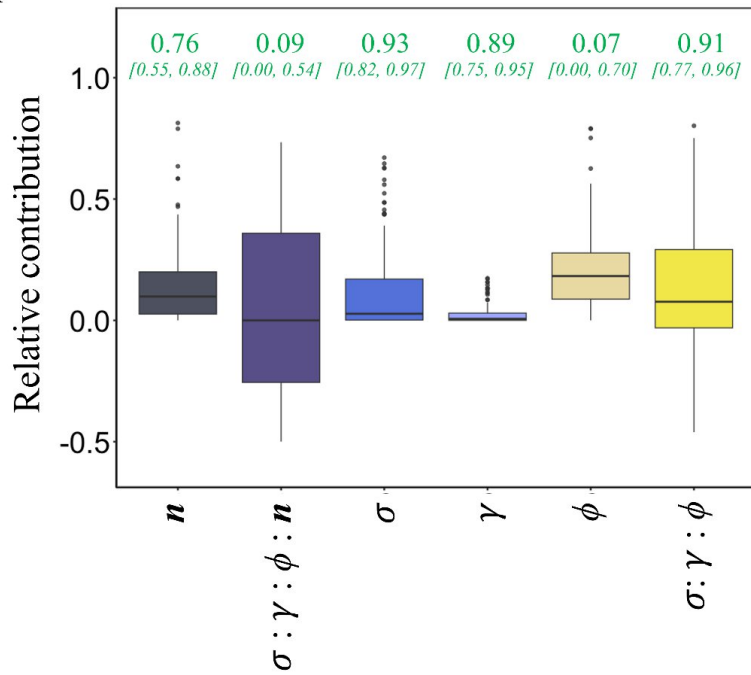
1197 **Figure S8. Spread of missing data across our demographic and life-history trait variables.** Inset
1198 percentages indicate the degree of missing data across each variable.

1199

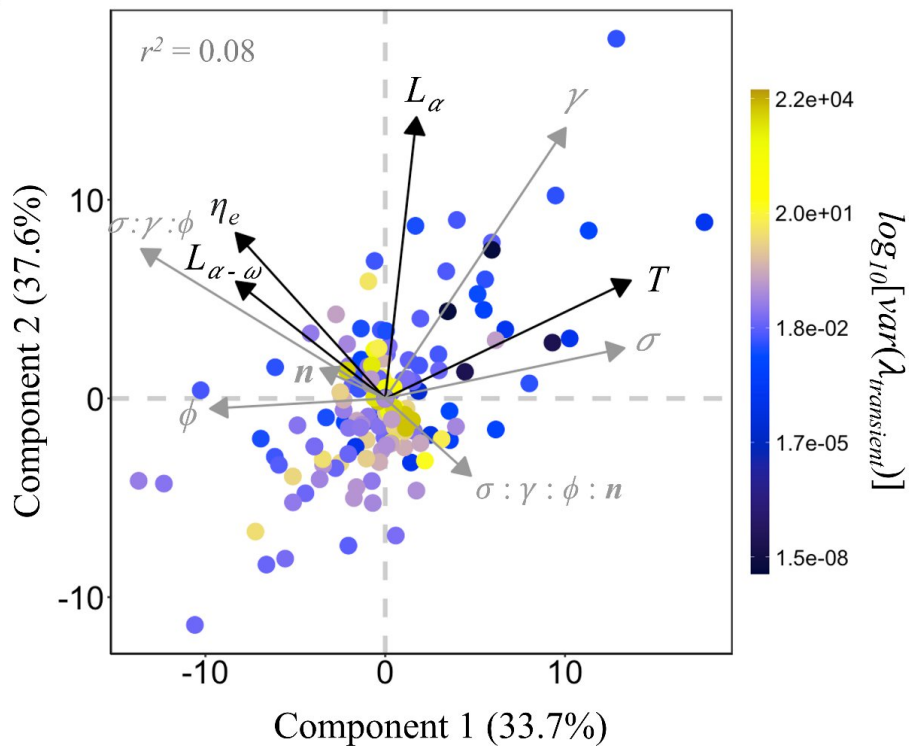
1200 To test the sensitivity of our findings to this imputation, we repeated the same
1201 methodological pipeline as described in the main manuscript, focusing on data sourced from
1202 population comparisons that required no imputation (183 population comparisons). Our brute-
1203 force sensitivity here also had no qualitative effect on our overall findings. First, contributions
1204 from changes in population structure (n) to observed variance in demographic performance
1205 remained comparable to those of changes in survival (σ) and fecundity (ϕ) (Fig. S9A). Second,
1206 contributions from changes in population structure (n) to observed variance in demographic
1207 performance were negatively correlated with generation time (T) (Fig. S9B). Instead, contributions
1208 from survival to demographic performance remained positively associated with generation time.
1209 Third, patterns in the relative balance between contributions from population structure *vs* vital
1210 rates towards observed variance in demographic performance (*i.e.*, contribution ratios, C_{ratio})
1211 across hemicryptophytes, chamaephytes, and phanerophytes remained consistent between our
1212 non-imputed (Fig. S10) and full data samples (Fig. 5). Finally, the patterns we observed between
1213 life cycle survival profile ($\Delta\sigma$) and degree of iteroparity with C_{ratio} across growth forms remain
1214 consistent across our non-imputed (Fig. S11 & S12) and full data samples (Figs. 3 & 4).

1215 **Figure S9. The relationship between the contribution of population structure to observed variation**
 1216 **in realised growth rates ($\lambda_{transient}$) and generation time remains intact even when focusing on the**
 1217 **assessment of only non-imputed data. (A)** The proportional effect distributions observed across different
 1218 sources of temporal variance in $\lambda_{transient}$: Population structure (\mathbf{n}), the interaction between population
 1219 structure and vital rates ($\sigma : \gamma : \phi : \mathbf{n}$), survival (σ), development (γ), fecundity (ϕ), and the interactions
 1220 between these vital rates ($\sigma : \gamma : \phi$). The inset values at the top illustrate the phylogenetic signal (Pagels λ)
 1221 associated with each contribution source, with error shown as 95% CIs. **(B)** Partial least squares biplot
 1222 showing the relationship between sources of temporal variation in $\lambda_{transient}$ and life-history traits: generation
 1223 time (T), mean life expectancy (η_e), age at reproductive maturity (L_α), and reproductive window ($L_{\alpha-\omega}$). The

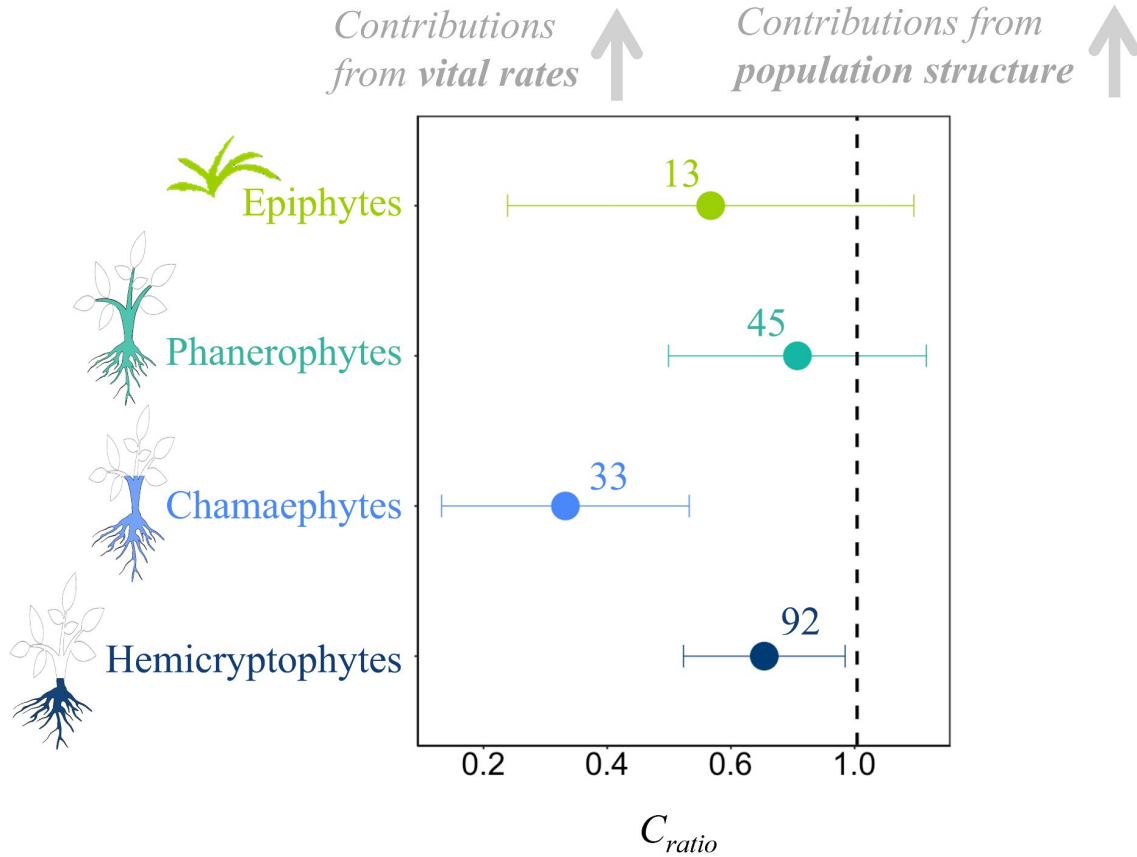
A



B

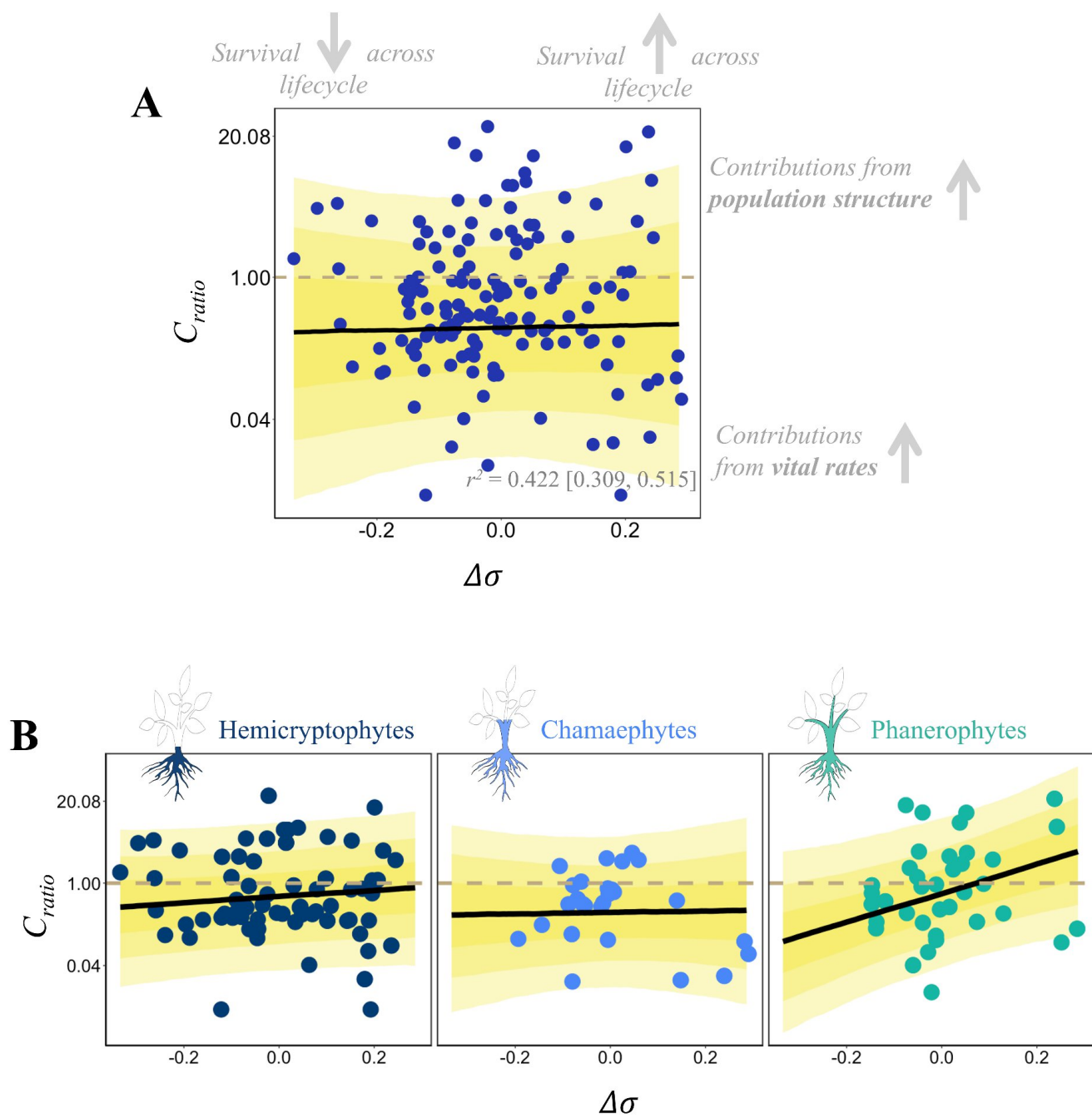


1224 colour scale depicts the log-transformed variance in $\lambda_{\text{transient}}$ recorded for population comparison.



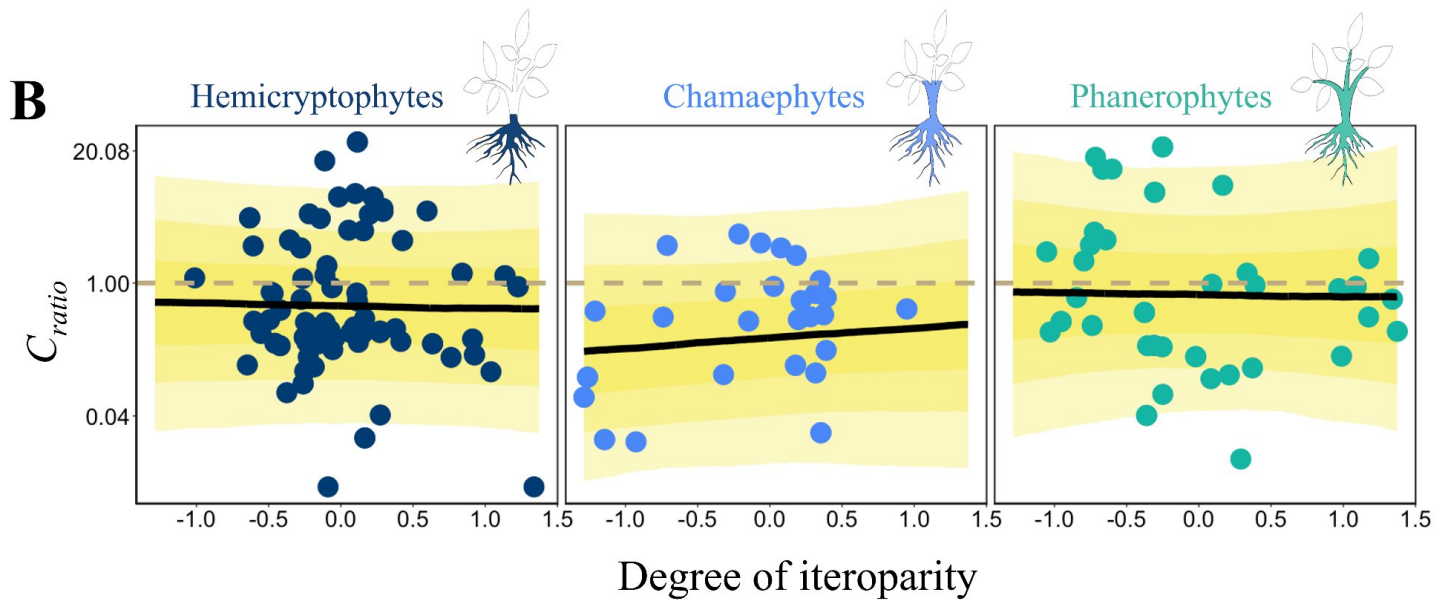
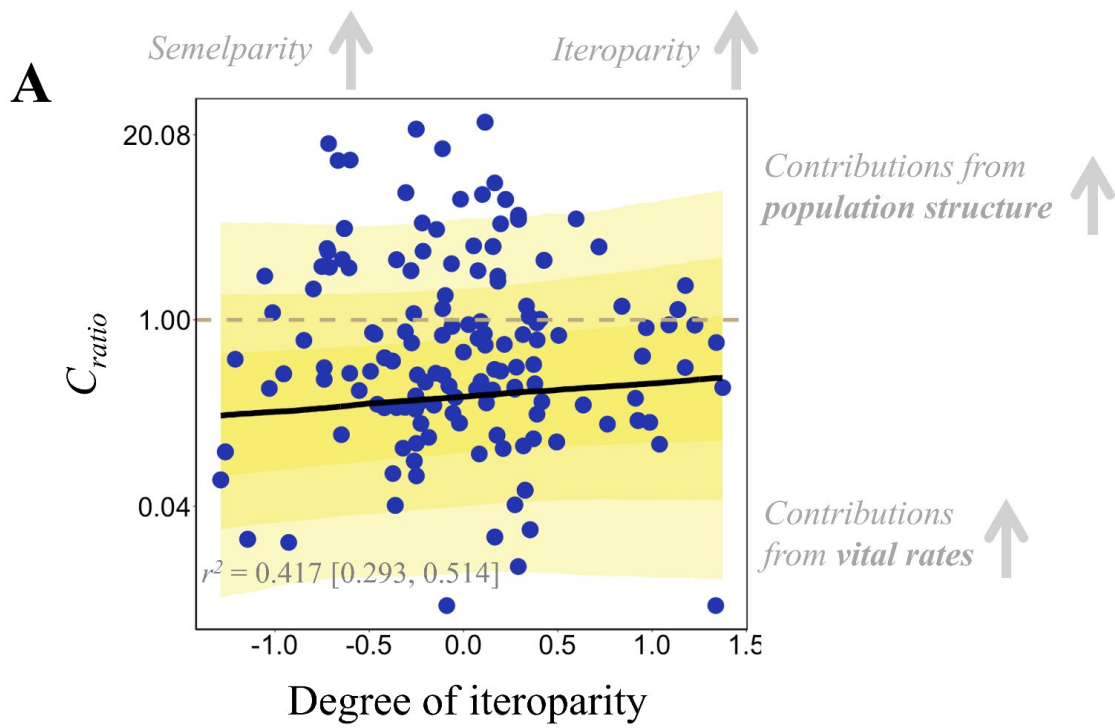
1225 **Figure S10. Variation in the relative importance of contributions of population structure vs. vital**
 1226 **rates to temporal variation in realised growth rates ($\lambda_{transient}$) across plant growth forms is maintained**
 1227 **when focusing on only non-imputed data.** Mean contribution ratio estimates (C_{ratio}) obtained for
 1228 populations of plants classified as epiphytes, phanerophytes, chamaephytes, and hemicryptophytes. The
 1229 vertical dashed line indicates equal contributions from vital rates and population structure to realised growth
 1230 rates, while $C_{ratio} < 1$ implies a greater contribution from vital rates. Error bars display 95% CIs. Inset values
 1231 indicate respective sample sizes used for each plant growth form.

Figure S11. The relationships between life cycle survival profile ($\Delta\sigma$) and the relative contributions of population structure to variation in realised growth rates ($\lambda_{transient}$) across growth forms are maintained even when assessing only non-imputed data. (A) The relationship between $\Delta\sigma$ (*i.e.*, rate of change in stage-specific survival across the lifecycle) and contribution ratio estimates (C_{ratio}) across 183 comparisons of plant population dynamics for which no data required imputing. C_{ratio} estimates reflect the relative balance of contributions of population structure (\mathbf{n}) and vital rates (\mathbf{A}) toward observed variance in demographic performance, with the dashed line at $C_{ratio} = 1$ indicating equal contributions. Error displayed using 95% CIs. **(B)** The relationship between $\Delta\sigma$ and contribution ratio estimates (C_{ratio}) shown separately for Hemicrypophytes, Chamaephytes, and Phanerophytes. Error is displayed as 95% CI with the horizontal



1241 dashed line indicating the 1:1 threshold between contributions from vital rates and population structure.

Figure S12. The relationships between degree of iteroparity and the relative contributions of population structure to variation in realised growth rates ($\lambda_{transient}$) across growth forms are maintained even when assessing only non-imputed data. (A) The relationship between degree of iteroparity (*i.e.*, spread of reproduction through the life cycle) and contribution ratio estimates (C_{ratio}) across 183 comparisons of plant population dynamics for which no data required imputing. C_{ratio} estimates reflect the relative balance of contributions of population structure (n) and vital rates (A) toward observed variance in demographic performance, with the dashed line at $C_{ratio} = 1$ indicating equal contributions. Error displayed using 95% CIs. **(B)** The relationship between degree of iteroparity and contribution ratio estimates (C_{ratio}) shown separately for Hemicryophytes, Chamaephytes, and Phanerophytes. Error is displayed as 95% CI with the horizontal dashed line indicating the 1:1 threshold between contributions



1252 from vital rates and population structure.

1253 **Table S1. A summary of the 268 population comparisons used in this study.** This table shows the
1254 number of matrix population models (MPMs) included for each comparison (# MPM), the number of life
1255 cycle stages represented by each model (# stages), whether the sets of MPMs describe comparisons across
1256 time, space, or experimental treatments, or combinations of these cases (LTRE comparison), and the
1257 Raunkiær growth form classification (6) assigned to the plant species associated with each comparison.

Species	# MPM	# stages	LTRE comparison	Raunkiær growth form
<i>Acer saccharum</i>	2	3	Time	Megaphanerophyte
<i>Actaea elata</i>	2	6	Time	Hemicryptophyte
<i>Actinostemon concolor</i>	2	3	Time	Nanophanerophyte
<i>Agrimonia eupatoria</i>	2	4	Space	Hemicryptophyte
<i>Agrimonia eupatoria</i>	2	4	Space	Hemicryptophyte
<i>Artemisia genipi</i>	2	5	Treatment	Chamaephyte
<i>Artemisia genipi</i>	2	5	Time	Chamaephyte
<i>Attalea humilis</i>	2	5	Treatment	Nanophanerophyte
<i>Boswellia papyrifera</i>	2	12	Treatment	Mesophanerophyte
<i>Bursera glabrifolia</i>	2	5	Time	Nanophanerophyte
<i>Chamaedorea elegans</i>	2	6	Treatment	Nanophanerophyte
<i>Chamaedorea elegans</i>	2	6	Treatment	Nanophanerophyte
<i>Chamaedorea elegans</i>	2	6	Treatment	Nanophanerophyte
<i>Chamaedorea elegans</i>	2	6	Time	Nanophanerophyte
<i>Chamaedorea elegans</i>	2	6	Time	Nanophanerophyte
<i>Clidemia hirta</i>	2	5	Space	Epiphyte

<i>Coryphantha werdermannii</i>	2	6	Multiple	Chamaephyte
<i>Coryphantha werdermannii</i>	2	6	Multiple	Chamaephyte
<i>Coryphantha werdermannii</i>	2	6	Multiple	Chamaephyte
<i>Danthonia sericea</i>	2	6	Time	Hemicryptophyte
<i>Danthonia sericea</i>	2	6	Time	Hemicryptophyte
<i>Danthonia sericea</i>	2	6	Time	Hemicryptophyte
<i>Dicerandra frutescens</i>	2	6	Treatment	Nanophanerophyte
<i>Eryngium cuneifolium</i>	2	6	Space	Hemicryptophyte
<i>Eryngium cuneifolium</i>	2	6	Multiple	Hemicryptophyte
<i>Eryngium cuneifolium</i>	2	6	Space	Hemicryptophyte
<i>Eryngium cuneifolium</i>	2	6	Multiple	Hemicryptophyte
<i>Fagus grandifolia</i>	2	9	Treatment	Megaphanerophyte
<i>Heliconia acuminata</i>	2	6	Treatment	Geophyte
<i>Heliconia acuminata</i>	2	6	Treatment	Geophyte
<i>Liatris scariosa</i>	2	5	Multiple	Hemicryptophyte
<i>Liatris scariosa</i>	2	5	Multiple	Hemicryptophyte
<i>Liatris scariosa</i>	2	5	Multiple	Hemicryptophyte
<i>Liatris scariosa</i>	2	5	Multiple	Hemicryptophyte
<i>Liatris scariosa</i>	2	5	Time	Hemicryptophyte
<i>Lomatium cookii</i>	2	6	Space	Hemicryptophyte
<i>Lomatium cookii</i>	2	6	Space	Hemicryptophyte
<i>Lomatium cookii</i>	2	6	Space	Hemicryptophyte

<i>Lomatium cookii</i>	2	6	Space	Hemicryptophyte
<i>Mammillaria crucigera</i>	2	8	Time	Chamaephyte
<i>Mammillaria huitzilopochtli</i>	2	5	Multiple	Chamaephyte
<i>Mammillaria huitzilopochtli</i>	2	5	Multiple	Chamaephyte
<i>Mammillaria huitzilopochtli</i>	2	5	Multiple	Chamaephyte
<i>Mammillaria huitzilopochtli</i>	2	5	Multiple	Chamaephyte
<i>Mammillaria huitzilopochtli</i>	2	5	Multiple	Chamaephyte
<i>Mammillaria magnimamma</i>	2	7	Multiple	Chamaephyte
<i>Mammillaria magnimamma</i>	2	7	Time	Chamaephyte
<i>Manilkara zapota</i>	2	9	Time	Mesophanerophyte
<i>Nardostachys jatamansi</i>	2	6	Treatment	Hemicryptophyte
<i>Nardostachys jatamansi</i>	2	6	Treatment	Hemicryptophyte
<i>Nardostachys jatamansi</i>	2	6	Treatment	Hemicryptophyte
<i>Nardostachys jatamansi</i>	2	6	Treatment	Hemicryptophyte
<i>Neobuxbaumia tetetzo</i>	2	10	Time	Mesophanerophyte
<i>Pachycereus pecten-</i>				
<i>aboriginum</i>	2	9	Time	Mesophanerophyte
<i>Primula veris</i>	2	6	Time	Hemicryptophyte
<i>Primula vulgaris</i>	2	5	Time	Hemicryptophyte
<i>Primula vulgaris</i>	2	5	Time	Hemicryptophyte
<i>Primula vulgaris</i>	2	5	Time	Hemicryptophyte
<i>Primula vulgaris</i>	2	5	Time	Hemicryptophyte

<i>Primula vulgaris</i>	2	5	Time	Hemicryptophyte
<i>Primula vulgaris</i>	2	5	Time	Hemicryptophyte
<i>Pseudophoenix sargentii</i>	2	5	Treatment	Mesophanerophyte
<i>Pterocereus gaumeri</i>	2	10	Space	Mesophanerophyte
<i>Pterocereus gaumeri</i>	2	10	Time	Mesophanerophyte
<i>Sanicula europaea</i>	2	5	Treatment	Hemicryptophyte
<i>Sanicula europaea</i>	2	5	Treatment	Hemicryptophyte
<i>Sanicula europaea</i>	2	5	Treatment	Hemicryptophyte
<i>Sanicula europaea</i>	2	5	Treatment	Hemicryptophyte
<i>Sanicula europaea</i>	2	5	Treatment	Hemicryptophyte
<i>Sanicula europaea</i>	2	5	Treatment	Hemicryptophyte
<i>Tillandsia recurvata</i>	2	4	Multiple	Epiphyte
<i>Tillandsia recurvata</i>	2	4	Multiple	Epiphyte
<i>Tillandsia recurvata</i>	2	4	Multiple	Epiphyte
<i>Tillandsia recurvata</i>	2	4	Time	Epiphyte
<i>Trollius europaeus</i>	2	6	Space	Hemicryptophyte
<i>Trollius europaeus</i>	2	6	Time	Hemicryptophyte
<i>Viola elatior</i>	2	5	Space	Hemicryptophyte
<i>Viola pumila</i>	2	5	Space	Hemicryptophyte
<i>Viola stagnina</i>	2	5	Space	Hemicryptophyte
<i>Vriesea sanguinolenta</i>	2	7	Space	Epiphyte
<i>Vriesea sanguinolenta</i>	2	7	Space	Epiphyte

<i>Vriesea sanguinolenta</i>	2	7	Space	Epiphyte
<i>Acer saccharum</i>	2	3	Time	Megaphanerophyte
<i>Acer saccharum</i>	3	3	Space	Megaphanerophyte
<i>Actaea elata</i>	2	6	Time	Hemicryptophyte
<i>Actinostemon concolor</i>	2	3	Time	Nanophanerophyte
<i>Agrimonia eupatoria</i>	2	4	Time	Hemicryptophyte
<i>Agrimonia eupatoria</i>	2	4	Space	Hemicryptophyte
<i>Agrimonia eupatoria</i>	2	4	Space	Hemicryptophyte
<i>Agrimonia eupatoria</i>	4	4	Time	Hemicryptophyte
<i>Ambrosia dumosa</i>	4	6	Treatment	Chamaephyte
<i>Ambrosia dumosa</i>	4	6	Treatment	Chamaephyte
<i>Boechera fecunda</i>	3	4	Space	Hemicryptophyte
<i>Boechera fecunda</i>	3	4	Space	Hemicryptophyte
<i>Boechera fecunda</i>	3	4	Space	Hemicryptophyte
<i>Boechera fecunda</i>	3	4	Space	Hemicryptophyte
<i>Boechera fecunda</i>	6	4	Time	Hemicryptophyte
<i>Boechera fecunda</i>	4	4	Time	Hemicryptophyte
<i>Boechera fecunda</i>	4	4	Time	Hemicryptophyte
<i>Araucaria laubenfelsii</i>	3	4	Treatment	Megaphanerophyte
<i>Artemisia genipi</i>	2	5	Treatment	Chamaephyte
<i>Artemisia genipi</i>	2	5	Time	Chamaephyte
<i>Astragalus scaphoides</i>	10	5	Time	Hemicryptophyte

<i>Astragalus scaphoides</i>	9	5	Time	Hemicryptophyte
<i>Astragalus scaphoides</i>	8	5	Time	Hemicryptophyte
<i>Astragalus tyghensis</i>	6	5	Time	Hemicryptophyte
<i>Astragalus tyghensis</i>	6	5	Time	Hemicryptophyte
<i>Astragalus tyghensis</i>	7	5	Time	Hemicryptophyte
<i>Astragalus tyghensis</i>	5	5	Time	Hemicryptophyte
<i>Astragalus tyghensis</i>	6	5	Time	Hemicryptophyte
<i>Attalea humilis</i>	2	5	Treatment	Nanophanerophyte
<i>Boswellia papyrifera</i>	2	12	Treatment	Mesophanerophyte
<i>Brassica insularis</i>	3	3	Space	Chamaephyte
<i>Brassica insularis</i>	4	3	Space	Chamaephyte
<i>Brassica insularis</i>	4	3	Space	Chamaephyte
<i>Brassica insularis</i>	3	3	Space	Chamaephyte
<i>Brassica insularis</i>	4	3	Space	Chamaephyte
<i>Brassica insularis</i>	4	3	Space	Chamaephyte
<i>Brassica insularis</i>	3	3	Space	Chamaephyte
<i>Brassica insularis</i>	3	3	Space	Chamaephyte
<i>Brassica insularis</i>	3	3	Space	Chamaephyte
<i>Brassica insularis</i>	8	3	Time	Chamaephyte
<i>Brassica insularis</i>	5	3	Time	Chamaephyte
<i>Brassica insularis</i>	9	3	Time	Chamaephyte
<i>Brassica insularis</i>	9	3	Time	Chamaephyte

<i>Bursera glabrifolia</i>	2	5	Time	Nanophanerophyte
<i>Castanea dentata</i>	6	8	Multiple	Mesophanerophyte
<i>Castanea dentata</i>	6	8	Multiple	Mesophanerophyte
<i>Castanea dentata</i>	6	8	Multiple	Mesophanerophyte
<i>Castanea dentata</i>	5	8	Multiple	Mesophanerophyte
<i>Castanea dentata</i>	4	8	Time	Mesophanerophyte
<i>Castanea dentata</i>	4	8	Time	Mesophanerophyte
<i>Castanea dentata</i>	4	8	Time	Mesophanerophyte
<i>Castanea dentata</i>	3	8	Time	Mesophanerophyte
<i>Castanea dentata</i>	4	8	Time	Mesophanerophyte
<i>Castanea dentata</i>	4	8	Time	Mesophanerophyte
<i>Centaurea horrida</i>	3	5	Space	Nanophanerophyte
<i>Chamaedorea elegans</i>	3	6	Treatment	Nanophanerophyte
<i>Clidemia hirta</i>	2	5	Space	Epiphyte
<i>Clidemia hirta</i>	2	5	Time	Epiphyte
<i>Coryphantha werdermannii</i>	3	6	Multiple	Chamaephyte
<i>Danthonia sericea</i>	2	6	Time	Hemicryptophyte
<i>Danthonia sericea</i>	2	6	Time	Hemicryptophyte
<i>Danthonia sericea</i>	2	6	Time	Hemicryptophyte
<i>Danthonia sericea</i>	3	6	Space	Hemicryptophyte
<i>Danthonia sericea</i>	5	6	Space	Hemicryptophyte
<i>Dicerandra frutescens</i>	2	6	Multiple	Nanophanerophyte

<i>Dicerandra frutescens</i>	3	6	Multiple	Nanophanerophyte
<i>Dicerandra frutescens</i>	3	6	Multiple	Nanophanerophyte
<i>Dicerandra frutescens</i>	3	6	Multiple	Nanophanerophyte
<i>Dicerandra frutescens</i>	3	6	Multiple	Nanophanerophyte
<i>Dicerandra frutescens</i>	2	6	Multiple	Nanophanerophyte
<i>Dicerandra frutescens</i>	3	6	Multiple	Nanophanerophyte
<i>Dicerandra frutescens</i>	2	6	Multiple	Nanophanerophyte
<i>Dicerandra frutescens</i>	2	6	Multiple	Nanophanerophyte
<i>Dicerandra frutescens</i>	5	6	Time	Nanophanerophyte
<i>Dicerandra frutescens</i>	3	6	Time	Nanophanerophyte
<i>Dicerandra frutescens</i>	3	6	Time	Nanophanerophyte
<i>Dicerandra frutescens</i>	5	6	Time	Nanophanerophyte
<i>Dicerandra frutescens</i>	11	6	Time	Nanophanerophyte
<i>Eriogonum longifolium</i>	4	5	Time	Hemicryptophyte
<i>Eriogonum longifolium</i>	4	5	Time	Hemicryptophyte
<i>Eriogonum longifolium</i>	8	5	Time	Hemicryptophyte
<i>Eryngium alpinum</i>	2	4	Multiple	Hemicryptophyte
<i>Eryngium alpinum</i>	3	4	Time	Hemicryptophyte
<i>Eryngium cuneifolium</i>	4	6	Time	Hemicryptophyte
<i>Eryngium cuneifolium</i>	3	6	Time	Hemicryptophyte
<i>Eryngium cuneifolium</i>	7	6	Time	Hemicryptophyte
<i>Eryngium cuneifolium</i>	8	6	Time	Hemicryptophyte

<i>Eryngium cuneifolium</i>	7	6	Time	Hemicryptophyte
<i>Eryngium cuneifolium</i>	2	6	Multiple	Hemicryptophyte
<i>Eryngium cuneifolium</i>	2	6	Space	Hemicryptophyte
<i>Eryngium cuneifolium</i>	5	6	Multiple	Hemicryptophyte
<i>Eryngium cuneifolium</i>	3	6	Space	Hemicryptophyte
<i>Eryngium cuneifolium</i>	5	6	Multiple	Hemicryptophyte
<i>Eryngium cuneifolium</i>	3	6	Space	Hemicryptophyte
<i>Eryngium cuneifolium</i>	4	6	Multiple	Hemicryptophyte
<i>Eryngium cuneifolium</i>	3	6	Space	Hemicryptophyte
<i>Eryngium cuneifolium</i>	4	6	Multiple	Hemicryptophyte
<i>Eryngium cuneifolium</i>	3	6	Space	Hemicryptophyte
<i>Eryngium cuneifolium</i>	2	6	Multiple	Hemicryptophyte
<i>Eryngium cuneifolium</i>	2	6	Space	Hemicryptophyte
<i>Eryngium cuneifolium</i>	2	6	Multiple	Hemicryptophyte
<i>Euterpe edulis</i>	3	7	Time	Mesophanerophyte
<i>Fagus grandifolia</i>	2	9	Treatment	Megaphanerophyte
<i>Guarianthe aurantiaca</i>	2	5	Time	Epiphyte
<i>Pyrrocoma radiata</i>	7	4	Time	Hemicryptophyte
<i>Pyrrocoma radiata</i>	7	4	Time	Hemicryptophyte
<i>Pyrrocoma radiata</i>	7	4	Time	Hemicryptophyte
<i>Pyrrocoma radiata</i>	6	4	Time	Hemicryptophyte
<i>Pyrrocoma radiata</i>	6	4	Time	Hemicryptophyte

<i>Pyrrocoma radiata</i>	6	4	Time	Hemicryptophyte
<i>Pyrrocoma radiata</i>	5	4	Time	Hemicryptophyte
<i>Pyrrocoma radiata</i>	5	4	Time	Hemicryptophyte
<i>Heliconia acuminata</i>	2	6	Treatment	Geophyte
<i>Heliconia acuminata</i>	2	6	Treatment	Geophyte
<i>Hypericum cumulicola</i>	4	6	Space	Hemicryptophyte
<i>Hypericum cumulicola</i>	4	6	Space	Hemicryptophyte
<i>Hypericum cumulicola</i>	3	6	Space	Hemicryptophyte
<i>Hypericum cumulicola</i>	3	6	Space	Hemicryptophyte
<i>Hypericum cumulicola</i>	4	6	Space	Hemicryptophyte
<i>Hypericum cumulicola</i>	4	6	Time	Hemicryptophyte
<i>Hypericum cumulicola</i>	3	6	Time	Hemicryptophyte
<i>Hypericum cumulicola</i>	3	6	Time	Hemicryptophyte
<i>Hypericum cumulicola</i>	4	6	Time	Hemicryptophyte
<i>Hypericum cumulicola</i>	4	6	Time	Hemicryptophyte
<i>Hypericum cumulicola</i>	4	6	Time	Hemicryptophyte
<i>Liatris scariosa</i>	2	5	Multiple	Hemicryptophyte
<i>Liatris scariosa</i>	3	5	Multiple	Hemicryptophyte
<i>Liatris scariosa</i>	3	5	Time	Hemicryptophyte
<i>Liatris scariosa</i>	2	5	Time	Hemicryptophyte
<i>Lomatium cookii</i>	4	6	Time	Hemicryptophyte
<i>Lomatium cookii</i>	4	6	Time	Hemicryptophyte

<i>Lomatium cookii</i>	2	6	Space	Hemicryptophyte
<i>Lomatium cookii</i>	2	6	Space	Hemicryptophyte
<i>Lomatium cookii</i>	2	6	Space	Hemicryptophyte
<i>Lomatium cookii</i>	2	6	Space	Hemicryptophyte
<i>Lomatium cookii</i>	4	6	Time	Hemicryptophyte
<i>Lomatium cookii</i>	4	6	Time	Hemicryptophyte
<i>Mammillaria crucigera</i>	2	8	Time	Chamaephyte
<i>Mammillaria huitzilopochtli</i>	5	5	Time	Chamaephyte
<i>Mammillaria huitzilopochtli</i>	2	5	Multiple	Chamaephyte
<i>Mammillaria huitzilopochtli</i>	2	5	Multiple	Chamaephyte
<i>Mammillaria huitzilopochtli</i>	2	5	Multiple	Chamaephyte
<i>Mammillaria huitzilopochtli</i>	2	5	Multiple	Chamaephyte
<i>Mammillaria huitzilopochtli</i>	2	5	Multiple	Chamaephyte
<i>Mammillaria huitzilopochtli</i>	5	5	Time	Chamaephyte
<i>Mammillaria magnimamma</i>	2	7	Multiple	Chamaephyte
<i>Mammillaria magnimamma</i>	2	7	Time	Chamaephyte
<i>Manilkara zapota</i>	2	9	Time	Mesophanerophyte
<i>Nardostachys jatamansi</i>	5	6	Treatment	Hemicryptophyte
<i>Neobuxbaumia</i>				
<i>macrocephala</i>	3	10	Time	Mesophanerophyte
<i>Neobuxbaumia mezcalaensis</i>	3	10	Time	Mesophanerophyte
<i>Neobuxbaumia tetetzo</i>	2	10	Time	Mesophanerophyte

<i>Pachycereus pecten-</i>				
<i>aboriginum</i>	2	9	Time	Mesophanerophyte
<i>Pediocactus bradyi</i>	4	3	Time	Chamaephyte
<i>Pediocactus bradyi</i>	5	3	Time	Chamaephyte
<i>Pediocactus bradyi</i>	4	3	Time	Chamaephyte
<i>Pediocactus bradyi</i>	13	3	Time	Chamaephyte
<i>Phyllanthus indofischeri</i>	5	7	Time	Mesophanerophyte
<i>Plantago media</i>	3	5	Time	Hemicryptophyte
<i>Plantago media</i>	2	5	Time	Hemicryptophyte
<i>Primula veris</i>	2	6	Time	Hemicryptophyte
<i>Primula vulgaris</i>	7	5	Space	Hemicryptophyte
<i>Primula vulgaris</i>	7	5	Space	Hemicryptophyte
<i>Primula vulgaris</i>	2	5	Time	Hemicryptophyte
<i>Primula vulgaris</i>	2	5	Time	Hemicryptophyte
<i>Primula vulgaris</i>	2	5	Time	Hemicryptophyte
<i>Primula vulgaris</i>	2	5	Time	Hemicryptophyte
<i>Primula vulgaris</i>	2	5	Time	Hemicryptophyte
<i>Primula vulgaris</i>	2	5	Time	Hemicryptophyte
<i>Pseudophoenix sargentii</i>	2	5	Treatment	Mesophanerophyte
<i>Pterocereus gaumeri</i>	2	10	Space	Mesophanerophyte
<i>Pterocereus gaumeri</i>	2	10	Time	Mesophanerophyte
<i>Ramonda myconi</i>	2	5	Space	Hemicryptophyte

<i>Sanicula europaea</i>	4	5	Treatment	Hemicryptophyte
<i>Tillandsia recurvata</i>	2	4	Multiple	Epiphyte
<i>Tillandsia recurvata</i>	2	4	Multiple	Epiphyte
<i>Tillandsia recurvata</i>	2	4	Multiple	Epiphyte
<i>Tillandsia recurvata</i>	6	4	Space	Epiphyte
<i>Tillandsia recurvata</i>	2	4	Time	Epiphyte
<i>Trollius europaeus</i>	3	6	Space	Hemicryptophyte
<i>Trollius europaeus</i>	2	6	Space	Hemicryptophyte
<i>Trollius europaeus</i>	2	6	Time	Hemicryptophyte
<i>Viola elatior</i>	2	5	Space	Hemicryptophyte
<i>Viola pumila</i>	2	5	Space	Hemicryptophyte
<i>Viola stagnina</i>	2	5	Space	Hemicryptophyte
<i>Vriesea sanguinolenta</i>	3	7	Space	Epiphyte

Supporting Information References

1. O. R. Jones, *et al.*, Rcompadre and Rage - two R packages to facilitate the use of the COMPADRE and COMADRE databases and calculation of life history traits from matrix population models. *Methods Ecol. Evol.* **13**, 770–781 (2022).
2. R. Salguero-Gómez, J. B. Plotkin, Matrix dimensions bias demographic inferences: Implications for comparative plant demography. *Am. Nat.* **176**, 710–722 (2010).
3. R. Salguero-Gómez, *et al.*, Fast–slow continuum and reproductive strategies structure plant life-history variation worldwide. *Proc. Natl. Acad. Sci.* **113**, 220–235 (2016).
4. P. Capdevila, *et al.*, Life history mediates the trade-offs among different components of demographic resilience. *Ecol. Lett.* **00**, 1–14 (2022).
5. T. D. James, R. Salguero-Gómez, O. R. Jones, D. Z. Childs, A. P. Beckerman, Bridging gaps in demographic analysis with phylogenetic imputation. *Conserv. Biol.* **35**, 1210–1221 (2021).
6. C. Raunkiaer, *The life forms of plants and statistical plant geography; being the collected papers of C. Raunkiaer.* (1934).